True eddy accumulation – Part 2: Theory and experiment of the short-time eddy accumulation method

Anas Emad and Lukas Siebicke
Bioclimatology, University of Göttingen, Büsgenweg 2, 37077 Göttingen, Germany

Correspondence: Anas Emad (anas.emad@uni-goettingen.de)

Received: 22 May 2022 – Discussion started: 30 May 2022
Revised: 5 December 2022 – Accepted: 13 December 2022 – Published: 4 January 2023

Abstract. A new variant of the eddy accumulation method for measuring atmospheric exchange is derived, and a prototype sampler is evaluated. The new method, termed short-time eddy accumulation (STEA), overcomes the requirement of fixed accumulation intervals in the true eddy accumulation method (TEA) and enables the sampling system to run in a continuous flow-through mode. STEA enables adaptive time-varying accumulation intervals, which improves the system’s dynamic range and brings many advantages to flux measurement and calculation.

The STEA method was successfully implemented and deployed to measure CO₂ fluxes over an agricultural field in Braunschweig, Germany. The measured fluxes matched very well against a conventional eddy covariance system (slope of 1.04, \( R^2 \) of 0.86). We provide a detailed description of the setup and operation of the STEA system in the continuous flow-through mode, devise an empirical correction for the effect of buffer volumes, and describe the important considerations for the successful operation of the STEA method.

The STEA method reduces the bias and uncertainty in the measured fluxes compared to conventional TEA and creates new ways to design eddy accumulation systems with finer control over sampling and accumulation. The results encourage the application of STEA for measuring fluxes of more challenging atmospheric constituents such as reactive species. This paper is Part 2 of a two-part series on true eddy accumulation.

1 Introduction

Monitoring the exchange of trace gases and energy between the Earth’s surface and the atmosphere is a key problem in ecology and climate science. The eddy covariance method (EC) has become the standard method for estimating the flux density on the scale of plant canopies (Baldocchi, 2014; Hicks and Baldocchi, 2020). The flux in the EC method is calculated as the covariance between the vertical wind velocity and the concentration of the atmospheric constituent (\( \geq 10 \text{ Hz} \)). This requirement limits the EC method to a few trace gases for which fast-response gas analyzers are available. For constituents for which only slow-response gas analyzers are available, several methods for measuring the fluxes exist (Rinne et al., 2021). Among these methods, true eddy accumulation (TEA) (Desjardins, 1977) is the most direct and the closest to EC. TEA is formulated using similar principles and assumptions as the EC method. However, unlike EC, the TEA method requires the scalar concentration measurements to be carried out once every averaging interval (30 min). For a long time, the development of the TEA method was hindered by the difficulty of fast airflow rate control and strict operational requirements (Businger and Oncley, 1990; Hicks and McMillen, 1984). A recent improvement to the TEA method used a new type of mass flow controller, online coordinate rotation, and several online treatments of the signal to overcome important limitations of the method’s applicability (Siebicke and Emad, 2019). The new system showed a good match with a reference eddy covariance system, with coefficients of determination of up to 86 % and a slope of 0.98. While this study demonstrated a successful proof of concept of TEA using modern sampling, it also showed that further research was required for continuous accumulation and long-term field operation, which we address with the current study.
The absence of high-frequency measurements of the scalar concentration creates unique challenges for the TEA method. The sampling decisions in TEA need to be made in real time without complete knowledge of the wind statistics of the averaging interval. The problem of nonzero mean vertical wind velocity, a direct consequence of this limitation, is discussed in the companion paper (Emad and Siebicke, 2023).

Furthermore, the lack of high-frequency scalar measurements implies that sample accumulation needs to happen on a timescale similar to the flux averaging interval (30 to 60 min). Therefore, a minimum limit is imposed on the sampling accumulation interval before the scalar concentration measurement can be conducted. This time limit imposes restrictive design considerations related to the size and function of sample accumulation reservoirs. It also dictates that the sampling apparatus needs to accommodate a large dynamic range (up to 5 $\sigma_w$) to cover the range of wind velocities during flux averaging intervals (Hicks and McMillen, 1984). The minimum time limit is also problematic if the sampled scalar changes in concentration over time, e.g., reactive species. Additionally, the accumulation for long time intervals and the discontinuous nature of sample collection are particularly sensitive to nonstationary conditions in the accumulation apparatus (Siebicke and Emad, 2019). Furthermore, the use of expandable bags in discrete sampling for the accumulation reservoirs was found to be unreliable and prone to mechanical fatigue (Siebicke and Emad, 2019). Therefore, a more flexible approach is needed whereby the accumulation interval can be adapted to the requirements of the sampling system and to the trace gas being measured.

In this paper, we address the limitations of fixed accumulation intervals in TEA by developing a novel method for eddy accumulation and providing a prototype implementation of such a system. First, we derive a new eddy accumulation method, which we call short-time eddy accumulation (STEA). The STEA method enables the sample accumulation to be carried out at variable shorter intervals, which brings many improvements to the TEA method including the ability to accumulate samples in a continuous flow-through mode and an increased dynamic range. Next, we discuss the effect of using buffer volumes on the concentration measurements and develop an empirical correction for the use of buffer volumes. Finally, we show a prototype and experimental measurements for CO$_2$ fluxes using the newly developed STEA method in the flow-through mode and compare the measured fluxes to reference EC measurements. We discuss the advantages and steps required to carry out flux measurements using the STEA method, different constraints, and operational requirements.

2 Theory

A detailed description of the TEA method derivation and assumptions is provided in the companion paper (Emad and Siebicke, 2023). Here we provide a brief overview of the TEA method and the assumptions that are required for the derivation of the short-time eddy accumulation method.

Under the assumptions of flow homogeneity and stationarity, the vertical exchange of the atmospheric scalar $c$ is the flux across the measurement plane at height $h$. The flux $F_c$ is expressed as (Finnigan et al., 2003; Gu et al., 2012)

$$ F_c = \overline{w c}. $$

(1)

Here, $w$ is the vertical wind velocity (m s$^{-1}$), $c$ is the scalar density (mol m$^{-3}$), and the overbar denotes time averages that follow Reynolds' averaging rules.

