

Retrieval of NO<sub>2</sub> over  
biomass burning  
regions

N. Bousseréz

# Space-based retrieval of NO<sub>2</sub> over biomass burning regions: quantifying and reducing uncertainties

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## Abstract

The quality of space-based nitrogen dioxide ( $\text{NO}_2$ ) retrievals from solar backscatter depends on a priori knowledge of the vertical profiles of  $\text{NO}_2$  and aerosol optical properties. This information is contained in an air mass factor (AMF), which accounts for atmospheric scattering and is used to convert the measured line-of-sight “slant” columns into vertical columns. In this study we investigate the impact of biomass burning emissions on the AMF in order to quantify  $\text{NO}_2$  retrieval errors in the Ozone Monitoring Instrument (OMI) products over these sources. Sensitivity analyses are conducted using the Linearized Discrete Ordinate Radiative Transfer (LIDORT) model and the GEOS-Chem chemistry-transport model with an improved daily biomass burning emission inventory. Aircraft in situ data collected during two field campaigns, Arctic Research of the Composition of the Troposphere from Aircraft and Satellites (ARCTAS) and Dust and Biomass-burning Experiment (DABEX), are used to evaluate the modeled aerosol optical properties and  $\text{NO}_2$  profiles over Canadian boreal fires and western Africa savanna fires respectively. Biomass burning aerosols increase the AMF by 3 to 15 % over boreal fires, while they decrease the AMF by  $-10$  to  $-30$  % over savanna fires. The presence of an elevated aerosol layer over west Africa due to the Harmattan front explains the negative aerosol effect over this area. The impact of fires on the AMF is driven by the  $\text{NO}_2$  shape profile perturbations, which decrease the AMF by  $-10$  to  $-60$  % over both regions. Aerosol and shape factor effects are most sensitive to surface reflectance and clouds. In particular, retrieval errors associated with shape factor uncertainties can increase by a factor of 2 due to the presence of clouds. In contrast with conclusions from previous studies, we demonstrate that in the presence of pre-existing clouds, the effect of aerosols on the AMF cannot be fully accounted for through the modified retrieved cloud parameters. Finally, a new method that uses slant column information to correct for shape factor error in the retrieval is proposed and tested over west African fires.

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## 1 Introduction

Nitrogen oxides ( $\text{NO}_x = \text{NO}_2 + \text{NO}$ ) play a key role in tropospheric chemistry by affecting ozone, atmospheric oxidation, and aerosol formation (Logan, 1983; Finlayson-Pitts and Pitts, 1986). Biomass burning is a significant source of  $\text{NO}_x$ , but due to the high space and time variability of fire events, this source remains difficult to quantify. Space-based observations of  $\text{NO}_2$  offer the possibility to infer global  $\text{NO}_x$  emissions with a spatial and time coverage that allows to capture such source variability (Leue et al., 2001; Martin et al., 2003). However, top-down estimates of those emissions depend upon the accuracy of the  $\text{NO}_2$  observations themselves.

Satellite retrievals of  $\text{NO}_2$  columns from solar backscatter are usually based on a two-step process in which slant columns are inferred from spectral fitting and subsequently converted into vertical columns using an air mass factor (AMF) (Richter and Burrows, 2002; Martin et al., 2002). The AMF corrects the measurements for the viewing geometry and the atmospheric scattering effects based on a priori knowledge of the  $\text{NO}_2$  profile shapes and atmospheric optical properties. The AMF is the largest source of error in  $\text{NO}_2$  retrievals over  $\text{NO}_x$  sources (Boersma et al., 2004; Martin et al., 2002) and  $\text{NO}_2$  profile uncertainty contributes significantly to the AMF error (Russel et al., 2011; Hains et al., 2010). The  $\text{NO}_2$  profiles are usually obtained from a chemical transport model (CTM). Because the high spatial and temporal variability of fires is generally not resolved in such models, significant uncertainties in the simulated  $\text{NO}_2$  profile are expected. Biomass burning aerosols provide an additional complication to the AMF by altering the scattering properties of the atmosphere (Martin et al., 2003; Leitão et al., 2010). Previous studies have investigated the impact of aerosols on the AMF using modeled profiles of aerosol characteristics (Leitão et al., 2010; Martin et al., 2003). However, it is still unclear how to account for aerosols in the  $\text{NO}_2$  retrieval, since cloud products used in AMF calculations are also sensitive to aerosols (Boersma et al., 2011, 2004). In this study, we assess the impact of biomass burning emissions on

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the AMF through the modified NO<sub>2</sub> and aerosol profiles and we analyze the interplay between clouds and aerosols in the retrieval algorithm.

The Arctic Research of the Composition of the Troposphere from Aircraft and Satellites (ARCTAS) and The Dust And Biomass EXperiment (DABEX) experiments provide extensive aircraft in situ measurements of NO<sub>2</sub> concentrations and aerosol optical properties over boreal and savanna fires respectively. ARCTAS-B took place in June–July 2008 with a major emphasis on boreal forest fire influences over Canada (Jacob et al., 2009). DABEX, part of the African Monsoon Multidisciplinary Analysis (AMMA) (Redelsperger et al., 2006), took place in January 2006 over West Africa. These two campaigns provide an unprecedented dataset to evaluate modeled aerosol and NO<sub>2</sub> profiles that are used in the AMF calculation. To our knowledge, a comprehensive analysis of NO<sub>2</sub> retrievals over biomass burning supported by detailed aerosol and NO<sub>2</sub> in situ dataset has not been conducted yet.

Section 2 presents the theoretical basis for the AMF algorithm. Section 3 provides a description of the ARCTAS-B and DABEX experiments. Section 4 evaluates the ability of the GEOS-Chem model to reproduce the observed NO<sub>2</sub> profiles and aerosol optical properties. Sections 5 to 8, we analyze the sensitivity of the AMF to biomass burning emissions and investigate the impact of clouds. Finally, in Sect. 9 we present a method to reduce AMF errors associated with NO<sub>2</sub> shape factor uncertainties by applying a correction factor based on slant column information.

## 2 AMF calculation

The AMF is defined as the ratio of the slant column, which is the total amount of trace gas along an average backscattered path observed by a satellite instrument, to the vertical column of this trace gas. Following Palmer et al. (2001), the AMF can be expressed as

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$$\text{AMF} = \text{AMF}_G \int_0^1 \omega(\eta) S(\eta) d\eta \quad (1)$$

where  $\text{AMF}_G$  is the geometric AMF accounting for the satellite viewing geometry,  $S(\eta)$  is the *shape factor*, a dimensionless normalized NO<sub>2</sub> vertical profile, and  $\omega(\eta)$  are the scattering weights describing the vertical sensitivity of the radiance observed by the instrument to NO<sub>2</sub>.

The AMF formulation we use takes into account cloud-contaminated pixels, as described in Martin et al. (2002). The effective cloud fraction and cloud pressure are retrieved from OMI using the O<sub>2</sub>-O<sub>2</sub> algorithm (Acarreta et al., 2004). Surface reflectivity was obtained from a climatology from OMI at 440 nm (Kleipool et al., 2008). Scattering weights are calculated using the Linearized Discrete Ordinate Radiative Transfer (LIDORT) model (Spurr et al., 2001; Spurr, 2002). Shape factors are obtained from a GEOS-Chem simulation described in Appendix A.

### 3 Observational dataset

#### 3.1 ARCTAS-B campaign

The summer ARCTAS-B campaign took place in late June and early July 2008 (Jacob et al., 2009). The primary goal of this experiment was to better understand the factors driving current changes in Arctic atmospheric composition and climate.

