

Validation and field application of a low-cost device to measure CO₂ and evapotranspiration (ET) fluxes

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Abstract. Mitigating the global climate crisis and its consequences, such as more frequent and severe droughts, is one of the major challenges for future agriculture. Therefore, identifying land use systems and management practices that reduce greenhouse gas (GHG) emissions and promote water use efficiency (WUE) is crucial. This, however, requires accurate and precise measurements of carbon dioxide (CO₂) fluxes and evapotranspiration (ET). Despite that, commercial systems to measure CO₂ and ET fluxes are expensive and thus often exclude research in ecosystems within the Global South. This is especially true for research and data of agroecosystems in these areas, which are to date still widely underrepresented. Here, we present a newly developed lowcost, non-dispersive infrared (NDIR)-based CO2 and ET flux measurement device (~ EUR 200) that provides reliable, accurate and precise CO2 and ET flux measurements in conjunction with manually operated closed chambers. To validate the system, laboratory and field validation experiments were performed, testing multiple different low-cost sensors. We demonstrate that the system delivers accurate and precise CO₂ and ET flux measurements using the K30 FR NDIR (CO₂) and SHT31 (RH, relative humidity) sensor. An additional field trial application demonstrated its longer-term stability (> 3 months) and ability to obtain valid net ecosystem C balances (NECBs) and WUE. This was the case, even

though environmental conditions at the field trial application site in sub-Saharan Africa were rather challenging (e.g., extremely high temperatures, humidity and rainfall). Consequently, the developed low-cost CO_2 and ET flux measurement device not only provides reasonable results but also might help with democratizing science and closing current data gaps.

1 Introduction

The global climate crisis is one of the most critical problems of our time, and identifying and implementing measures to mitigate or adapt to its consequences, such as more frequent and severe drought, is a key challenge. Solving this challenge requires first and foremost a substantial reduction of anthropogenic greenhouse gas (GHG) emissions in all sectors (IPCC, 2019). While agriculture is a significant contributor to these anthropogenic GHG emissions (FAO, 2020), it might also offer the potential to mitigate the climate crisis by increasing soil carbon (C) sequestration (Lal, 2004). Specifically, land use systems and management practices which not only promote a net C uptake but also promote an efficient water use are needed. They might help to increase soil C stocks and crop productivity, reducing GHG emissions while simultaneously sustaining yield, despite intensifying climate stressors, such as more frequent and severe droughts. Hence, it is crucial to evaluate land use systems regarding their potential to sequester additional C and effectively utilize water. Common parameters used to assess both are the net ecosystem C balance (NECB; Smith et al., 2010) and the agronomic and ecosystem water use efficiency (WUE; Beer et al., 2009). Their determination, however, requires accurate and precise measurement of carbon dioxide (CO_2) and evapotranspiration (ET) fluxes (Chapin et al., 2016; Xu et al., 2019).

Measurement of CO₂ and ET fluxes are commonly performed using eddy covariance or chamber-based systems (Baldocchi et al., 1996; Smith et al., 2010; Wang et al., 2017; Yang et al., 2014), while the latter are especially well suited for direct treatment comparisons (Dubbert et al., 2014; Hoffmann et al., 2018; Kübert et al., 2020). In the case of a remote study site or limitations in power supply, manually operated closed-chamber measurements are particularly used to measure the CO₂ exchange and ET fluxes (Rochette and Hutchinson, 2015). However, the relatively high costs of needed measurement equipment (particularly gas analyzers) strongly limits their accessibility and often excludes research in ecosystems within the Global South. This resulted in a pronounced underrepresentation of regions, land use systems, and management practices from subtropical and tropical South America, south Asia, and Africa, even though the quantification of, for example, CO₂ fluxes in these regions might reduce disparities in the global CO₂ budget (Canadell et al., 2011; Gurney et al., 2002; Kondo et al., 2015).

Recent efforts to solve this financial constraint focus on developing low-cost, yet reliable, measurement devices. This was catalyzed by the growing availability of relatively inexpensive microcontrollers, which are increasingly utilized for scientific environmental research (Blackstock et al., 2019; Capri et al., 2021). An additional contribution came from the improvement in accuracy and precision of low-cost relative humidity (RH) and especially nondispersive infrared (NDIR) CO₂ sensors. Evaluation of commercially available NDIR CO2 sensors (Keimel, 2019; Martin et al., 2017; Pandey et al., 2007; Yasuda et al., 2012) showed that they have acceptable precision and accuracy in measuring CO₂ concentrations, especially when proper calibration methods are applied. Although low-cost NDIR CO₂ sensors are commonly used in air quality monitoring studies (Araújo et al., 2020; Wastine et al., 2022), these sensors have also been applied in environmental research (Bastviken et al., 2015; Brown et al., 2020). For example, multiple studies have demonstrated the applicability of using low-cost NDIR CO₂ sensors for reliable measurements of soil CO₂ efflux (Brändle and Kunert, 2019; Curcoll et al., 2022; Harmon et al., 2015) and water crop use determination (Capri et al., 2021). However, in the case of RH sensors, the inversely increased measurement uncertainty of total water vapor concentration with decreasing RH (e.g., a typical low-cost RH sensor has a measurement accuracy of 1 %-3 % in relative but not absolute humidity) might constitute a problem. Despite first studies showing the potential of using low-cost sensors as an alternative to more-expensive commercial counterparts, there is still little evidence that in situ closed-chamber CO₂ and ET flux measurements using both are comparable in precision and accuracy.

Here, we present the hardware and software implementation, as well as laboratory and in situ validation, of a new low-cost CO2 and ET flux measurement device with open-source technology. We hypothesize that by using the device in conjunction with a closed chamber (1) CO_2 and ET fluxes can be reliably and accurately measured, and that (2) measured CO_2 and ET fluxes can be used to obtain valid estimates of net ecosystem C balance (NECB) and WUE, even under challenging environmental conditions such as extremely high air temperatures, humidity and precipitation. To test these hypotheses, we first validated the accuracy and precision of four different low-cost NDIR CO2 sensors (K30 FR, SCD30, MH-Z14 and MH-Z19) under controlled laboratory conditions. Afterwards, the NDIR sensors passing laboratory validation as well as two different RH sensors were validated in the field. During field validation, ET and CO₂ fluxes (ecosystem respiration (R_{eco}) and net ecosystem exchange (NEE)), as well as temperature-dependent R_{eco} and photosynthetically active radiation (PAR)-dependent gross primary production (GPP) parameters, were compared to the results obtained simultaneously with a reference infrared gas analyzer (IRGA; LI-850, LI-COR, USA). Finally, the ability of the developed low-cost CO₂ and ET flux measurement device to obtain reliable NECB and WUE as well as its practicability and stability was tested. Therefore, multiple devices were used during a field trial application in northern Ghana to obtain seasonal CO₂ exchange and ET, as well as NECB and WUE for four different fertilizer treatments in a maize cultivation.