The true eddy accumulation method is formulated by partitioning the average $\overline{w c}$ using the direction of the vertical wind velocity. Therefore, we write the flux as the expected value of the random variable $wc$ conditional on the sign of the vertical wind velocity, sign$(w)$,

$$ \overline{w c} = \overline{w c \uparrow} P(w \uparrow) + \overline{w c \downarrow} P(w \downarrow), $$

(2)

where the arrows denote the direction of the vertical wind velocity, with $\uparrow$ for updrafts and $\downarrow$ for downdrafts. $P(w \uparrow)$ is the probability that the observed wind velocity is in the respective direction. The TEA method makes use of this simple partitioning by physically realizing the terms $w \uparrow c \uparrow$ and $w \downarrow c \downarrow$ using sample accumulation instead of measuring individual realizations of $w$ and $c$. For the practical implementation of a TEA system, a parameter $A$ is necessary to relate the sampling flow rate to the measured $w$.

2.1 Short-time eddy accumulation

The original formulation of the true eddy accumulation method requires the samples to be accumulated for the entire averaging interval $\Delta t$ before the concentration measurement is ready for flux calculation. This formulation poses a challenge for the practical implementation of the TEA method. First, the longer averaging times require the sampling apparatus to cover a larger range of wind speeds (high dynamic range). And, second, the fixed averaging times limit the flexibility of the sampling system to adapt to changing conditions, therefore making it more prone to flow nonstationarities.

To achieve a higher dynamic range for the sampling system and realize a more robust flow-through eddy accumulation system, we propose a modification to the TEA method whereby samples can be accumulated for a sequence of shorter intervals $\tau_i$ that add up to the averaging period $\Delta t$. Therefore, the flux $\overline{w c}$, and consequently the sample accumulation, is partitioned based on two conditions: the sign of the vertical wind velocity and the variable $I$ that divides the averaging interval into a sequence of shorter intervals $\tau_i$.

This formulation can be achieved by applying the law of total expectation to the random variable $cw$ with respect to a partitioning variable $I$ that divides the averaging period $\Delta t$ into multiple non-overlapping partitions with the length $\tau_i$. 

This partitioning scheme is applied independently to updrafts and downdrafts, i.e., after partitioning with the direction of vertical wind velocity, therefore allowing the short intervals for updrafts and downdrafts to be different. We write the expectation of \( c^\dagger w^\dagger \) as

\[
c^\dagger w^\dagger = \left( (c^\dagger w^\dagger) | I \right) = \sum_i (c^\dagger w^\dagger) | I_i \ P(I_i). \tag{3}
\]

The previous equation is similarly valid for the downdraft flux \( c^\dagger w^\dagger \). The measured concentration during a short averaging interval \( i \) is given by

\[
C_i = \frac{cw|I_i}{|w|}. \tag{4}
\]

The probability of the short averaging interval can be obtained easily as \( P(I_i) = \tau_i / \Delta t \).

\( V_i \) is the volume accumulated during the short interval \( i \), defined as

\[
V_i = A_i \int_i^{i+\tau_i} |w| \ dr. \tag{5}
\]

The concentration in either updraft or downdraft reservoirs for the averaging interval \( \Delta t \) is the weighted mean of the short interval concentration measurements, \( C_i \):

\[
C^\dagger_{\text{acc}} \equiv \frac{1}{|w| \Delta t} \sum_i j C^\dagger_i |w| |\tau_i|, \tag{6}
\]

We notice here that \( |w| |\tau_i| = V_i / A_i \) and \( \bar{\omega} \Delta t = \sum_{i=1} C^\dagger_i / A_i \).

The obtained \( C^\dagger_{\text{acc}} \) and \( C^\dagger_{\text{acc}} \) can be used to calculate the STEA flux (Emad and Siebicke, 2023) as

\[
F_{\text{STEA}} = \frac{C^\dagger_{\text{acc}} \ V^\dagger \left( |w| - \bar{w} \right) - C^\dagger_{\text{acc}} \ V^\dagger \left( |w| + \bar{w} \right)}{|w| - \alpha_e \bar{w}} \times \frac{|w|}{V_{\text{total}}}, \tag{7}
\]

where \( F_{\text{STEA}} \) is the kinematic flux density (mol m s\(^{-1}\)). \( C^\dagger_{\text{acc}} \) and \( C^\dagger_{\text{acc}} \) are the mean molar densities (mol m\(^{-3}\)) of the scalar \( c \) in updraft and downdraft reservoirs for the whole accumulation period \( \Delta t \) as calculated from Eq. (6). \( V^\dagger \) and \( V^\dagger \) are the accumulated sample volumes (m\(^3\)) in updraft and downdraft reservoirs during the averaging period. It is important here to use \( V^\dagger \) and \( V^\dagger \) as \( |w| \Delta t \) since the parameter \( A \) was not constant for different short intervals, \( |w| \) is the mean of the absolute vertical wind velocity (m s\(^{-1}\)) during the averaging period, \( \bar{w} \) is the mean of the vertical wind velocity, \( \alpha_e \) is the transport asymmetry coefficient for the scalar \( c \) (dimensionless) and is defined as the ratio of the covariance \( c^\dagger |w|^\dagger \) to the flux \( c^\dagger \bar{w} \bar{v} \). Methods for estimating \( \alpha_e \) and the derivation of the TEA flux equation are discussed in the companion paper (Emad and Siebicke, 2023).

### 2.2 Effect of buffer volumes

The short-time eddy accumulation method can be achieved in at least two ways, either using expandable buffer volumes (e.g., bags), which are emptied after each short interval measurement \( C_i \), or using a flow-through system with rigid buffer volumes. The flow-through system has practical operational benefits but requires additional correction to reverse the effect of buffer volumes on the scalar concentration signal. Buffer volumes act as low-pass filters (Cescatti et al., 2016). They attenuate the magnitude of the high-frequency part and shift the phase of the signal. The buffer concentration at time step \( n \) is dependent on the new input sample concentration and the buffer concentration from the previous step \( y[n - 1] \). Thus, the buffer volume concentration \( y_n \) response to an input \( C_i \) can be described with the following linear difference equation:

\[
y[n] = C_i[n] \dot{q} + (1 - \dot{q}) y[n-1], \tag{8}
\]

where \( \dot{q} \) is a dimensionless flow rate that is defined as the ratio between the sample mass to the total mass of air in the buffer volume at each time step \( n \). Therefore, the dimensionless flow rate is the fraction of the air mass in the buffer volume that is replaced by the new sample mass.

\[
\dot{q}_n = \frac{V_i \rho_i}{V_b \rho_b} \tag{9}
\]

Here \( V_i \) and \( \rho_i \) are the volume and density of the accumulated sample during the interval \( i \), respectively. \( V_b \) and \( \rho_b \) are the volume and the air density of the air in the buffer volume, respectively. Equation (8) characterizes a first-order linear filter.

