



Satellite remote-sensing capability to assess tropospheric-column ratios of formaldehyde and nitrogen dioxide: case study during the Long Island Sound Tropospheric Ozone Study 2018 (LISTOS 2018) field campaign

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Abstract. Satellite retrievals of tropospheric-column formaldehyde (HCHO) and nitrogen dioxide (NO₂) are frequently used to investigate the sensitivity of ozone (O₃) production to emissions of nitrogen oxides and volatile organic carbon compounds. This study inter-compared the systematic biases and uncertainties in retrievals of NO₂ and HCHO, as well as resulting HCHO–NO₂ ratios (FNRs), from two commonly applied satellite sensors to investigate O₃ production sensitivities (Ozone Monitoring Instrument, OMI, and TROPOspheric Monitoring Instrument, TROPOMI) using airborne remote-sensing data taken during the Long Island Sound Tropospheric Ozone Study 2018 between 25 June and 6 September 2018. Compared to aircraft-based HCHO and NO₂ observations, the accuracy of OMI and TROPOMI were magnitude-dependent with high biases in clean environments and a tendency towards more accurate comparisons to even low biases in moderately polluted to polluted regions. OMI and TROPOMI NO₂ systematic biases were similar in magnitude (normalized median bias, NMB = 5 %–6 %; linear regression

slope \approx 0.5–0.6), with OMI having a high median bias and TROPOMI resulting in small low biases. Campaign-averaged uncertainties in the three satellite retrievals (NASA OMI; Quality Assurance for Essential Climate Variables, QA4ECV OMI; and TROPOMI) of NO₂ were generally similar, with TROPOMI retrievals having slightly less spread in the data compared to OMI. The three satellite products differed more when evaluating HCHO retrievals. Campaign-averaged tropospheric HCHO retrievals all had linear regression slopes \sim 0.5 and NMBs of 39 %, 17 %, 13 %, and 23 % for NASA OMI, QA4ECV OMI, and TROPOMI at finer (0.05° × 0.05°) and coarser (0.15° × 0.15°) spatial resolution, respectively. Campaign-averaged uncertainty values (root mean square error, RMSE) in NASA and QA4ECV OMI HCHO retrievals were \sim 9.0 × 10¹⁵ molecules cm⁻² (\sim 50 %–55 % of mean column abundance), and the higher-spatial-resolution retrievals from TROPOMI resulted in RMSE values \sim 30 % lower. Spatially averaging TROPOMI tropospheric-column HCHO, along with NO₂ and FNRs, to resolutions similar to the OMI reduced the uncertainty in

these retrievals. Systematic biases in OMI and TROPOMI NO₂ and HCHO retrievals tended to cancel out, resulting in all three satellite products comparing well to observed FNRs. However, while satellite-derived FNRs had minimal campaign-averaged median biases, unresolved errors in the indicator species did not cancel out in FNR calculations, resulting in large RMSE values compared to observations. Uncertainties in HCHO retrievals were determined to drive the unresolved biases in FNR retrievals.

1 Introduction

Tropospheric ozone (O₃) is a harmful pollutant, and near-surface concentrations of this species have detrimental impacts on human and environmental health (Kampa and Castanas, 2008; Van Dingenen et al., 2009). The production and destruction rates of tropospheric O₃ are controlled by complex chemical reactions involving the primary precursor species of nitrogen oxides (NO_x: nitric oxide and nitrogen dioxide (NO + NO₂)) and volatile organic compounds (VOCs) (Sillman, 1999; Lelieveld and Dentener, 2000). It is critical to understand precursor species emissions and subsequent atmospheric chemistry controlling surface-level O₃ production rates since the United States (US) Environmental Protection Agency (EPA) enforces concentration limits of criteria pollutants (e.g., O₃, NO₂, carbon monoxide, particulate matter, and sulfur dioxide) under the National Ambient Air Quality Standards (NAAQS) (US EPA, 2015). To reduce and maintain surface-level O₃ concentrations below NAAQS thresholds, many regions have implemented emission control strategies for precursor species. To design effective emission reduction strategies, knowledge about the non-linear sensitivity of O₃ formation to NO_x and VOCs is critical (Crutzen, 1973; Sillman, 1999). Based on the relative concentrations of NO_x and VOCs, O₃ formation is sensitive to perturbations of either NO_x (NO_x-limited regimes) or VOC emissions (NO_x-saturated or VOC- or radical-limited regimes). These O₃ sensitivity regimes are separated by a transitional regime where O₃ formation is sensitive to both NO_x and VOC emissions.

To understand the non-linear relationship of O₃ formation to NO_x and VOC emissions in complex chemical environments (e.g., polluted regions and areas of heterogeneous emissions of NO_x and VOCs), spatiotemporally dense in situ measurements or airborne remote-sensing observations of precursor species are desired (e.g., Souri et al., 2020). Since these measurements are often spatiotemporally sparse, satellite retrievals of chemical proxies for NO_x (i.e., NO₂) and VOCs (i.e., formaldehyde, HCHO) have been demonstrated to provide insight into the O₃–NO_x–VOC relationship (Tonnesen and Dennis, 2000; Martin et al., 2004; Duncan et al., 2010; Souri et al., 2017; Jin et al., 2017, 2020). The HCHO–NO₂ concentration ratio (hereinafter FNR) has been shown to provide information to monitor the local sensitivity

of O₃ production from the chemical loss of HO₂ + RO₂ and chemical loss of NO_x controlling O₃–NO_x–VOC chemistry (Tonnesen and Dennis, 2000; Kleinman et al., 2001).

Multiple past and current space-based spectrometers have the capability to retrieve simultaneous NO₂ and HCHO tropospheric columns to calculate FNRs including the Global Ozone Monitoring Experiment (GOME; Martin et al., 2004), GOME-2 (Choi et al., 2012), Ozone Monitoring Instrument (OMI; Duncan et al., 2010), SCanning Imaging Absorption spectroMeter for Atmospheric CHartographY (SCIAMACHY; Jin et al., 2020), and TROPOspheric Monitoring Instrument (TROPOMI; Chan et al., 2020; Souri et al., 2021). In addition to these low Earth orbiting (LEO) satellites, Tropospheric Emissions: Monitoring of Pollution (TEMPO) is a upcoming National Aeronautics and Space Administration (NASA) geostationary satellite mission which will retrieve hourly NO₂ and HCHO tropospheric columns over North America (Zoogman et al., 2017; Chance et al., 2019). This geostationary sensor is part of a constellation of air quality spaceborne sensors including the Geostationary Environment Monitoring Spectrometer (GEMS) instrument onboard the Korea Aerospace Research Institute GEO-KOMPSAT-2B satellite (Kim et al., 2020) and the European Space Agency (ESA) Sentinel-4 mission (ESA, 2017). Satellite retrievals of NO₂ and HCHO have been applied to determine the sensitivity of O₃ formation to NO_x and VOC emissions at coarse spatial and temporal scales (e.g., Martin et al., 2004; Duncan et al., 2010) to finer spatiotemporal scales and focusing on long-term trends (e.g., Choi et al., 2012; Jin and Holloway, 2015; Choi and Souri, 2015; Schroeder et al., 2017; Souri et al., 2017; Jin et al., 2017, 2020). However, uncertainties remain in how accurately satellites can retrieve information needed to study planetary boundary layer (PBL) O₃–NO_x–VOC relationships. These uncertainties stem from (a) the exact thresholds of FNRs that separate NO_x-limited, transition, and VOC-limited regimes; (b) the ability of tropospheric-column retrievals to represent PBL chemical composition due to variability in the vertical structure of NO₂ and HCHO concentrations and satellite sensitivity throughout the entire troposphere; (c) whether HCHO is an effective proxy for total VOC reactivity; (d) satellite spatial representation errors; and (e) the accuracy and uncertainty in satellite retrievals of tropospheric-column HCHO and NO₂. Of all these sources of uncertainty, systematic and random biases due to noise in satellite retrievals may be the largest source of error for retrieving FNRs (Souri et al., 2023).

This study is designed to demonstrate the effectiveness of two frequently applied satellites for evaluating O₃–NO_x–VOC relationships (i.e., OMI and TROPOMI) to accurately retrieve tropospheric HCHO and NO₂ columns and FNRs. OMI and TROPOMI retrievals have been evaluated in numerous studies (e.g., Judd et al., 2020; Vigouroux et al., 2020; Zhu et al., 2020; Lamsal et al., 2021), typically focusing on a specific sensor and species (e.g., evaluating OMI or TROPOMI and NO₂ or HCHO separately), however, not

for the accuracy to retrieve tropospheric-column FNRs. Here we validate OMI and TROPOMI retrievals of HCHO and NO₂, as well as subsequent FNRs, with airborne spectrometer data obtained during the Long Island Sound Tropospheric Ozone Study 2018 (LISTOS 2018) field campaign conducted during the summer of 2018 in the northeastern region of the US. Furthermore, this work investigates the capability of OMI and TROPOMI to capture the spatiotemporal variability in observed FNRs and discusses the possible causes of systematic error and uncertainties in these retrievals.

2 Methods

This study focuses on the spatial domain and time period (25 June to 6 September 2018) of the LISTOS 2018 field campaign (<https://www.nescaum.org/documents/listos>, last access: 8 May 2023). This campaign was chosen due to the overlap of the TROPOMI and OMI missions; the availability of airborne spectrometer retrievals (i.e., Geostationary Trace gas and Aerosol Sensor Optimization, GeoTASO, and GEOstationary Coastal and Air Pollution Events (GEOCAPE) Airborne Simulator, GCAS) of tropospheric-column HCHO and NO₂, which are effective satellite validation data (e.g., Judd et al., 2020); and the large spatiotemporal coverage of the airborne spectrometer data. Studies have applied stationary sources of ground-based remote-sensing data to validate OMI and TROPOMI (e.g., multi-axis differential optical absorption spectroscopy, MAX-DOAS; Fourier-transform infrared, FTIR; and Pandora); however, using the airborne data allows for the evaluation of satellite retrievals in variable environments (i.e., clean to heterogeneous and polluted regions).

2.1 OMI remote-sensing products

The nadir-viewing OMI sensor, on the polar-orbiting NASA Aura satellite launched in 2004, is an ultraviolet–visible (UV–Vis) spectrometer (Levelt et al., 2006). Retrievals are made from three wavelength channels between 260 and 510 nm (UV-1: 264 to 311 nm, UV-2: 307 to 383 nm, Vis: 349 to 504 nm). Aura OMI has a local equatorial overpass time of $\sim 13:45$ LT (local time) with nearly complete daily global surface coverage due to the large ~ 2600 km swath width. Level-2 (L2) tropospheric vertical column density (VCD) OMI NO₂ retrievals from the NASA version 4 standard product (OMNO2; Lamsal et al., 2021) and the NASA operational OMI HCHO version 3 product using the Smithsonian Astrophysical Observatory (SAO) retrieval algorithm (OMHCHO; González Abad et al., 2015, 2016) were applied in this study. To investigate the impact of different retrieval algorithms, we also apply tropospheric-column OMI NO₂ and HCHO data derived in the Quality Assurance for Essential Climate Variables (QA4ECV) project.

Starting in 2007, OMI experienced a field-of-view blockage known as the “row anomaly”, which affects the data quality at all retrieval wavelengths for some rows (Dobber et al., 2008; Schenkeveld et al., 2017). The row anomaly in NO₂ and HCHO retrievals was avoided using data quality flags to filter out pixels flagged by the row anomaly detection algorithm. OMI also has systematically biased retrievals in a striped pattern running in 60 cross-track fields of view. A “destriping” correction is applied to NO₂ data (Boersma et al., 2011), and the reference sector method corrects this artifact in HCHO data (De Smedt et al., 2015; González Abad et al., 2015; Zara et al., 2018).