2 Material and methods

2.1 Hardware and software implementation

The developed highly portable CO₂ and ET flux measurement device consists of a logger and sensor unit, both assembled from combination of various low-cost, off-the-shelf components. A complete list of used components, distributors and prices is given in Table 1. Figure 1 shows the assembled logger and attachable sensor unit, together with a schematic representation of the wiring. The logger unit consists of an Arduino Uno-like microcontroller (Atmega328, AZ-Delivery Vertriebs GmbH, Germany) with attached logger shield module (AZ-Delivery Vertriebs GmbH, Germany), including an SD card reader and SD card (2 GB) to store sensor readings and a real time clock (RTC), which helps to keep the time and date even when the system is switched off. A BME280 air temperature (± 1 °C), air humidity (± 3 %) and air pressure sensor $(\pm 1 \text{ hPa}; \text{Reichelt electronics GmbH})$ Germany) as well as an LCD (AZ-Delivery Vertriebs GmbH, Germany) and HC-05 Bluetooth module are part of the logger unit and are connected to the microcontroller. The logger unit is fitted into a weather- and shock-resistant outdoor housing (B&W outdoor case type 500, Overhaul Media GmbH, Germany). It easily connects to end-user devices using the Bluetooth module, so data can be visualized inter alia with a smartphone in real time without the need to open the weather- and shock-resistant outdoor housing. The external sensor unit consists of an NDIR-based CO2 (0-10000 ppm, $\pm 30 \text{ ppm} \pm 3\%$ accuracy; K30 FR, Senseair AB, Sweden) and an air humidity (RH) and air temperature sensor (SHT31, $\pm 2\%$ accuracy, Sensirion AG, Switzerland or DHT22, $\pm 2\%$ to 5 % accuracy, Aosong Electronics Co., Ltd., China). Both sensors were connected through a seven-core cable to the logger unit using UART (K30 FR) and I2C (SHT31) data communication, respectively. The power supply of the microcontroller is energized by six rechargeable AA nickelmetal hydride (Ni-MH) batteries (1.2 V; 2600 mAh) in a $6 \times AA$ battery holder, which supplies 7.2 V. Due to the power requirements of the external sensor unit (K30 FR and SHT31), an additional $6 \times AA$ battery holder is attached to the housing directly. The software implementation was done using Arduino IDE 2.0.3.

2.2 Laboratory validation

To identify the NDIR sensor most suitable for in situ, dynamic closed-chamber measurements, four different NDIRbased sensors were tested and validated regarding their precision and accuracy during a laboratory validation experiment. The sensors tested were (1) MH-Z19 (Winsen Electronics Technology Co., Ltd., China), (2) MH-Z14 (Winsen Electronics Technology Co., Ltd., China), (3) SCD30 (Sensirion AG, Switzerland) and (4) K30 FR (Senseair AB, Sweden). Sensors were placed separately into a sealed, ventilated, cylindrical vessel (Fig. 2; volume: 1425.5 cm³) and connected to the developed low-cost logger system.

All sensors were calibrated in ambient air prior to use according to manufacturer instructions. Afterwards, different distinct amounts (5 to 30 mL, in 5 mL steps; each step was repeated five times) of a technical gas containing 10 000 ppm CO₂ (Linde, Germany) were injected into the sealed vessel using a syringe. In between injections, the vessel was flushed with ambient air by two pumps (1.5 L min^{-1}) connected to the vessel (semi-automatic measurement mode of the developed device). Finally, CO₂ concentration increases inside the vessel, measured with a 5 s interval by the NDIRbased sensors, from before to after injection (Δ CO₂ in ppm) were compared against mixing-induced CO₂ concentration increases. Sensors that performed best in terms of accuracy and precision were subsequently validated during the field validation experiment.

2.3 Field validation

Field validation of the low-cost CO₂ and ET flux measurement device was performed through parallel manually operated taken closed-chamber measurements using an infrared gas analyzer (IRGA; LI-850, LI-COR, USA) and NDIR sensors (CO₂) passing previous laboratory validation, as well as two different RH sensors (ET). Measurements were conducted at the "PatchCrop" experimental field, managed by the Leibniz Centre for Agricultural Landscape Research (Fig. 3; ZALF). PatchCrop features multiple smaller patches (72 × 72 m), with diverse and site-specific crop rotations, aiming to create synergies and interactions between fields.

The experimental field PatchCrop is located near the village of Tempelberg, northeast Germany (52°26'49.2" N, $14^{\circ}09'00.0''$ E). The temperate climate is characterized by a mean annual air temperature of 9.7 °C and mean annual precipitation of 544 mm (ZALF weather station, 2010-2019). The medium loamy, sand-textured soil can be classified as Luvisol (IUSS Working Group WRB, 2015). CO2 exchange (NEE and R_{eco}) and ET measurements were conducted for a mixture of Phacelia and Guizotia abyssinica at three repetitive plots, established at one of the patches by installing PVC collars (area: 0.5625 m^2 , 5 cm deep) in the beginning of October 2022. Measurements started shortly after sunrise and lasted to late afternoon during 2 consecutive days, using a dynamic, (non-)flow-through non-steady-state ((N)FT-NSS) closed-chamber system. The used transparent (86% light transmission; NEE flux measurements) and opaque (R_{eco} flux measurements) cube-shaped PVC chambers had a total volume of $0.296 \,\mathrm{m}^3$ and were equipped with a fan for efficient headspace mixing. CO2 and H2O concentrations, as well as RH, during chamber deployment were recorded in parallel using a LI-850 IRGA and the developed lowcost measurement device, equipped with a K30 FR, SCD30, SHT31 and DHT22 sensors, respectively. NEE, R_{eco} and ET fluxes were measured by alternately deploying the opaque and transparent chambers on the three pre-installed PVC frames. During individual 4 min measurements, CO2 and H₂O concentration changes in the chamber headspace, as well as RH, air temperature inside and outside the chamber, soil temperature and humidity (TMS-4, TOMST, Czech Republic) as well as PAR (outside the chamber; Skye, UK) were recorded at a 3s (LI-850) and 5s interval (NDIR and RH sensors). To validate the low-cost CO₂ and ET flux measurement device, measured R_{eco} , NEE and ET fluxes, as well as the derived temperature and PAR dependency functions for R_{eco} and GPP, respectively, were directly compared with results obtained in parallel with the LI-850.