The mixing in the buffer volume is assumed to be instantaneous and perfect. Additionally, the accumulated short samples in the STEA method are considered individually separable homogeneous parcels of air as they are forwarded to the gas analyzer. This discrete behavior is best modeled with a discrete-time system as shown in Eq. (8). The system response is characterized by the dimensionless flow rate or the time constant \( \tau \). The time constant of the system is defined as the required time for the system to reach \( 1/e \) from a step increase and relates to \( \dot{q} \) by

\[
\tau = -\frac{\Delta s}{\ln(1 - \dot{q})}. \tag{10}
\]

(Taylor et al., 2013) where \( \Delta s \) is the length of the sampling interval.

Figure 1 shows the filter’s magnitude and phase responses. The magnitude response \( |H| \) plot shows how the magnitudes of different frequencies are attenuated. The smaller the dimensionless flow rate, the larger the time constant and the stronger the attenuation.
3 Methods

3.1 Experimental site

Flux measurements were performed over a flat agricultural field of the Thünen Institute, located at 52.297° N, 10.449° E in Braunschweig, Germany. The site has an altitude of 76 m above sea level. During the measurement period, the fields south and north of the tower were planted with oats and corn, respectively. Both crops had a similar height of approximately 50 cm above the ground at the start of the comparison period.

3.2 Experiment period

Fluxes were measured throughout the year 2020. We selected 6 weeks of good quality in summer based on instrument performance and weather conditions, spanning from 18 June 2020 to 31 July 2020, to compare the different methods. Meteorological conditions (Fig. 3) during the experimental period from 18 June to 31 July 2020 were characterized by warm weather conditions with net radiation peaking around 600 W m\(^{-2}\) at noon. Air temperature was predominantly above 10° and averaged 18°. Several precipitation events were observed during the experiment period. Precipitation totaled 66 mm, with several high precipitation events, in particular starting from the second and third weeks of the experiment. Precipitation data were obtained from the German weather service (DWD) station i Braunschweig (number: 00662), which is located 600 m from the measurement tower. Soil water content tracked precipitation events except for one distinct occasion on 27 June when precipitation is not registered at the DWD station. Wind direction was dominated by southerly and easterly winds.

3.3 Instruments

EC and STEA measurement complexes were mounted at 5 m height above the ground (Fig. 2). The instruments used in the experiment for flux measurements and data analysis are listed in Table 1. Meteorological variables were logged using a Sutron 9210 XLite logger (Sterling, USA). All the raw data needed for flux processing were synchronized on the STEA computer and remote servers for real-time processing.

The EC system comprised a dedicated sonic anemometer (uSonic-3 Omni H) and an open-path infrared gas analyzer (IRGA). Wind and scalar density data were acquired at 20 Hz frequency. Relative to the Class-A sonic anemometer used for STEA, the northward, eastward, and vertical separation of the IRGA was \(-17, 26,\) and \(-15\) cm, respectively. The Class-A sonic anemometer had a north offset azimuth of 90°. Relative to the Omni sonic anemometer used for EC, the northward, eastward, and vertical separation of the IRGA was \(20, -15.3,\) and \(-20\) cm. The north offset of the Omni sonic anemometer was 169°.

3.4 STEA system description

The STEA system used in the experiment is based on an earlier system of Siebicke and Emad (2019). The new system used the same mass flow controllers and shared most of the operating software. It has, however, several differences and improvements. One major difference is the use of fixed stainless-steel buffer volumes instead of expandable bags. The system was initially developed as a hybrid TEA–EC method to run the TEA method in a continuous flow-through mode (Siebicke, 2016). The system was set up to operate in the STEA continuous flow-through mode. A constant duration for the short intervals (\(\tau_i\)) equal to 1 min was used. The STEA system is comprised of two identical sampling lines, one for updrafts and one for downdrafts. Each of the sam-
Table 1. Variables and instruments. Manufacturer key: METEK GmbH (Elmshorn, Germany), LI-COR Environmental Inc. (Lincoln, Nebraska, USA), LGR, (Los Gatos Research Inc., USA), Bosch (Bosch Sensortec GmbH, Germany), Vaisala (Helsinki, Finland), Kipp & Zonen (Delft, the Netherlands), Delta-T Devices Ltd (UK), Stevens Water Monitoring Systems, Inc (Oregon, USA), Texas Electronics (Dallas, USA).

<table>
<thead>
<tr>
<th></th>
<th></th>
<th></th>
<th></th>
<th></th>
</tr>
</thead>
<tbody>
<tr>
<td>Wind u, v, w</td>
<td>uSonic-3 Omni H</td>
<td>METEK</td>
<td>EC</td>
<td>20 Hz</td>
</tr>
<tr>
<td>Sonic temp. T</td>
<td>uSonic-3 Omni H</td>
<td>METEK</td>
<td>EC</td>
<td>20 Hz</td>
</tr>
<tr>
<td>Wind u, v, w</td>
<td>uSonic-3 Class A</td>
<td>METEK</td>
<td>TEA</td>
<td>10 Hz</td>
</tr>
<tr>
<td>Sonic temp. T</td>
<td>uSonic-3 Class A</td>
<td>METEK</td>
<td>TEA</td>
<td>10 Hz</td>
</tr>
<tr>
<td>CO₂ density</td>
<td>LI-7500A</td>
<td>LI-COR</td>
<td>EC</td>
<td>10 Hz</td>
</tr>
<tr>
<td>H₂O density</td>
<td>LI-7500A</td>
<td>LI-COR</td>
<td>EC</td>
<td>10 Hz</td>
</tr>
<tr>
<td>CO₂ ppm</td>
<td>FGGA-24r-EP</td>
<td>LGR</td>
<td>TEA</td>
<td>1 Hz</td>
</tr>
<tr>
<td>H₂O ppm</td>
<td>FGGA-24r-EP</td>
<td>LGR</td>
<td>TEA</td>
<td>1 Hz</td>
</tr>
<tr>
<td>CH₄ ppm</td>
<td>FGGA-24r-EP</td>
<td>LGR</td>
<td>TEA</td>
<td>1 Hz</td>
</tr>
<tr>
<td>Air pressure P</td>
<td>BME280</td>
<td>Bosch</td>
<td>TEA</td>
<td>50 Hz</td>
</tr>
<tr>
<td>Air temperature</td>
<td>BME280</td>
<td>Bosch</td>
<td>TEA</td>
<td>50 Hz</td>
</tr>
<tr>
<td>Air humidity</td>
<td>HMP155</td>
<td>Vaisala</td>
<td>Meteo</td>
<td>10 min</td>
</tr>
<tr>
<td>Air temperature</td>
<td>HMP155</td>
<td>Vaisala</td>
<td>Meteo</td>
<td>10 min</td>
</tr>
<tr>
<td>Net radiation</td>
<td>CNR4</td>
<td>KIPP</td>
<td>Meteo</td>
<td>10 min</td>
</tr>
<tr>
<td>Global radiation</td>
<td>BF5</td>
<td>DELTA-T</td>
<td>Meteo</td>
<td>10 min</td>
</tr>
<tr>
<td>Soil heat flux</td>
<td>HFP01</td>
<td>LI-COR</td>
<td>Meteo</td>
<td>10 min</td>
</tr>
<tr>
<td>Soil moisture</td>
<td>SDI-12</td>
<td>Stevens</td>
<td>Meteo</td>
<td>10 min</td>
</tr>
<tr>
<td>Precipitation</td>
<td>TR-525M</td>
<td>Texas Elec.</td>
<td>Meteo</td>
<td>10 min</td>
</tr>
</tbody>
</table>

