Elemental analysis of oxygenated organic coating on black carbon particles using a soot-particle aerosol mass spectrometer

Mutian Ma¹, Laura-Hélèna Rivellini², YuXi Cui¹, Megan D. Willis³, Rio Wilkie⁴, Jonathan P. D. Abbatt⁴, Manjula R. Canagaratna⁵, Junfeng Wang⁶, Xinlei Ge⁶, and Alex K. Y. Lee¹,²,a

¹Department of Civil and Environmental Engineering, National University of Singapore, Singapore
²NUS Environmental Research Institute, National University of Singapore, Singapore
³Department of Chemistry, Colorado State University, Fort Collins, CO, USA
⁴Department of Chemistry, University of Toronto, Toronto, ON, Canada
⁵Aerodyne Research, Inc., Billerica, MA, USA
⁶School of Environmental Science and Engineering, Nanjing University of Information Science and Technology, Nanjing, China

Correspondence: Alex K. Y. Lee (alex.lee@canada.ca)

Received: 27 July 2020 – Discussion started: 26 August 2020
Revised: 16 February 2021 – Accepted: 17 February 2021 – Published: 12 April 2021

Abstract. Chemical characterization of organic coatings is important to advance our understanding of the physio-chemical properties and environmental fate of black carbon (BC) particles. The soot-particle aerosol mass spectrometer (SP-AMS) has been utilized for this purpose in recent field studies. The laser vaporization (LV) scheme of the SP-AMS can heat BC cores gradually until they are completely vaporized, during which organic coatings can be vaporized at temperatures lower than that of the thermal vaporizer (TV) used in a standard high-resolution time-of-flight aerosol mass spectrometer (HR-ToF-AMS) that employs flash vaporization. This work investigates the effects of vaporization schemes on fragmentation and elemental analysis of known oxygenated organic species using three SP-AMS instruments. We show that LV can reduce fragmentation of organic molecules. Substantial enhancement of $C_2H_3O^+/CO_2^+$ and $C_2H_4O_2^+$ signals was observed for most of the tested species when the LV scheme was used, suggesting that the observational frameworks based on the use of HR-ToF-AMS field data may not be directly applicable for evaluating the chemical evolution of oxygenated organic aerosol (OOA) components coated on ambient BC particles. The uncertainties of $H:C$ and $O:C$ determined using the improved-ambient (I-A) method for both LV and TV approaches were similar, and scaling factors of 1.10 for $H:C$ and 0.89 for $O:C$ were determined to facilitate more direct comparisons between observations from the two vaporization schemes. Furthermore, the I-A method was updated based on the multilinear regression model for the LV scheme measurements. The updated parameters can reduce the relative errors of $O:C$ from $-26.3\%$ to $5.8\%$, whereas the relative errors of $H:C$ remain roughly the same. Applying the scaling factors and the updated parameters for the I-A method to ambient data, we found that even though the time series of OOA components determined using the LV and TV schemes are strongly correlated at the same location, OOA coatings were likely less oxygenated compared to those externally mixed with BC.

1 Introduction

Atmospheric black carbon (BC) particles have significant impacts on climate and human health. Studies have shown that BC particles have negative impact on vascular, cardiopulmonary, respiratory, and chronic diseases (Brook et al., 2010; Heal et al., 2012; Heinzerling et al., 2016; Nel, 2005). They are strongly light-absorbing, resulting in positive radiative forcing which is equivalent to $\sim 55\%$ of the radiative forcing caused by carbon dioxide (Bond and Bergstrom, 2007;
Ambient BC particles are largely internally mixed with organic aerosols (OAs). While BC and hydrophobic organic coating can be co-emitted from combustion processes, BC can be coated by oxygenated organic compounds formed via oxidation of organic vapors and heterogeneous processing (China et al., 2013; Lee et al., 2019b; Wu et al., 2019). Organic coatings can alter the physio-chemical properties of BC particles, such as enhancements of light absorptivity (Liu et al., 2017; Peng et al., 2016), cloud formation potential (Kuwata et al., 2009; Liu et al., 2013), and subsequently atmospheric transport and lifetime of BC particles (Bond et al., 2013; Laborde et al., 2013; McMeeking et al., 2011). However, characterization of organic coatings on BC particles remains a great analytical challenge due to their chemical complexity.

The soot-particle aerosol mass spectrometer (SP-AMS), which is a standard high-resolution time-of-flight aerosol mass spectrometer (HR-ToF-AMS) equipped with an additional infrared (IR) laser vaporizer (Onasch et al., 2012), has been utilized to characterize refractory BC (rBC) and its coatings (Collier et al., 2018; Lee et al., 2017; Massoli et al., 2015; Willis et al., 2016; Wu et al., 2019). The laser vaporization scheme of the SP-AMS can heat rBC cores gradually (up to ~4000 K) until they are completely vaporized, during which non-refractory coatings can be vaporized at a temperature lower than the operating temperature of thermal vaporizer (600°C) typically used in HR-ToF-AMSs for flash vaporization (DeCarlo et al., 2006; Onasch et al., 2012) (Fig. 1). Different fragmentation patterns of oxygenated OA (OOA) components have been observed between the two vaporization schemes in both laboratory and field studies (Canagaratna et al., 2015b; Massoli et al., 2015). In particular, positive matrix factorization (PMF) of field measurements has shown that mass spectra of OOA materials, which usually represents secondary OA (SOA) components in the atmosphere, determined by the laser vaporization (LV) scheme of the SP-AMS were dominated by C$_2$H$_5$O$^+$ signals (Collier et al., 2018; Wu et al., 2019). In contrast, the co-located HR-ToF-AMS measurements (i.e., thermal vaporization (TV) scheme) generated mass spectra of more-oxidized OOA components with the most intense signals from CO$^+$ and CO$_2$ fragments (Chen et al., 2018; Lee et al., 2017; Massoli et al., 2015) as illustrated in Fig. 1a and b. Previous field studies have also shown that there were some mass spectral differences (e.g., C$_3$H$_5^+$/C$_3$H$_7^+$ and C$_5$H$_7^+$/C$_5$H$_5^+$ ratios) between the hydrocarbon-like OA (HOA) factors, which usually represents primary OA (POA) components emitted from traffic or fossil fuel combustion processes, identified using the TV and LV schemes at the same location (Chen et al., 2018; Lee et al., 2017; Massoli et al., 2015; Wang et al., 2019, 2020, 2017; Xie et al., 2019a, b).