The campaign included the DC-8 research aircraft to measure chemical and radiative properties of the atmosphere. From 26 June to 14 July 2008, the aircraft was based in Cold Lake, Canada (54° N, 110° W), where it sampled many fresh and aged biomass burning plumes (Alvarado et al., 2010). The DC-8 included highly sensitive high-frequency (1–10 s) measurements of key ozone precursors (NO<sub>y</sub>, HO<sub>x</sub>, PAN,

HCHO) as well as aerosol optical and physical properties. Measurement characteristics for the species considered in this study are summarized Table 1.

Figure 1 shows the DC-8 aircraft flight paths during the ARCTAS-B experiment. Aircraft flights over central Canada mostly targeted biomass burning plumes and sampled a wide range of altitudes, providing valuable information on the influence of fires on tropospheric chemistry. Chemical and optical properties of biomass burning plumes sampled by the DC-8 aircraft during ARCTAS-B were analyzed in detail in Singh et al. (2010) and Alvarado et al. (2010).

### 3.2 DABEX campaign

The African Monsoon Multidisciplinary Analysis (AMMA) is a major international campaign designed to improve our understanding of the west African monsoon (Redelsperger et al., 2006). The Dust And Biomass burning EXperiment (DABEX) was part of the Special Observing Period (SOP) dry season (January–February 2006). It was carried out over West Africa by the UK Met Office, which used the FAAM BAe-146 aircraft to perform high-quality in situ measurements of the chemical, optical and physical properties of biomass burning and dust aerosols over western Africa.

Observed optical properties of dust and biomass burning aerosols during DABEX were analyzed in detail in several studies (Johnson et al., 2008a,b; Osborne et al., 2008; Capes et al., 2008). Although valuable information could be obtained from this dataset, characterization of optical properties for fresh biomass burning aerosols were challenging since fresh plumes were usually sampled near the surface where they were also mixed with mineral dust.

Table 2 summarizes the measurement characteristics for the data considered in this study. Figure 2 shows the FAAM BAe-146 flight tracks and altitude during DABEX. The area sampled covered a narrow domain spanning from western Niger, where mineral dust is predominant, to center Benin, where savanna fires occurred.

To complete our model evaluation, we also used aerosol optical depth (AOD) and single scattering albedo (SSA) surface in situ data from the Aerosol Robotic Network

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(AERONET) network, which consist of sun photometer measurements (Dubovik et al., 2002). Figure 2 shows the location of the stations considered.

#### 4 Model evaluation

The GEOS-Chem model described in Appendix A is a central tool in our analysis of the AMF, since it provides the aerosol and NO<sub>2</sub> profiles used in its calculation. Here we evaluate the simulated aerosol and NO<sub>2</sub> profiles against aircraft in situ measurements from the DABEX and ARCTAS experiments. For this purpose, model outputs were sampled along the flight tracks and filtered to retain only measurements influenced by biomass burning.

For ARCTAS, measurements impacted by biomass burning emissions were identified by enhancements in HCN and CO concentrations greater than 20% of the regional background.

For DABEX, measurements influenced by biomass burning were defined according to Johnson et al. (2008b), who derived the following criteria:

- scattering at 550 nm (green)  $> 30 \cdot 10^{-6}$  m
- Angström exponent  $> 1$

We then performed a sensitivity simulation to ensure that the selected locations were also associated with biomass burning influences in the model. Modeled and observed profiles were subsequently derived by vertically binning the data over the entire ARCTAS (150–80° W; 49–70° N) and DABEX (2° W–8° E; 8–20° N) domains.

Figure 3 shows the shape factor, aerosol extinction and SSA mean profiles during ARCTAS for the DC-8 measurements and the model. There is an overall good agreement between the model and the measurements. The NO<sub>2</sub> shape factor shows a sharp enhancement near the surface due to fire emissions. Model analyses (not shown) suggest that the gradient observed at the top of the troposphere and not seen in the model

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is due to an underestimation of lightning  $\text{NO}_x$  emissions. Aerosol extinction also increases in the boundary layer as a result of biomass burning emissions. The shape and magnitude of the extinction is similar in the model and the measurements. The mean SSA profile is also reasonably well reproduced by the model, with relatively homogeneous values of 0.90–0.95.

Figure 4 shows the profiles of the shape factor, aerosol extinction and aerosol scattering at 450 nm from GEOS-Chem and the DC-8 measurements during DABEX. The modeled and observed shape factors show a significant increase near the surface, similar to the one observed over boreal fires. Aerosol extinction values over savanna fires ( $\sim 0.4\text{--}0.8 \text{ km}^{-1}$ ) are significantly greater than those observed over boreal fires during ARCTAS, which is due to more intense fire emissions over Africa during DABEX. Aerosol extinction and scattering profile shapes are well reproduced by GEOS-Chem. The model underestimation near the surface may be attributed to representativeness errors. Note that infrequent sampling of the boundary layer during DABEX limits confidence in the results near the surface.

Table 3 contains the AOD and SSA columns for GEOS-Chem and the AERONET measurements at several stations in western Africa. The modeled SSA agrees with observations within 0.02 in average, and the modeled AOD within 23 %.

In conclusion, the evaluation of the GEOS-Chem  $\text{NO}_2$  profiles and aerosol optical properties over boreal and savanna fires shows that this model can be used to quantify the impact of biomass burning emissions on the AMF.

## 5 Impact of biomass burning emissions on the AMF

Generally the high spatial and temporal variability of fire emissions is not accurately represented in CTMs. Errors in the simulated aerosol and  $\text{NO}_2$  profiles will translate into errors in the AMF estimates. Such errors can be quantified by doing an AMF sensitivity analysis, where aerosols and  $\text{NO}_2$  emitted by fires are excluded from the model simulation. Here we define a total biomass burning AMF correction factor ( $\text{tot}_{\text{cor}}$ ) as the

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and savanna fires. Over boreal fires, the aerosol and NO<sub>2</sub> layers are co-located and extend from the surface to ~2 km, whereas over savanna fires the top of the aerosol layer reaches ~4 km altitude, well above the surface NO<sub>2</sub> layer. The simulated vertical extension of the aerosol layer over west Africa is in agreement with the CALIPSO observations for the same period, as shown in Fig. 6. As explained in Haywood et al. (2008), the presence of this elevated aerosol layer is due to the uplift of biomass burning emissions by the Harmattan front. The narrower NO<sub>2</sub> layer stems from the shorter lifetime of NO<sub>x</sub> compared to aerosols.

The difference in the relative vertical distribution of aerosols and NO<sub>2</sub> between boreal and savanna fires explains the opposite sign in the aerosol correction. Indeed, surface aerosols mixed with NO<sub>2</sub> tend to increase measurement sensitivity due to light scattering, while aerosols above NO<sub>2</sub> produce a “shielding” effect that decreases measurement sensitivity (Leitão et al., 2010).

In the next two sections we investigate in more detail the sensitivity of the AMF to aerosols and NO<sub>2</sub> shape factor over west African fires.

## 6 Aerosol correction sensitivity analysis

In this section, we assess the influence of aerosols on the AMF and its dependence to aerosol characteristics (optical properties, profile shape) as well as key retrieval parameters (satellite viewing geometry, surface reflectance, clouds). The reference aerosol and NO<sub>2</sub> profiles for the sensitivity experiment are taken from the GEOS-Chem DABEX simulation over west Africa. Since the model resolution (2° × 2.5°) is much coarser than OMI, and in order to minimize the representativeness error for the NO<sub>2</sub> profiles, we chose model profiles associated with highest fire emissions. Figure 7 shows the simulated NO<sub>2</sub> and aerosol profiles at OMI overpass time (13:30 LT – local time).