The STEA sampling inlets were installed near the sonic anemometer’s center of the measurement volume. The horizontal separation was 22 cm, while the vertical separation between the two inlets was 2 cm. Upon sampling, the collected samples were carried using 6 mm Teflon tubes to the first set of buffers. The sampling can be summarized in the following steps (see a detailed description of the system operation and sampling in Siebicke and Emad, 2019).

1. 3D wind measurements are acquired from the sonic anemometer (uSonic-3 Class A) with a 10 Hz sampling frequency.

2. Wind coordinates are rotated into the streamline coordinates using the planar fit method without an intercept (Dijk et al., 2004). The fit is performed online as a running window operation with a window width of 2 d and an update frequency of once every 30 min.

3. The mean vertical wind from the previous 30 min interval is removed to minimize $\overline{\bar{v}}$. This is equivalent to applying a high-pass filter to the vertical wind velocity measurements.

4. The active sampling line is determined (updraft or downdraft) based on the direction of the rotated vertical wind velocity component.

5. The sampling scaling factor $A_i$ is calculated based on wind conditions in the near past and the calibration coefficients of the mass flow controllers. The scaling factor should be constant during the short accumulation intervals.

6. Air samples are collected, and the controllers are adjusted to collect an air sample with a volume equal to $A_i |\bar{w}|$.

7. When enough sample volume is accumulated in the respective buffer volume, samples are forwarded to the gas analyzer for analysis. The amount of sample volume needed is determined based on the required flow rate for the gas analyzer and the time needed to flush the tubes and the measurement cell as well as to perform enough repeated measurements.

8. Mean concentrations of accumulated samples are measured. The slow gas analyzer (LGR FGGA-24r-EP) alternates measuring the concentrations $C_i$ of the accumulated samples for updraft and downdraft. The accumulation time for the short intervals was set to a fixed interval of 1 min instead of an adaptive interval duration. During each short interval, the gas analyzer performs repeated measurements for the gas concentration. The observed variability for repeated measurements in the short averaging intervals was SD = 0.501 ppm, which was similar to the measured repeatability of the gas analyzer for a similar time interval.
Figure 3. Meteorological conditions and turbulent energy fluxes during the experiment period from 18 June to 31 July 2020: air temperature, relative humidity, soil water content, cumulative daily precipitation, wind velocity, wind direction, sensible heat flux ($H$), and latent heat flux (LE). Precipitation data were obtained from the German weather service (DWD) station in Braunschweig (number: 00662), which is located 600 m from the measurement tower.

3.5 STEA flux computations

This section describes the steps followed to obtain the final and corrected STEA flux. Firstly, we discuss the effect of water vapor on the measured concentrations of other scalars and how we corrected the remaining water cross-sensitivity. Then, we present the procedure of data quality screening. Next, we detail the steps of calculating the final STEA flux. Finally, we present the buffer volume empirical correction we applied.

3.5.1 Water vapor correction

The gas analyzer used for the STEA measurements (LGR FGGA-24r-EP) reports the molar fraction of CO$_2$ and CH$_4$ of moist air in parts per million (ppm). The measurements of CO$_2$ cannot be used directly, as they are affected by the presence of water vapor. The presence of varying water vapor concentrations in the sample affects the measurements of CO$_2$ and CH$_4$ in cavity ring-down spectroscopy instruments in at least two ways: (i) the dilution effect and (ii) spectroscopic line broadening (Rella, 2010). Rella (2010) proposed a quadratic equation to correct for the combined effect of line broadening and water vapor dilution. The correction involves estimating the parameters ($a$) and ($b$) in the equation