Table 1. Sensor components and cost (in Euro) at the time of writing, including weather and shock-proof housing and energy supply (rechargeable batteries). Components needed for optional semi-automatic mode are listed in addition. The last access date for section "Cost of other NDIR sensors tested" is 8 December 2023; the last access date for other URLs cited in the table is 21 March 2023.

Component	Amount	Description	Price	Distributor
B&W outdoor case type 500	1	Outdoor case for housing electrical components	EUR 28.75	https://www.profikoffer.de/
PVC hard-foam plate	1	PVC 5 mm hard-foam plate to create interior of housing for electronic components	EUR 1.5	https://www.amazon.de/
Luster terminals	12	Luster terminals for wiring electrical components within housing	EUR 0.6	https://www.amazon.de/
0.2 mm^2 , 24 AWG electrical wire		Electrical wires for wiring electrical components within housing		https://www.amazon.de/
Seven-pin aviation connector	2	Aviation connector to connect logger unit within weatherproof housing with passive NDIR sensor installed in the closed chamber to be attached	EUR 2.9	https://www.amazon.de/
Seven-core rubber cable (1.5 m)	1	Cable to connect logger unit within weatherproof housing with passive NDIR sensor installed in the closed chamber to be attached	EUR 3.75	https://www.conrad.de/
WS R13-112 AAAA rocker switch	1	Rocker switch for switching on and off	EUR 1	https://www.reichelt.de/
ATmega 328	1	Arduino Uno-like microcontroller	EUR 5	https://www.az-delivery.de/
Datalogger module	1	Logger shield for Arduino Uno-like microcontroller with SD card reader and RTC unit	EUR 4.6	https://www.az-delivery.de/
Hama class-4 SD memory card, 2 GB, 10 MB s^{-1}	1	SD memory card to save sensor readings	EUR 6	https://www.saturn.de/
HC-05 Bluetooth wireless RF transceiver module RS232	1	Bluetooth module for wireless communication	EUR 5.2	https://www.az-delivery.de/
16×2 LCD or OLED display with I2C adapter	1	LCD or OLED display for data visualization	EUR 3.7	https://www.az-delivery.de/
BMP280	1	Air pressure, air humidity and air temperature sensor	EUR 1.7	https://www.reichelt.de/
DHT22 or SHT31 module	1	Air temperature and air humidity sensor	EUR 6.4	https://www.az-delivery.de/
Senseair K30 FR (fast response)	1	CO_2 measuring module with fast response time; measuring range: 0 to 5000 ppm CO_2 , operating range: 0 to 50 °C	EUR 85	https://www.driesen-kern.de/
Goobay 11467 $6 \times (4 \times)$ Mignon (AA) battery holder	2 (1)	Battery holder for $6 \times \text{Ni-MH}$ rechargeable Mignon (AA) batteries	EUR 4.6	https://www.conrad.de/
Conrad energy HR06 Mignon (AA)-Akku Ni–MH 2600 mAh 1.2 V	12 (16)	Ni-MH rechargeable Mignon (AA) batteries	EUR 38	https://www.conrad.de/
4.5 V metal brush air pump	2	Air pump for flushing headspace of small chambers	EUR 9.45	https://www.berrybase.de/
IRLZ44N MOSFET	1	MOSFET to control power supply to pumps	EUR 0.75	https://www.reichelt.de/
Cost of other NDIR sensors tested				
Sensirion SCD30 module	1	NDIR gas sensor for CO_2 (0–10 000 ppm) integrated with humidity and temperature sensor in the same module	EUR 63.50	https://www.berrybase.de/
MH-Z14 CO ₂ sensor module	1	NDIR gas sensor for accurately measuring the CO ₂ concentration (0–10 000 ppm)	EUR 55.60	https://www.kaufland.de/
MH-Z19 CO ₂ sensor module	1	NDIR gas sensor for accurately measuring the CO ₂ concentration (0–10 000 ppm)	EUR 28.50	https://www.reichelt.de/
Total cost			EUR 199.7	



Figure 1. (a) Logger unit in weather- and shock-resistant housing, (b) external sensor unit attached to a transparent non-flow-through non-steady-state (NFT-NSS) closed chamber and (c) schematic representation of wiring.

2.4 Field trial application

The developed low-cost measurement device has been tested for applicability and reliability under challenging environmental conditions in an experimental field managed by the Council for Scientific and Industrial Research at the Savanna Agricultural Research Institute (Fig. 4; CSIR-SARI). The experimental field $(21 \times 54 \text{ m})$, located near the city of Nyankpala, northern Ghana $(9^{\circ}24'15.9'' \text{ N}, 01^{\circ}00'12.1'' \text{ W})$, featured a split-plot design $(3 \times 6 \text{ m}; n = 3)$, with the main plot assigned to tillage practice (conventional vs. reduced tillage) and the subplot assigned to a factorial combination of organic and mineral fertilizers. The tropical region around Nyankpala is characterized by a mean annual air temperature of $26 \,^{\circ}\text{C}$ and a unimodal rainfall pattern with a distinct rainy season from June to October followed by a dry season from November to May (Alua et al., 2018), resulting in a mean annual precipitation of 1100 mm (CSIR-SARI weather station, 1995–2013). The soil is sandy loam textured and classified as Acrisol (IUSS Working Group WRB, 2015). CO₂ exchange (NEE and R_{eco}) and ET measurements were conducted for maize (*Zea mays*) from July to October 2022 at four out of the nine treatments with reduced tillage (bullock plow), namely, (1) Fertisoil (5 tha⁻¹; commercial organic fertilizer in northern Ghana; denoted FT), (2) farmyard manure (5 tha⁻¹; denoted FM), (3) Fertisoil + NPK (5 tha⁻¹ + 90–60–60 kg ha⁻¹; denoted FT + MIN) and (4) farmyard manure + NPK (5 tha⁻¹ + 90– 60–60 kg ha⁻¹; denoted FM + MIN). Measurement campaigns took place every 2 weeks from sunrise to late evening



Figure 2. Experimental setup of the performed laboratory validation experiment for four different NDIR CO₂ sensors connected to the developed low-cost CO₂ and ET flux measurement device (MH-Z14, MH-Z19, SCD30 and K30 FR). Validation was performed by injecting distinct amounts of technical gas (Linde, Germany; 10 000 ppm CO₂) into the air-tight, sealed, cylindrical vessel.