2.1.1 OMI – NASA OMNO2 and OMHCHO

The primary OMI data used in this study are the L2 tropospheric VCD OMNO2 and OMHCHO retrievals provided at ~ 13 km \times 24 km near the nadir to ~ 24 km \times 160 km towards the edge of the swath. Lamsal et al. (2021) describe the OMNO2 retrieval algorithm (referred to as NASA OMI NO₂) in detail, and it is explained here briefly. The NASA OMI NO₂ retrieval uses a differential optical absorption spectroscopy (DOAS) approach, with a fitting window between 405 and 465 nm, to derive slant column densities (SCDs). Tropospheric NO₂ columns are separated from the entire atmospheric column using an observation-based stratosphere–troposphere separation scheme (Bucsela et al., 2013). Tropospheric SCDs are converted to tropospheric VCDs using an air mass factor (AMF) calculated with a radiative transfer model and simulated atmospheres from a chemical transport model (CTM). Specifics for the data used in AMF calculations for NASA OMI NO₂ are presented in Table S1 in the Supplement. Tropospheric AMFs are calculated in NASA OMI NO₂ retrievals using monthly averaged a priori profiles from the NASA Global Modeling Initiative (GMI) model at $1.0^\circ \times 1.25^\circ$ spatial resolution, clouds from the OMI O₂–O₂ algorithm (Acarreta et al., 2004; Veefkind et al., 2016; Vasilkov et al., 2018), and surface albedo from geometry-dependent surface Lambertian-equivalent reflectivity (GLER) data (Vasilkov et al., 2017; Qin et al., 2019; Fasnacht et al., 2019). The uncertainty in the NASA OMI NO₂ product has been shown to vary with cloudiness and pollution levels and is in the range of $\sim 20\%$ to $\sim 60\%$ (Bucsela et al., 2013), with contributions from errors in spectral fitting, stratospheric correction, and AMF calculations.

González Abad et al. (2015, 2016) describe the OMHCHO retrieval algorithm in detail (referred to here as NASA OMI HCHO). Retrievals of HCHO SCDs are obtained by spectrally fitting OMI radiances using the basic optical absorption spectroscopy (BOAS) method (Chance, 1998) with a fitting window between 328.5 and 356.5 nm. HCHO SCDs are converted to VCDs, applying AMFs using GEOS-Chem (Goddard Earth Observing System) a priori profiles at $2.0^\circ \times 2.5^\circ$ spatial resolution, cloud information (Martin et al., 2002;

Acarreta et al., 2004), and surface albedo data (Kleipool et al., 2008). Information about the input data for NASA OMI HCHO AMF calculations is presented in Table S1. Finally, postprocessing across-track bias corrections are applied by comparing daily HCHO VCDs with VCDs simulated with the GEOS-Chem CTM over a clean region (known as the reference sector). NASA OMI HCHO uncertainties have been shown to vary with pollution levels ranging from ~ 45 % to ~ 105 %, with the largest contributions from the spectral fitting and AMF calculations (González Abad et al., 2015, 2016).

2.1.2 OMI – QA4ECV NO₂ and HCHO

We also evaluated OMI NO₂ and HCHO retrievals from the QA4ECV project (<https://knmi.sitearchief.nl/?subsite=qa4ecv#archive>, last access: 8 May 2023). Retrievals from the QA4ECV NO₂ version 1.1 and QA4ECV HCHO version 1.2 data products are provided daily at the same spatial resolution as the NASA OMI products. Zara et al. (2018) describe the QA4ECV OMI NO₂ and HCHO slant column retrievals, and Boersma et al. (2018) and De Smedt et al. (2018) describe the entire QA4ECV OMI NO₂ and HCHO retrieval algorithms, respectively, in detail. They are summarized here briefly.

QA4ECV retrievals of NO₂ SCDs are obtained by linear fits of optical depths to the observed optical depth using the DOAS technique with a fitting window between 405 and 465 nm (Boersma et al., 2018). The QA4ECV NO₂ retrieval differs from the NASA OMI NO₂ retrieval in some of the retrieval steps (Compernelle et al., 2020). To calculate tropospheric AMFs, the QA4ECV NO₂ retrieval algorithm uses the surface albedo (Kleipool et al., 2008) and cloud products (Veefkind et al., 2016) from the previous NASA OMI NO₂ version 3 product (see Lamsal et al., 2021), however, using a priori profiles from the Tracer Model version 5 (TM5) CTM at 1.0° × 1.0°. Tropospheric VCDs are separated from the entire column using the global TM5 assimilation model. For detailed information on the differences in spectral fitting between the NASA OMI NO₂ and QA4ECV NO₂ retrieval algorithms, we refer you to Zara et al. (2018). For details about differences between AMF calculations in the NASA and QA4ECV OMI algorithms, see Lorente et al. (2017). QA4ECV NO₂ data have been shown to perform relatively well in clean to moderately polluted regions and have a low bias in highly polluted regions (Compernelle et al., 2020). Retrievals of QA4ECV HCHO SCDs are conducted in a manner similar to QA4ECV NO₂ using the DOAS technique and optical depths with a fitting window between 328.5 and 359.0 nm (Zara et al., 2018; De Smedt et al., 2018). For information about the inputs used in AMF calculations for QA4ECV OMI NO₂ and HCHO retrievals, see Table S1. QA4ECV HCHO retrievals show minimal bias in clean to moderately polluted regions and low biases in polluted locations (e.g., De Smedt et al., 2021).

2.2 TROPOMI remote-sensing products

The TROPOMI hyperspectral spectrometer (including eight bands in the UV, VIS, near-infrared, and shortwave infrared wavelengths) is on the Sentinel-5 Precursor (S5P) satellite which was launched in October 2017. TROPOMI is in orbit with a local equatorial overpass time (~ 13:30 local time) similar to OMI. TROPOMI has a swath width of ~ 2600 km and a ground pixel size of 3.5 km × 7.0 km at nadir during LISTOS 2018 (since 6 August 2019 TROPOMI data have been available at 3.5 km × 5.5 km), which is > 12 times finer than OMI. TROPOMI retrievals have been used in numerous recent studies investigating processes controlling NO₂ concentrations and trends (e.g., Goldberg et al., 2021) and FNRs (Wu et al., 2022), taking advantage of the high spatial resolution of the sensor, along with being validated thoroughly (e.g., Judd et al., 2020; De Smedt et al., 2021). The high-spatial-resolution information provided by TROPOMI, compared to past UV–Vis spaceborne sensors, reduces the representation error in each retrieved NO₂ and HCHO pixel (Souri et al., 2022). In this study, we apply daily TROPOMI tropospheric-column NO₂ v2.3.1 (van Geffen et al., 2022) and HCHO v1.1.5 retrievals (De Smedt et al., 2018). For TROPOMI NO₂ data we used the product provided by the Product Algorithm Laboratory (PAL). The retrievals of both species use QA4ECV methods described above with spectral fitting windows between 405.0 and 465.0 nm for NO₂ (Boersma et al., 2018) and 328.5 and 359.0 nm for HCHO (De Smedt et al., 2018). TROPOMI retrievals are similar to those from the QA4ECV OMI product as it applies the same a priori profiles, albedo data, and cloud fraction information. TROPOMI NO₂ v2.3.1 retrievals do differ from QA4ECV OMI NO₂ products as it uses cloud pressure input from the O₂–A band following the FRESKO+ (Fast Retrieval Scheme for Clouds from the Oxygen A band) wide approach (van Geffen et al., 2022). TROPOMI HCHO v1.1.5 retrievals differ from the QA4ECV OMI HCHO data by applying the S5P ROCINN (Retrieval of Cloud Information using Neural Networks) algorithm which uses the O₂–A for cloud pressures (Loyola et al., 2018). For more information about the input data used in AMF calculations for TROPOMI NO₂ and HCHO retrievals, see Table S1.

2.3 Airborne spectrometers

The primary evaluation data set used in this study is from the UV–Vis airborne remote-sensing data product from GeoTASO and GCAS. Since no bias-corrected tropospheric-column HCHO data are available during LISTOS 2018 from the Pandora network, this ground-based remote-sensing network is not applied here. Both the GeoTASO and GCAS instruments and retrievals are very similar and together provide a consistent evaluation data set (see details on the instruments and NO₂ and HCHO retrievals in Kowalewski and Janz, 2014; Leitch et al., 2014; Nowlan et al., 2016, 2018; and Judd

et al., 2020). GeoTASO and GCAS data were obtained from a nominal flight altitude of 9 km above ground level (a.g.l.) covering the majority of the troposphere. Airborne data from 13 flight days during LISTOS 2018 (see Table 1) are provided with a native spatial resolution of 250 m × 250 m. To reduce noise in the raw GeoTASO and GCAS retrievals, the data were averaged to a 1 km × 1 km spatial resolution. In total, measurements from 8 and 12 flight days were spatiotemporally co-located with OMI and TROPOMI overpasses, respectively.

The airborne GeoTASO and GCAS retrievals used as the reference data set for validating all satellite data are not without error. A nearly identical airborne NO₂ data set used here was applied in Judd et al. (2020) and was evaluated with Pandora systems. Judd et al. (2020) demonstrated that the airborne NO₂ retrievals had a median bias of ~ 1 %, with no magnitude-dependent biases, and uncertainty within ±25 %. Due to limited availability of Pandora HCHO data, airborne GeoTASO and GCAS retrievals of this species have had minimal evaluation. Nowlan et al. (2018) did evaluate GCAS tropospheric HCHO retrievals using airborne in situ measurements and determined GCAS had generally good performance with a < 10 % bias (minimal magnitude dependence in bias) and high correlation with observations. Overall, the satisfactory comparison of airborne GeoTASO and GCAS tropospheric-column NO₂ and HCHO with independent observations provides confidence that this data can be applied as a reference data set to validate OMI and TROPOMI. However, it should be kept in mind that there is some level of error associated with the GeoTASO and GCAS data (e.g., Nowlan et al., 2016, 2018; Judd et al., 2020).

The GeoTASO and GCAS data taken during LISTOS 2018 provided a novel opportunity to use airborne observations to validate both OMI and TROPOMI coincidentally. This airborne data differ from many of the recent satellite validation studies which use longer-term information from networks of point-source measurements (e.g., Pandora, MAX-DOAS) (e.g., Compernelle et al., 2020; Vigouroux et al., 2020; Verhoelst et al., 2021; Lamsal et al., 2021; Sourì et al., 2023). The airborne sensors allowed for evaluation of OMI and TROPOMI over large areas which equates to having tens to hundreds of clustered ground-based systems on each flight day. Having long-term observations for robust temporal validation of satellite sensors is ideal; however, this case study is unique in that it provides information about the performance of coincident retrievals from OMI and TROPOMI over variable emission source regions (urban to rural areas) and scenes with differing geophysical characteristics (e.g., surface albedo, tropospheric compositions, clouds, aerosol amounts) during a single flight, which is another novel aspect. Even though there are limited observations available from the flights in LISTOS 2018 (Table 1), all correlation statistics presented in this study (Tables 2 and 3) are significant to a 95 % confidence interval (p value < 0.05).

2.4 Community Multiscale Air Quality Model (CMAQ) model simulation

Prior vertical profiles play a major role in satellite retrievals of tropospheric chemistry (e.g., Palmer et al., 2001; Boersma et al., 2007; Johnson et al., 2018). Past research has demonstrated that using a well-constrained, high-spatial-resolution air quality model as the a priori profile source for satellite retrievals can improve VCD results (e.g., Laughner et al., 2019). To compare NASA OMI and TROPOMI tropospheric NO₂, HCHO, and FNR retrievals using a common a priori profile data set, we conduct sensitivity tests using high-spatial-resolution (4 km × 4 km) model-simulated vertical profiles of NO₂ and HCHO from the Community Multiscale Air Quality Model (CMAQ) (version 5.3) to reprocess these satellite retrievals.

CMAQ simulations were driven offline using the meteorological fields simulated by the Weather Research and Forecasting (WRF) model (version 4.1). The WRF-CMAQ spatial domain setup is shown in Fig. S1. The outer WRF domain covers the contiguous United States (CONUS) at a horizontal grid spacing of 12 km × 12 km, and the inner domain, encompassing the entire LISTOS 2018 domain, has a horizontal grid spacing of 4 km × 4 km. Both the outer and inner model domains use 35 vertical levels between the surface and 50 hPa. The WRF configuration follows Appel et al. (2017) and employs four-dimensional data assimilation every 6 h to limit the growth of meteorological errors (WRF configuration details in Table S2). A 15 d spinup period was used for the WRF-CMAQ simulations. Anthropogenic emissions of trace gases and aerosols are based on the National Emissions Inventory (NEI) representative of 2014, as this was the latest available inventory at the time of emission preparation. NEI 2014 emissions were processed using the Sparse Matrix Operator Kernel Emissions (SMOKE) model. WRF simulations were used to drive SMOKE for generating meteorology-dependent anthropogenic emissions. Biogenic emissions of trace gases are calculated online using the Biogenic Emission Inventory System (BEIS). Gas-phase chemistry is represented using Carbon Bond 6 (CB06) version r3. Chemical lateral boundary conditions for the outer domain used the idealized profiles available in CMAQ but are dynamically provided to the inner domain every hour based on the outer-domain simulations.