$$r_c = \frac{\chi_c}{1 + a \chi_w + b \chi_w^2},$$

where $r_c$ is the dry mole fraction of the species $c$, $\chi_c$ is the wet mole fraction measured by the instrument, and $\chi_w$ is the water mole fraction measured by the instrument. For CO$_2$ measured by the LGR gas analyzer in parts per million, Hiller et al. (2012) experimentally estimated these coefficients as $a = -1.219 \times 10^{-6}$ and $b = 1.229 \times 10^{-12}$. We found that using the same parameters could not control for all the effects of water vapor on measured CO$_2$ signals. A linear slope different from zero was still found when supplying the gas analyzer with air of varying water concentration and of constant CO$_2$. This suggested a remaining cross-sensitivity of CO$_2$ to the presence of water vapor. To control for this small remaining cross-sensitivity, we conducted a field experiment in which we measured the CO$_2$ concentration in air of varying water concentration and then used the results in a linear fit to obtain a correction slope. We were not able to source the necessary equipment and gas cylinders to supply the gas analyzer with air of known CO$_2$ concentration and varying water vapor in the field. Instead, we used the system’s buffer volumes to collect and pressurize ambient air from the atmosphere, closed the inlets, and supplied the gas analyzer with enough sample flow rate for measurement. This procedure utilizes the effect of air drying due to decompression to deliver a varying water vapor content. The experiment involved collecting ambient humid air near saturation (RH $\approx$ 90%, $T = 21^\circ$C) in the system’s buffer volumes to a pressure of 2.6 bar. As a result of the high pressure, the water partial pressure in the pressurized buffer volumes will become higher than the saturation vapor pressure and water will precipitate, leading to dryer air. Air is then decompressed and forwarded to the analyzer. As the buffer pressure is decreasing, water vapor content will increase to reach the same level of atmospheric humidity. Using this method we were able to modulate the water vapor content in the air from 6000 to 14 000 ppm. The accumulated sample was enough to supply the gas analyzer for ca. 10 min. We repeated the measurements several times and used the obtained dataset to correct the remaining cross-sensitivity using a linear fit.
Figure 4. Functional and pneumatic schematic of the implemented flow-through STEA system showing components, layout, properties, and operation conditions. Air samples are collected at the input and travel in distinct sampling lines for updrafts and downdrafts. Samples travel through tubes (lengths are shown) and through filters and are then collected into two sets of buffer volumes shown here as “first BV” and “second BV” separated by two vacuum pumps. The “output flow valves” followed by mass flow controllers (MFCs) control the output flow rate from the second set of buffers to the gas analyzers. Finally, samples can optionally be forwarded to a set of mass flowmeters (MFMs) used for calibration purposes. The colored bottom bar below shows the range of pressure values at each stage.

3.5.2 Raw data quality screening

Raw measurements of the wind velocity and scalar concentration were screened for outliers due to measurement errors and instrument malfunction. This included the following steps.

- Statistical screening: despiking, dropout removal (Vickers and Mahrt, 1997), and plausibility limits of the raw gas analyzer and wind measurements (Sabbatini et al., 2018).

- Flushing time removal: measurement of the short interval events involves regularly switching the sampling line coming to the gas analyzer between updraft and downdraft reservoirs. This caused subsequent samples to get contaminated. We experimentally chose a 25 s threshold at the start of each short interval event to account for the flushing time. The measurements falling before the threshold were discarded. Figure 5 shows an example of discarded flushing times at the start of each averaging interval.

- Detection of sample contamination: periods during which the flow rate to the gas analyzer is smaller than 400 mL min\(^{-1}\) are flagged. Under these conditions, ambient air might enter the system and contaminate the collected samples. When the number of flagged data points exceeded 10% of the total points in the sampling interval, data in the sampling interval were discarded.

3.5.3 STEA flux calculation

After measurements are quality-checked and erroneous data points are excluded, the final STEA flux is calculated as follows.

- Short interval statistics: for each short interval sample, the gas analyzer will have several repeated measurements for the concentrations \(C_i\). However, only one value is needed for the flux calculation. We use the median to obtain the representative value in order to minimize uncertainty and exclude outliers. Figure 5 shows an example of data quality checking and choice.

- Calculate air molar volume: the molar volume of air is needed to express the flux in units of mol m\(^{-2}\) s\(^{-1}\). The molar volume is calculated using sonic temperature, pressure, and humidity measurements.

- Calculate short interval weights: following Eq. (6), the measured short interval concentration should be weighted by the ratio of the accumulated volume during that interval to the total buffer volume.

- Calculate values of \(\alpha_\theta\): values of the transport asymmetry coefficient \(\alpha_\theta\) are calculated using vertical wind velocity and sonic temperature measurements. Values of \(\alpha_\theta\) larger than 1 are discarded as they indicate a problem with the measurement as discussed in the companion paper (Emad and Siebicke, 2023).

- Calculate updraft and downdraft mean concentrations: \(C_{\text{acc}}^\uparrow\) and \(C_{\text{acc}}^\downarrow\) are calculated for the averaging period \(\Delta t\).

- Calculate the flux: the STEA flux equation shown in Eq. (7) is used to obtain the final flux.
3.5.4 Buffer volume empirical correction

Buffer volumes act on the signal as a low-pass filter and introduce systematic bias to the fluxes. We used Eq. (10) to estimate the time constant of the buffer volumes used in our experiment. For each of the buffer volumes, a measurement point is acquired every 2 min. The mean dimensionless mass flow rate to the gas analyzer was estimated from the pressure, the volume, and the estimated volumetric flow rate to the gas analyzer. We simulated the effect of buffer volumes on the high-frequency sonic temperature signal and parameterized the flux loss by artificially degrading the sonic temperature in a procedure similar to Goulden et al. (1996) and Berger et al. (2001).

3.6 EC reference flux measurements and computations

The raw data from the two sonic anemometers and the high-frequency gas density measurements from the IRGA were used to compute eddy covariance fluxes for water vapor and CO₂ in the period from 1 April 2020 to 1 November 2020 using EddyPro® software (LI-COR Env. Inc. USA) version 7.0.4. The flux processing steps were chosen to be as similar as possible to the STEA processing scheme. The calculation of EC fluxes involved statistical screening for the data quality issues following Vickers and Mahrt (1997), mean removal by block averaging, compensation for the time lag between the wind and the scalar time series using covariance maximization, tilt correction using the planar fit method without an intercept (Dijk et al., 2004) similar to STEA, and analytical high- and low-frequency corrections to correct for the spectral attenuation of the IRGA (Moncrieff et al., 2005, 1997).

3.6.1 Density fluctuation correction

Due to using a closed-path gas analyzer, the TEA and STEA methods do not require the WPL (Webb, Pearman, and Leuning) correction (Webb et al., 1980). WPL accounts for the effect of density fluctuations due to changes in temperature, humidity, and pressure. In TEA and STEA, after samples are collected and mixed in buffer volumes, the mean mixing ratio is measured. Therefore, no correction for density effects is needed as long as accurate mass flow sampling of air is maintained. The measured TEA and STEA flux is equivalent to the flux measured with mixing ratios \( \frac{2}{3}w^2 \).

3.7 Data selection for method comparison

To compare the fluxes calculated from both methods, we selected averaging intervals according to the following criteria.

- Spike removal: this is done following Vickers and Mahrt (1997) using a window width of 6 h and a threshold of 2 standard deviations. This was mainly to account for unreliably elevated CO₂ concentrations recorded by the open-path gas analyzer due to water condensation.