Elemental ratios of OAs (i.e., hydrogen-to-carbon, H : C, and oxygen-to-carbon, O : C) measured by HR-ToF-AMSs have been widely used to investigate the physical and chemical properties of OAs (Koop et al., 2011; Kuwata et al., 2012; Lambe et al., 2013; Massoli et al., 2010; Wong et al., 2011), and their evolution depends on the types of primary emissions and aging processes (Heald et al., 2010; Jimenez et al., 2009; Kroll et al., 2011). The elemental ratios obtained from HR-ToF-AMS mass spectra can be potentially biased by vaporization and ion fragmentation processes as described in detail by Canagaratna et al. (2015a). To account for such measurement uncertainties, the calibration factors between experimentally measured and theoretical elemental compositions of known organic compounds were reported by Aiken et al. (2007, 2008). The “Aiken-ambient” (A-A) method was developed for elemental analysis of ambient OAs using empirically estimated CO$^+$ and H$_2$O$^+$ ion signals to avoid interferences from ambient air (Aiken et al., 2008). Canagaratna et al. (2015a) further established the “improved-ambient” (I-A) method that uses specific ion fragments as markers to reduce composition-dependent systematic biases. Both A-A and I-A methods have been fully integrated into the standard procedure for analyzing ambient OAs measured by HR-ToF-AMSs. However, the two methods were developed based on OA mass spectra generated by the TV approach; hence their direct applications for determining elemental compositions of OAs vaporized using the LV scheme (i.e., organic coating on BC particles) may not be appropriate (Canagaratna et al., 2015b).

To improve the accuracy of elemental analysis for OOA materials coated on ambient BC particles using SP-AMSs, this work compares the elemental ratios (H : C and O : C) of known oxygenated organic compounds determined by both TV and LV schemes. This work focuses on investigating oxygenated organic compounds as significant mass spectra differences were observed from ambient OOA components detected using TV and LV schemes. The I-A method was applied to determine the elemental composition of OOA materials for all our laboratory experiments. Such comparison has been conducted by Canagaratna et al. (2015b), in which only a small number of organic species was tested by a single SP-AMS. In this study, we extend this type of investigation in three ways: (1) increasing the number of organic species with different functional groups to be tested, (2) deploying two additional independent SP-AMSs from different research groups to conduct our measurements, and (3) generating new fitting parameters based on the approach for developing the I-A method to enhance the accuracy of elemental analysis for organic coatings detected using the LV scheme of the SP-AMS. The results can be used to evaluate the robustness of applying the combination of the laser vaporization approach and the I-A method for determining H : C, O : C, and OSc of ambient OOA coated on BC particles, and to provide insight into the potential discrepancies between ambient OOA materials that are externally and internally mixed with BC particles.
2 Experiment

2.1 Particle generation

A total of 30 oxygenated organic species, including dicarboxylic acids, polycarboxylic acids, alcohols, and multifunctional compounds, were included in this study (Table 1). Small amounts of a standard organic compound were dissolved in ultrapure water, which was subsequently used to generate pure organic particles using a constant output atomizer (Model 2076, TSI). For generating rBC–organic mixed particles, REGAL black pigment (REGAL 400R, Cabot Corp) and a standard organic compound were mixed in the bulk solution for atomization. REGAL black was used in this study because it has been suggested as an rBC standard for calibrating the LV scheme of the SP-AMS (Onasch et al., 2012). Atomized particles were subsequently dried by a diffusion dryer using silica gel to minimize the interference of particle-phase water to $\text{H}_2\text{O}^+$ signals and other related fragments (i.e., O$^+$ and HO$^+$). Pure argon gas was used for atomization and dilution to minimize the interference of gaseous N$_2$ and CO$_2$ to the quantification of CO$^+$ and CO$_2^+$ signals in organic mass spectra (Canagaratna et al., 2015a; Corbin et al., 2014; Willis et al., 2014).

2.2 Soot-particle aerosol mass spectrometer measurements

The details of the SP-AMS (Aerodyne Research, Inc.) have been reported in detail by Onasch et al. (2012). In brief, the SP-AMS is equipped with a thermal vaporizer (i.e., a heated tungsten surface) and a laser vaporizer (i.e., a continuous-wave intra-cavity 1064 nm Nd:YAG laser). While the thermal vaporizer operated at 600 °C can vaporize non-refractory particulate matter (NR-PM, including organic, sulfate, nitrate, ammonium, and chloride), the laser vaporizer is designed for vaporizing rBC-containing particles at which rBC cores can be gradually heated up to ∼4000 K. During this heating process, organic coatings can be vaporized at a lower range of temperature (likely <600 °C) depending on the volatility of each organic compound. The vaporized analytes are ionized using 70 eV electron impact (EI) ionization, and the ions are subsequently detected by a high-resolution time-of-flight mass spectrometer operated in V mode (Canagaratna et al., 2007; DeCarlo et al., 2006).

The SP-AMS instruments were operated in two different vaporization schemes for characterizing pure organic particles and rBC–organic mixed particles, respectively. When the laser vaporizer of the SP-AMS was off, the instrument was operated as a standard HR-ToF-AMS to facilitate flash-vaporization of pure organic particles (TV scheme). For the second part of the experiments, the thermal vaporizer was removed from the SP-AMS and the laser vaporizer was turned on for measuring standard organic compounds coated on REGAL black exclusively (LV scheme; Fig. 1c). Note that some pure organic particles might be generated through atomization of the rBC-organic mixture, but they cannot be detected using the LV scheme. Observations from three SP-AMSs were reported in this study, and they are labeled as SP-AMS 1, 2, and 3. SP-AMS 1 and 2 were used to generate new data for 18 and 20 organic species, respectively. Data of SP-AMS 3 (10 organic species) were extracted from Canagaratna et al. (2015b). Table 1 summarizes the tested species for each SP-AMS. The three SP-AMSs were operated by different researchers from the National University of Singapore (SP-AMS 1), University of Toronto (SP-AMS 2), and Aerodyne Research (SP-AMS 3).