The sensitivity of the aerosol correction to aerosol profile shape has been assessed by comparing the correction with and without the elevated aerosol layer, as represented

Fig. 7. As a result of the shielding effect, the aerosol correction decreases from +20 to -30 % in the presence of the elevated aerosol layer.

Figure 8 shows the influence of SSA, solar zenith angle (SZA) and surface reflectance on the aerosol correction for the two types of profile considered. For each of these parameters, the sensitivity of the aerosol correction is similar whether or not the elevated aerosol layer is included. The AMF being a vertical integration of the scattering coefficients weighted by the NO<sub>2</sub> shape factor, its sensitivity is driven by optical properties near the surface where the NO<sub>2</sub> shape factor peaks, which explains the similarities observed.

The aerosol correction is most sensitive to surface reflectance. We observe an exponential decay, which reflects the competing effects of the measurement sensitivity increase due to scattering above the aerosol layer and the shielding effect below it. We see that the relative impact of the shielding effect increases quickly until it reaches an asymptotic limit for surface reflectance > 0.1.

A linear increase of the aerosol correction with SSA is observed, while the viewing geometry (SZA) has negligible impact on the correction, which is consistent with findings by Leitão et al. (2010) over urban areas.

The impact of emissions on the aerosol correction has been assessed by re-scaling the aerosol and NO<sub>2</sub> concentrations within the same surface layer (0–1 km). Figure 8 d) shows that the effect of aerosols on the AMF is almost insensitive to emission intensity. This is due to the AMF increase caused by aerosol scattering effects being counterbalanced by the AMF decrease due to NO<sub>2</sub> shape factor enhancements at the surface. A direct consequence is that the aerosol correction will be only weakly impacted by representativeness errors within one model grid box as long as the aerosol/NO<sub>2</sub> emission factors are homogeneous inside that domain.

Clouds can also indirectly modify the effect of aerosols on the AMF. Figure 9 shows the aerosol correction as a function of cloud irradiance fraction for different AODs, and for the two aerosol profiles considered in Fig. 7. Here the cloud top pressure was set to 460 hPa, which corresponds to the clouds being above the elevated aerosol layer. For

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any AOD value, the aerosol correction increases linearly with cloud irradiance fraction, since decreasing the sensitivity below clouds increases the relative impact of aerosols scattering on the AMF. Note that increasing the optical thickness of the elevated aerosol layer does not modify the sensitivity of the aerosol correction to cloud irradiance fraction. This is expected since an elevated aerosol layer will essentially play the same role as an increase in the cloud irradiance fraction in the retrieval. However in the case where NO<sub>2</sub> and aerosols are co-located near the surface, the aerosol correction sensitivity to cloud irradiance fraction increases with AOD. As a result, for AODs > 1, commonly measured over biomass burning events, we see from Fig. 9 that a cloud irradiance fraction of 20 % can lead to an overestimation of a factor 2 in the retrieved NO<sub>2</sub> columns when aerosol effects are not accounted for. It is important to realize that although in our pseudo-experiment the effect of clouds can be separated from the effect of aerosols, in real retrievals they are difficult to disentangle. This important aspect is analyzed in the next section.

### 7 Do clouds implicitly account for aerosol effects?

Aerosols may impact the retrieval of cloud parameters. Boersma et al. (2011, 2004) suggested that the effect of aerosols on the AMF was implicitly accounted for through the modified cloud parameters. They recommended to not use an explicit aerosol correction in the AMF, since it could potentially be detrimental to the quality of the NO<sub>2</sub> retrievals. However the results presented in their studies were derived using specific assumptions on the retrieved scene (clear sky, exponentially decreasing shape profiles for aerosols and NO<sub>2</sub>), and this conclusion may not be valid in all cases. A particular case is the presence of pre-existing clouds in the scene, which as we showed in Sect. 6 modifies the effect of aerosols on the AMF.

Let us assume as Boersma et al. (2004) that under clear-sky conditions the aerosol-induced perturbation of the cloud parameters produces a similar effect on the AMF than an explicit aerosol correction. From our previous results, if true clouds are added

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to the scene, the aerosol correction will increase. Since this aerosol correction perturbation is not associated with any perturbation of the retrieved cloud parameters (it is an artifact of clouds), we see that in the presence of true clouds the implicit aerosol correction cannot fully account for the explicit aerosol effects. This result suggests that the impact of aerosols should be explicitly accounted for in the AMF calculation. In practice though, this would require both a good knowledge of the aerosol vertical distribution and accurately quantifying its impact on the OMI cloud retrieval.

### 8 Shape factor correction sensitivity analysis

The impact of biomass burning emissions on the AMF through the modified NO<sub>2</sub> shape factor is also altered by changes in other retrieval parameters. Figure 10 shows the sensitivity of the shape factor correction to cloud irradiance fraction and surface reflectance. The shape factor correction decreases almost linearly with cloud fraction. For a cloud fraction of 30 %, which is the threshold currently used in operational OMI NO<sub>2</sub> retrievals, AMF shape factor errors can be amplified by a factor 2. Therefore, using a lower threshold for the cloud screening over regions with uncertain NO<sub>x</sub> emissions (like biomass burning areas) might be critical in order to reduce retrieval errors.

Similarly to the aerosol correction, the shape factor correction is highly sensitive to surface reflectance at low values. For surface reflectances typically found over savanna or boreal forest areas (0.03–0.05), our results show that retrieval errors associated with shape factor uncertainties can be as high as 60 %. This highlights the need for using high spatial resolution surface albedo climatology in the AMF calculation, as recently included in an improved OMI NO<sub>2</sub> product by Boersma et al. (2011).

## 9 Measurement-based shape factor correction

### 9.1 Methodology

In the previous section we showed that errors in the simulated  $\text{NO}_2$  shape factor are the dominant source of error in the AMF over biomass burning regions. Of concern is that current operational OMI  $\text{NO}_2$  retrievals use climatological  $\text{NO}_2$  profiles that cannot represent the true spatial and temporal variation needed for accurate retrievals. Here we investigate a method to infer shape factor corrections using measurement information. As explained in Sect. 2, measurement information is contained in the slant columns, whereas AMFs are calculated using a priori information, usually from a CTM. Over high  $\text{NO}_x$  sources like fires, most of the  $\text{NO}_2$  column will be concentrated in the boundary layer. Local variability of the slant columns will therefore be primarily affected by variability of the lower part of the  $\text{NO}_2$  profile.

Here we use this property to establish a relationship between the shape factor correction and a slant column perturbation, defined as the relative increase of the slant column with respect to a case without biomass burning emissions. Slant columns and shape factor corrections are computed using a pseudo-retrieval experiment, in which OMI  $\text{NO}_2$  averaging kernels are applied to the GEOS-Chem  $\text{NO}_2$  profiles (Eskes et al., 2003). Two GEOS-Chem simulations, one using daily biomass burning emissions ( $\text{BB}_{\text{sim}}$ ), and one with biomass burning emissions turned off ( $\text{NOBB}_{\text{sim}}$ ) are used to generate three types of pseudo  $\text{NO}_2$  retrievals:

- $A$ , of  $\text{BB}_{\text{sim}}$  using  $\text{NO}_2$  shape factors from  $\text{NOBB}_{\text{sim}}$ ;
- $B$ , of  $\text{NOBB}_{\text{sim}}$  using  $\text{NO}_2$  shape factors from  $\text{NOBB}_{\text{sim}}$ ;
- $C$ , of  $\text{BB}_{\text{sim}}$  using  $\text{NO}_2$  shape factors from  $\text{BB}_{\text{sim}}$ .