Table 2. Summary of data quality checks for STEA and EC fluxes used in the EC–STEA flux intercomparison showing the number of averaging intervals that were excluded and the ratio of the excluded averaging intervals to the total for each criterion. Details on the criteria and the thresholds used are provided in Sect. 4.3.

<table>
<thead>
<tr>
<th>Criteria</th>
<th>Averaging intervals</th>
<th>Ratio (%)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Spikes</td>
<td>3</td>
<td>0.2</td>
</tr>
<tr>
<td>EC missing value</td>
<td>16</td>
<td>0.9</td>
</tr>
<tr>
<td>Technical failure</td>
<td>38</td>
<td>2.1</td>
</tr>
<tr>
<td>Rain</td>
<td>91</td>
<td>5.0</td>
</tr>
<tr>
<td>STEA low flow rate</td>
<td>107</td>
<td>5.9</td>
</tr>
<tr>
<td>Flux quality flag 2</td>
<td>195</td>
<td>10.7</td>
</tr>
<tr>
<td>Flux quality flag 1</td>
<td>382</td>
<td>20.9</td>
</tr>
<tr>
<td>OK data</td>
<td>992</td>
<td>54.4</td>
</tr>
</tbody>
</table>

- Rainy period exclusion: data records during rainy weather conditions were excluded.

- Flux quality flags: periods when the flux quality flag is 1 or 2 according to Foken et al. (2005) were excluded.

- STEA low flow rate: averaging intervals flagged with the low flow rate flag described earlier were discarded.

After applying the above criteria, 992 averaging intervals remained. They accounted for 54.4 % of the whole comparison period. Nighttime data were the majority of excluded values with only 33 % of averaging intervals valid during nighttime compared to 70 % during daytime. The open-path gas analyzer used for EC produced unreliable measurements during high humidity conditions at night due to water condensation. Table 2 shows a summary of data quality-check results.

To compare the overall difference between the two methods, we used the coefficient of determination \( R^2 \) and the slope of the orthogonal distance regression (ODR) (also known as major-axis regression and model II regression). ODR considers the errors in \( x \) and \( y \) as opposed to ordinary least squares (OLS) regression, which assumes that the error in \( x \) is negligible (Wehr and Saleska, 2017).

4 Results and discussion

We first discuss the newly proposed short-time eddy accumulation method. Then, we discuss some results and aspects of the STEA flux calculations. Afterward, we present the flux intercomparison between STEA and EC. Finally, we discuss the effect of using fixed buffer volumes on the fluxes and the proposed empirical correction.

4.1 Short-time eddy accumulation

Using the STEA method reduced the dynamic range requirement for eddy accumulation sampling. For a short averaging
interval of 1 min, the range was on average 60% of the range required for the conventional eddy accumulation. As a result, the upper bound of the required dynamic range for $w$ reported by Hicks and McMillen (1984) as $5 \sigma_w$ is lowered to $3 \sigma_w$. The reduction of the required dynamic range improves the accuracy and performance of the STEA system. The accumulation on shorter timescales brings many advantages. First, it allows adapting to the local range of vertical wind velocity values, which improves the resolution and dynamic range of the system. This can be achieved by exploiting the autocorrelation of the wind velocity signal to predict a scaling parameter, $A_i$, better adapted to the local velocity field for each interval. For a short interval, the range that the sampling apparatus needs to cover will be smaller on average than the range of the whole averaging interval.

Additionally, the accumulation on varying intervals means the measurement frequency can be adjusted to match that of the gas analyzer or the precision requirements. This can be useful for reactive species and other trace gases, for which relatively fast gas analyzers are available but not fast enough for EC.

Figure 6 demonstrates how the STEA method works. In this example, the high-frequency samples are collected at 5 Hz frequency for a 30 min averaging interval. The averaging interval is divided into 30 short intervals with a duration varying from 70 to 190 s. The flux in this example equals $-14.24 \mu$mol m$^{-2}$ s$^{-1}$.

Finally, the STEA method facilitates using the STEA system in a continuous flow-through mode using rigid reservoirs. The operation in flow-through mode requires two sets of buffer volumes in a series as shown in Fig. 4, with two buffer volumes for each sampling line. The ideal operation of such a system can be achieved as follows.

1. Wind velocity is measured and rotated, and the value of the scaling parameter $A_i$ is updated based on wind statistics and the flow calibration parameters.
2. For each sampling line, air samples are collected into the respective set of buffer volumes continuously according to the sign of the vertical wind velocity and proportional to its magnitude and the value of $A_i$ until a predefined accumulated volume is reached.
3. When the predefined accumulated volume is reached, the second buffer volume in the sampling line is disconnected from the first. Sample accumulation time, $\tau_i$, and accumulated mass are recorded. Then, samples are forwarded to the gas analyzer.
4. The slow gas analyzer alternates measuring scalar concentration for each interval $C_i$ from the second set of buffer volumes for updraft and downdraft.

The successful use of this scheme requires keeping the mass flow rate of air from the second set of buffer volumes to the gas analyzer constant for consecutive short intervals since the model used to represent the buffer volumes in Eq. (8) assumes the flow rate to be constant with respect to time.

### 4.2 STEA fluxes computations

In this section, we discuss some aspects related to the calculation of the STEA fluxes. We first discuss the effects of water vapor on CO$_2$ concentration measurements. Then, we discuss the effect of coordinate rotation on the fluxes.

#### 4.2.1 Water vapor correction

Treatment of the residual cross-sensitivity of CO$_2$ to water vapor content using a linear fit produced a small slope of
Figure 6. Sample accumulation using the STEA method. An example of 30 min of measurements: (a) samples \( w_c \) are collected based on wind direction and proportional to its magnitude. (b) Short intervals are accumulated. The variable short interval duration guarantees equal accumulated volume for consecutive short intervals. Points are the concentrations \( C_i \) measured by the gas analyzer. The area of each rectangle represents the accumulated sample volume in arbitrary units and is equal to the relative weight for each concentration measurement. The sum of all measurements \( C_i \) weighted by the relative sample volume will equal the covariance. Data are from 20 June 2020.

\[ -1.17 \times 10^{-4} \] shown in Fig. (7). Thus, a difference in water concentration of 4000 ppm between updraft and downdraft reservoirs, typically observed in extreme conditions, will lead to a difference on the order of 0.5 ppm for CO\(_2\). Applying the water vapor correction using the quadratic fit and the slope correction reduced the magnitude of STEA fluxes in comparison to the direct calculation of mixing ratios. However, it improved the fit between the STEA and the reference EC flux (slope decreased from 1.18 to 1.04, and \( R^2 \) increased from 0.80 to 0.86).