2.3 Data analysis

The raw data of SP-AMS measurements were processed by the AMS data analysis software (Squirrel for unit mass resolution (UMR) data and PIKA 1.21b for high-resolution peak fitting; Sueper, 2015), and statistical analyses were processed...
Table 1. Summary of the true values of H : C, O : C, and OS_C of the oxygenated organic species tested by the three SP-AMSs.

<table>
<thead>
<tr>
<th>Class</th>
<th>Name</th>
<th>Formula</th>
<th>H : C</th>
<th>O : C</th>
<th>OS_C</th>
<th>SP-AMS</th>
</tr>
</thead>
<tbody>
<tr>
<td>Multifunction</td>
<td>cis-Pinonic acid</td>
<td>C_{10}H_{16}O_{5}</td>
<td>1.60</td>
<td>0.30</td>
<td>-1.00</td>
<td>2, 3</td>
</tr>
<tr>
<td></td>
<td>Citric acid</td>
<td>C_{6}H_{8}O_{7}</td>
<td>1.33</td>
<td>1.17</td>
<td>1.00</td>
<td>1, 2, 3</td>
</tr>
<tr>
<td></td>
<td>Glutamic acid</td>
<td>C_{5}H_{6}NO_{4}</td>
<td>1.80</td>
<td>0.80</td>
<td>-0.20</td>
<td>2</td>
</tr>
<tr>
<td></td>
<td>Glycolic acid</td>
<td>C_{2}H_{4}O_{3}</td>
<td>2.00</td>
<td>1.50</td>
<td>1.00</td>
<td>1, 2</td>
</tr>
<tr>
<td></td>
<td>Ketogluutaric acid</td>
<td>C_{5}H_{6}O_{5}</td>
<td>1.20</td>
<td>1.00</td>
<td>0.80</td>
<td>3</td>
</tr>
<tr>
<td></td>
<td>Ketopimelic acid</td>
<td>C_{7}H_{10}O_{5}</td>
<td>1.43</td>
<td>0.71</td>
<td>0</td>
<td>3</td>
</tr>
<tr>
<td></td>
<td>Levulinic acid</td>
<td>C_{2}H_{5}O_{3}</td>
<td>1.60</td>
<td>0.60</td>
<td>-0.40</td>
<td>2</td>
</tr>
<tr>
<td></td>
<td>Malic acid</td>
<td>C_{4}H_{6}O_{5}</td>
<td>1.50</td>
<td>1.25</td>
<td>1.00</td>
<td>1, 2</td>
</tr>
<tr>
<td></td>
<td>Pyruvic acid</td>
<td>C_{3}H_{4}O_{3}</td>
<td>1.33</td>
<td>1.00</td>
<td>0.67</td>
<td>2</td>
</tr>
<tr>
<td></td>
<td>Tartaric acid</td>
<td>C_{4}H_{8}O_{6}</td>
<td>1.50</td>
<td>1.50</td>
<td>1.50</td>
<td>1, 2, 3</td>
</tr>
<tr>
<td>Diacids</td>
<td>Adipic acid</td>
<td>C_{6}H_{10}O_{4}</td>
<td>1.67</td>
<td>0.67</td>
<td>-0.33</td>
<td>1, 2</td>
</tr>
<tr>
<td></td>
<td>Azelaic acid</td>
<td>C_{9}H_{16}O_{4}</td>
<td>1.78</td>
<td>0.44</td>
<td>-0.89</td>
<td>1, 2, 3</td>
</tr>
<tr>
<td></td>
<td>Glutaric acid</td>
<td>C_{5}H_{6}O_{4}</td>
<td>1.60</td>
<td>0.80</td>
<td>0</td>
<td>1, 2, 3</td>
</tr>
<tr>
<td></td>
<td>Maleic acid</td>
<td>C_{4}H_{4}O_{4}</td>
<td>1.00</td>
<td>1.00</td>
<td>1.00</td>
<td>2</td>
</tr>
<tr>
<td></td>
<td>Malonic acid</td>
<td>C_{3}H_{4}O_{4}</td>
<td>1.33</td>
<td>1.33</td>
<td>1.33</td>
<td>1, 2, 3</td>
</tr>
<tr>
<td></td>
<td>Oxalic acid</td>
<td>C_{2}H_{2}O_{4}</td>
<td>1.00</td>
<td>2.00</td>
<td>3.00</td>
<td>1, 2</td>
</tr>
<tr>
<td></td>
<td>Phthalic acid</td>
<td>C_{8}H_{6}O_{4}</td>
<td>0.75</td>
<td>0.50</td>
<td>0.25</td>
<td>1</td>
</tr>
<tr>
<td></td>
<td>Pimelic acid</td>
<td>C_{7}H_{12}O_{4}</td>
<td>1.71</td>
<td>0.57</td>
<td>-0.57</td>
<td>1, 3</td>
</tr>
<tr>
<td></td>
<td>Suberic acid</td>
<td>C_{8}H_{14}O_{4}</td>
<td>1.75</td>
<td>0.50</td>
<td>-0.75</td>
<td>1</td>
</tr>
<tr>
<td></td>
<td>Succinic acid</td>
<td>C_{4}H_{6}O_{4}</td>
<td>1.50</td>
<td>1.00</td>
<td>0.50</td>
<td>1, 2, 3</td>
</tr>
<tr>
<td>Polycyclic</td>
<td>Tricarballylic acid</td>
<td>C_{9}H_{8}O_{6}</td>
<td>1.33</td>
<td>1.00</td>
<td>0.67</td>
<td>2</td>
</tr>
<tr>
<td>Alcohols</td>
<td>Arabitol</td>
<td>C_{5}H_{12}O_{5}</td>
<td>2.40</td>
<td>1.00</td>
<td>-0.40</td>
<td>1</td>
</tr>
<tr>
<td></td>
<td>Phenol</td>
<td>C_{6}H_{6}O</td>
<td>1.00</td>
<td>0.17</td>
<td>-0.67</td>
<td>2</td>
</tr>
<tr>
<td></td>
<td>Xylool</td>
<td>C_{5}H_{12}O_{5}</td>
<td>2.40</td>
<td>1.00</td>
<td>-0.40</td>
<td>1, 3</td>
</tr>
<tr>
<td></td>
<td>1,5-Pentanediol</td>
<td>C_{4}H_{12}O_{2}</td>
<td>2.40</td>
<td>0.40</td>
<td>-1.60</td>
<td>2</td>
</tr>
<tr>
<td></td>
<td>Dextrose</td>
<td>C_{6}H_{12}O_{6}</td>
<td>2.00</td>
<td>1.00</td>
<td>0</td>
<td>2</td>
</tr>
<tr>
<td></td>
<td>Glucose</td>
<td>C_{6}H_{12}O_{6}</td>
<td>2.00</td>
<td>1.00</td>
<td>0</td>
<td>1</td>
</tr>
<tr>
<td></td>
<td>Sucrose</td>
<td>C_{12}H_{22}O_{11}</td>
<td>1.83</td>
<td>0.92</td>
<td>0</td>
<td>1, 3</td>
</tr>
<tr>
<td></td>
<td>Levoglucosan</td>
<td>C_{6}H_{10}O_{5}</td>
<td>1.67</td>
<td>0.83</td>
<td>0</td>
<td>1</td>
</tr>
<tr>
<td>Esters</td>
<td>Bis(2-ethylhexyl) sebionate</td>
<td>C_{26}H_{50}O_{4}</td>
<td>1.92</td>
<td>0.15</td>
<td>-1.62</td>
<td>2</td>
</tr>
</tbody>
</table>