Note that using a perfect shape factor in  $B$  and  $C$  (the one corresponding to the column retrieved) results in  $B$  and  $C$  being the actual GEOS-Chem columns.

We further define:

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–  $\Delta\text{NO}_2 = \frac{A-B}{B}$ , the relative  $\text{NO}_2$  column increase compared to the case without biomass burning;

–  $\text{NO}_2$  correction factor =  $\frac{C}{A}$ .

The dependence of  $\Delta\text{NO}_2$  to measurement information can be made explicit by  
5 rewriting:  $\Delta\text{NO}_2 = \frac{\text{Sl}_A - \text{Sl}_B}{\text{Sl}_B}$ .

Figure 11 shows the relationship between  $\Delta\text{NO}_2$  and the  $\text{NO}_2$  correction factor over western Africa fires during DABEX. The correction factor increases with  $\Delta\text{NO}_2$  with a “saturation effect” for  $\Delta\text{NO}_2 > 4$ , where the  $\text{NO}_2$  correction is almost constant (= 3). Further increasing  $\text{NO}_x$  emissions above this threshold does not change the shape  
10 factor since almost the totality of the  $\text{NO}_2$  column is already concentrated in the mixed layer. The relationship between the  $\text{NO}_2$  correction factor and  $\Delta\text{NO}_2$  can be well approximated by the following logarithmic fit:

$$\text{NO}_2 \text{ correction factor} = \log(3.4 \Delta\text{NO}_2 + 2.2). \quad (4)$$

The above formula can be applied to real retrievals by replacing the pseudo  $\text{NO}_2$   
15 slant columns in  $A$  by the measured  $\text{NO}_2$  slant columns and assuming the reference column in  $B$  is a good approximation of what would be the measured slant column without the fire source. Such measurement-based shape factor correction is intended to reduce uncertainties related to both representativeness errors and misrepresentation of the location and time of fires in the model simulation.

## 20 9.2 Results

Figure 12 shows the average  $\text{NO}_2$  tropospheric column distribution over west Africa during DABEX from the OMI DP GC product (Lamsal et al., 2010), which consists in replacing the TM4  $\text{NO}_2$  shape profile by the GEOS-Chem  $\text{NO}_2$  shape profile in the DOMINO product (Boersma et al., 2007). Here  $\text{NO}_2$  profiles are obtained from

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the DABEX simulation, using the FLAMBE daily biomass burning emissions inventory. MODIS AOD over potential biomass burning NO<sub>x</sub> sources (OMI DP GC NO<sub>2</sub> columns > 10<sup>15</sup> molec cm<sup>-2</sup>) as well as the percentage change in the OMI DP GC NO<sub>2</sub> columns associated with the measurement-based correction are also shown. Corrections to the original OMI NO<sub>2</sub> retrieval can reach more than 40 % over fires, where MODIS AODs are maximum. The fact that only positive correction are observed reflects a low bias in the modeled shape factor at the surface, which is likely related to an overall underestimation of biomass burning emissions over west Africa.

The standard deviation of the NO<sub>2</sub> correction in figure (Fig. 12) can be interpreted as the error associated with the logarithmic fit approximation (Eq. 4). Although large uncertainties can be observed, one sees that they are all significantly smaller than the retrieval error associated with missing biomass burning sources. We therefore recommend using the proposed correction for NO<sub>2</sub> retrievals over biomass burning regions. Note that generalization of this formula to retrievals over other areas (e.g. urban) would require more investigation, since it is possible that very different characteristics in parameters such as surface reflectance could result in a different relationship. Also, due to the lack of aircraft measurements in the boundary layer during DABEX, our approach cannot be rigorously evaluated using in situ-derived NO<sub>2</sub> columns. A study using in situ data from other aircraft campaigns (e.g. INTEX-A, INTEX-B and PAVE) would allow to properly evaluate the robustness of the above formula as well as to better characterize the errors.

## 10 Conclusions

In this work, we analyze the sensitivity of the air mass factor (AMF) to biomass burning emissions in order to better characterize and understand NO<sub>2</sub> retrieval uncertainties over fire areas. The dependence of the AMF to aerosols and NO<sub>2</sub> shape profiles is investigated in details using outputs from the GEOS-Chem model with daily-resolved fire emissions (FLAMBE) over boreal and savanna fires. Simulated aerosol optical prop-



erties and NO<sub>2</sub> profiles were validated against aircraft in situ measurements from the ARCTAS-B (Canadian fires, summer 2008) and DABEX (western Africa fires, winter 2006) campaigns.

Biomass burning aerosols increase the AMF by 3 to 15% over Canadian boreal fires, and decrease it by -10 to -35% over western Africa savanna fires. Co-localized aerosol and NO<sub>2</sub> layers at the surface over boreal fires explains the increase in measurement sensitivity due to aerosol scattering. Over west Africa, uplift of biomass burning emissions by the Harmattan front generates an elevated aerosol layer that decreases measurement sensitivity by shielding effect. Over both regions, the impact of fire-induced shape factor perturbations on the AMF range from -10 to -60%.

We further analyzed the influence of several retrieval parameters (surface reflectance, clouds, viewing geometry) on the AMF sensitivity to aerosols and shape factors. Both aerosol and shape factor effects are most sensitive to surface reflectance. Our results show that using accurate surface reflectances is critical to reduce retrieval errors over biomass burning areas. Our sensitivity study also shows that for a given NO<sub>2</sub>/aerosols emission ratio the effect of aerosols on the AMF is not sensitive to the intensity of emissions, which suggests that representativeness errors are not significant in this case.

We found that the impact of biomass burning emissions on the AMF is very sensitive to clouds. In particular, shape factor-related AMF errors can increase by a factor of 2 due to the presence of clouds in the retrieval. On the other hand, the OMI cloud retrieval is sensitive to aerosols, and previous studies suggested that the aerosol effect on the AMF was implicitly accounted for through the modified cloud parameters in the NO<sub>2</sub> retrieval. Here we demonstrate that in the case of pre-existing clouds, this implicit cloud correction cannot fully account for an explicit aerosol correction. Therefore, more investigations are needed in order to better understand in what cases including an explicit aerosol correction reduces retrieval errors compared to a retrieval that does not account for aerosol effects.

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Finally, we developed a new method to reduce NO<sub>2</sub> retrieval errors associated with misrepresentations of the NO<sub>2</sub> shape factors. A measurement-based shape factor correction, based on the local sensitivity of the NO<sub>2</sub> slant column to NO<sub>2</sub> sources, has been proposed and applied to real OMI NO<sub>2</sub> measurements over west Africa. The results show an increase of up to 40 % compared to the original tropospheric NO<sub>2</sub> column retrievals over highest fire emissions. In practice, this methodology could be extended to any type of NO<sub>x</sub> source, with possible adaptations of the formula over areas with very different retrieval characteristics (e.g. surface reflectance). Future studies using in situ-derived NO<sub>2</sub> columns over urban and biomass burning regions would allow a proper evaluation of this method and to assess its relevance for existing operational OMI NO<sub>2</sub> retrievals.

## Appendix A

### GEOS-Chem simulations

GEOS-Chem is a global 3-D chemical transport model driven by assimilated meteorological observations from the Goddard Earth Observing System (GEOS) of the NASA Global Modeling and Assimilation Office (GMAO) ([www.geos-chem.org](http://www.geos-chem.org); Bey et al., 2001). We use the GEOS-5 meteorological data regridded to 2°, latitude × 2.5°, longitude, with 47 vertical layers. For both the DABEX and ARCTAS-B period we use the NRT-ARCTAS GEOS-Chem simulation, which was originally developed in support of the ARCTAS mission and is based on GEOS-Chem v8-01-01 with minor modifications (Mao et al., 2010). The simulation includes ozone-NO<sub>x</sub>-HO<sub>x</sub>-VOC-aerosol chemistry as described in Park et al. (2005).