4.2.2 Coordinate rotation

The online coordinate rotation produced rotation angles with low variability over the experiment period. The eddy covariance fluxes calculated using the Class-A sonic anemometer with a 2-month-long dataset (1 June 2020 to 1 August 2020) produced the following rotation angles: \( x - \) pitch = 0.6\(^\circ\); \( y - \) roll = \(-4.3\)\(^\circ\) (using the YXZ Euler convention), whereas for the TEA moving-window online rotation, larger pitch angles were observed with a mean of 3.6\(^\circ\) and values slowly climbing from 1.2 to 6\(^\circ\) during the 6-week comparison period. The roll angle ranged from \(-0.9\) to \(-0.24\)\(^\circ\) with an average of \(-0.4\)\(^\circ\).

The use of online rotation with a moving window of 2 d minimized the residual mean vertical wind in comparison to using the whole period of the experiment. This is likely due to a better adaptation to the local wind field. Furthermore, the distribution of normalized mean vertical wind velocity of the short moving window had less spread and thinner tails, and it showed more symmetry around the mean compared to the whole-dataset rotation. The residual mean magnitude of rotated \( w \) for the short moving window was 0.04 \( \sigma_w \). The first and third quartiles were \(-0.03\) and 0.03 \( \sigma_w \), whereas for the whole-dataset rotation, the mean magnitude was 0.17 \( \sigma_w \) and the first and third quartiles were \(-0.07\) and 0.22 \( \sigma_w \), respectively.

To estimate the effect of the online rotation method on the fluxes, we calculated EC fluxes using the two different rotation approaches while keeping other treatments constant. The comparison revealed that the online rotation with a moving window had a minimal effect on the fluxes: a slope of approximately 1 and an \( R^2 \) of 0.98 were obtained when using a linear fit. Nevertheless, this comparison only included data of good quality from an ideal site. These results might differ for nonideal conditions at a more complex site.
4.3 STEA–EC flux intercomparison

The measured CO₂ fluxes using the STEA method in flow-through mode showed a good match with the reference EC fluxes (Fig. 8).

The time series of measured CO₂ fluxes in Fig. 8(a) shows that the STEA method was able to reproduce the daily dynamics of CO₂ flux very well. The estimated fluxes using the STEA method appear to have fewer spikes and are smoother in general; this is likely due to the smoothing effect of buffer volumes and the lower sensitivity of the closed-path gas analyzer to rain and high humidity, in particular during nighttime. The correction for nonzero mean vertical wind velocity using \( \alpha \) was on average less than 1.5% of the flux magnitude. This is due to the ideal topography of the site and the online rotation of the coordinates. The correction at less ideal sites with more complex topography may differ.

The mean diurnal cycle estimates from the two methods match very well (Fig. 8b). However, a small time shift can be observed in the mean diurnal cycle as a result of the phase shift introduced by the low-pass-filtering effect of the buffer volumes.

The linear regression in Fig. 8c shows that the measured CO₂ fluxes using the STEA method in flow-through mode have very good agreement with the reference EC fluxes. The magnitude of STEA fluxes was comparable to EC fluxes (ODR slope = 1.04). This indicates that the STEA method does not introduce systematic error to the fluxes.

Cumulative fluxes for the entire 6-week experiment period show generally good agreement between STEA and EC (Fig. 9). The cumulative CO₂ flux was estimated to be \( -168.3 \pm 1.42 \text{ gr C m}^{-2} \) using EC and \( 179.3 \pm 1.42 \text{ gr C m}^{-2} \) using STEA. The difference between the two methods is \( -10.9 \text{ gr C m}^{-2} \), which is about 6% of the cumulative flux. While this difference falls slightly outside the uncertainty of the two methods, we note that the uncertainty estimates are based on the sampling error only and do not include other sources of uncertainty such as the different gas analyzers, the different data treatments, and the additional uncertainty contributions of sampling and buffer volumes in STEA. Furthermore, the difference between the two methods seems to increase in the last 3 weeks of the experiment, which coincides with intermittent rainfall and more variance in wind direction. Therefore, we believe this difference does not indicate a systematic error in the flux estimates.

The coefficient of determination of the measured fluxes during the whole experiment period \( R^2 \) was 0.86. The remaining 13% of unexplained variance is the joint contribution of the uncertainties of the two flux estimates from the EC and STEA methods. The observed uncertainty from the two methods calculated as the standard deviation of the difference was 4.36 \( \mu \text{mol m}^{-2} \text{s}^{-1} \). We suggest three different mechanisms contributing to the observed uncertainty leading to the unexplained variance between the two estimates. First, the random sampling error arising from the stochasticity of turbulence (Hollinger and Richardson, 2005). The
mean random sampling error of EC fluxes calculated following Finkelstein and Sims (2001) was 1.58 µmol m$^{-2}$ s$^{-1}$. The standard deviation of the difference between the two methods can be estimated to be 2.34 µmol m$^{-2}$ s$^{-1}$ if STEA fluxes are assumed to have a similar random sampling error. Therefore, the random sampling error of the two methods accounts for more than half of the observed variance. The difference between the two methods also shows heteroscedasticity, with the error increasing along with the absolute magnitude of the flux; a similar behavior of the random sampling error was observed by Hollinger and Richardson (2005) when comparing two tower estimates. The second source of uncertainty is the use of different gas analyzers for STEA and EC. Polonik et al. (2019) compared five different analyzers for measuring CO$_2$ fluxes. They showed that the root mean square error (RMSE) was in the range of 1 to 3.35 µmol m$^{-2}$ s$^{-1}$ depending on the analyzer type and the spectral correction method applied, with larger discrepancies observed when comparing open-path to closed-path sensors. Our results have an RMSE value of 4.39 µmol m$^{-2}$ s$^{-1}$. While our result is slightly higher, it should be noted that RMSE is not an ideal metric for cross-study comparison. A relative metric, such as $R^2$, would be more comparable but was unavailable. The third source of uncertainty is the use of buffer volumes in the STEA method. Figure 10a demonstrates the increase in scatter in the measured fluxes due to the use of buffer volumes. Finally, the different processing steps between the two methods can contribute to the uncertainty, in particular the effects of time-lag compensation, spectral corrections, and statistical screening. We determined the combined effect of these processing treatments by calculating the EC flux with and without the treatments and found that the systematic errors in the flux were negligible.