by R (version 3.6). Given that pure argon gas was used for particle generation and dilution in all the experiments, CO⁺ signals could be quantified in the high-resolution aerosol mass spectra. Note that only small N₂O signals were detected in some of the experiments using SP-AMS 1 due to the residual N₂ desorbed from the desiccant. For the results of rBC-organic mixed particles, the interference of refractory CO⁺ signals (rCO₊, i.e., CO⁺ and CO₂⁺) formed during rBC vaporization to organic mass spectra was corrected in the fragmentation table using C⁺ : C_f⁺, CO⁺ : C_f⁺, and CO₂⁺ : C_f⁺ ratios obtained from pure REGAL black particles. REGAL black calibrations were performed for each instrument, and thus the applied corrections for the fragmentation table were instrument specific. Elemental analysis was performed using the I-A method (Canagaratna et al., 2015b) to calculate H : C and O : C ratios of each organic compound. The average carbon oxidation state of each organic compound was also calculated based on the method reported in Kroll et al. (2011) (i.e., OS_C = 2 × O : C − H : C).

3 Results and discussion

3.1 Different fragmentation patterns between TV and LV schemes

Figure 2 shows the normalized mass spectra of azelaic acid (a dicarboxylic acid) measured by the SP-AMS 1 for illustrating different fragmentation patterns generated using the TV and LV schemes. The comparison for azelaic acid shows that the LV approach can produce larger organic fragments (e.g., m / z 60, 69, 73, 83, and 84) compared to the TV approach. Similarly, arabitol (a sugar alcohol) and levoglucosan (a chemical marker for biomass burning OAs) show less fragmentation when they were vaporized with rBC particles by the laser vaporizer (Figs. S1 and S2). The normalized cumulative histograms of m / z for the organic compounds measured by the SP-AMS 1 are presented in Fig. 2c (i.e., the average of 18 species), which clearly shows that the curve shifts toward larger m / z when the compounds were vaporized by the LV scheme. The green dashed line represents the differences be-
tween respective fragments from the TV (blue line) and LV (red line) schemes. It shows a decreasing trend within the red shaded region, illustrating that the LV scheme could generate more organic fragments starting from \( m/z \) 55 on average. Overall, our observations support the general hypothesis and previous observations (Canagaratna et al., 2015b) that a thermal vaporizer operated at 600 °C tends to generate organic mass spectra with smaller molecular fragments compared to the LV scheme, in which organic vaporization and fragmentation can occur at a lower range of temperature, for the same organic compound. It is worth mentioning that although organo-nitrate compounds, which have significant contribution to SOA mass in some locations (Fry et al., 2018; Lee et al., 2019a; Xu et al., 2015, 2017), were not tested in this work, changes in fragmentation for ammonium nitrate (i.e., AMS calibration standard) due to the vaporization scheme were observed. Our observations show that the \( \text{NO}^+ / \text{NO}_2^+ \) ratios of ammonium nitrate measured using the TV and LV schemes are equal to 1.9 and 0.7, respectively, suggesting that the quantification of organo-nitrate compounds based on the \( \text{NO}^+ / \text{NO}_2^+ \) ratios approach can also be affected using the vaporization scheme (Farmer et al., 2010; Xu et al., 2015).

### 3.2 Changes in \( f_{\text{C}_2\text{H}_5\text{O}^+} / f_{\text{CO}_2^+} \) ratio

The organic fragments of \( \text{C}_2\text{H}_3\text{O}^+ \) and \( \text{CO}_2^+ \) are the two dominant peaks observed in ambient OOA components identified by the PMF analysis of standard HR-ToF-AMS measurements (i.e., TV scheme) (Ng et al., 2011). The relative importance of the two fragments varies between OOA components identified at the same locations, and the ratio of \( f_{\text{C}_2\text{H}_5\text{O}^+} / f_{\text{CO}_2^+} \) (\( f_i^+ = \) a mass fraction of \( m/z \) \( i^+ \) to total organic) usually decreases with the degree of oxidative aging (Ng et al., 2011, 2010). Based on the observations from SP-AMS 1 and 2, Fig. 3a shows that most of the organic species detected using the LV scheme gave higher \( f_{\text{C}_2\text{H}_5\text{O}^+} / f_{\text{CO}_2^+} \) ratios compared to the TV scheme. Note that dicarboxylic acids and multifunctional organic compounds show stronger enhancement compared to alcohols. The average of \( f_{\text{C}_2\text{H}_5\text{O}^+} / f_{\text{CO}_2^+} \) ratios measured by the TV and LV schemes are 0.48 (±0.52) and 3.07 (±3.59), respectively, indicating less thermal-induced decarboxylation with the gradual vaporization.