A daily biomass burning emission inventory with 1° × 1° resolution obtained from Fire Locating and Monitoring of Burning Emissions (FLAMBE) (Reid et al., 2009) is used to capture the high spatial and temporal variability of the fire emissions. The FLAMBE product is constrained by GOES and MODIS fires count data and provides daily global

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smoke and carbon emissions. Emissions for black carbon and organic carbon species in GEOS-Chem are obtained by applying emission factors to the FLAMBE smoke emissions, while emissions for other species, including NO<sub>x</sub>, are derived by applying emission factors to the carbon emissions. Emission factors are those proposed by Andreae and Merlet (2001). Further information about the biomass burning emissions in the model can be found in Fisher et al. (2010).

Dust aerosol concentrations can be significant over west Africa due to transport from the Sahara desert. The mineral dust module in GEOS-Chem simulates the mobilization of dust from the Earth's surface, gravitational settling, and wet and dry deposition as described in Fairlie et al. (2007).

For our ARCTAS simulation, a 70 % decrease of carbon emissions and a 60 % decrease of smoke emissions in the FLAMBE inventory, as well as the use of an extra-tropical forest fires NO<sub>x</sub> emission factor of 0.42 g NO/kg DM (Alvarado et al., 2010) were required in order to match the aircraft in situ observations.

*Acknowledgements.* This work was supported by the Canadian Foundation for Climate and Atmospheric Sciences.

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**Table 1.** DC-8 research payload during ARCTAS-B for the species considered in this study.

Species, parameters	Method	Investigator	Reference
NO <sub>2</sub>	TD-LIF*	R. Cohen, UC Berkeley	Cleary et al. (2002)
Aerosol optical properties	Nephelometer, PSPAP	B. Anderson, NASA LaRC	Anderson et al. (1998)
Black carbon aerosol	SP2**	Y. Kondo, U. Tokyo	Moteki and Kondo (2007, 2008)

\* TD-LIF = Thermal-Dissociation Laser Induced Fluorescence, \*\* SP2 = Single Particle Soot Photometer.



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**Table 2.** FAAM BAe-146 research payload during DABEX for the species considered in this study.

Species, parameters	Instrument	Comment
NO <sub>2</sub>	TECO 42	detection limit < 2 ppbv
Aerosol absorption	Radiance Research PSAP	wavelength, $\lambda = 0.568 \mu\text{m}$
Dry aerosol scattering	TSI 3563 Nephelometer	$\lambda = 0.45, 0.55, 0.70 \mu\text{m}$

**Table 3.** Simulated and observed mean SSA and AOD columns at several AERONET sites during DABEX.

	SSA	AOD
Banizoumbou (13° N, 2° E)		
AERONET	0.84	0.75
GC	0.84	0.63
Djougou (9° N, 1° E)		
AERONET	0.82	1.05
GC	0.84	1.19
IER-Cinzana (13° N, 5° W)		
AERONET	0.83	0.65
GC	0.85	0.61
Ilorin (8° N, 4° E)		
AERONET	0.86	1.37
GC	0.89	0.80
DMN-Maine Soroa (13° N, 12° E)		
AERONET	0.88	0.80
GC	0.82	1.23
Dakar (14° N, 16° W)		
AERONET	0.83	0.84
GC	0.85	0.95

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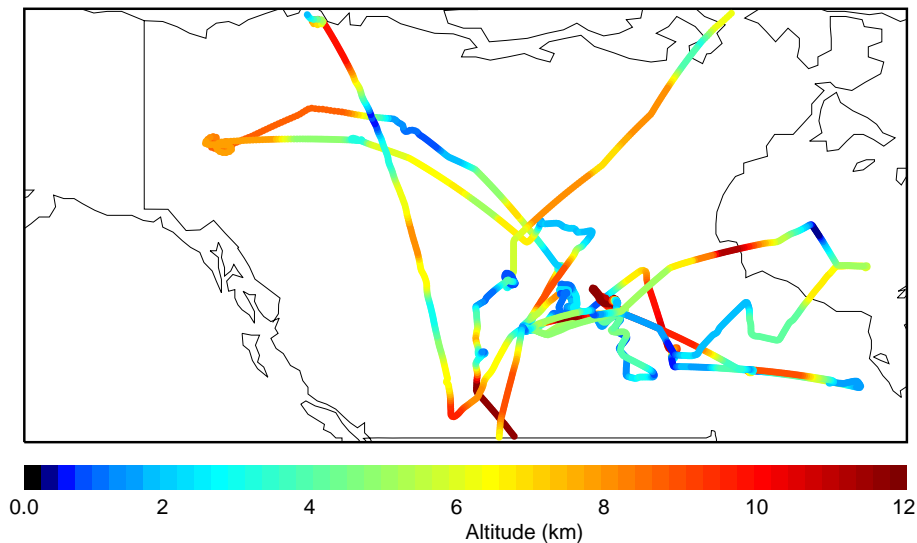
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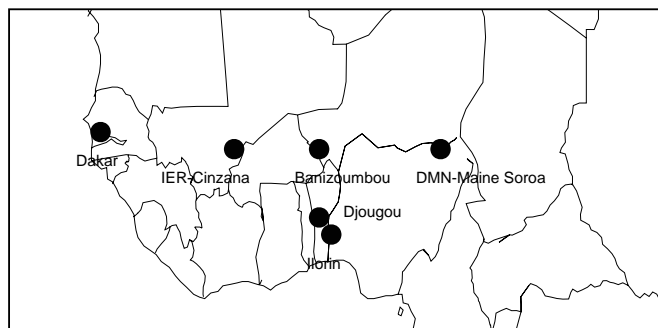
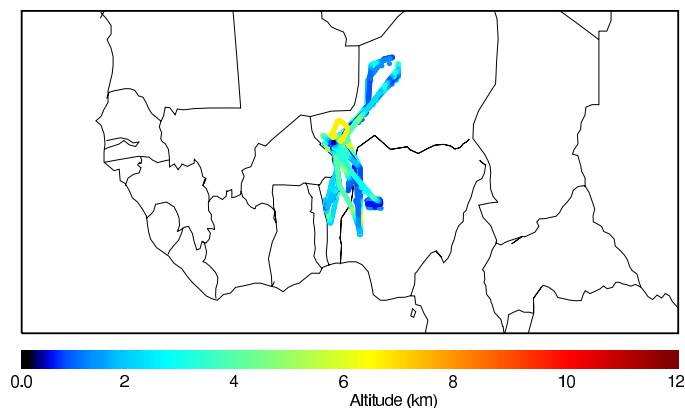
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**Fig. 1.** NASA DC-8 flight tracks during ARCTAS-B over Canada. The color indicates the altitude of the aircraft.

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**Fig. 2.** Top panel: FAAM BAe-146 flight tracks during DABEX. The color indicates the altitude of the aircraft. Bottom panel: locations and names of the AERONET stations considered in this study.