Figure 3. Parallel opaque (R_{eco}) manually operated closedchamber measurements with a LI-COR 850 IRGA (LI-850, LI-COR, USA) and the developed low-cost CO₂ and ET flux measurement device at ZALF experimental field near the village of Tempelberg, northeast Germany ($52^{\circ}26'49.2''$ N, $14^{\circ}09'00.0''$ E). The developed system was equipped with a K30 FR and SCD30 NDIR, as well as SHT31 and DHT22 sensors.

using a dynamic, NFT-NSS closed-chamber system. The used transparent (86% light transmission; NEE flux measurements) and opaque (R_{eco} flux measurements) cubeshaped PVC chambers had a total volume of 1.56 m^3 and were equipped with a fan for efficient headspace mixing. CO₂ concentration and RH changes during chamber deployment were recorded using the developed low-cost measurement device, equipped with a K30 FR and DHT22 sensor. During each measurement campaign, NEE, R_{eco} and ET fluxes were measured by alternately deploying the opaque and transparent chambers on pre-installed frames (area: 0.96 m^2) at each of the measured plots.



Figure 4. Transparent (NEE) manually operated closed-chamber measurement at the CSIR-SARI experimental field, used for field trial application of the developed low-cost CO_2 and ET flux measurement device, near the city of Nyankpala, northern Ghana (9°24'15.9" N, 01°00'12.1" W).

2.5 Data processing

2.5.1 CO₂ and ET flux calculation, separation and gap-filling

For laboratory validation, the changes in CO₂ concentrations in the vessel, expressed as ΔCO_2 in ppm, were calculated as the mixing ratio of measured ambient air and injected technical gas CO₂ concentration (10000 ppm). These were compared with the ΔCO_2 obtained for the four different NDIR sensors as the difference in mean CO₂ concentrations measured for 1 min right before and 2 min after injection. For the field validation, measured CO2 and ET fluxes were calculated using a modular R script, described in detail by Hoffmann et al. (2015; CO₂) and Dahlmann et al. (2023; ET), respectively. Prior to CO₂ and ET flux calculation, underlying data were trimmed by removing the first and last 10 % of each chamber measurement dataset. This was conducted to eliminate data noise caused by turbulence and pressure fluctuations due to chamber deployment (Hoffmann et al., 2015) and to mitigate biases arising from the time needed to homogenize chamber headspace air (Vaidya et al., 2021). CO₂ concentrations measured using the LI-850 were additionally corrected for changes in water vapor during chamber measurements (Webb et al., 1980; McDermitt et al., 1993). Unlike the LI-850 which provided H₂O as mole fraction, used low-cost RH sensors (DHT22 and SHT31) required addi-

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tional postprocessing. RH measurements were converted into a mass concentration following Hamel et al. (2015; Eq. 1):

$$H_2 O = \frac{RH \cdot e^s}{100 \cdot P}, \qquad (1)$$

where RH is the relative humidity, P is the gas pressure (Pa) and e^s is the saturated vapor pressure (Pa) that was calculated according to Allen et al. (1998). Thereafter, CO₂ and ET fluxes were calculated based on the ideal gas law using a linear regression approach (Eq. 2):

$$f = \frac{MpV}{RTA} \cdot \frac{\Delta c}{\Delta t},\tag{2}$$

where M denotes the molar mass of the gas $(g \mod^{-1})$, p denotes the ambient air pressure (Pa) and V denotes the chamber volume (m³). Since plants accounted for < 0.1 % of the total chamber volume, a static chamber volume was assumed. *R* denotes the gas constant $(8.314 \text{ m}^3 \text{ Pa} \text{ K}^{-1} \text{ mol}^{-1})$, T denotes temperature inside the chamber (K), A denotes the basal area (m⁻²), and $\Delta c/\Delta t$ denotes the linear CO₂ (e.g., Leiber-Sauheitl et al., 2014) and H₂O concentration changes over time (e.g., Dahlmann et al., 2023). The variables T and, more importantly, $\Delta c / \Delta t$, were obtained by applying a variable (window size 0.5 to 4 min) moving-window filter to each chamber measurement. Thus, resulting multiple ET and CO₂ fluxes per measurement (based on generated variable moving window data subsets) were further evaluated according to the following criteria: (1) fulfilled prerequisites for applying a linear regression (normality (Lilliefors adaption of the Kolmogorov-Smirnov test), homoscedasticity (Breusch–Pagan test) and linearity), (2) regression slope $(p \le 0.1, t \text{ test}), (3)$ range of within-chamber air temperature not larger than ± 1.5 K and a PAR deviation (only transparent chamber measurements) not larger than $\pm 20\%$ of the average to ensure stable environmental conditions within the chamber throughout the respective measurement window, and (4) no outliers present ($\pm 6 \times IQR$, interquartile range). Calculated CO₂ and ET fluxes that did not meet all criteria were discarded. In cases where more than one flux per measurement met all criteria, the CO₂ and ET fluxes with the steepest slope and closest in time to chamber closure were chosen. For field validation and field trial application, CO₂ fluxes were additionally separated into its flux components $R_{\rm eco}$, GPP and NEE and gap-filled through deriving empirical models. In the case of R_{eco} , a temperature-dependent Arrhenius-type function was used and fitted for air as well as soil temperatures measured at different depths (Lloyd and Taylor, 1994; Eq. 3).