The observed enhancement of \( f_{\text{C}_2\text{H}_5\text{O}^+} / f_{\text{CO}_2^+} \) ratios in this work suggests that the \( f_{\text{a4}} \) vs. \( f_{\text{a3}} \) (or \( f_{\text{CO}_2^+} \) vs. \( f_{\text{C}_2\text{H}_5\text{O}^+} \)) observational framework developed based on the HR-ToF-AMS datasets worldwide by Ng et al. (2010) may not be directly applicable for evaluating the degree of aging and characteristics of OOA coatings measured using the LV scheme of the SP-AMS. Figure 3a compares the ambient OOA components determined using the TV and LV schemes (i.e., concurrent HR-ToF-AMS and SP-AMS LV scheme measurements) at the same location if their time series of mass concentrations are strongly correlated (i.e., \( R > 0.75 \)). Table S1 summarizes the characteristics of the OOA components observed in Beijing summer (Xie et al., 2019a; Xu et al., 2019); Beijing winter (Wang et al., 2019; Xie et al., 2019b); Tibet (Wang et al., 2017; Xu et al., 2018); and Fontana, CA (Chen et al., 2018; Lee et al., 2017), that are used for our comparison. It can be found that most of the OOA factors determined using the LV scheme gave significantly higher \( f_{\text{C}_2\text{H}_5\text{O}^+} / f_{\text{CO}_2^+} \) ratios compared to those determined using the TV scheme by factors of 6–18, which is similar to those observed for dicarboxylic acids and multifunctional organic compounds (Fig. 3a). Therefore, the differences in \( f_{\text{C}_2\text{H}_5\text{O}^+} / f_{\text{CO}_2^+} \) ratios between the ambient OOA components determined using the two vaporization schemes can be partially explained by the laboratory observation reported in this...
section. Nevertheless, we cannot rule out the possibility that the chemical compositions of the OOA coatings on BC particles were different than those externally mixed with BC particles despite their strong temporal correlations (see more discussion on this topic in Sect. 3.6 based on the elemental analysis).

3.3 Changes in $f_{C_{2}H_{4}O_{2}}$

$C_{2}H_{4}O_{2}$ is a tracer fragment often associated with biomass burning OAs (BBOA), and $f_{C_{2}H_{4}O_{2}}$ is commonly used to identify the presence of BBOA and to evaluate their degree of aging (Bozzetti et al., 2017; Cubison et al., 2011; Milic et al., 2017). Figure 3b shows that most of the species from SP-AMS 1 and 2 (i.e., 97 % of the tested species) showed the enhancement of $f_{C_{2}H_{4}O_{2}}$ when they were detected using the LV scheme regardless of their functional moieties. The average enhancement factor of $f_{C_{2}H_{4}O_{2}}$ for alcohol, dicarboxylic acids, and multifunctional groups are $2.62 \pm 0.92$, $2.90 \pm 0.78$, and $2.69 \pm 1.11$, respectively. Levoglucosan, a known cellulose-derived compound produced during biomass burning (Simoneit et al., 1999), gives an enhanced $f_{C_{2}H_{4}O_{2}}$ by a factor of 2.33. Different to $f_{C_{2}H_{4}O_{2}}/f_{CO_{2}}$ ratios that show scattered data between vaporization schemes, a strong linear correlation was obtained for all the tested species ($R = 0.95$) with the slope equal to 2.45. This strong linear correlation suggests that the $f_{C_{2}H_{4}O_{2}}$ increased to a similar extent for most of the tested oxygenated organic species.

Together with the changes in $f_{CO_{2}}$ caused by the vaporization schemes, our observation suggests that the $f_{44}$ vs. $f_{60}$ (or $f_{CO_{2}}$ vs. $f_{C_{2}H_{4}O_{2}}$) observational framework developed by Cubison et al. (2011) has to be used cautiously for evaluating potential influences of biomass burning emissions on the chemical composition of OOA coatings. For example, Rivellini et al. (2020) observed that the laser vaporization approach led to the enhancement of $f_{C_{2}H_{4}O_{2}}$ for total OAs, less oxidized OOA (LO-OOA), and more oxidized OOA (MO-OOA) in an urban environment by comparing their laser-off and laser-on measurements using a dual-vaporizers scheme (i.e., the SP-AMS switched between the TV scheme and TV + LV scheme during operation). Wang et al. (2019) and Xie et al. (2019b) identified BBOA factors in Beijing winter from their concurrent SP-AMS and HR-ToF-AMS measurements, respectively, and a higher $f_{C_{2}H_{4}O_{2}}$ value was observed for the BBOA factor determined using the TV scheme. However, it is important to note that the two BBOA factors were weakly correlated ($R = 0.42$), and thus they likely represented BBOA materials from different origins.

3.4 Elemental analysis of organic coating

Elemental analysis of pure OAs and organic coatings on rBC particles was performed based on the I-A method. Tables S2–S4 summarize the H : C, O : C, and OSC values of all the organic compounds measured by the three SP-AMSs. Figure 4 compares the measured H : C, O : C, and OSC values of all the organic compounds generated using the LV scheme to their true values. While the measured H : C ratios are scattered around the 1 : 1 line (Fig. 4a), the measured O : C and OSC values are generally lower than their corresponding true values (Fig. 4b and c). Note that some organic species, such as glycolic acid and glutamic acid, give relatively large discrepancies between the measured and the true O : C values that can be due to their low contribution of CHO$^+$ and CO$^+$ signals to the total organic mass (e.g., 1 %–2 %). The average relative errors of H : C and O : C ratios for individual SP-AMSs varied from −3.4 % to 11.5 % (mean = 6.6 %) and from −37.1 % to −22.0 % (mean = −26.3 %), respectively (Table 2). Note that there are no statistical differences (ANOVA, $p<0.05$) between the relative errors of elemental ratios determined by the three instruments, suggesting that the elemental analysis is not strongly instrument depen-
dent. For the thermal vaporization approach, the average relative errors of $H: C$ (mean = $-5.2\%$) and $O: C$ (mean = $-21.5\%$) ratios determined in this work are similar to the measurement uncertainties of HR-ToF-AMSs previously reported by Canagaratna et al. (2015a).