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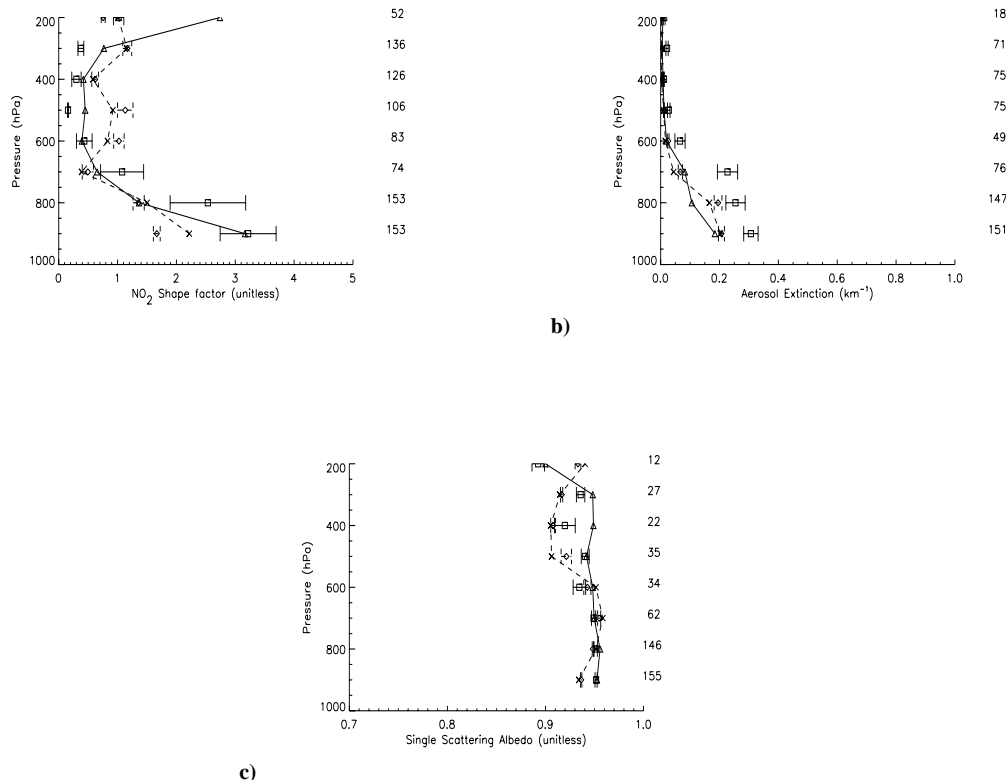
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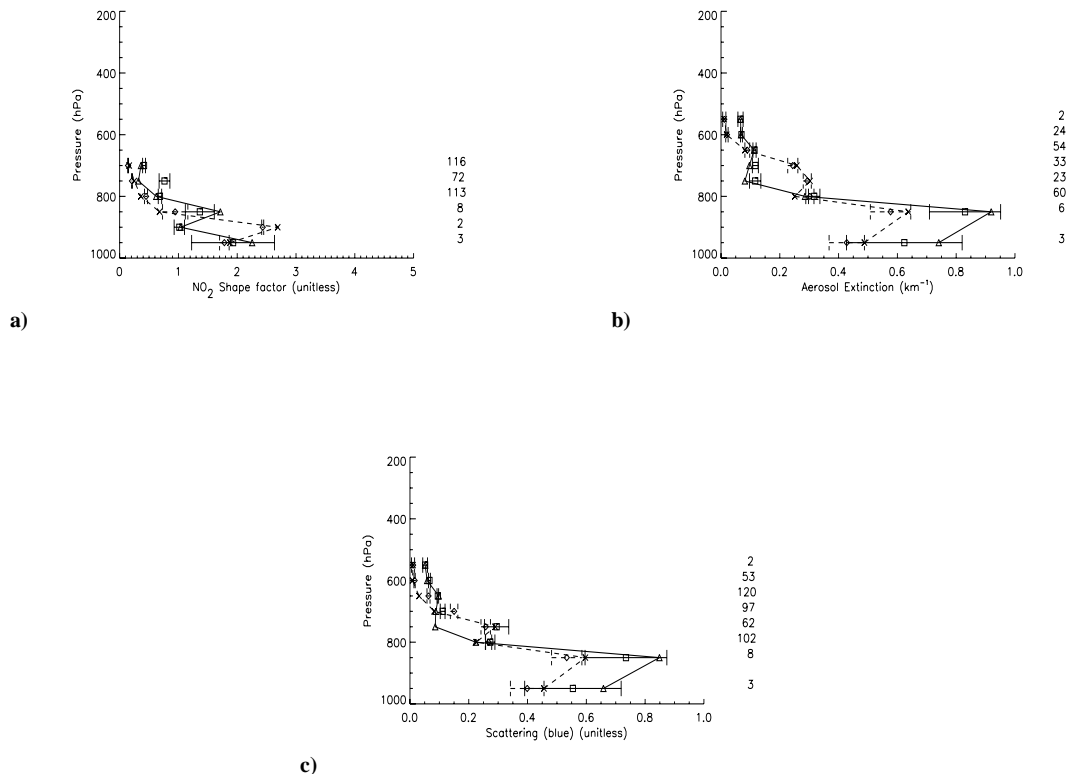
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**Fig. 3.** Simulated and observed vertical profiles of NO<sub>2</sub> shape factors **(a)**, aerosol extinction **(b)**, and single scattering albedo **(c)** during ARCTAS. Model median values (dashed line) are represented by crosses and measurements median values (solid line) by triangles. Model mean values are represented by diamonds and measurements mean values by square. Bars indicate the standard error of the mean at each altitude level.

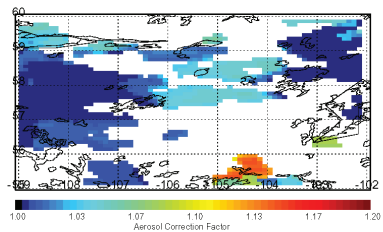
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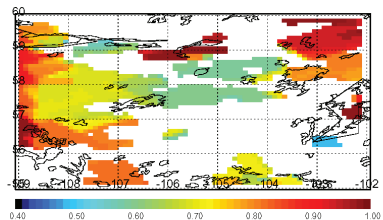


**Fig. 4.** Simulated and observed vertical profiles of  $\text{NO}_2$  shape factors (**a**), aerosol extinction (**b**), and aerosol scattering at 450 nm (blue) (**c**) during DABEX. Model median values (dashed line) are represented by crosses and measurements median values (solid line) by triangles. Model mean values are represented by diamonds and measurements mean values by square. Bars indicate the standard error of the mean at each altitude level.

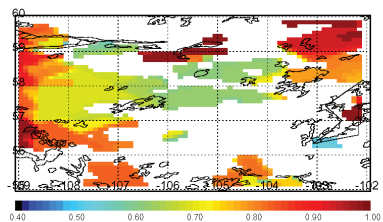
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Aerosol correction

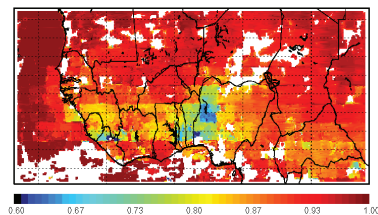


Shape factor correction

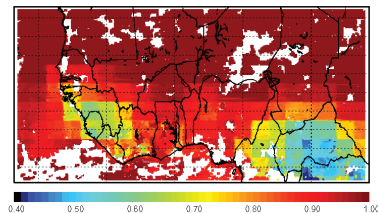


Total correction

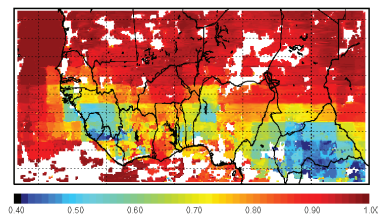
## Africa



Aerosol correction



Shape factor correction



Total correction

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**Fig. 5.** Mean AMF aerosol correction, shape factor correction and total correction over Canadian fires during ARCTAS (30 June–10 July 2008) (left column) and western Africa fires during DABEX (13 January–2 February 2006) (right column). Retrievals with a cloud irradiance fraction > 5 % have been excluded.