2.5 Evaluation techniques

To perform a systematic, direct comparison of daily satellite products to airborne retrievals, OMI, GeoTASO, and GCAS data were spatially averaged to 0.15° × 0.15° (~ 15 km × 15 km, similar to OMI nadir spatial resolution) for evaluating OMI. TROPOMI and airborne observations were spatially averaged at 0.05° × 0.05° (~ 5 km × 5 km, similar to the TROPOMI nadir spatial resolution) for evaluating TROPOMI. To investigate the impact of the higher spatial

Table 1. Airborne (GeoTASO and GCAS) flight information (date, flight times, number of co-located satellite and airborne FNR grids) used in this study.

Flight day number	Date	Time (UTC)	OMI FNR co-locations ^a	TROPOMI FNR co-locations ^b
1	25 June 2018	Morning: 12:30–15:42 Afternoon: 16:48–20:18	12	201
2	30 June 2018	Morning: 12:12–15:36 Afternoon: 16:42–20:24	37	251
3	2 July 2018	Morning: 11:24–16:36 Afternoon: 17:54–21:30	6	66
4	19 July 2018	Morning: 11:24–15:18 Afternoon: 16:54–20:54	0	155
5	20 July 2018	Morning: 11:24–15:18 Afternoon: 17:06–21:06	5	136
6	5 August 2018	Morning: 12:30–16:30 Afternoon: 17:48–22:18	5	0
7	6 August 2018	Morning: 11:42–16:00 Afternoon: 17:12–21:30	0	67
8	15 August 2018	Morning: 11:12–15:30 Afternoon: 17:00–21:36	0	150
9	18 August 2018	Morning: 11:18–15:18 Afternoon: 17:18–21:30	0	108
10	24 August 2018	Morning: 10:54–15:18 Afternoon: 16:36–21:00	20	147
11	28 August 2018	Morning: 11:18–15:18 Afternoon: 16:36–20:18	8	150
12	29 August 2018	Morning: 11:12–15:06 Afternoon: 16:36–20:48	0	166
13	6 September 2018	Morning: 11:54–15:48 Afternoon: 17:12–21:24	8	96

^a OMI FNR co-locations at 0.15° × 0.15° spatial resolution. ^b TROPOMI FNR co-locations at 0.05° × 0.05° spatial resolution.

resolution of TROPOMI, NO₂, HCHO, and FNR retrievals from this sensor were also averaged to 0.15° × 0.15° for inter-comparison with OMI. To smooth and reduce the noise of satellite data, as well as reduce spatial representation errors of the satellite compared to GeoTASO and GCAS, we apply a point oversampling technique (e.g., McLinden et al., 2012) when spatially averaging the retrievals. This method uses a larger grid box radius, compared to the averaging resolution, to bin individual retrievals. When averaging satellite data to the 0.15° × 0.15° spatial resolution (standard radius of 0.075°), we employed a radius twice the standard size equal to 0.15°. Similarly, when averaging satellite data to the 0.05° × 0.05° spatial resolution (standard radius of 0.025°), we applied a radius of 0.05°.

Given that the nominal flight altitude during LIS-TOS 2018 was 9 km a.g.l., in order to directly com-

pare to satellite tropospheric-column retrievals, we scaled airborne tropospheric-column NO₂ values by multiplying the observed values by the ratio of the total tropospheric NO₂ column abundance over the tropospheric-column NO₂ abundance below 9 km a.g.l. (i.e., $\frac{\text{Tropospheric NO}_2 \text{ (surface to tropopause)}}{\text{Tropospheric NO}_2 \text{ (surface to 9 km a.g.l.)}}$). This scaling factor for NO₂, which showed that typically 60 % to 99 % of tropospheric NO₂ is below 9 km a.g.l., was derived for each co-located GeoTASO and GCAS retrieval, using WRF-CMAQ simulations. Airborne tropospheric-column HCHO data were not scaled since typically > 95 % of HCHO was determined to be below the nominal aircraft flight altitude.

GeoTASO and GCAS data were co-located to OMI and TROPOMI using a temporal threshold of ±60 min. GeoTASO and GCAS HCHO and NO₂ data were first filtered to remove airborne retrievals where the radiance flag was > 0.5

Table 2. Statistical evaluation of NASA OMI, QA4ECV OMI, and TROPOMI retrievals of tropospheric-column NO₂ and HCHO and resulting FNRs. Statistics presented are the number of co-located grids (*N*), mean concentration ± standard deviation from satellite (sat. conc.) and airborne (air. conc.) retrievals, median bias ± bias standard deviation, NMB (%), RMSE, coefficient of determination (*R*²^a), and linear regression slope.

	NASA OMI (0.15° × 0.15°)			QA4ECV OMI (0.15° × 0.15°)		
	FNR	HCHO ^b	NO ₂ ^b	FNR	HCHO ^b	NO ₂ ^b
<i>N</i>	101	101	116	82	85	106
Sat. conc.	4.4 ± 4.3	17.5 ± 7.5	6.3 ± 5.3	3.7 ± 3.5	16.5 ± 9.1	5.9 ± 3.9
Air. conc.	3.6 ± 2.1	13.2 ± 7.1	6.1 ± 6.6	3.4 ± 2.2	13.3 ± 7.4	6.1 ± 6.9
Bias	0.4 ± 3.8	5.1 ± 7.8	0.4 ± 4.1	−0.2 ± 3.3	2.3 ± 8.9	0.4 ± 4.5
NMB	11.0	38.7	6.3	−5.4	17.3	6.8
RMSE	3.8	8.9	4.1	3.3	9.4	4.5
<i>R</i> ²	<i>0.23</i>	<i>0.19</i>	<i>0.62</i>	<i>0.17</i>	<i>0.19</i>	<i>0.62</i>
Slope	1.0	0.46	0.63	0.67	0.54	0.44
	TROPOMI (0.15° × 0.15°)			TROPOMI (0.05° × 0.05°)		
	FNR	HCHO ^b	NO ₂ ^b	FNR	HCHO ^b	NO ₂ ^b
<i>N</i>	261	261	261	1693	1741	1802
Sat. conc.	3.6 ± 1.8	15.9 ± 4.7	5.9 ± 4.2	4.0 ± 2.6	16.2 ± 7.0	5.7 ± 4.6
Air. conc.	3.2 ± 1.7	12.8 ± 6.3	6.0 ± 6.1	3.4 ± 2.0	14.6 ± 6.7	6.6 ± 6.9
Bias	0.3 ± 1.4	2.9 ± 4.9	0.3 ± 3.3	0.4 ± 2.3	1.9 ± 6.7	−0.3 ± 3.7
NMB	9.3	23.1	5.8	13.0	12.9	−4.8
RMSE	1.4	5.6	3.3	2.3	6.7	3.9
<i>R</i> ²	<i>0.48</i>	<i>0.40</i>	<i>0.74</i>	<i>0.29</i>	<i>0.28</i>	<i>0.75</i>
Slope	0.75	0.47	0.59	0.70	0.55	0.58

^a Correlation values which are presented in italics are statistically significant to a 95 % confidence interval.

^b Concentration, bias, and RMSE units are × 10¹⁵ molecules cm^{−2}.

as they are considered to be influenced by clouds or glint. We initially applied a temporal threshold of ±30 min; however, this resulted in < 50 total co-locations with OMI retrievals. Therefore, the longer-temporal-threshold criterion was necessary to achieve enough co-locations for statistical evaluation. The longer temporal threshold of ±60 min resulted in only slightly larger median biases compared to when applying the ±30 min threshold. Similar bias statistics using temporal offsets of 30 and 60 min agree with other studies which show minimal dependence on temporal offsets between 0 and 60 min (e.g., Tack et al., 2021). It should be noted that the temporal threshold of ±60 min, as well as spatial averaging methods applied in this study, resulted in a slightly larger spread in TROPOMI NO₂ data when evaluated with GeoTASO and GCAS data compared to the results in Judd et al. (2020) which used a 30 min threshold.

Satellite retrievals with high quality were isolated for use by using individual OMI retrievals that had quality flags of qa = 0 for HCHO and NO₂. This qa value is suggested in OMI data user manuals for the application of the highest-quality data and for the removal of OMI pixels impacted by the row anomaly. For TROPOMI, individual NO₂ and HCHO retrievals that had qa > 0.75 and qa > 0.5 were used, respectively, as recommended by the TROPOMI data user manuals. To avoid anomalous OMI and TROPOMI retrieval

values of HCHO, we remove VCDs with lower and upper bounds of -8.0×10^{15} and 7.6×10^{16} molecules cm^{−2}, respectively. These bounds were determined from typical HCHO VCD values and a threshold of 3 times the fitting uncertainty in OMI retrievals presented by Zhu et al. (2020). Similarly, to avoid anomalous OMI and TROPOMI retrieval values of NO₂, we remove VCDs with lower and upper bounds of -1.08×10^{15} and 8.07×10^{16} molecules cm^{−2}, respectively (OMI NO₂ algorithm team, personal communication, 2021). Both OMI and TROPOMI retrievals with solar zenith angles > 70° and effective cloud fractions > 30 % and > 50 %, respectively, were also removed. These additional thresholds were chosen based on guidance from the OMI and TROPOMI data user manuals. Finally, only co-located spatially averaged grids that had 75 % spatial coverage by GeoTASO and GCAS and airborne remote-sensing NO₂ VCDs > 1.0×10^{15} molecules cm^{−2} were used.

The statistical evaluation of OMI and TROPOMI retrievals with co-located GeoTASO and GCAS data was primarily done using the bias (median); variability in bias represented by the standard deviation of bias (referred to as bias standard deviation); normalized median bias (NMB), normalized to the magnitude of observed data; root mean square error (RMSE); and simple linear regression statistics (slope; *y* in-

Table 3. Statistical evaluation of NASA OMI and TROPOMI retrievals of tropospheric-column NO₂ and HCHO and resulting FNRs when reprocessed with high-spatial-resolution WRF-CMAQ a priori information. Statistics presented are the number of co-located grids (*N*), mean concentration ± standard deviation from satellite (sat. conc.), median bias ± bias standard deviation, NMB (%), RMSE, coefficient of determination (*R*²), and linear regression slope.

	NASA OMI (0.15° × 0.15°)			Scaled NASA OMI (0.15° × 0.15°) ^a		
	FNR	HCHO ^b	NO ₂ ^b	FNR	HCHO ^b	NO ₂ ^b
<i>N</i>	101	101	116	101	101	116
Sat. conc.	3.5 ± 4.3	20.4 ± 8.9	10.5 ± 8.5	3.1 ± 3.6	15.8 ± 7.1	6.3 ± 6.2
Bias	−0.3 ± 3.9	8.6 ± 7.8	3.1 ± 5.1	0.5 ± 3.2	4.4 ± 7.1	−0.3 ± 3.9
NMB	−9.4	65.7	50.0	16.7	35.6	−4.2
RMSE	3.9	10.6	6.7	3.5	7.8	3.9
<i>R</i> ²	0.17	0.30	0.65	0.21	0.25	0.67
Slope	0.85	0.70	1.03	1.05	0.50	0.76
	TROPOMI (0.15° × 0.15°)			TROPOMI (0.05° × 0.05°)		
	FNR	HCHO ^b	NO ₂ ^b	FNR	HCHO ^b	NO ₂ ^b
<i>N</i>	261	261	261	1693	1741	1802
Sat. conc.	3.2 ± 1.7	12.8 ± 4.4	6.0 ± 4.3	3.4 ± 2.6	14.6 ± 6.8	6.6 ± 5.2
Bias	−0.3 ± 1.4	−1.2 ± 5.1	0.1 ± 3.8	0.2 ± 2.2	−0.1 ± 6.3	−0.4 ± 4.1
NMB	−9.1	−9.4	2.0	4.7	−0.3	−6.4
RMSE	1.4	5.2	3.8	2.2	6.3	4.1
<i>R</i> ²	0.43	0.35	0.61	0.32	0.32	0.67
Slope	0.67	0.41	0.55	0.74	0.58	0.61

^a Reprocessed with “scaled” CMAQ a priori profiles. ^b Concentration, bias, and RMSE units are × 10¹⁵ molecules cm^{−2}.

tercept; and coefficient of determination, *R*²) based on ordinary least squares.