Figure 5 compares the elemental ratios of standard organic compounds determined using the LV and TV schemes. The data points from the three independent SP-AMSs are well-aligned with each other even though a large fraction of the tested organic species ($\sim 50\%$) were not repeated between the different SP-AMS instruments. The linear fits of all the measured data demonstrate that $O: C$ and $H: C$ ratios determined by the LV scheme differ from their corresponding values determined using the TV scheme by factors of 0.83 and 1.16, respectively. Canagaratna et al. (2015b) conducted similar comparisons for 10 organic species (i.e., data from SP-AMS 3 in this work) based on the results obtained from the A-A method, reporting that the $O: C$ and $H: C$ values determined using the LV scheme differ from their corresponding values determined using the TV scheme by factors of 0.83 and 1.16, respectively. The uncertainty of $O: C$ and $H: C$ ratios is further reduced to approximately $\pm 10\%$ between the TV and LV schemes on average in this study. Figure S4 shows that the I-A method with the scaling factor applied can improve the accuracy of elemental ratios of oxygenated species in general except for the $H: C$ ratios of alcohol group. While the fragmentation of oxygenated organic species due to the TV and LV scheme can be significantly different, this work illustrates that their elemental compositions can be comparable to the I-A method applied for laboratory-generated particles with a single oxygenated organic species.

### 3.5 Improved-ambient method for the LV scheme (I-ASP)

The I-A method has been widely used for the elemental analysis of ambient OAs measured using the TV scheme of HR-ToF-AMSs, and this approach has been described in detail by Canagaratna et al. (2015a). In brief, chemical standards, including dicarboxylic acids, multifunctional acids, and alcohols, were tested using the TV scheme. The elemental ratios were first determined by the A-A method developed by Aiken et al. (2008), and they were subsequently corrected by a multi-linear regression (MLR) model based on the fraction contributions of $\text{CHO}^+$ and $\text{CO}_2^+$ fragments (i.e., $f_{\text{CHO}^+}$ and $f_{\text{CO}_2^+}$) to address the composition dependence in the OA fragmentation, which is referred to as the I-A method. Note that $f_{\text{CHO}^+}$ and $f_{\text{CO}_2^+}$ can be used as surrogates for alcohol and acid groups, respectively, and they are major peaks observed in ambient OOA (Canagaratna et al., 2015a; Duplissy et al., 2011; Takegawa et al., 2007). As illustrated in the previous sections, OA fragmentation can be significantly different between the LV and TV schemes, and hence an updated multilinear regression parameter is conducted to check whether the model accuracy can be improved for analyzing data generated using the LV scheme. Following the approach for developing the I-A method (Canagaratna et al., 2015a), updated multiple linear regression parameters for determining $H: C$ and $O: C$ ratios of organic coatings were obtained based on the data from two SP-AMSs (1 and 2) as shown in Eqs. (1) and (2), respectively.

$$
\begin{align*}
\text{H}: C_{\text{I-ASP}} &= H: C_{\text{A-A}} \times [0.90 + 1.02 \times f_{\text{CHO}^+} + 2.78 \times f_{\text{CO}_2^+}], \\
\text{O}: C_{\text{I-ASP}} &= O: C_{\text{A-A}} \times [1.74 - 2.50 \times f_{\text{CHO}^+} + 1.93 \times f_{\text{CO}_2^+}],
\end{align*}
$$

where $H: C_{\text{I-ASP}}$ and $O: C_{\text{I-ASP}}$ are the elemental ratios obtained from the new fitting parameters (denoted as I-ASP method hereafter) and $H: C_{\text{A-A}}$ and $O: C_{\text{A-A}}$ are the elemental ratios obtained by the A-A method.

Given that most of the tested organic species were not the same for each SP-AMS, a direct inter-instrument comparison was not possible. Hence the average relative errors obtained from the two instruments were used to evaluate the performance of the I-ASP method across all the tested species. The relative errors of I-ASP method for $H: C$, $O: C$, and $\text{OS}_C$ were $6.3\%$, $5.8\%$, and $-9.8\%$, respectively, for the LV scheme data. While the I-ASP method leads to a substantial improvement of the relative errors of $O: C$ compared to the I-A method (i.e., from $-26.3\%$ to $5.8\%$), the relative average errors of $H: C$ ratio obtained from I-ASP and I-A method are comparable. As shown in Fig. 4d–f, the $H: C$, $O: C$, and $\text{OS}_C$ values calculated using the I-ASP method are better aligned with the 1:1 line compared to those determined using the I-A method (Fig. 4a–c) with smaller root mean squared error (RMSE) reported in Table S5. However, it is worth noting that the I-ASP method gives the highest positive bias for the $O: C$ ratio for alcohol species compared to the results from the I-A method (with and without applying the scaling factor) as illustrated in Fig. S4. For the $H: C$ ratios, the I-ASP method can reduce relative error for alcohol species but generate a larger range of errors for dicarboxylic acids and multifunctional species.

### 3.6 Insight into ambient OOA characteristics

There is an increasing number of field studies operating a standard HR-ToF-AMS and an SP-AMS concurrently (i.e., total OAs measured by the TV scheme vs. organic coatings measured using the LV scheme) to investigate the mixing state of BC particles and the effects of primary emissions and atmospheric processing on the formation of organic coatings on BC particles (Lee et al., 2017; Massoli et al., 2015; Wang et al., 2020). In particular, whether SOA materials condensed on BC particles have similar chemical characteristics to those externally mixed with BC particles remains poorly understood. Previous observations in urban environments have reported that the mass spectral features of ambient OOA components identified by the PMF analysis were
Table 2. Relative errors of H : C, O : C, and OS\text{C} measured by the three independent SP-AMSs.