$$R_{\rm eco} = R_{\rm ref} \cdot e^{E_0^{\left[\frac{1}{T_{\rm ref} - T_0} - \frac{1}{T - T_0}\right]}},$$
(3)

where R_{ref} is the respiration rate at the reference temperature (T_{ref} : 283.15 K), E_0 is an activation energy-like parameter, T_0 is the starting temperature constant (227.13 K), and *T* is the mean air or soil temperature during the flux measurement. Out of the four obtained R_{eco} models (one model for air temperature inside the chamber, one for air temperature outside the chamber, and soil temperatures at 2 and 5 cm depths), the model with the lowest Akaike information criterion (AIC) was finally used. In the case of GPP, a PAR-dependent, rectangular hyperbolic light-response function, based on the Michaelis–Menten kinetic, was used (Elsgaard et al., 2012; Hoffmann et al., 2015; Wang et al., 2013; Eq. 4). Since GPP cannot be measured directly, GPP fluxes were calculated as the difference between measured NEE and modeled R_{eco} fluxes, using campaign-specific previously derived parameters R_{ref} and T_0 .

$$GPP = \frac{GP_{max} \cdot \alpha \cdot PAR}{\alpha \cdot PAR + GP_{max}},$$
(4)

where GP_{max} is the maximum rate of C fixation at infinite PAR (μ mol CO₂ m⁻² s⁻¹), α is the light use efficiency (μ mol CO₂ μ mol⁻¹ photons) and PAR is the photon flux density (corrected for chamber light transmission) of the photosynthetically active radiation (μ mol⁻¹ photons m⁻² s⁻¹). In cases where the rectangular hyperbolic light-response function did not result in significant parameter estimates, a non-rectangular hyperbolic light-response function was used (Gilmanov et al., 2007, 2013). Reco and GPP parameter sets were evaluated and discarded in the case of non-significant parameter estimates. If no fit or a non-significant fit was achieved, averaged flux rates were used for R_{eco} and GPP instead. Reco, GPP and NEE were modeled in half-hourly steps for the entire period based on continuously monitored temperature and PAR. For ET, campaign-wise average daily ET fluxes (for nighttime ET fluxes measured before, for daytime ET fluxes measured after 08:00) were determined and linearly interpolated between campaigns for the entire crop growth period.

2.5.2 NECB and WUE

NECB for the field trial application experiment was calculated as the sum of cumulated NEE, C output such as harvested biomass C and C input due to organic fertilizer application (Eq. 5; Smith et al., 2010).

$$NECB = NEE + C_{input} - C_{output}$$
(5)

Several minor NECB components have not been considered, such as, C input from seeding and methane emissions. However, due to their relatively low magnitude (e.g., no methane emissions in mineral soil under aerobic conditions), their influence on the NECB of our study is neglectable. Values for R_{eco} , GPP, NEE, harvested biomass C and NECB are given using the atmospheric sign convention (Ceschia et al., 2010), where positive values indicate C losses from the plant–soil system and negative values indicate C uptake. Thus, NECB refers to the total change in belowground C. WUE was calculated as the agricultural WUE (WUE_{agro}; Eq. 6; Hatfield and Dold, 2019).

$$WUE = \frac{DM}{ET},$$
(6)

where DM denotes harvested dry biomass (in g m^{-2}), and ET is cumulative evapotranspiration (in mm).

2.5.3 Error calculation and statistical analysis

To test for normal distribution of the data obtained from laboratory and field validation measurements, a Kolmogorov-Smirnov test (p < 0.05) was performed. In the case of a normal distribution, significant differences between ΔCO_2 (in ppm) or R_{eco} , NEE, or ET fluxes measured from lowcost sensors and mixing ratio ΔCO_2 or IRGA-based R_{eco} , NEE, and ET fluxes were determined using a one-sample t test (p < 0.05). The error calculation for CO₂ fluxes, as well as crop season CO₂ exchange, was quantified using a comprehensive error prediction algorithm described in detail by Hoffmann et al. (2015). The approach utilizes bootstrapping alongside k-fold subsampling to estimate uncertainties for each flux measurement as well as subsequent R_{eco} and GPP parameterization and final gap-filling. An adaptation of this approach was used to calculate errors in ET fluxes (Dahlmann et al., 2023). Seasonal ET flux errors were then estimated based on $1.96 \times SD$ of daily average ET fluxes.

3 Results and discussion

3.1 CO₂ sensor laboratory validation

Differences in accuracy and precision among the four different low-cost tested NDIR sensors are shown in Fig. 5a-d as 1:1 agreement plots between mixing ratio (calculated) and measured ΔCO_2 . While accuracy can be assessed as deviation from the 1:1 agreement line, precision is determined by the residual standard deviation (SD) and the coefficient of determination (r^2) of the linear regression fitted on calculated versus measured ΔCO_2 . The K30 FR (Fig. 5d) showed the highest accuracy among all tested NDIR sensors, reflecting well the increase in CO_2 concentration (ΔCO_2) derived through mixing ratio. Correspondingly, no significant difference (one sample t test, p = 0.80) was found between calculated and measured ΔCO_2 . The SCD30 (Fig. 5c), even though fairly accurate at lower values, failed to reflect higher calculated ΔCO_2 values and generally tends to overestimate triggered ΔCO_2 . Neither the MH-Z14 (Fig. 5b) nor the MH-Z19 (Fig. 5a) was sufficiently accurate or able to reflect triggered ΔCO_2 . While the MH-Z14 showed a rather constant offset from the 1:1 agreement by 28 ppm, the MH-Z19 tended to increasingly overestimate higher ΔCO_2 values derived through mixing ratio. Hence, unlike the K30 FR, all other NDIR sensors measured significantly higher ΔCO_2 when compared to mixing ratio ΔCO_2 (one sample



Figure 5. The 1:1 agreement between mixing ratio and measured ΔCO_2 (in ppm) from the four low-cost sensors (K30 FR, SCD30, MH-Z14 and MH-Z19). The dashed black line indicates the 1:1 agreement. The dotted green line shows the linear regression through the average ΔCO_2 for each injection step (n = 5), calculated from the repetitive measurements per step. Error bars indicate ± 1.96 SD. The grey shaded area represents the respective confidence band of the regression line.

t test, p < 0.01). Unlike the accuracy, overall precision and measurement repeatability among all four NDIR sensors was generally high and fairly comparable, showing a residual SD of 2.78, 4.23, 2.52 and 3.58 ppm, respectively. Regarding the response time (defined as mean time from injection to measured initial CO₂ concentration increase), all four NDIR sensors differed substantially, with only 44 s for the K30 FR and more than 280 s for the MH-Z14. The same was true for the response strength (defined as the mean time from beginning to end of the injection triggered CO₂ concentration increase, which represents its steepness), with 61, 160 and 265 s for the K30 FR, SCD30 and MH-Z19, respectively. In the case of the MH-Z14, response strength could not be evaluated, since no clear saturation after injection-induced CO₂ concentration increase could be observed.