3 Results

3.1 Campaign-averaged tropospheric FNRs

Airborne observations during the summer of 2018 suggest that during the midday hours large areas of FNRs ≤ 1.0 occurred over the urban regions surrounding the city of New York (NYC). The term “urban” here is used qualitatively as the region close in proximity to the center of NYC, where elevated tropospheric-column NO₂ values were frequently observed. The opposite is true for the usage of “rural”. Figure 1 shows the campaign-averaged FNRs from OMI (NASA and QA4ECV) and TROPOMI, averaged to 0.15° × 0.15° and 0.05° × 0.05° spatial resolution, compared to co-located airborne remote-sensing products. These regions of FNRs ≤ 1.0 likely have O₃ production which is limited by VOC emissions. Outside of the VOC-limited region, airborne observations show a clear transition zone of FNRs between 1.0 and 2.0 and NO_x-limited regimes (FNR > 2.0) in the rural regions of the northeastern US. It should be noted these FNR thresholds follow the assumptions of Duncan et al. (2010); however, there are uncertainties in the exact thresholds separating O₃ sensitivity production regimes, and they can be spatiotemporally variable (e.g., Lu and Chang, 1998; Schroeder

et al., 2017; Souri et al., 2020; Jin et al., 2020; Ren et al., 2022). For simplicity, we use the constant FNR ratio thresholds defined by Duncan et al. (2010) for discussion.

Satellite retrievals also displayed the same general regional patterns of FNRs in the northeastern US that were observed by airborne remote sensing (see Fig. 1). However, all satellite products show higher FNRs (between 1.0 and 3.0) in the areas where airborne observations observed NO_x-saturated regimes. In general, TROPOMI FNRs at the 0.05° × 0.05° spatial resolution have the lowest values over NYC in better agreement with airborne observations. TROPOMI retrieval data also better capture the spatial pattern and urban–rural interface of observed O₃ sensitivity production regimes compared to OMI data. TROPOMI FNR retrievals and airborne observations display a clear urban–rural interface; however, OMI products result in noisier spatial patterns. When averaged to a resolution similar to the native resolution of OMI (0.15° × 0.15°), TROPOMI data suggest higher FNRs ≥ 2.0 in the vicinity of NYC, in line with OMI retrieval products. It should be noted that satellite- and airborne-retrieved FNRs are dependent on both tropospheric NO₂ and HCHO data. Median and unresolved biases in FNRs can then be driven by errors in retrievals of NO₂ and/or HCHO. Therefore, the following sections of this work investigate the statistical evaluation of NASA OMI, QA4ECV OMI, and TROPOMI tropospheric NO₂, HCHO, and resulting FNRs.

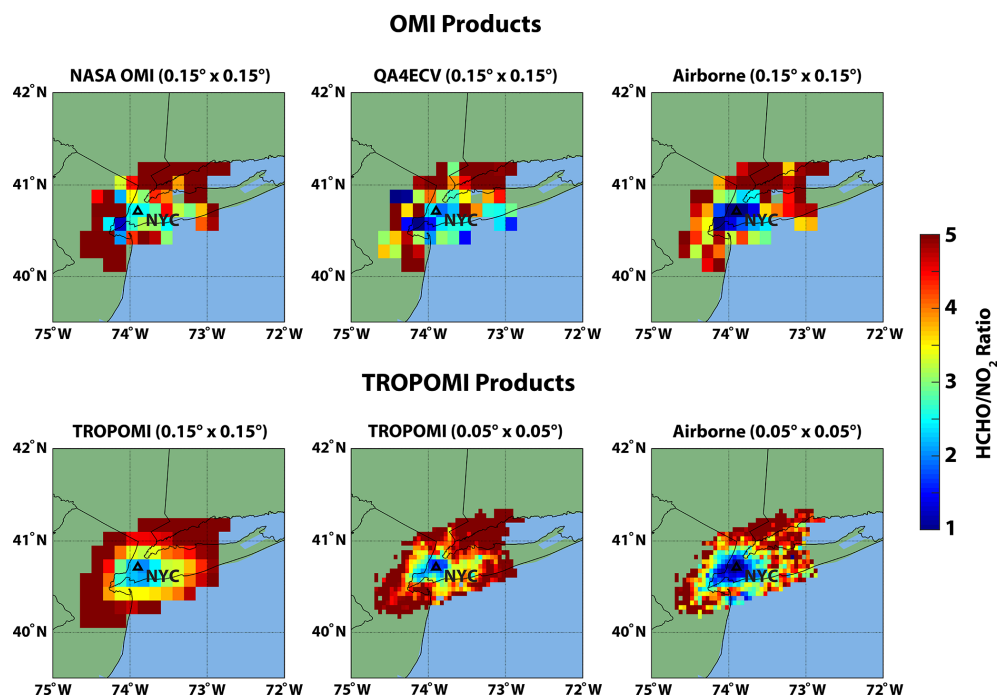


Figure 1. NASA OMI, QA4ECV OMI, TROPOMI, and airborne tropospheric-column FNR retrievals averaged for all flights conducted during the LISTOS 2018 field campaign. All co-located OMI and airborne remote-sensing tropospheric-column FNR values are averaged at $0.15^\circ \times 0.15^\circ$, and TROPOMI co-locations are averaged at $0.05^\circ \times 0.05^\circ$ and $0.15^\circ \times 0.15^\circ$ spatial resolution. The black triangle indicates the location of NYC.

3.2 Statistical evaluation of OMI and TROPOMI

3.2.1 Tropospheric-column NO₂ systematic bias and uncertainty

The spatial pattern of campaign-averaged tropospheric-column NO₂ retrieved by the satellites and airborne sensors highlight the large pollution region around the urban areas of NYC (see Fig. S2). Tropospheric NO₂ columns over NYC from both satellite and airborne observations frequently exceed 1.0×10^{16} molecules cm⁻². However, while airborne NO₂ values in the rural regions surrounding NYC were frequently $< 2.0 \times 10^{15}$ molecules cm⁻², satellite retrievals had larger NO₂ columns between 2.0×10^{15} and $> 4.0 \times 10^{15}$ molecules cm⁻². This suggests OMI and TROPOMI retrievals have a high bias in clean-region tropospheric NO₂ columns (spatial distribution of satellite NO₂ bias shown in Fig. S4). This high bias in satellite tropospheric-column NO₂ values in clean regions can possibly be linked to an underestimated abundance of free-tropospheric NO₂ in CTMs used as a priori profile data sets for OMI and TROPOMI retrievals resulting in AMFs which are too low (e.g., Silvern et al., 2019). Studies have shown that the coarse spatial resolution of the CTMs used to derive a priori NO₂ profiles for OMI and TROPOMI cannot resolve the sharp gradients of NO₂ at the urban–rural interface and lead to the overestimation of satellite retrievals in low-pollution regions (Lamsal

et al., 2014; Tack et al., 2021). Finally, other aspects of the satellite retrievals such as biases in stratospheric NO₂ concentrations and separation from the troposphere, aerosol interference, and surface albedo could contribute to these overestimations in clean regions (e.g., Lamsal et al., 2021).

Figure 2 shows the comparison of co-located NASA OMI, QA4ECV OMI, and TROPOMI retrievals of tropospheric NO₂ columns with observed data from all flights (statistical evaluation in Table 2). The high bias of tropospheric NO₂ columns in clean regions retrieved by the satellite sensors outside the urban regions of NYC resulted in linear regression slopes < 0.65 and positive y intercepts. Some of this high bias in clean regions is offset in the campaign-averaged median biases by the fact that the satellite retrievals have a low bias compared to NO₂ values observed over polluted regions ($> 1.0 \times 10^{16}$ molecules cm⁻²). The magnitude dependence of satellite-retrieved NO₂ bias agrees with past validation studies (e.g., Zhao et al., 2020; Lamsal et al., 2021; Verhoelst et al., 2021). This magnitude dependence has been shown to be driven by uncertainties in AMF values used in the retrievals (Martin et al., 2002; Boersma et al., 2004). TROPOMI at its near-native spatial resolution has the least high bias of clean-region tropospheric NO₂ demonstrated by the lower y -axis intercept compared to all OMI and TROPOMI data products at the coarser spatial resolution. Overall, NASA OMI and QA4ECV displayed small campaign-averaged median biases (NMB %) of $\sim 0.4 \times$

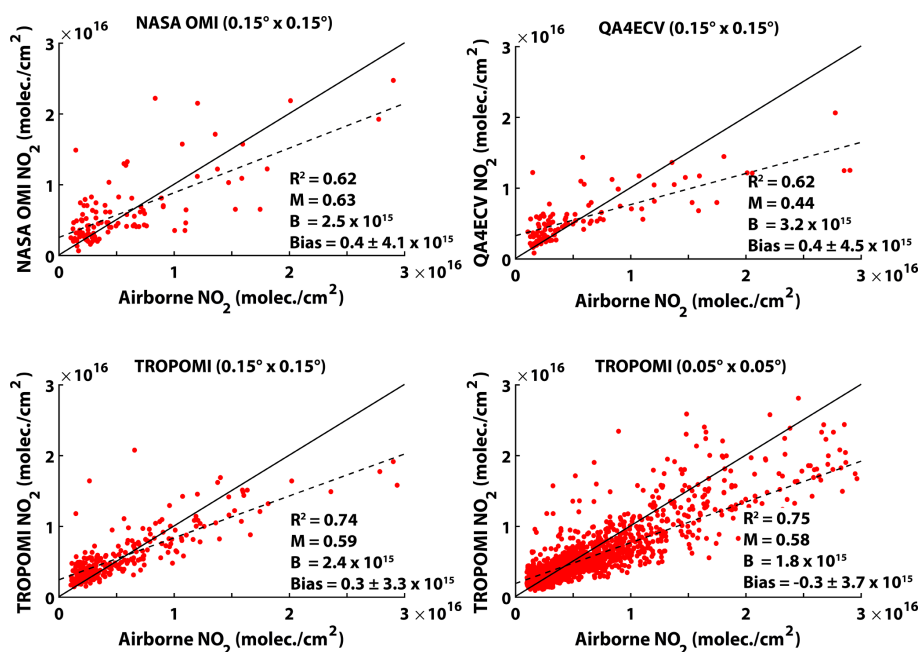


Figure 2. Comparison of satellite- (NASA OMI, QA4ECV OMI, and TROPOMI) and airborne-retrieved tropospheric NO₂ (molecule cm⁻²) for each co-located measurement taken during the field campaign. All co-located OMI and airborne remote-sensing tropospheric-column NO₂ values are averaged at 0.15° × 0.15° resolution, and TROPOMI co-located data are averaged at 0.15° × 0.15° and 0.05° × 0.05° spatial resolution. The solid black line shows the 1 : 1 comparison, and the dashed line shows the linear regression fit. The figure inset shows the main statistics (coefficient of determination (R^2), slope (M), y intercept (B), and median bias and bias standard deviation) of the comparison.

10¹⁵ molecules cm⁻² (~ 6.5 %). TROPOMI retrievals had a campaign-averaged median bias of -0.3 molecules cm⁻² (-4.8 %) and 0.3×10^{15} molecules cm⁻² (5.8 %) when averaged at 0.05° × 0.05° and 0.15° × 0.15° spatial resolution, respectively. It should be noted that the TROPOMI low bias in tropospheric-column NO₂ is improved with the newer retrieval algorithm used in this study compared to early versions of the data product (e.g., v1.2.2 had a campaign-averaged median low bias of -1.3×10^{15} molecules cm⁻²) primarily due to better cloud pressure input data (FRESCO+ wide) now used in TROPOMI retrievals (Riess et al., 2022).