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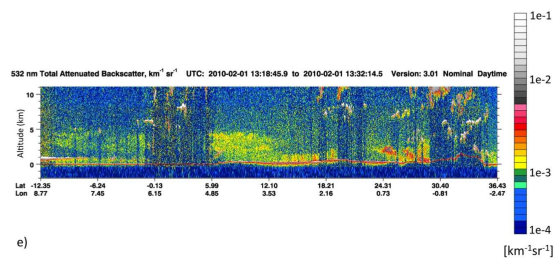
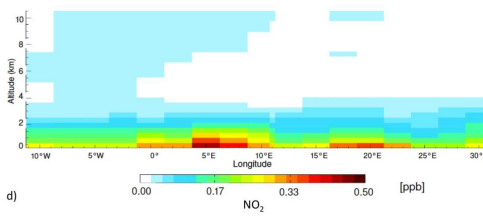
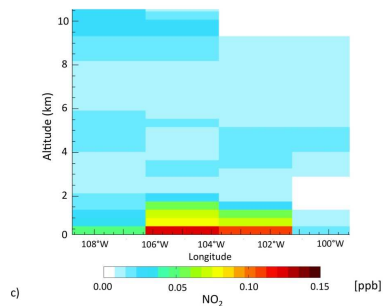
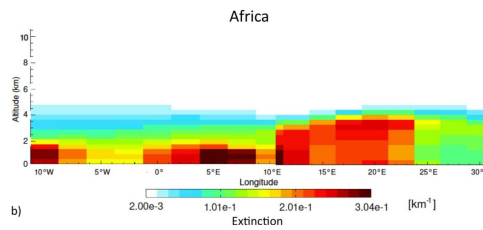
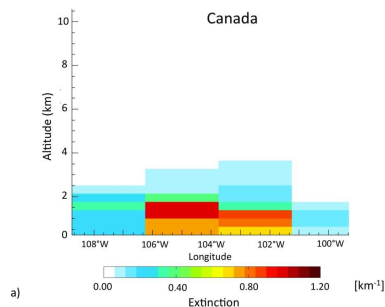
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**Fig. 6.** Mean vertical distribution of the simulated aerosol extinction and NO<sub>2</sub> mixing ratio over Canada (109° W–99° W, 56° N) during ARCTAS (left column) and west Africa (10° W–30° E, 6° N–12° N) during DABEX (right column). Model outputs have been sampled at OMI overpass time (01:30 LT – local time) and averaged over land only. Bottom panel: a transect of the CALIPSO total attenuated backscatter (532 nm) on 1 February 2010 showing the vertical extension of the aerosol layer over west African fires.

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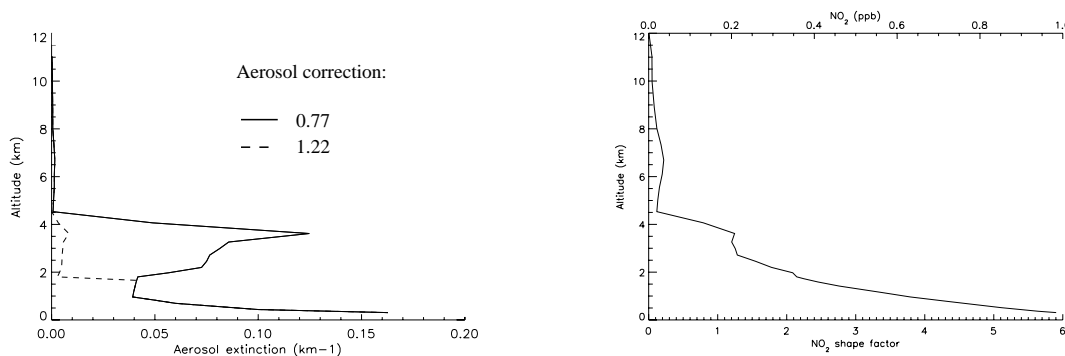
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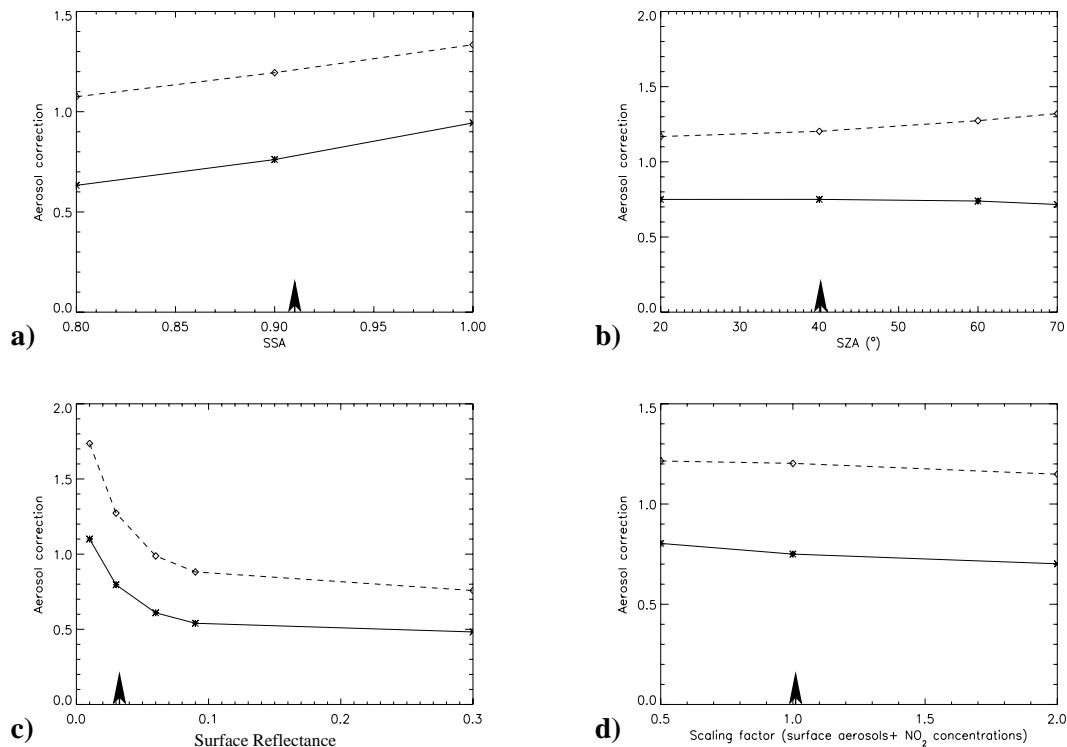


**Fig. 7.** Left panel: GEOS-Chem aerosol extinction profile with (solid) and without (dashed) the elevated aerosol layer for a single grid cell over west Africa fires at OMI overpass time (01:30 LT). The aerosol correction values corresponding to each type of aerosol profile are reported on the figure. Right panel: GEOS-Chem NO<sub>2</sub> profile shape for the same grid cell.

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**Fig. 8.** Aerosol correction sensitivity to: **(a)** single scattering albedo, **(b)** solar zenith angle, **(c)** surface reflectance, and **(d)** emissions (aerosol and  $\text{NO}_2$  surface concentration rescaling). Solid and dashed lines correspond respectively to the case with and without the elevated aerosol layer. For each parameter, the arrow indicates the value corresponding to the reference case from the DABEX simulation.