While accuracy and precision are of course highly relevant, response time and response strength in particular play a key role in determining the extent to which the tested NDIR sensors can be used for in situ NFT-NSS closed-chamber measurements. With a response time of almost 2 and 5 min, respectively, as well as low response strength, MH-Z19 and MH-Z14 would likely fail to correctly reflect ΔCO_2 during short-time (< 4 min) closed-chamber measurements, regardless of their low accuracy, which makes them additionally

unsuitable. Therefore, only the K30 FR (and to a much lower extent the SCD30) with its fast response time and high response strength passed laboratory validation and met all necessary requirements for accurate and precise in situ measurements of CO₂ exchange. Our findings, comparing accuracy and precision of four different NDIR sensors during the laboratory setup, are in good agreement with previous studies performing laboratory validation of single sensors. Brändle and Kunert (2019), who compared the MH-Z14A NDIR sensor against a GFS-3000 (Heinz Walz GmbH, Germany) during a laboratory validation, observed a similar response time and a general measurement offset of approx. +40 ppm. Based on this and an additionally conducted field validation, Brändle and Kunert (2019) also suggested that the MH-Z14A is not suitable for short-term measurements (< 5 min). Also, findings by González Rivero et al. (2023), who tested the ability of the SCD30 to reflect calibration gas concentrations and concluded an acceptable accuracy and response time, are in good agreement with results of the present study. The most widely tested NDIR sensors so far, however, are those of the K series (e.g., Ali et al., 2016; Blackstock et al., 2019; Brown et al., 2020; Mendes et al., 2015). Laboratory validation performed by Blackstock et al. (2019) using a K30 1% sensor to measure a span of different CO₂ concentrations verified that it well reflects CO₂ concentrations within the accuracy stated by the manufacturer. Similarly, laboratory tests performed by Mendes et al. (2015) found that the K30 sensor has nearly perfect linear response against calibration gas CO₂ concentrations. Lastly, the laboratory experiment by Ali et al. (2016) also highlighted the accuracy of the K30 1% sensor when compared against measurements of an SBA-5 CO2 gas analyzer (PP Systems, USA). During their experiment, both sensors showed a strong correlation and no offset when K30 1 % sensor self-calibration was used, highlighting the self-calibration capabilities of the K-series sensors that contribute to their stable performance and high measurement repeatability with minimal maintenance compared to other NDIR sensors.

3.2 Field validation

3.2.1 In situ ET flux validation

Two low-cost RH sensors (ET; SHT31 and DHT22) were tested in parallel with NDIR sensors passing the laboratory validation (CO₂; K30 FR and SCD30) against LI-850 as reference. To avoid the systematic impact of opaque chambers on plant transpiration via stomatal closure upon darkening, in the case of ET fluxes, only transparent chamber measurements were taken into account (Larcher, 2003). Out of the 20 NEE measurements, 13 valid ET fluxes could be calculated in the case of the LI-850. Compared to that, 18 and 17 valid ET fluxes were obtained for the SHT31 and DHT22, respectively. Differences in accuracy and precision for ET fluxes calculated based on RH measurements (Fig. 6c–d)

compared to ET fluxes calculated based on LI-850 are shown as 1:1 agreement plots in Fig. 6. No significant difference (mean diff. -0.01 mm d^{-1} ; one sample *t* test, p = 0.89) was found between ET fluxes calculated from H₂O concentration and RH measurements, using the LI-850 and SHT31, respectively (Fig. 6c). Together with an r^2 of 0.72, this indicates a reasonable accuracy of SHT31-derived ET flux estimates. Compared to that, ET fluxes, determined through RH measurements using the DHT22 (Fig. 6d), were significantly smaller (mean diff. 0.28 mm d⁻¹; one sample t test, p <0.05) than LI-850-based ET fluxes and with an r^2 of 0.64, i.e., less accurate. This is consistent with sensor accuracy for measuring relative humidity specified by their corresponding manufacturers, which are $\pm 2\%$ accuracy for SHT31 and $\pm 2\%$ -5% accuracy for DHT22. Since these low-cost sensors were only capable of measuring at this level of accuracy, a higher uncertainty at lower RH concentrations and consequently derived ET fluxes might occur, even though not directly detected within this study. The overall precision of SHT31- and DHT22-derived ET fluxes were fairly similar but with a residual SD of 0.36 and 0.39 mm d^{-1} , i.e., rather high.

3.2.2 In situ CO₂ flux validation

A total of 41 closed-chamber measurements (R_{eco} : 21; NEE: 20) have been conducted during the 2 consecutive days of field validation, using the LI-850 as reference for both NDIR sensors passing the laboratory validation (CO2; K30 FR and SCD30). While 41 valid CO₂ fluxes (R_{eco} : 21; NEE: 20) could be calculated for the LI-850 sensor, 35 (R_{eco} : 21; NEE: 14) and 36 (Reco: 21; NEE: 15) valid fluxes were obtained for the K30 FR and SCD30 sensors, respectively. Differences in accuracy and precision for CO₂ fluxes calculated based on NDIR (Fig. 6a-b) compared to CO₂ and ET fluxes calculated based on LI-850 are shown as 1:1 agreement plots in Fig. 6. While the comparison between R_{eco} and NEE fluxes calculated from LI-850 and K30 FR measurements (Fig. 6a) was in accordance with the laboratory validation and showed again the overall accuracy and precision of this NDIR sensor, a small positive offset was found. Hence, CO₂ fluxes for the K30 FR were significantly higher $(R_{\rm eco} \text{ mean diff. } 1.12 \,\mu\text{mol}\,\text{m}^{-2}\,\text{s}^{-1}; \text{ one sample } t \text{ test, } p < t$ 0.05) and less negative (NEE mean diff. 1.41 μ mol m⁻² s⁻¹; one sample t test, p < 0.05) when compared to LI-850. No such systematic offset was found in the case of the SCD30 (Fig. 6b), which showed significantly lower R_{eco} (mean diff. $-1.33 \,\mu\text{mol}\,\text{m}^{-2}\,\text{s}^{-1}$; one sample t test, p < 0.05) and much less negative NEE fluxes (mean diff. $-4.18 \,\mu\text{mol}\,\text{m}^{-2}\,\text{s}^{-1}$; one sample t test, p < 0.05) compared to LI-850. Since both NDIR sensors did not show a similar offset and an overestimation was not found for the K30 FR during the laboratory validation already, it can be assumed that the detected offset in the case of the K30 FR is neither a direct result of microclimatic effects (e.g., increasing humidity) nor in-