<table>
<thead>
<tr>
<th>Vaporization scheme</th>
<th>SP-AMS Elemental analysis method</th>
<th>Average relative errors (%) H : C</th>
<th>O : C</th>
<th>OS\text{C}</th>
</tr>
</thead>
<tbody>
<tr>
<td>Laser vaporization</td>
<td>1 I-A</td>
<td>8.5 %</td>
<td>-22.0 %</td>
<td>-33.2 %</td>
</tr>
<tr>
<td></td>
<td>2 I-A</td>
<td>11.5 %</td>
<td>-23.3 %</td>
<td>-18.3 %</td>
</tr>
<tr>
<td></td>
<td>3 I-A</td>
<td>-3.4 %</td>
<td>-37.1 %</td>
<td>-36.9 %</td>
</tr>
<tr>
<td></td>
<td>1, 2 &amp; 3 I-A</td>
<td>6.6 %</td>
<td>-26.3 %</td>
<td>-27.2 %</td>
</tr>
<tr>
<td></td>
<td>1 &amp; 2 I-A</td>
<td>6.3 %</td>
<td>5.8 %</td>
<td>-9.8 %</td>
</tr>
<tr>
<td>Thermal vaporization</td>
<td>1 I-A</td>
<td>-0.6 %</td>
<td>-21.5 %</td>
<td>-46.7 %</td>
</tr>
<tr>
<td></td>
<td>2 I-A</td>
<td>-5.2 %</td>
<td>-17.8 %</td>
<td>-49.5 %</td>
</tr>
<tr>
<td></td>
<td>3 I-A</td>
<td>-12.0 %</td>
<td>-27.0 %</td>
<td>-32.0 %</td>
</tr>
<tr>
<td></td>
<td>1, 2 &amp; 3 I-A</td>
<td>-5.2 %</td>
<td>-21.5 %</td>
<td>-44.3 %</td>
</tr>
</tbody>
</table>

Figure 4. Comparisons between the measured and true values of H : C, O : C, and OS\text{C} determined using the three SP-AMSs using the LV scheme. Panels (a–c) and (d–f) show the results calculated using the I-A method and I-A\text{SP} method, respectively. Red circles, blue triangles, and green crosses represent data measured by SP-AMS 1, 2, and 3, respectively. The dashed lines represent 1 : 1 line.

significantly different between the two co-located measurements even though their temporal variabilities strongly correlated to each other (Chen et al., 2018; Lee et al., 2017; Liu et al., 2019; Massoli et al., 2015; Wang et al., 2020; Xu et al., 2019; Zhao et al., 2019). Given our observations that vaporization scheme plays a critical role in the fragmentation process of oxygenated organic species, the LV elemental analysis scaling factors (0.89 for H : C and 1.10 for O : C) and the I-A\text{SP} method obtained in this work can facilitate more direct and robust comparison between the two types of measurements based on elemental analysis. Nevertheless, it is important to note that rBC and organics were likely more homogeneously mixed within our laboratory-generated particles, which can be very different from the morphology of ambi-
ent organically coated BC particles. Therefore, such inter-
instrument comparisons assume that BC morphology is not a
key factor to affect organic fragmentation observed using the
LV scheme.

Figure 6 shows the elemental ratios of OOA components
observed from previous field studies conducted in California
Research at the Nexus of Air Quality and Climate Change
(CalNex) 2010 campaign (Massoli et al., 2015), Fontana,
California, in 2015 (Chen et al., 2018; Lee et al., 2017), Tibet
in 2015 (Wang et al., 2016; Lu et al., 2018; Xu et al., 2018), Beijing winter
in 2016 (Wang et al., 2019; Xie et al., 2019b), and Beijing
summer in 2017 (Wang et al., 2020; Xie et al., 2019a; Xu
et al., 2019). Correlations of hourly-averaged mass concen-
trations between different PMF factors identified by the TV
and LV schemes were investigated, and only strongly corre-
lated OOA factors (R > 0.85) were included in Fig. 6. If an
OOA component correlated well to multiple PMF OA fac-
tors identified by another vaporization scheme, the compar-
ison was only performed for a pair of OA factors that gave
the strongest correlation (see Table S1). Note that a trans-
ported BBOA factor identified in Tibet, which was associ-
atied with a significant amount of OOA materials (R = 0.96),
was also included in this comparison. Figure 6a and b show
the adjusted H : C and O : C ratios measured using the LV
scheme (y axis) using the inter-conversion factor and the I-
Asp method, respectively. The error bars of each data point
represent the average absolute errors of the I-A and I-A
method obtained in this study. Both methods likely over-
estimate O : C ratios of organics coated on ambient BC particles because of the contributions of refractory COx+ fragments (e.g., CO+ and COx2+) usually
remain unresolved.