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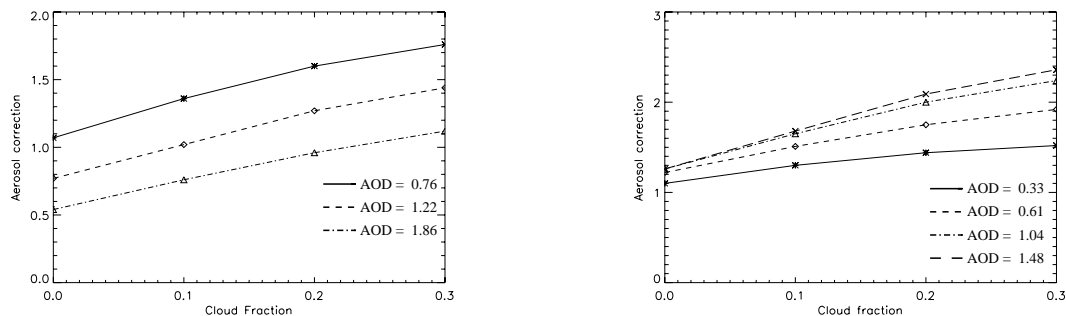
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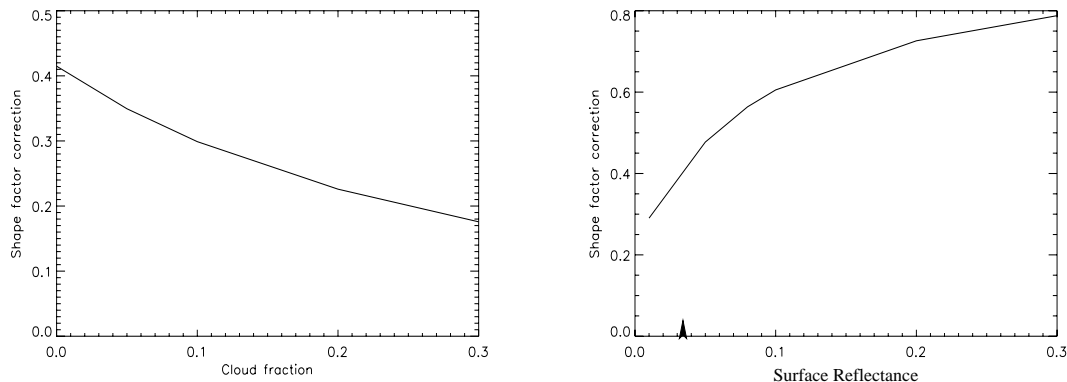


**Fig. 9.** Aerosol correction sensitivity to cloud irradiance fraction for different AODs. The left panel corresponds to the case with an elevated aerosol layer, while the right panel corresponds to a surface-only aerosol layer. Different AODs were obtained by rescaling either the elevated aerosol layer (left panel) or the surface aerosol layer (right panel).

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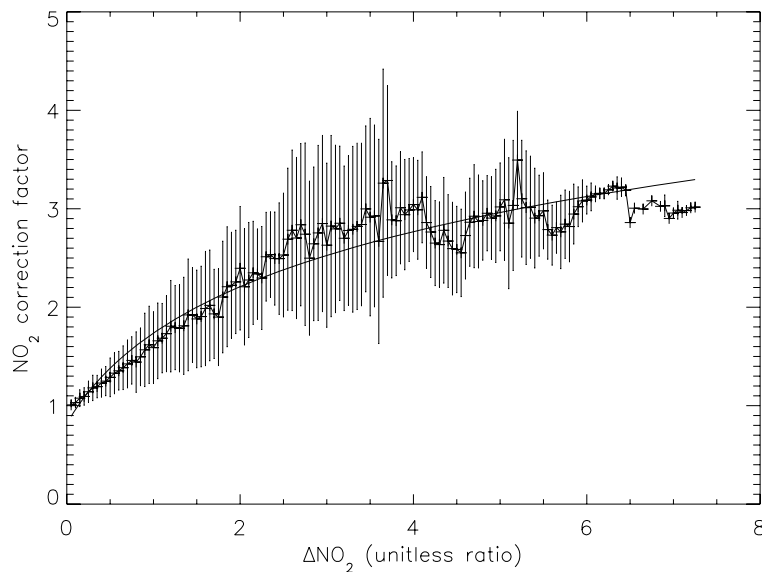


**Fig. 10.** Shape factor correction sensitivity to cloud irradiance fraction (left panel) and surface reflectance (right panel). Arrows indicate values corresponding to the reference case from the DABEX simulation.

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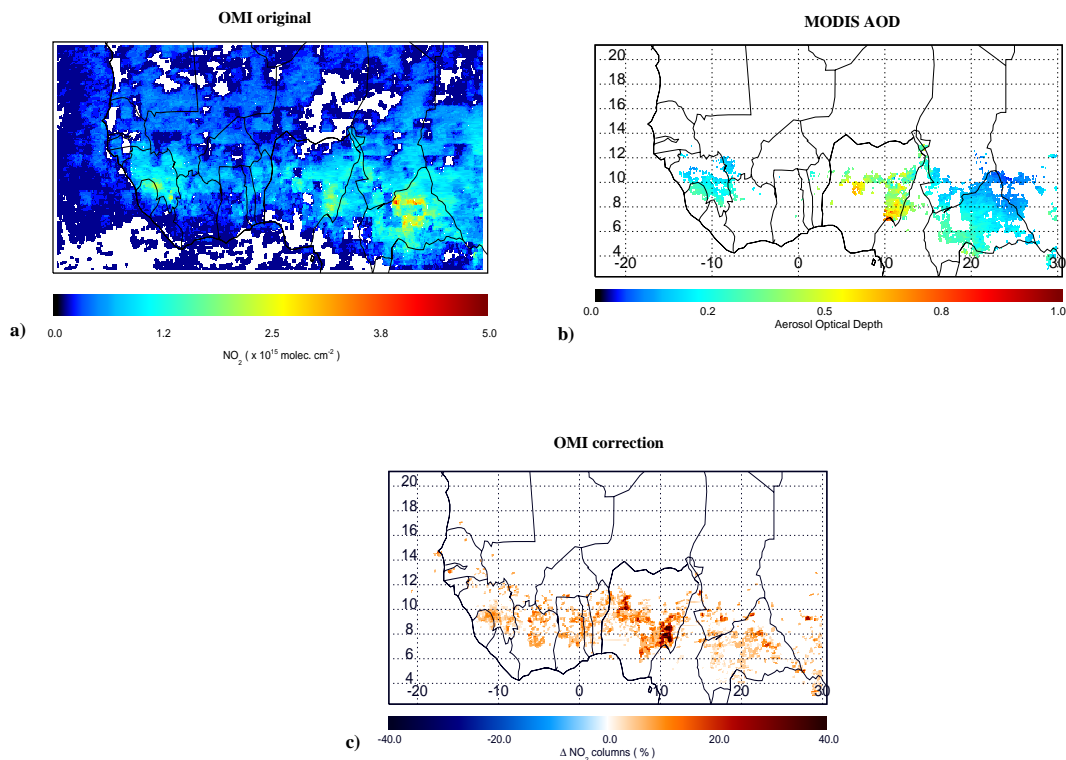


**Fig. 11.** Relationship between the NO<sub>2</sub> correction factor and  $\Delta\text{NO}_2$ . Vertical bars represent the standard deviation of the NO<sub>2</sub> correction factor within each  $\Delta\text{NO}_2$  bin. The solid black line represents the logarithmic fit to the curve.

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**Fig. 12.** Mean OMI DP GC tropospheric  $\text{NO}_2$  column (original product) **(a)**, mean MODIS aerosol optical depth **(b)**, and mean OMI  $\text{NO}_2$  correction factor (%) **(c)** during DABEX. MODIS AODs and  $\text{NO}_2$  correction factors are shown only over areas with OMI DP GC  $\text{NO}_2$  columns  $> 10^{15}$  molec  $\text{cm}^{-2}$ .