Figure 6. The 1 : 1 agreement between (**a–b**) CO₂ (R_{eco} : dark red points; NEE: dark green points) and (**c–d**) ET fluxes measured with infrared gas analyzer (IRGA; LI-850, LI-COR, USA) and low-cost NDIR sensors (K30 FR and SCD30), as well as low-cost RH sensors (SHT 31 and DHT22), respectively. The dashed black line indicates the 1 : 1 agreement. The dotted green/blue line shows the linear regression through the measured CO₂/ET fluxes. The grey/blue shaded area represents the respective confidence band of the regression line. Error bars indicate calculated flux error (CI: 95 %; p < 0.05).

correct sensor readings. Instead, inter alia differences within the chamber headspace and the position of the NDIR sensor right below the chamber top, approx. 10 cm above the LI-850 inlet and outlet, might help to explain it. Nonetheless, the K30 FR NDIR sensor still exhibited higher accuracy than the SCD30 sensor when validated against LI-850 flux measurements. The root mean squared error (RMSE), mean squared error (MSE) and mean absolute error (MAE) obtained from the K30 FR (RMSE: $1.77 \,\mu mol \, m^{-2} \, s^{-1}$, MSE: $3.16 \,\mu\text{mol}\,\text{m}^{-2}\,\text{s}^{-1}$, MAE: $1.34 \,\mu\text{mol}\,\text{m}^{-2}\,\text{s}^{-1}$) were lower in comparison to the SCD30 (RMSE: $3.97 \,\mu$ mol m⁻² s⁻¹, MSE: 15.77 μ mol m⁻² s⁻¹, MAE: 2.80 μ mol m⁻² s⁻¹). Compared to the K30 FR, the NEE fluxes obtained by the SCD30 were also especially characterized by a very low precision. The reason for this is certainly the lower CO₂ concentrations (< 400 ppm) in the NEE measurements, which are clearly outside the measurement range specified by the manufacturer (400 to 10 000 ppm). This also explains the decreasing precision with increased negative NEE fluxes obtained by SCD30, since these are likely related to CO₂ concentration measurements well below 400 ppm. The general underestimation of R_{eco} and NEE fluxes derived from SCD30, however, is probably a result of its rather long response time and lower response strength when compared to the K30 FR (see Sect. 3.1).

3.2.3 Temperature and PAR dependency of measured CO₂ fluxes

Figure 7 shows temperature-dependent R_{eco} (Fig. 7a–b) and PAR-dependent GPP (Fig. 7c–d) parameter estimates for flux measurements performed with the LI-850 compared to K30

FR (Fig. 7a, c) and SCD30 (Fig. 7b, d), respectively. Since the R_{eco} and GPP parameters are based on the fluxes presented in Fig. 6, similar differences between LI-850, K30 FR and SCD30 could be obtained. With an R_{ref} and E_0 of 4.60 and 212.71, respectively, the K30 FR had similar but slightly higher R_{eco} parameters (Fig. 7a) when compared to the LI-850 (R_{ref} : 4.14; E_0 : 195.01). This indicates not only in general higher R_{eco} fluxes but, more importantly, also a stronger increase of R_{eco} fluxes with rising temperature. In the case of the SCD30 (R_{ref} : 2.54; E_0 : 270.07), differences in R_{eco} parameters were, however, much more pronounced. The same tends to be true for obtained GPP parameters, which were highly comparable for LI-850 (α : -0.048; GP_{max}: -39.83) and K30 FR (α : -0.042; GP_{max}: -38.42) but distinctly different for SCD30 (α : -0.029; GP_{max}: -31.83). As a result, the fitted K30 FR PAR dependency function was fully within the confidence band of the LI-850 PAR dependency function. In summary, the K30 FR well represented R_{eco} and GPP fluxes measured with the LI-850 and thereon based parameter estimates for R_{eco} and GPP. Unlike the K30 FR, the SCD30 was only able to reflect LI-850 R_{eco} and GPP fluxes measured within the manufacturer-specified concentration range. Correspondingly, accurate parameter estimates, especially with GPP, were not obtained. Our findings are further supported by studies that compared the accuracy of K-series sensors against commercial sensor counterparts and its accuracy for field CO₂ flux measurements (Curcoll et al., 2022). They integrated a K30 STA sensor into NFT-NSS chamber measurements and were able to accurately measure CO2 fluxes for a grassland ecosystem. Adding to that, the average CO₂ flux obtained during our study using a K30 FR $(0.4 \,\mu\text{mol}\,\text{m}^{-2}\,\text{s}^{-1})$ falls within the range of reported daily average NEE values (4 to $-6 \,\mu \text{mol}\,\text{m}^{-2}\,\text{s}^{-1}$) in the study by Emmel et al. (2018) for a field site in Switzerland which was also covered with Phacelia cover crop. Based on the performed field validation, the developed low-cost measurement device equipped with the K30 FR and SHT31 is likely to accurately measure CO₂ and ET fluxes in situ, using NFT-NSS closed chambers.