Noise in the satellite retrievals resulting in unresolved errors (RMSE) is important for accurate retrievals of the spatially resolved daily tropospheric-column NO₂, HCHO, and FNRs. At the near-native spatial resolution of the three satellite NO₂ retrievals, RMSE values were similar (~ 3.5 – 4.5×10^{15} molecules cm⁻²) with QA4ECV OMI data having the largest bias standard deviation and RMSE values and TROPOMI having the least noise in the data (see Table 2). To determine if the higher spatial resolution and less noise of TROPOMI retrievals resulted in more favorable comparisons to observations, we further compared TROPOMI tropospheric-column NO₂ values to OMI results. TROPOMI data averaged to match OMI spatial resolution displayed the lowest RMSE values. At both spatial resolutions, TROPOMI tropospheric NO₂ data have less spread in the data compared to OMI products. The larger noise in OMI tropospheric NO₂

SCDs compared to TROPOMI NO₂ SCDs has been shown in recent studies (van Geffen et al., 2020, 2022) and has been attributed to its higher spatial resolution (better by a factor of > 12) and similar, to even better, signal-to-noise ratios. Furthermore, TROPOMI NO₂ at 0.05° × 0.05° better reproduces the spatial patterns of observed tropospheric-column NO₂. This is emphasized by the higher correlation and lower RMSE values when evaluating TROPOMI tropospheric NO₂ columns with observations in comparison to the other satellite products and visually more clearly separating the urban–rural interface seen in tropospheric NO₂ (see Fig. S2). Finally, TROPOMI NO₂ data averaged to the coarser spatial resolution of OMI have a similarly high median bias as both OMI retrieval algorithms; however, displayed RMSE values are nearly twice as small as NASA and QA4ECV OMI, further emphasizing the importance of spatial resolution for retrieving tropospheric NO₂ columns.

3.2.2 Tropospheric-column HCHO systematic bias and uncertainty

The spatial pattern of campaign-averaged tropospheric-column HCHO retrieved by the satellites and airborne sensors demonstrate the large HCHO concentrations in both urban and rural regions during the summer of 2018 (see Fig. S3). Airborne observations of tropospheric-column HCHO concentrations show that over NYC the concentra-

tions are on average $\sim 1.5 \times 10^{16}$ molecules cm⁻² and can exceed 2.5×10^{16} molecules cm⁻² during the afternoon hours (see Fig. S3). Both OMI and TROPOMI retrieval products have smaller gradients between HCHO concentrations in the urban and rural regions in comparison to airborne observations.

Figure 3 shows the scatterplot comparison of co-located NASA OMI, QA4ECV OMI, and TROPOMI retrievals of tropospheric HCHO columns compared to observed data (statistical evaluation in Table 2). This figure and Table 2 illustrate the high bias of clean-region tropospheric HCHO columns retrieved by satellites (spatial distribution of HCHO bias in OMI and TROPOMI shown in Fig. S5). All satellite products have a high bias for tropospheric-column HCHO values of $\leq 1.5 \times 10^{16}$ molecules cm⁻², linear regression slopes of < 0.60 , and positive y intercepts. Both OMI retrieval products and TROPOMI data better replicate the larger HCHO concentrations (between 1.5×10^{16} and 3.0×10^{16} molecules cm⁻²) with some small low bias in more polluted regions ($> 3.0 \times 10^{16}$ molecules cm⁻²). On average, NASA OMI had the largest campaign-averaged median high bias of 5.1×10^{15} molecules cm⁻² (38.7%). QA4ECV OMI data result in a lower campaign-averaged median high bias of 2.3×10^{15} molecules cm⁻² (17.3%). Finally, TROPOMI retrievals had the lowest campaign-averaged median high bias of 1.9×10^{15} molecules cm⁻² (12.9%) at $0.05^\circ \times 0.05^\circ$ spatial resolution and 2.9×10^{15} molecules cm⁻² (23.1%) when averaged at $0.15^\circ \times 0.15^\circ$.

The results of the validation shown in Fig. 3 and Table 2 are consistent with recent validation studies such as the work of Vigouroux et al. (2020) and De Smedt et al. (2021) which also show OMI and TROPOMI retrievals are biased high in clean conditions and in regions of high tropospheric HCHO columns are generally consistent with some moderate low bias. In order to provide more of a quantitative comparison with recent validation studies of OMI and TROPOMI HCHO (Vigouroux et al., 2020; De Smedt et al., 2021), we separated our co-located satellite–airborne data points into clean ($< 5.0 \times 10^{15}$ molecules cm⁻²) and polluted ($\geq 8.0 \times 10^{15}$ molecules cm⁻²) scenes. We chose a slightly higher threshold for separating clean HCHO columns to optimize the number of co-locations for statistics. We also added a highly polluted threshold ($> 16.0 \times 10^{15}$ molecules cm⁻²) to further emphasize our results (see Table S3). While the positive tropospheric HCHO column biases derived in our study are higher compared to Vigouroux et al. (2020) and De Smedt et al. (2021), the magnitude dependence is similar. We show that clean-region satellite HCHO columns are larger than observations for all satellite products and transition to a low bias in highly polluted regions.

The NASA and QA4ECV OMI HCHO retrievals had RMSE values $\sim 9.0 \times 10^{15}$ molecules cm⁻² with QA4ECV data having a slightly larger data spread. The higher spatial resolution and sufficient signal-to-noise ratio of TROPOMI resulted in HCHO RMSE values $\sim 25\%$ – 30% lower com-

pared to OMI. Spatially averaging TROPOMI tropospheric-column HCHO to coarser grids to increase the signal-to-noise ratio aided in further reducing RMSE values (see Table 2). While both TROPOMI and OMI tropospheric HCHO retrievals display large noise, TROPOMI has correlation values better compared to OMI, with R^2 values being higher by a factor of 2 at the same spatial resolution. Vigouroux et al. (2020) and De Smedt et al. (2021) agree with our analysis that TROPOMI HCHO has lower RMSE values and higher correlations with observations, compared to OMI products. The larger spread in tropospheric HCHO from OMI compared to TROPOMI is likely due to the weaker signal-to-noise ratio in OMI and potentially the fewer co-located data points for statistical analysis.

All three satellite HCHO products have larger RMSE values and low correlations when compared to the statistical evaluation of satellite NO₂ retrievals. TROPOMI SCD retrievals of HCHO have been shown in recent work (e.g., De Smedt et al., 2021) to have less noise compared to OMI due to the higher spatial resolution and at least the same signal-to-noise ratio. Furthermore, UV–Vis retrievals at shorter wavelengths (~ 340 nm) have much smaller sensitivity to HCHO compared to longer wavelengths (~ 440 nm) employed for NO₂ retrievals (Lorente et al., 2017). The sensitivity of UV–Vis retrievals to HCHO is lower throughout the middle and lower troposphere compared to NO₂, due to stronger Rayleigh scattering at shorter wavelengths, approaching twice as low near the surface (Lorente et al., 2017). The higher sensitivity of NO₂ retrievals in the lower troposphere, compared to HCHO, is important as the largest spatiotemporal variability in both NO₂ and HCHO concentrations occur lower in the troposphere leading to the higher correlation and lower RMSE values in the tropospheric-column NO₂ statistical evaluation.

3.2.3 Tropospheric-column FNR systematic bias and uncertainty

The comparison of satellite- and airborne-retrieved FNRs is shown in Fig. 4 and Table 2. NASA OMI and TROPOMI (at $0.05^\circ \times 0.05^\circ$ and $0.15^\circ \times 0.15^\circ$) display campaign-averaged median biases of 0.3–0.4, and QA4ECV OMI data resulted in a campaign-averaged median bias of -0.2 . Regardless of how tropospheric-column NO₂ and HCHO compared to observations, all satellite products evaluated here resulted in campaign-averaged FNR median biases ≤ 0.4 , suggesting that the systematic biases in the individual proxy species for OMI and TROPOMI offset to result in accurate median campaign-averaged FNR values. Visual inspection of TROPOMI and QA4ECV OMI retrievals suggests that these two products have the best ability to replicate the lowest observed FNRs over NYC during the field campaign (see Fig. 1). However, besides NASA OMI retrievals, the satellite products have linear regression slopes < 1.0 , indicating a

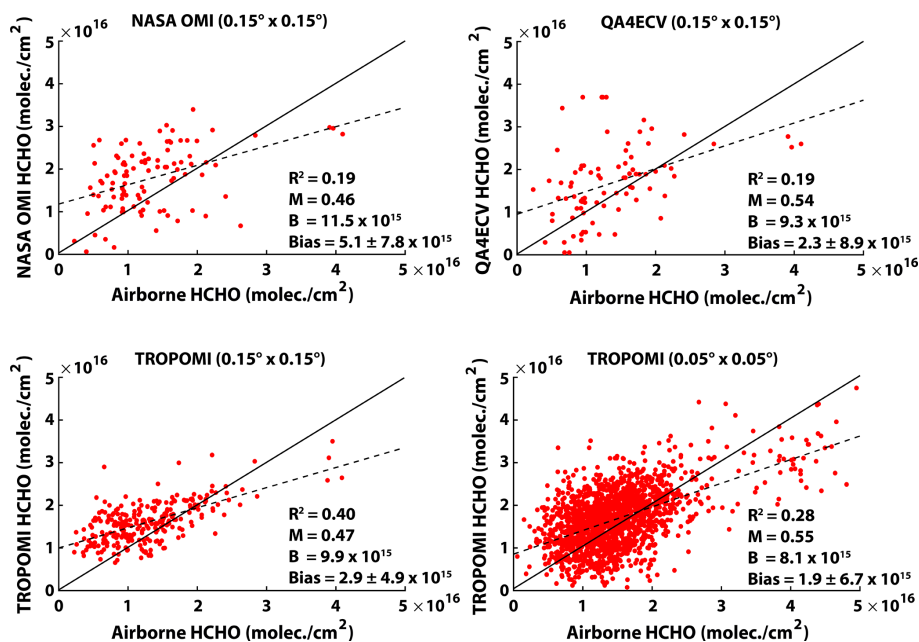


Figure 3. Comparison of satellite- (NASA OMI, QA4ECV OMI, and TROPOMI) and airborne-retrieved tropospheric HCHO (molecule cm⁻²) for each co-located measurement taken during the field campaign. All co-located OMI and airborne remote-sensing tropospheric-column HCHO values are averaged at 0.15° × 0.15° resolution, and TROPOMI co-located data are averaged at 0.15° × 0.15° and 0.05° × 0.05° spatial resolution. The solid black line shows the 1 : 1 comparison, and the dashed line shows the linear regression fit. The figure inset shows the main statistics (coefficient of determination (R^2), slope (M), y intercept (B), and median bias and bias standard deviation) of the comparison.

high bias for lower FNR values and some small low bias for higher observed FNRs.

All three satellite products displayed high correlation with tropospheric-column NO₂ observations, suggesting these spaceborne sensors can accurately assess the spatial and temporal patterns of this species. However, all the satellite products had very low correlation and high RMSE values when compared with observations of HCHO. In fact, the rank in correlation levels of all four FNR satellite products evaluated here directly matches the rank in correlation levels of tropospheric HCHO. This leads to the conclusion that given bias variability and RMSE in satellite tropospheric HCHO are large, and they directly drive the uncertainty in FNR retrievals; satellite HCHO observations are the limiting factor of using spaceborne retrievals to accurately assess daily FNRs.

An interesting finding of this study is that the median bias of OMI and TROPOMI HCHO and NO₂ tropospheric columns tend to cancel out, resulting in low median biases for FNRs; however, the unresolved biases in HCHO and NO₂ retrievals do not cancel out. This is clear as the RMSE values for FNRs are still large. Furthermore, biases for HCHO and NO₂ retrievals from OMI and TROPOMI are not correlated with R^2 values < 0.05 for all three satellite products. The uncertainty in HCHO and NO₂ retrievals resulted in FNR RMSE values for NASA OMI, QA4ECV

OMI, and TROPOMI (at near-native spatial resolutions) of 3.8, 3.3, and 2.3, respectively. Spatially averaging TROPOMI tropospheric-column HCHO data was shown to reduce the noise in the data, resulting in FNR RMSE values for TROPOMI at the coarser spatial resolution lower by nearly a factor of 2 compared to OMI. Overall, the large noise and unresolved error in tropospheric HCHO retrievals directly result in the uncertainty in FNR retrievals. It should be noted that the HCHO validation data from GeoTASO and GCAS are also hindered by weak absorption signatures in the shorter UV–Vis wavelengths and could add to the bias variability and RMSE values derived in this study. However, the level of uncertainty in tropospheric-column HCHO data from OMI and TROPOMI derived in this study is generally consistent with other recent studies (e.g., Vigouroux et al., 2020; De Smedt et al., 2021); therefore, we feel the conclusions drawn here are robust.