In Fig. 6a, although a few H : C and O : C ratios were well-
aligned onto the 1 : 1 line, it can be found that the majority of
O : C ratios measured by the LV scheme with the LV el-
emental analysis scaling factors applied (i.e., LV-OOA from
CalNex, OOA-2 from Fontana, BBOA from Tibet, OOA-1,
OOA-2 from Beijing winter and OOA-1 from Beijing sum-
mer) were still lower than those measured using the TV
scheme after considering the uncertainties of I-A method.
In addition, some H : C ratios of OOA coatings were higher
than the OOAs measured using the TV scheme. When the
I-Asp method was used, the O : C ratios generally increased
compared to those determined using the I-A method with the
inter-conversion factors applied (Figs. 6b and S3). In partic-
ular, LO-OOA from CalNex gave the largest enhancement of
O : C ratio, followed by OOA-2 from Fontana and SV-OOA
from CalNex. The H : C ratios remained roughly the same
between the two methods. Combining the results presented in
Fig. 6a and b, it can be concluded that the OOA materials
associated with rBC particles were likely less oxygenated (i.e.,
lower O : C) compared to the total OOAs measured using the
TV scheme at some of the sampling locations. This observa-
tion may be due to the fact that OOA materials formed from hetero-
geneous oxidation of POA such as HOA that are co-emitted
and internally mixed with BC are likely to be less and/or non-
oxigenated in nature. For example, our comparison shows
the largest differences of O : C ratios between the two va-
porization schemes for the observations from Beijing (OOA-
2 in winter and OOA-1 in summer), where the air quality
was expected to be significantly influenced by local combus-
tion sources. The O : C ratio of an aged/transported BBOA
factor detected using the LV scheme of the SP-AMS in Ti-
bet is also noticeably lower than that of MO-OOA detected
by the HR-ToF-AMS even though they are strongly corre-
lated (R = 0.96). Overall, our observations indicate that even
though the time series of OOA factors determined by the TV
and LV scheme are strongly correlated (e.g., R > 0.9), sug-
gesting that they were likely co-emitted or formed through
similar aging processes during transport, they might contain
multiple types of OA materials and their relative distribution
between rBC and non-BC particles might be significantly dif-
f
Comparisons between ambient OOA data measured by co-located SP-AMSs and HR-ToF-AMSs in the Nexus of Air Quality and Climate Change (CalNex) 2010 campaign (Massoli et al., 2015); Fontana, California, in 2015 (Chen et al., 2018; Lee et al., 2017); Tibet in 2015 (Wang et al., 2017; Xu et al., 2018); and Beijing in 2016 (Wang et al., 2019; Xie et al., 2019b) and 2017 (Wang et al., 2020; Xie et al., 2019a; Xu et al., 2019) (see Table S1 for the detail of ambient data comparison). For (a), the H : C and O : C ratios determined by the laser vaporization approached were corrected by the LV elemental analysis scaling factors (1.10 for H : C and 0.89 for O : C). For (b), the H : C and O : C ratios were determined using the I-ASP method. The LV scheme of the SP-AMS was used to detect rBC-containing particles exclusively. The error bar represents the average absolute errors of I-A (TV in a and b, and LV in a) and I-ASP (LV in b) for determining the H : C and O : C ratios. The dashed black line represents the 1 : 1 line and the dashed grey line indicates ±0.25 of 1 : 1 line.

4 Summary

Elemental ratios, in particularly H : C and O : C ratios, have been widely used to investigate the chemical properties of OAs such as particle viscosity (Chen et al., 2011; Zhang et al., 2015), particle phase transition (Pye et al., 2017), aromatic structure or sorption properties (Xiao et al., 2016), light absorption properties (Kumar et al., 2018), and hygroscopicity (Massoli et al., 2010) in many field and laboratory studies. Enhancing the accuracy of elemental analysis of OAs is important to improve understanding of their physiochemical properties and aging mechanisms. Although the I-A method has been widely utilized to quantify H : C and O : C ratios of OAs measured by a standard HR-ToF-AMS (i.e., TV scheme), the applicability of the I-A method for the elemental analysis of organic coatings that are measured using the LV scheme of the SP-AMS remains uncertain, especially for ambient OOA components that are always referred to as freshly formed and aged SOA materials based on their degree of oxygenation.

To address this knowledge gap, this work examined 30 oxygenated organic species with different functional moieties, which were characterized by both TV and LV schemes of three SP-AMS instruments operated in different laboratories. The results demonstrate that the LV scheme can retain larger fragment ions during OA fragmentation compared to the TV scheme. Changes in OA fragmentation due to the LV scheme can significantly impact the $f_{C_2H_5O^+}/f_{CO_2^+}$ ratio and $f_{C_2H_4O_3}$ of organic mass spectra. Therefore, the application of the observational-based framework developed based on these three organic fragments may not be straightforward for evaluating the chemical characteristics and aging of SOA and BBOA materials coated on ambient BC particles (Cubison et al., 2011; Ng et al., 2010). The I-A method is robust for determining elemental compositions of OOA materials detected by both TV and LV schemes, and the LV elemental analysis scaling factors of 1.10 and 0.89 for H : C and O : C ratios, respectively, were determined to further improve the accuracy. The I-ASP method is developed in this work based on the updated multilinear regression model for the LV scheme measurements. Compared to the I-A method, the I-ASP method can further reduce the relative errors of O : C ratio, which were from −26.3 % to 5.8 % on average for our tested species, and the average of relative errors for H : C ratios remain roughly the same. Nevertheless, it is worth noting that the I-ASP method may overestimate the O : C ratio of alcohol species and lead to more scattered H : C ratios for dicarboxylic acids and multifunctional species. Applying the LV elemental analysis scaling factors and the I-ASP method to ambient data, this work demonstrates that the formation mechanisms and chemical characteristics of OOA coatings on BC particles can be different than OOA materials externally mixed with BC at the same location. Lastly, a significant difference between the NO$^+$/NO$_2^+$ ratios for ammonium nitrate particles measured by the TV and LV scheme was observed in this work. As the NO$^+$/NO$_2^+$ ratios have been widely used for quantifying particle-phase organonitrate compounds, this observation suggests that changes in fragmentation can have an impact on quantifying organonitrate compounds coated on BC.
Data availability. The dataset for this publication is available upon contacting the corresponding authors.

Supplement. The supplement related to this article is available online at: https://doi.org/10.5194/amt-14-2799-2021-supplement.

Author contributions. AKYL supervised the project. YXC, RW, MDW, and MRC carried out the experiments. MM, LHR, RW, MDW, and MRC analyzed the lab data. MM, JW, and XG analyzed the field data. MM and AKYL wrote the paper with support and comments from all the co-authors.

Competing interests. The authors declare that they have no conflict of interest.

Acknowledgements. The authors would like to thank Max G. Adam for his assistance on initial stage of the experiments conducted at the National University of Singapore.

Financial support. This work is supported by the National Environmental Agency (NEA) of Singapore (NEA, R-706-000-043-490). The content does not represent NEA’s view. Measurements conducted at the University of Toronto were supported by the Natural Science and Engineering Research Council (NSERC) of Canada.

Review statement. This paper was edited by Mingjin Tang and reviewed by three anonymous referees.

References

Collier, S., Williams, L. R., Onasch, T. B., Cappa, C. D., Zhang, X., Russell, L. M., Chen, C.-L., Sanchez, K. J., Worsnop, D. R., and