3.3 Field trial application

During the measurement period, half-hourly air temperatures at the field site near Nyankpala, northern Ghana, reached as high as 46 °C, with daily average air temperatures ranging from 24 to 32 °C. Daily rainfall varied strongly between the rainy and dry season, with single heavy rain events of up to $115 \,\mathrm{mm}\,\mathrm{d}^{-1}$. Consequently, average monthly air humidity was highest (65% to 85%) during the rainy season and as low as 23 % during the dry season. Irrespective of these harsh environmental conditions, the reliability of the developed low-cost measurement device could be proven during the field trial application. Periodically performed diurnal CO₂ measurement campaigns resulted in consistent R_{eco} and NEE fluxes, showing throughout the entire crop growth a clear light (PAR) dependency for derived GPP fluxes (data not shown). The maximum daily R_{eco} (3.9 g C m⁻² d⁻¹) and GPP $(-6.9 \text{ g C m}^{-2} \text{ d}^{-1})$ fluxes derived for the non-mineralfertilized treatments were well within the range (4.0 and $-7.0 \,\mathrm{g}\,\mathrm{C}\,\mathrm{m}^{-2}\,\mathrm{d}^{-1}$) of eddy covariance (EC) derived maximum daily Reco and GPP fluxes reported by Quansah et al. (2015), who measured a mixed fallow and cropping system in northern Ghana, dominated by tall grasses. When adjusted for observation length, cumulative NEE, GPP and R_{eco} values obtained during the same study (27, -195 and $222 \,\mathrm{g}\,\mathrm{C}\,\mathrm{m}^{-2}$) were found to be consistent with the average cumulative NEE, GPP and R_{eco} values obtained from the non-mineral-fertilized treatments during our field trial application experiment $(-58 \pm 8, -355 \pm 1)$ and $297 \pm 7 \text{ g C m}^{-2}$). Also, EC measurements of an unfertilized cropland system (including maize) in Cameroon resulted with 218.5 g C m⁻² in a comparable cumulative R_{eco} (Verchot et al., 2020). Regarding ET, the highest cumulative ET of our study (FM + MIN; 229 ± 23 mm) was similar to the measured ET flux (238 mm) of a field site in northern Benin, which was dominated by C₄ plants (Mamadou et al., 2016). In general, obtained cumulative ET values (Fig. 8d) for all four treatments were furthermore in good agreement with ET obtained for northern Ghana from average monthly actual evapotranspiration (FAO, 2019), corrected using phenologyspecific crop factors for grain maize (263 mm; Brouwer and Heibloem, 1986). Cumulative R_{eco} and GPP fluxes recorded for the four different treatments reflected well the difference in harvested biomass $(529 \pm 59 \text{ g C m}^{-2} \text{ for FT} + \text{MIN} \text{ and}$ 534 ± 143 g C m⁻² for FM + MIN), with higher cumulative Reco and GPP for higher crop biomass (Fig. 8a-b). Consequently, also NEE and thereon based NECB was higher for additionally mineral-fertilized treatments compared to nonmineral-fertilized treatments, with differences between additionally mineral and non-mineral-fertilized treatments being more pronounced for FM when compared to FT (Fig. 8c and e). Similar tendencies were found for ET and thereon based WUE, with additionally mineral-fertilized treatments showing a higher ET and WUE compared to non-mineralfertilized treatments (Fig. 8d and f). This is in alignment with results reported by Mo et al. (2017) for maize in Kenya, where WUE increased with higher grain yield due to increasing mineral N fertilization. Besides the reliability of the developed low-cost measurement system, also its practicability was proved during the field trial application. Despite of the rather demanding environmental conditions, the system showed that it is uncomplicated and easy to operate even for untrained staff. After a short training session, even nontechnical trained staff can conduct minor repairs of the system directly in the field. However, the missing user interface currently still prevents direct input of information, such as names of measurement location and soil temperatures, which made data postprocessing more tedious.



Figure 7. Comparison of R_{eco} temperature dependency (dotted red lines) and GPP PAR dependency functions (dotted green lines) between LI-850 (dark red/green) and K30 FR and SCD30 (light red/green), respectively. Shaded red/green areas indicate confidence band around functions. Dots represent measured R_{eco} and derived GPP fluxes. Error bars indicate calculated flux error (95% CI; p < 0.05).

4 Conclusions and implications for further use

Performed experiments showed that CO_2 and ET fluxes can be measured reliably and in a stable manner over time using inexpensive NDIR and RH sensors in conjunction with a closed-chamber system. Out of the various low-cost CO_2 and RH sensors that were validated, the K30 FR and SHT31 proved to be the most accurate in measuring CO_2 and ET fluxes, respectively. Additionally, the developed low-cost measurement device was shown to be both practical and applicable to use even in environmentally challenging agroecosystems, as demonstrated by the field trial application in northern Ghana, sub-Saharan Africa. There within, seasonal CO_2 and ET fluxes turned out to be reliable and could be used to obtain valid NECB and WUE estimates. Since the developed system is battery-powered (solar rechargeable), based on open-source technology and all its components are lowcost, it can become easily accessible to a broad range of researchers. Its low weight and low power consumption, with the 12 rechargeable Ni-MH batteries lasting for as long as 8h, make the system especially suitable for in situ closedchamber measurements in remote tropical areas. Compared to Li-ion batteries, the rechargeable Ni-MH batteries are furthermore relatively safe to use at high temperatures. This opens many potential applications, especially in the Global South, regarding the evaluation and identification of various land use systems and management practices, in terms of their C sequestration potential, water consumption and WUE. Therefore, the developed measurement device can be a valuable tool in evaluating and assessing global carbon and water flux models, ultimately expanding the network for C budget and evapotranspiration research that are both critical for climate crisis adaptation and mitigation.



Figure 8. Cumulative (**a-d**) R_{eco} , GPP, NEE (g C m⁻²) and ET fluxes (mm) as well as thereon based estimates of (**e-f**) NECB (g C m⁻²) and WUE (g mm⁻¹) for the four different fertilizer treatments, namely, (1) Fertisoil (5 tha⁻¹; commercial organic fertilizer in northern Ghana; denoted FT), (2) farmyard manure (5 tha⁻¹; denoted FM), (3) Fertisoil + NPK (5 tha⁻¹ + 90–60–60 kg ha⁻¹; denoted FT + MIN) and (4) farmyard manure + NPK (5 tha⁻¹ + 90–60–60 kg ha⁻¹; denoted FM + MIN). Error bars indicate calculated flux error (90% CI; p < 0.1).

Code and data availability. The data and code referred to in this study are publicly accessible at https://doi.org/10.4228/zalf-hdqh-br42 (Macagga et al., 2024).

Author contributions. MH and RM conceptualized and developed the system and code. RM, DA and GS carried out the laboratory and field validation experiments. MA conducted the field trial application. RM, MH and MD wrote and prepared the manuscript with contributions by all co-authors. All authors have reviewed and agreed to the final version of the paper.

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