### 4.4 Effect of buffer volumes

Using fixed buffer volumes attenuates the signal. To understand the effect of buffer volume use on the measured scalar concentration, we carried out a simulation on a surrogate signal generated from sonic temperature. The simulation showed that buffer volumes caused a decline that can reach up to 10% of the fluxes under operation ranges similar to those of our experiment (for $\tau = 11$ min) (Fig. 10). The empirical correction was consistently able to mitigate most of the attenuation when the filter properties are assumed to be constant (i.e., the flow rate needs to be constant for consecutive short intervals). This assumption was difficult to maintain using the 1 min switching regime. The simulation showed that the empirical correction for the buffer volumes worked best when the correction factor was obtained using a linear fit, as opposed to taking a ratio of the attenuated flux to the true flux for each averaging interval. The correction factor, in this case, is the reciprocal of the slope of the linear regression between the attenuated flux and the true flux. The correction factor calculated using Eq. (10) shows good agreement between sensible heat flux and CO$_2$. However, the uncertainty of the correction factor increased with increasing buffer volume time constant. For our experiment, the average time constant for the first-order linear filter used to model the buffer volume was estimated to be $\tau = 700$ s. This value was used to simulate the loss on the fluxes using the sensible heat flux calculated from the sonic anemometer. The correction factor was obtained from the slope of the attenuated flux and was equal to 1.18.

The empirical correction scheme achieved here by simulating the loss on sensible heat flux lacks a proper treatment of the phase shift introduced by the first-order filter. Additionally, the similarity of transport between sensible heat flux and CO$_2$ is an approximation that is expected to be valid only on average as evidenced by the large scatter around the regression line in Fig. 10. Therefore, a direct correction that can estimate the original signal before attenuation is needed. Such a correction has since been developed and is the topic of a future publication (Emad, 2022).

### 5 Conclusions

In this paper, we proposed a new variety of the eddy accumulation method as an alternative to eddy covariance for the measurement of ecosystem-level fluxes. The new method, referred to here as short-time eddy accumulation (STEA), allows the sample accumulation to be carried out on shorter intervals of varying length. The STEA method offers more flexibility than the conventional TEA method and has many potential benefits. Most importantly, STEA provides a higher dynamic range and better accuracy than the TEA method and enables operating sample accumulation under a flow-through scheme using fixed buffer volumes. The flexibility introduced by the STEA method offers new ways to design eddy accumulation systems that are particularly suited for specific atmospheric constituent gas analyzers. For example,
the accumulation time can be tailored to measure reactive species with lifetimes shorter than a conventional flux integration interval or to distribute the gas analyzer time to measure fluxes at different heights.

Furthermore, we presented a prototype evaluation of the STEA method under the flow-through regime. We described the details of the system design and operation. We compared flux measurements from our new system against a reference EC system over a flat agricultural field. The fluxes from the two methods were in very good agreement. We highlighted the importance of different processing and design aspects between the two methods and their potential effects on the fluxes.

Finally, we analyzed the effect of buffer volumes in the flow-through operational mode on the fluxes and proposed an empirical correction to correct for the underestimation resulting from the low-pass-filtering behavior of the buffer volumes.

In summary, the new STEA method provides a direct flux measurement method that complements the state-of-the-art EC method. It extends the coverage of micrometeorological methods to new trace gases and atmospheric constituents beyond the scope of the EC method.

Appendix A: Symbols and subscripts with units

<table>
<thead>
<tr>
<th>Symbols</th>
<th>Units</th>
<th>Description</th>
</tr>
</thead>
<tbody>
<tr>
<td>$c$</td>
<td>mol m$^{-3}$</td>
<td>Molar density of a scalar</td>
</tr>
<tr>
<td>$w$</td>
<td>ms$^{-1}$</td>
<td>Vertical wind velocity</td>
</tr>
<tr>
<td>$\Delta t$</td>
<td>s</td>
<td>Flux averaging interval</td>
</tr>
<tr>
<td>$A$</td>
<td>–</td>
<td>TEA sampling scaling factor</td>
</tr>
<tr>
<td>$V$</td>
<td>m$^3$</td>
<td>Volume</td>
</tr>
<tr>
<td>$C$</td>
<td>mol m$^{-3}$</td>
<td>Mean concentration of accumulated samples</td>
</tr>
<tr>
<td>$\alpha_c$</td>
<td>–</td>
<td>Transport asymmetry coefficient for scalar $c$</td>
</tr>
<tr>
<td>$\rho$</td>
<td>–</td>
<td>Correlation coefficient</td>
</tr>
<tr>
<td>$\dot{q}$</td>
<td>–</td>
<td>Dimensionless mass flow rate</td>
</tr>
<tr>
<td>$\tau$</td>
<td>s</td>
<td>Time constant of the buffer volume</td>
</tr>
<tr>
<td>$r_c$</td>
<td>ppm</td>
<td>Mixing ratio in dry air for a scalar, $c$</td>
</tr>
</tbody>
</table>

<table>
<thead>
<tr>
<th>Subscripts</th>
<th>Description</th>
</tr>
</thead>
<tbody>
<tr>
<td>acc</td>
<td>Accumulated samples</td>
</tr>
<tr>
<td>$\uparrow$</td>
<td>Updraft buffer volume</td>
</tr>
<tr>
<td>$\downarrow$</td>
<td>Downdraft buffer volume</td>
</tr>
<tr>
<td>$c$</td>
<td>Atmospheric constituent</td>
</tr>
</tbody>
</table>

Financial support. This research has been supported by the Niedersächsische Ministerium für Wissenschaft und Kultur (Wissenschaft.Niedersachsen.Weltoffen grant), the H2020 European Research Council (grant no. 682512 – OXYFLUX), and the Deutsche Forschungsgemeinschaft (grant no. INST 186/1118-1 FUGG).

This open-access publication was funded by the University of Göttingen.

Review statement. This paper was edited by Hartwig Harder and reviewed by Christoph Thomas and one anonymous referee.

References