There are two main reasons HCHO retrievals are noisier compared to NO₂: (a) optical depths for HCHO peak in the UV range (< 380 nm) at the same wavelengths coinciding with large Rayleigh scattering and optical depths of O₃, leading to a weak and noisy signal, and (b) the stronger NO₂ optical depths in the visible wavelength range (400–500 nm), where there are higher signal-to-noise ratios, permit retrievals with less noise. To further evaluate the comparison of uncertainty in the remote sensing of

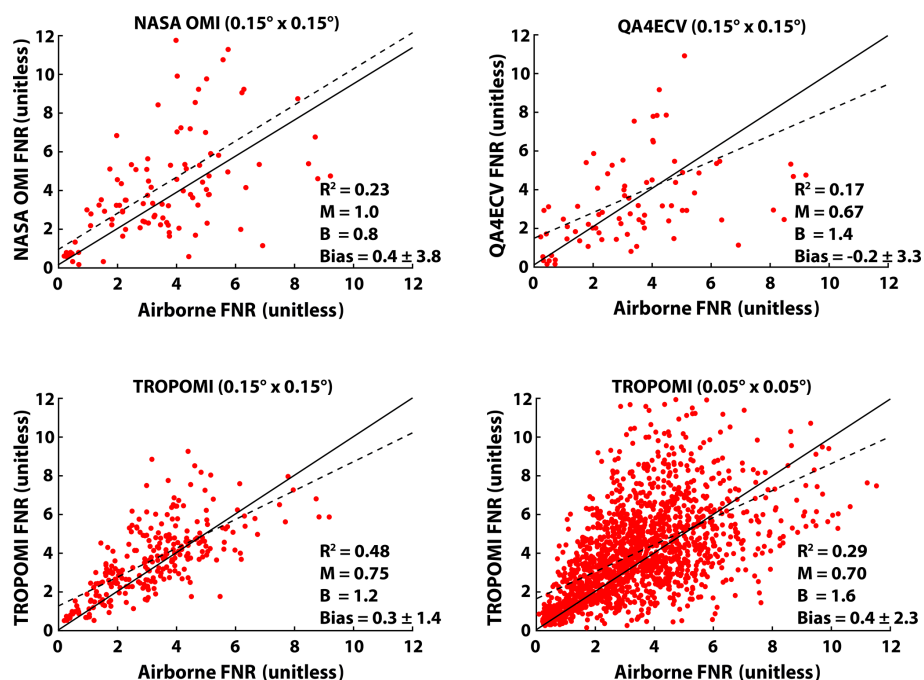


Figure 4. Comparison of satellite- (NASA OMI, QA4ECV OMI, and TROPOMI) and airborne-retrieved tropospheric FNR (unitless) for each co-located measurement taken during the field campaign. All co-located OMI and airborne remote-sensing tropospheric-column FNR values are averaged at $0.15^\circ \times 0.15^\circ$ resolution, and TROPOMI co-located data are averaged at $0.15^\circ \times 0.15^\circ$ and $0.05^\circ \times 0.05^\circ$ spatial resolution. The solid black line shows the 1 : 1 comparison, and the dashed line shows the linear regression slope. The figure inset shows the main statistics (coefficient of determination (R^2), slope (M), y intercept (B), and median bias and bias standard deviation) of the comparison.

NO₂ and HCHO, we compared GCAS and GeoTASO precision levels for the two species. Nowlan et al. (2018) derived the precision of the airborne remote-sensing systems used for NO₂ and HCHO retrievals in this study. Nowlan et al. (2018) quantified precisions of 1.0×10^{15} molecules cm⁻² and 1.9×10^{16} molecules cm⁻² at a fine spatial resolution of $250 \text{ m} \times 500 \text{ m}$ for NO₂ and HCHO, respectively. Averaging the data to the spatial resolution of $0.05^\circ \times 0.05^\circ$ improves these precision levels to 6.4×10^{13} and 1.2×10^{15} molecules cm⁻² for NO₂ and HCHO, respectively. The campaign-averaged tropospheric-column NO₂ and HCHO abundances from GCAS and GeoTASO at $0.05^\circ \times 0.05^\circ$ were 6.6×10^{15} and 1.5×10^{16} molecules cm⁻², respectively. Comparing the precision values of Nowlan et al. (2018) to the mean abundances during LISTOS at the same spatial resolution results in mean precision levels of 1 % and 8 % for NO₂ and HCHO, respectively. Overall, from this analysis it is expected that the HCHO retrievals should have more noise by a factor of 5–10 compared to NO₂.

3.3 Common a priori sensitivity test

This section analyzes the impact of using common, high-spatial-resolution ($4 \text{ km} \times 4 \text{ km}$), WRF-CMAQ-predicted NO₂ and HCHO vertical profiles as a priori information in NASA OMI and TROPOMI retrievals. GeoTASO and GCAS retrievals were not reprocessed in order to have a con-

sistent reference data set for the evaluation of the standard and reprocessed satellite retrievals. While reprocessing the airborne data with the higher-spatial-resolution model output would in itself be interesting (as done in Judd et al., 2020), the direct evaluation of the improvements in the reprocessed satellite data compared to the standard retrieval would not be possible. Figure 5 shows the campaign-averaged FNRs from NASA OMI and TROPOMI retrievals, when reprocessed with WRF-CMAQ NO₂ and HCHO a priori vertical profiles, compared to co-located airborne remote-sensing products (scatterplot comparison displayed in Fig. S6; statistical evaluation shown in Table 3). Comparing NASA OMI FNRs from this figure to Fig. 1, it is evident that using high-spatial-resolution WRF-CMAQ-predicted vertical profiles as a priori information resulted in FNR retrievals that are better able to capture the low FNR values ($\text{FNR} \leq 1.0$). Reprocessed TROPOMI FNRs also have lower values around NYC; however, they were reduced less compared to OMI retrievals.

Comparing standard retrieval products from NASA OMI (see Fig. S2 for NO₂ and Fig. S3 for HCHO) to reprocessed retrievals (see Fig. S7 for NO₂ and Fig. S8 for HCHO), it is clear that in general the higher-spatial-resolution model data resulted in larger tropospheric-column NO₂ and slightly larger tropospheric-column HCHO values. For TROPOMI, reprocessing the retrievals with WRF-CMAQ a priori in-

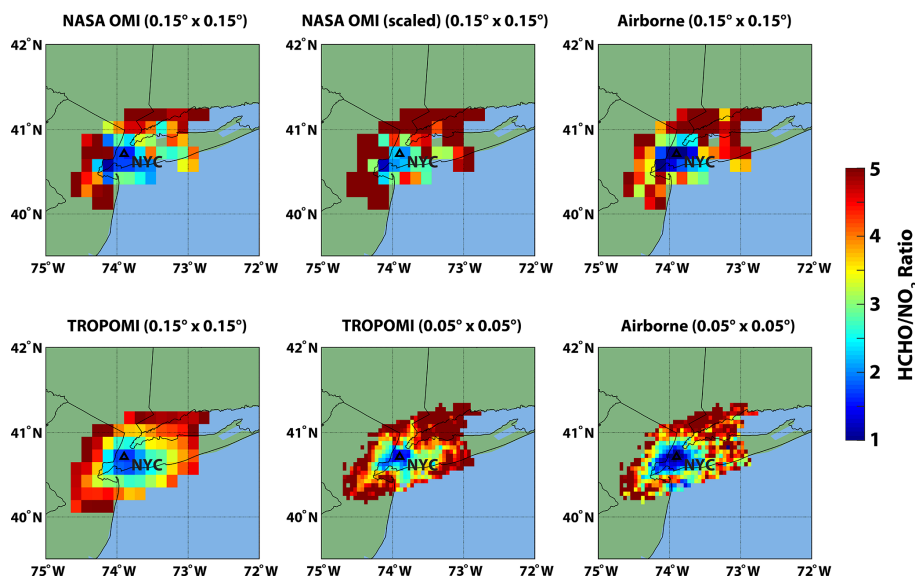


Figure 5. NASA OMI and TROPOMI reprocessed tropospheric-column FNR retrievals compared to airborne FNR observations averaged for all flights. All co-located OMI and airborne remote-sensing tropospheric-column FNR values are averaged at $0.15^\circ \times 0.15^\circ$, and TROPOMI co-locations are averaged at both $0.15^\circ \times 0.15^\circ$ and $0.05^\circ \times 0.05^\circ$ spatial resolution. The OMI FNR retrievals calculated with the scaled WRF-CMAQ profiles are identified as “scaled” in the figure panel titles. The black triangle indicates the location of NYC.

formation caused increases in tropospheric-column NO₂ over polluted regions but small decreases over rural areas. Tropospheric-column HCHO data for the reprocessed TROPOMI data were slightly lower in more polluted urban regions near NYC and much lower in the rural areas dominated by low concentrations compared to standard retrievals.

The increases in NASA OMI tropospheric NO₂ columns resulted in a small negative bias in FNR retrievals (-0.3), compared to a small positive bias in the standard products (0.4). When compared to airborne observations, the reprocessed NASA OMI NO₂ data display a large positive median bias which was not evident in the standard retrieval products. Similarly, reprocessed NASA OMI tropospheric-column HCHO data had a higher positive bias compared to standard retrievals. It should be noted, as previously discussed, that median biases in both reprocessed NASA OMI NO₂ and HCHO retrievals offset, resulting in median FNR values that compared relatively well to observations. However, the uncertainty in reprocessed satellite HCHO and NO₂ retrievals did not cancel out, resulting in FNR RMSE values which were still large for NASA OMI (3.9) and TROPOMI (3.5).

The larger tropospheric NO₂ columns in reprocessed NASA OMI data using high-spatial-resolution model data as a priori information were also shown in past studies (e.g., Souri et al., 2016; Goldberg et al., 2017). Both our study and the work by Goldberg et al. (2017) show that high-spatial-resolution CMAQ-predicted NO₂ a priori profiles result in OMI tropospheric NO₂ columns that are as high as larger by a factor of 2 than the standard retrievals. This high bias

is caused by smaller AMFs calculated due to the shape factor of high-spatial-resolution CMAQ-predicted NO₂ concentrations having a too steep NO₂ gradient. The change in HCHO shape factors when using WRF-CMAQ a priori profiles resulted in slightly higher tropospheric HCHO columns when compared to standard products for the same reason as tropospheric-column NO₂. Similar to Goldberg et al. (2017), we used airborne in situ observations of NO₂ and HCHO from LISTOS 2018 and the Ozone Water–Land Environmental Transition Study 2 (OWLETS-2, <https://www-air.larc.nasa.gov/missions/owlets/>, last access: 8 May 2023) field campaigns to correct the model-predicted a priori profiles for use in NASA OMI retrievals and is discussed later in this section.

Tropospheric NO₂ columns in reprocessed TROPOMI retrievals resulted in slightly lower median biases (-0.4×10^{15} molecules cm⁻²) compared to the standard products (-0.3×10^{15} molecules cm⁻²) with slightly larger RMSE values in the reprocessed NO₂ retrievals. Reprocessing TROPOMI retrievals of tropospheric-column HCHO resulted in smaller concentrations and improved median biases (-0.1×10^{15} molecules cm⁻²) and RMSE values (6.3×10^{15} molecules cm⁻²) compared to the median bias (1.9×10^{15} molecules cm⁻²) and RMSE (6.7×10^{15} molecules cm⁻²) in the standard products. The good performance of both reprocessed TROPOMI NO₂ and HCHO resulted in FNR values with a smaller median bias (0.2) compared to standard products (0.4) and slightly lower RMSE values.

Following methods similar to Goldberg et al. (2017) we used the University of Maryland Cessna 402B airborne observations to apply in situ data observational constraints on the NO₂ and HCHO a priori profiles applied in NASA OMI retrievals. The evaluation of WRF-CMAQ-predicted NO₂ (14 flights during LISTOS 2018 and OWLETS-2) and HCHO (7 flights during LISTOS 2018) vertical profiles using airborne data is displayed in Fig. S9. Compared to measured NO₂ values, the model displays a high bias below 1 km a.g.l. of ~ 0.4 ppb which was often $> 50\%$ larger than observations. This is in stark contrast to the model performance above 2 km a.g.l., where the model has a low bias of -0.2 to -0.4 ppb often approaching 100% lower than observations. For the WRF-CMAQ comparison to airborne in situ HCHO data, the model has a low bias throughout the lower troposphere, with larger low biases near the surface (-3.0 ppb between 0 and 1 km a.g.l.) and smaller low biases in the free troposphere (~ -1.3 ppb above 2 km a.g.l.). These low biases range between -50% to -100% compared to measured values. In addition to physiochemical parameterizations applied in CTMs, meteorological predictions by WRF, such as wind speed and direction, must have limited errors in order to accurately predict the horizontal and vertical distribution of NO₂ and HCHO concentration (e.g., Laughner et al., 2016; X. Liu et al., 2021). Compared to the airborne in situ observations, WRF wind speed and direction predictions during this study performed relatively well with median correlation (R) and bias values of 0.70 and 0.63 and ≤ 1.0 m s⁻¹ in the u and v wind components, respectively.

We applied approximated scaling factors to the a priori profiles to reprocess NASA OMI data (hereinafter referred to as scaled). Separate scaling factors were applied above and below the PBL, approximated to be at 1.5 km a.g.l., where noticeable differences in model performance were evident. For NO₂, we apply a scaling factor of 0.5 to WRF-CMAQ a priori NO₂ profiles in the PBL and 5.0 above the PBL. For HCHO, we applied a scaling factor of 2.0 to WRF-CMAQ a priori profiles in the PBL and 5.0 above the PBL. These scaling factors are approximations of the model performance and are simply applied to determine the impact of “raw” and “scaled” WRF-CMAQ-simulated a priori profiles in NASA OMI NO₂ and HCHO retrievals.

The spatial distribution of FNRs derived from the scaled NASA OMI reprocessed NO₂ and HCHO retrievals is shown in Fig. 5 (scatterplot comparison displayed in Fig. S6; statistical evaluation in Table 3). From Table 3 and Fig. 5 it can be seen that the scaled WRF-CMAQ a priori profiles result in higher FNR values and improved tropospheric-column NASA OMI NO₂ and HCHO retrievals compared to reprocessed products using the raw model output (see Figs. S7 and S8). Scaled NASA OMI tropospheric-column NO₂ and HCHO retrievals had smaller median biases of -0.3×10^{15} and 4.4×10^{15} molecules cm⁻² and much lower RMSE values of 3.9×10^{15} and 7.8×10^{15} molecules cm⁻², respectively, compared to the retrievals with raw WRF-CMAQ predic-

tions. Finally, the improved accuracy of tropospheric-column NO₂ and HCHO retrievals using scaled WRF-CMAQ predictions resulted in a slightly higher magnitude of FNR median bias (0.5), however, with lower RMSE values, compared to reprocessed data using raw CMAQ predictions. Compared to standard NASA OMI products, the reprocessed satellite data using scaled WRF-CMAQ data had similar median biases in FNR values and lower median biases for HCHO (4.4×10^{15} molecules cm⁻²) and NO₂ (-0.3×10^{15} molecules cm⁻²). All reprocessed data variables using scaled model-simulated shape factors, due to the reduction in uncertainty in retrieving HCHO and NO₂ data, had lower RMSE values, higher correlation (except for FNR), and similar or better linear regression slopes compared to standard satellite retrievals.

3.4 Discussion of satellite sensor errors and capabilities

3.4.1 Relative error in FNR retrievals

A recent study by Sourì et al. (2023) showed that satellite retrievals errors, in particular the unresolved error in HCHO products, are the largest source of uncertainty in using satellite FNRs to investigate O₃ sensitivity. Here we propagate the uncertainty (RMSE) calculated from NASA OMI, QA4ECV OMI, and TROPOMI to FNR calculations using Eq. (15) from Sourì et al. (2023) and created maps of the relative error (see Fig. 6). From this figure it can be seen that satellite retrieval errors in HCHO and NO₂ contribute significantly to satellite-derived FNR relative errors. In the largest NO_x emission source regions of NYC, where combined column abundances of HCHO and NO₂ are largest, the lowest relative errors in FNRs occur. For TROPOMI, which has the smallest values of uncertainty, relative errors are as low as $\sim 40\%$. Away from the emission region of NYC these relative error values reach as high as $\sim 80\%$. Similar patterns of relative error in FNRs from NASA and QA4ECV OMI retrievals are derived; however, the lowest relative error values over NYC are $\sim 50\%$ and reach values up to 100% . The largest relative errors are seen outside the source region of NYC in QA4ECV OMI retrievals due to having the largest uncertainty in HCHO and lower column abundances of this species in the rural regions of the domain. In addition to the fact that the less noisy retrievals from TROPOMI result in lower relative errors in FNR data, Fig. 6 further demonstrates the larger uncertainty in OMI as the relative error patterns are more heterogeneous. The spatial averaging of TROPOMI data results in the lowest relative errors in all four satellite products, as TROPOMI at the coarser ($0.15^\circ \times 0.15^\circ$) spatial resolution had relative errors as low as 35% , only increasing to $\sim 60\%$ outside of the source location of NYC.

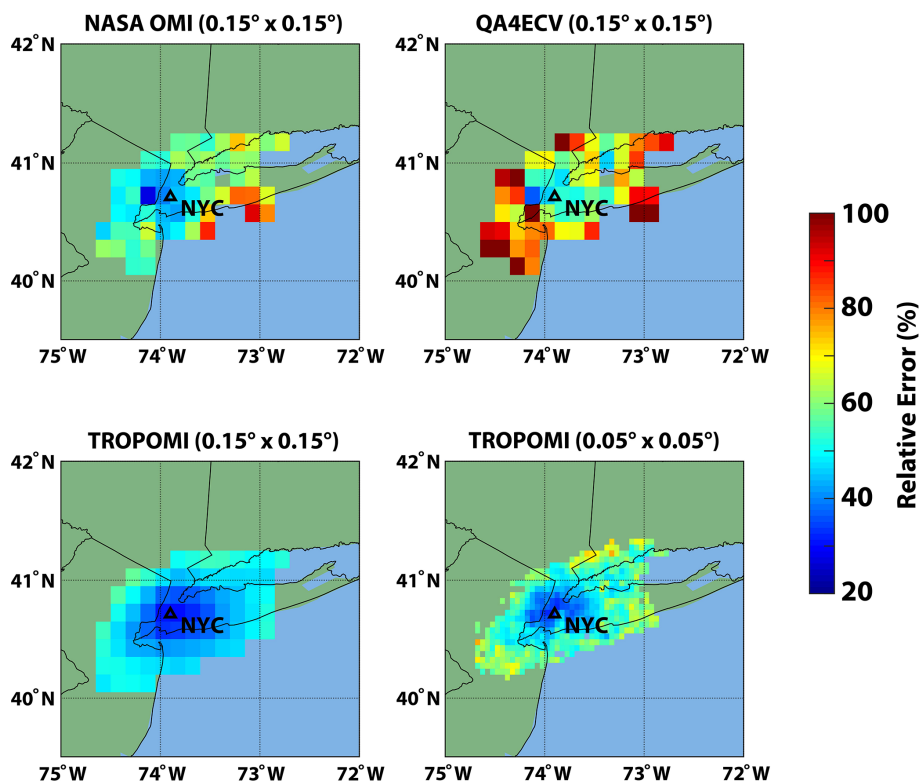


Figure 6. Campaign-averaged relative error in FNR products from standard NASA OMI, QA4ECV OMI, and TROPOMI retrievals due to uncertainty in HCHO and NO₂ retrievals. All co-located OMI and airborne remote-sensing tropospheric-column FNR values are averaged at 0.15° × 0.15°, and TROPOMI co-locations are averaged at 0.05° × 0.05° and 0.15° × 0.15° spatial resolution. The black triangle indicates the location of NYC.

3.4.2 Spatial and temporal capabilities of satellite FNR retrievals

Given the limited spatiotemporal data coverage provided by the LISTOS campaign, a robust understanding of the temporal capabilities of OMI and TROPOMI to retrieve FNRs is not possible. LEO satellites obtain, at best, a single snapshot of both HCHO and NO₂ each day, so one could only hope to obtain daily variability in FNRs from these spaceborne systems. To determine whether OMI and TROPOMI capture the variability in the daily mean tropospheric-column quantities of NO₂, HCHO, and FNRs over the entire LISTOS domain from airborne data, we compared these daily mean values from the satellite products to the airborne observations. For NASA OMI, daily correlation (R^2) values were 0.85 ($p = 0.001$), 0.58 ($p = 0.03$), and 0.26 ($p = 0.20$) for NO₂, HCHO, and FNRs, respectively. For QA4ECV OMI, daily correlation values were 0.85 ($p = 0.001$), 0.80 ($p = 0.002$), and 0.47 ($p = 0.06$) for NO₂, HCHO, and FNRs, respectively. For TROPOMI, daily correlation values were 0.92 ($p \leq 0.001$), 0.85 ($p \leq 0.001$), and 0.41 ($p = 0.03$) for NO₂, HCHO, and FNRs, respectively. All daily correlation statistics for HCHO and NO₂ were significant to a 95 % confidence interval and suggest that both OMI and TROPOMI

can capture the overall inter-daily magnitudes of FNR indicator species. However, only TROPOMI could observe the daily variability in domain-wide FNRs within a 95 % confidence interval. This suggests that unresolved errors in either HCHO or NO₂ retrievals from OMI, using both the NASA and QA4ECV algorithms, are too large to confidently capture the inter-daily variability in FNRs.

The same analysis was conducted for NASA and QA4ECV OMI except just for retrievals near the large anthropogenic source regions in NYC (within 0.35° of the city center) where relative errors due to satellite retrievals for FNR calculations were the lowest (see Fig. 6). Daily correlation values for FNR retrievals near the source region of NYC for NASA OMI (0.13; p value = 0.39) were reduced compared to domain-wide means, and QA4ECV OMI (0.66; p value = 0.01) correlations were improved near the source region of NYC. Indicator species correlation values from NASA OMI were degraded compared to the domain-wide analysis, suggesting that this satellite product may not be able to capture inter-daily variability in FNRs even in large source regions. However, this analysis suggests that QA4ECV OMI data have the capability to retrieve daily variability in FNRs in large emission regions such as NYC to a statistically significant level. Overall, TROPOMI retrievals at both fine and

coarse spatial resolutions evaluated in this study are able to capture daily variability in tropospheric FNRs over the entire domain and emission source regions better compared to OMI products.

Recent studies have shown that averaging OMI data (especially HCHO retrievals) for longer temporal periods can reduce the noise and uncertainty in this data product. For example, in the recent paper by Sourì et al. (2023), it was shown that unresolved errors in OMI HCHO can be reduced in monthly averages compared to daily retrievals by $\sim 33\%$, while there was little improvement in uncertainty statistics of NO₂ retrievals from OMI. However, recent studies (e.g., Schroeder et al., 2017) have also shown that for trend studies, monthly averaging column FNR data can mask FNR temporal gradients that exist within that period. This could hinder the results of trend studies of pollution on O₃ exceedance days and days of lower pollution.

To understand the extent to which OMI and TROPOMI retrievals lose spatial information (variance) compared to airborne data, we applied the algorithm named SpaTial Representation Error EstimaTor (STREET) (Sourì, 2022). This method creates semivariograms determining the changes in spatial variability with distance for a defined variable (we used HCHO and NO₂). The maximum variance at which the modeled semivariogram levels off is defined as a sill, and data sets with larger sill values possess richer spatial information. Figure S10 shows semivariograms, as well as the fitted stable Gaussian function, applied to TROPOMI and NASA OMI compared to airborne NO₂ columns. Concerning the comparison of TROPOMI and airborne data at $0.05^\circ \times 0.05^\circ$ resolution, we observe airborne semivariograms as high as 20×10^{15} molecules cm⁻², larger by a factor of 2 than what OMI achieves. At a ~ 20 km length scale, TROPOMI can only observe $\sim 40\%$ of the airborne spatial variance, indicating that the spatial representation error in TROPOMI is $\sim 60\%$ at this scale. NASA OMI fails to recreate $> 50\%$ of the maximum variance observed in airborne data at $0.15^\circ \times 0.15^\circ$ resolution. At ~ 20 km length scale, the spatial loss of OMI is $> 70\%$.

Figure S10 depicts the semivariograms and fitted exponential curves applied to TROPOMI and airborne HCHO. Immediately evident is that both semivariograms level off at longer distances compared to the analysis of NO₂. This stems from the fact that HCHO columns tend to be spatially more homogeneous in the region of the LISTOS domain. For most length scales, TROPOMI can replicate the spatial variance observed in airborne data ($\sim 70\%$). We do not present the semivariogram for NASA OMI HCHO columns as the underlying unresolved biases in OMI are very large, introducing artifacts that cannot be solely attributable to unresolved spatial scales. Since TROPOMI is able to capture the observed HCHO variability to a sufficient degree and can sufficiently retrieve NO₂ spatial variability, this suggests that TROPOMI has better capability to retrieve FNR spatial variability compared to OMI.

3.4.3 Reasons for systematic bias and uncertainty in FNRs

As demonstrated in this study, median biases of OMI and TROPOMI HCHO and NO₂ retrievals tend to cancel out when calculating tropospheric-column FNRs. Figures S4 and S5 show that the median bias spatial distribution of all satellite HCHO and NO₂ retrievals are similar, with a small low median bias in column abundances near the source region of NYC and high biases in clean regions. Table S1 shows that AMF calculations from NASA OMI, QA4ECV OMI, and TROPOMI use many of the same input data sets for geophysical variables (e.g., surface albedo, cloud fraction, cloud radiance), resulting in campaign-averaged AMFs of HCHO, NO₂, and the ratios of these products (AMF FNRs) which are relatively similar across the LISTOS domain (see Fig. S11). For all satellite products, HCHO and NO₂ AMFs have much less variability compared to AMFs derived for airborne data which along with SCD biases may contribute to the median high biases in clean HCHO and NO₂ retrievals. A primary reason for the inability of satellites to capture AMF variability over the LISTOS domain is likely the shape factors being used for these calculations having spatial resolutions from $1.0^\circ \times 1.0^\circ$ to even coarser grids. Furthermore, while TROPOMI and QA4ECV OMI retrievals used daily model data, NASA OMI uses monthly products which make it challenging to capture the large spatiotemporal variability in tropospheric HCHO and NO₂ vertical profiles. Finally, coarse geophysical input data sets used in satellite AMF calculations will not capture the spatial distribution of these variables as well as the high-spatial-resolution geophysical data sets used in airborne AMF calculations.

The more interesting aspect found in this study is that unresolved errors in HCHO and NO₂ retrievals do not cancel out in FNR calculations. While there are some reasons why uncertainty in HCHO and NO₂ retrievals could stem from opposite impacts of geophysical parameters in AMF calculations, such as AMF uncertainties in HCHO and NO₂ having opposite trends with increasing surface reflectance (comparing Fig. 10 from De Smedt et al., 2018, and Fig. 20 from S. Liu et al., 2021), these differences are minor and AMF calculations for both species in NASA OMI, QA4ECV OMI, and TROPOMI have similar input data sets. A portion of the uncertainty in HCHO and NO₂ retrievals not canceling out stems from the AMF calculations shown in Fig. S11. In order for HCHO and NO₂ AMFs to have no impact on VCD uncertainty cancelations, AMF FNRs would need to be a constant or similar value at all locations. However, from Fig. S11 it is shown that AMF FNRs, while having smooth spatial variability, are not a constant value. Therefore, some of the unresolved error residual in the FNR calculations will be due to differences in HCHO and NO₂ AMF calculations. This is emphasized in NASA OMI AMF FNR plots in Fig. S11, where different CTMs, at different spatial resolutions (see Table S1), are used to derive HCHO and NO₂ shape factors

leading to noticeable differences in the respective AMF calculations. This likely is one of the reasons that NASA OMI FNRs have the largest uncertainty (highest bias standard deviation and RMSE values) compared to airborne data (see Table 2).

The rest of the remaining unresolved error in FNR calculations is likely due to the SCD retrievals from OMI and TROPOMI sensors. SCD retrievals of HCHO from TROPOMI have been shown to have less noise compared to OMI (De Smedt et al., 2021). The larger uncertainty in OMI retrievals of HCHO compared to TROPOMI directly leads to the higher bias standard deviation and RMSE values for derived FNRs in OMI compared to TROPOMI (see Table 2). This is further emphasized in the spatially averaged TROPOMI data (at $0.15^\circ \times 0.15^\circ$ to match OMI data), where HCHO and FNR retrievals have an RMSE lower by a factor of 2–3 compared to OMI products. TROPOMI NO₂ SCDs have also been shown to have less noise compared to OMI retrievals (van Geffen et al., 2020, 2022). This is also shown in Table 2 when averaging TROPOMI data to match the OMI spatial resolution. Overall, HCHO and NO₂ SCD noise contributes to uncertainty in OMI, and TROPOMI VCDs and are not canceled out in FNR calculations; however, the reduced noise in TROPOMI SCD retrievals leads to improved VCDs of HCHO and NO₂ abundances.

4 Conclusions

This study presents a statistical evaluation and inter-comparison of tropospheric FNR retrievals from two commonly applied LEO sensors for investigating O₃ production sensitivity regimes (i.e., OMI and TROPOMI). The evaluation of NASA OMI, QA4ECV OMI, and TROPOMI retrievals of tropospheric NO₂ and HCHO, as well as resulting FNRs, was conducted with airborne remote-sensing observations (GeoTASO and GCAS) during LISTOS 2018. Past studies have focused on the evaluation of satellite retrievals of tropospheric-column NO₂ and HCHO, individually; however, this is the first study to validate the ability of multiple satellite platforms and retrieval algorithms to retrieve tropospheric FNRs. The quantification of satellite-retrieved tropospheric FNRs errors is currently an important, but relatively unknown, uncertainty when applying spaceborne remote-sensing products to investigate O₃ production regimes.

The statistical evaluation of NASA OMI, QA4ECV OMI, and TROPOMI illustrated that all three retrievals have a high bias of clean-region tropospheric-column NO₂ and HCHO concentrations. The satellite retrievals compare more accurately to larger tropospheric-column NO₂ and HCHO values observed in more polluted areas. The magnitude-dependent biases for OMI and TROPOMI NO₂ and HCHO derived in this study agree with other recent validation projects (e.g., Judd et al., 2020; Vigouroux et al., 2020; Zhao et al., 2020; Compennolle et al., 2020; Lamsal et al., 2021; De Smedt et

al., 2021; Verhoelst et al., 2021). Both OMI and TROPOMI retrievals compared well to observed NO₂ throughout the campaign; however, the statistical comparison with observed HCHO data resulted in larger and more variable biases. Overall, daily averaged and campaign-averaged comparisons of the satellite HCHO data to observations displayed large RMSE values, emphasizing the large noise in these retrieval products which hinders the accuracy of FNRs from spaceborne sensors. While all three satellite products at near-native spatial resolutions had low systematic campaign-averaged FNR median biases, suggesting median biases in HCHO and NO₂ data cancel out, the RMSE values for FNRs remained large, primarily due to uncertainty in HCHO and NO₂ retrievals not offsetting. Given the limited measurement sensitivity of shorter UV–Vis wavelengths to HCHO in the middle to lower troposphere, improved information (in situ, remote sensing, or models) of the vertical profiles of HCHO to be used as a priori information would benefit satellite remote-sensing capabilities for observing HCHO and FNRs.

The higher spatial resolution of TROPOMI, along with a good signal-to-noise ratio, allows for this sensor to better capture the spatiotemporal variability and urban–rural interface of tropospheric-column NO₂ and HCHO values and resulting FNRs. This satellite data had the highest correlations with observed NO₂, HCHO, and FNRs throughout the campaign, along with the lowest RMSEs of all three satellite products. The added benefit of TROPOMI spatial resolution is important as this sensor has now been operational for 5+ years and can be applied in trend analysis along with case studies. Future studies of FNR trends should include both OMI and TROPOMI retrievals and determine best practices to fuse the two data sets.

Applying multiple retrieval algorithms to the radiances of a single satellite sensor is of interest in order to determine how input variables (e.g., information on a priori vertical profiles, clouds, surface albedo) impact the retrieval performance. This study evaluated results of OMI retrievals, applying two well-known retrieval algorithms (i.e., the NASA version 4 product and output from the QA4ECV project). Results from the two retrievals were similar for NO₂ but differed primarily in tropospheric-column HCHO, where NASA OMI data had a median bias larger by a factor of 2 compared to QA4ECV. Both retrieval algorithms resulted in large RMSE values indicative of the noise in tropospheric HCHO retrievals. While NASA OMI data displayed less accurate retrievals in HCHO, as well as similar performance for NO₂, NASA OMI data resulted in FNR values with similar biases and uncertainties. Given that both the NASA and QA4ECV retrievals of tropospheric HCHO resulted in noisy data products, this emphasizes the need for an improved signal-to-noise ratio and calibration and improved a priori vertical profile information of HCHO to negate the low measurement sensitivity of HCHO in the middle to lower troposphere for future satellite sensors.

Our study investigated the impact of high-spatial-resolution WRF-CMAQ-predicted NO₂ and HCHO a priori profiles on OMI and TROPOMI retrievals of FNRs. TROPOMI reprocessed data had improved performance when using the higher-spatial-resolution WRF-CMAQ data as the a priori product compared to standard retrievals which apply coarser-resolution TM5 output. In comparison to standard OMI products the reprocessed satellite data using optimized WRF-CMAQ a priori information had similar median biases in FNR values and lower median biases in both indicator species. All reprocessed OMI data variables using optimized simulated shape factors, due to the reduction in unresolved error in retrieved HCHO and NO₂ data, had lower RMSE, higher correlation (except for FNR), and similar or better linear regression slopes compared to observations. These results emphasize the importance of accurate a priori information. Future studies should investigate the impact of a priori profile data sets of various spatial resolutions, ranging from the $\sim 1^\circ \times 1^\circ$ GMI and TM5 model data used for OMI and TROPOMI, respectively, to much higher-resolution air quality model simulations, on the results of reprocessed satellite NO₂ and HCHO retrievals.

Overall, the systematic biases and uncertainties presented in this study can be used in future studies when interpreting the accuracy of OMI and TROPOMI retrievals of FNRs, as well as the two indicator species used for investigating O₃ sensitivity regimes. A main take away from this study is that it is necessary to statistically evaluate both the tropospheric FNRs and the NO₂ and HCHO products individually, as large median biases in both NO₂ and HCHO satellite products can offset, resulting in accurate median FNR values. However, this study emphasizes that uncertainty in NO₂ and HCHO satellite retrievals do not offset in OMI or TROPOMI products, greatly hindering the accuracy of daily scenes of FNRs from these sensors. The large unresolved biases in tropospheric-column HCHO retrievals appear to be the controlling and limiting factor of daily FNR accuracy. While both TROPOMI and OMI captured some of the spatiotemporal variability in observed NO₂ within the LISTOS domain, only TROPOMI is able to capture spatiotemporal HCHO variability with uncertainty low enough for potentially capturing daily FNR variability. The unresolved error in HCHO retrievals from OMI is too large and likely limits the application of these data on a daily basis near the native spatial resolution of the sensor. Overall, the individual satellite products display varying degrees of capability to retrieve tropospheric FNRs, and it is necessary to further validate OMI and TROPOMI retrievals using other field campaign or stationary network data in different regions of the world to identify the primary controlling factors of systematic biases and uncertainty.

Data availability. The primary data sources were the NASA OMI NO₂ (<https://doi.org/10.5067/Aura/OMI/DATA2017>,

Krotkov et al., 2019) and SAO HCHO (<https://doi.org/10.5067/Aura/OMI/DATA2015>, Chance, 2007), QA4ECV OMI NO₂ (<https://doi.org/10.21944/qa4ecv-no2-omi-v1.1>, Boersma et al., 2017) and HCHO (<https://doi.org/10.18758/71021031>, De Smedt et al., 2017; <https://doi.org/10.5194/amt-11-2395-2018>, De Smedt et al., 2018), and TROPOMI PAL NO₂ (<https://data-portal.s5p-pal.com/>, last access: 20 December 2020; Jos van Geffen et al., 2021) and operational HCHO (<https://doi.org/10.5270/S5P-tjlxfd2>, ESA Copernicus Sentinel-5P, 2019) satellite data. For evaluating these satellite products we use airborne remote-sensing data from GeoTASO and GCAS which were downloaded from the LISTOS 2018 campaign data repository (<https://www-air.larc.nasa.gov/missions/listos/index.html>, Janz, 2020).

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Competing interests. At least one of the (co-)authors is a member of the editorial board of *Atmospheric Measurement Techniques*. The peer-review process was guided by an independent editor, and the authors also have no other competing interests to declare.

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