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# High concentrations of $N_2O_5$ and $NO_3$ observed in daytime with a TD-CIMS: chemical interference or a real atmospheric phenomenon?

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Dinitrogen pentoxide (N<sub>2</sub>O<sub>5</sub>) and the nitrate radical (NO<sub>3</sub>) play important roles in atmospheric chemistry, yet accurate measurements of their concentrations remain challenging. A thermal dissociation chemical ionization mass spectrometer (TD-CIMS) was deployed to an urban site in Hong Kong to measure the sum of N<sub>2</sub>O<sub>5</sub> and NO<sub>3</sub> in autumn 2010. To our surprise, very high concentrations of  $N_2O_5 + NO_3$  were frequently observed in daytime, with mixing ratios in the range of 200-1000 pptv. To investigate this unusual phenomenon, various interference tests and measurements with different instrument configuration were conducted. It was found that peroxy acetyl nitrate (PAN) contributed to measurable signals at 62 amu, and more importantly, this interference increased significantly with co-existence of NO<sub>2</sub>. Nitric acid (HNO<sub>3</sub>), on the other hand, had little interference to the detection of N<sub>2</sub>O<sub>5</sub>/NO<sub>3</sub> via the NO<sub>3</sub> ion in our TD-CIMS. According to the test results, the interference from PAN and NO<sub>2</sub> could have contributed to 30–50 % of the average daytime (12:00–16:00 LT)  $N_2O_5 + NO_3$  signal at our site. However, evidence exists for the presence of elevated daytime N<sub>2</sub>O<sub>5</sub>, in addition to the daytime signal at 62 amu. This includes: (1) daytime N<sub>2</sub>O<sub>5</sub> measured via the  $I(N_2O_5)^-$  cluster ion with an unheated inlet, which subjects to minimum interferences, and (2) observation of elevated daytime CINO<sub>2</sub> (a product of N<sub>2</sub>O<sub>5</sub> hydrolysis) during a follow-up study. In view of the difficulty in accurately quantifying the contribution from the interferences of PAN and NO<sub>2</sub> and un-tested potential interfering chemicals in the real atmosphere, we caution the use of 62 amu in the TD-CIMS for measuring ambient N<sub>2</sub>O<sub>5</sub> in a high NO<sub>x</sub> environment like Hong Kong. Additional studies are needed to re-examine the daytime issue using other measurement techniques.

#### 1 Introduction

The nitrate radical (NO<sub>3</sub>) and dinitrogen pentoxide (N<sub>2</sub>O<sub>5</sub>) play important roles in the nocturnal tropospheric chemistry. NO<sub>3</sub> is among the most important oxidants in the

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atmosphere, particularly for biogenic hydrocarbons and sulfur-containing compounds (Atkinson, 1991). N<sub>2</sub>O<sub>5</sub> has long been recognized as a key intermediate in the transformation of nitrogen oxides  $(NO_x = NO + NO_2)$  to aerosol nitrates (Riemer et al., 2003; Aldener et al., 2006; Chang et al., 2011; Brown and Stutz, 2012). Recent studies have 5 also demonstrated an important role of N<sub>2</sub>O<sub>5</sub> hydrolysis at night in chlorine activation and the subsequent effect on the next-day's ozone formation (Osthoff et al., 2008; Simon et al., 2009; Thornton et al., 2010).

Due to the low ambient abundances and high reactivity, accurate measurements of atmospheric N<sub>2</sub>O<sub>5</sub> and NO<sub>3</sub> have been challenging. Based on the strong absorption of NO<sub>3</sub> in the visible spectrum at 662 nm, several optical techniques have been developed to measure the ambient NO3, including long-path differential optical absorption spectroscopy (DOAS) (Platt et al., 1980; Atkinson et al., 1986), cavity ring-down spectroscopy (CRDS) (Brown et al., 2001, 2002), laser-induced fluorescence (LIF) (Wood et al., 2003, 2005; Matsumoto et al., 2005), and cavity enhanced absorption spectroscopy (CEAS) (Venables et al., 2006; Langridge et al., 2008). N<sub>2</sub>O<sub>5</sub> is determined by using a heated channel to decompose it into NO<sub>3</sub> or from the calculation according to the fast equilibrium between N<sub>2</sub>O<sub>5</sub> with NO<sub>3</sub> and NO<sub>2</sub>.

Another emerged technique for detecting ambient N<sub>2</sub>O<sub>5</sub> and NO<sub>3</sub> is the chemical ionization mass spectrometry (CIMS) which combines the ion-molecule chemistry with mass spectrometry detection. This technique was originally used in the laboratory to study the heterogeneous uptake kinetics of N2O5 and CINO2 (e.g., Hu and Abbatt, 1997; Thornton et al., 2003; Thornton and Abbatt, 2005), and later on was applied in field measurements (Slusher et al., 2004; Zheng et al., 2008). The fundamental of this method is the reaction of I<sup>-</sup> (the reagent ion) with N<sub>2</sub>O<sub>5</sub> (and/or NO<sub>3</sub>) forming the NO<sub>3</sub> ion that can be detected at 62 amu. Previous laboratory studies suggested that the N<sub>2</sub>O<sub>5</sub>/NO<sub>3</sub> measurement at 62 amu may be subject to interferences from other Ncontaining trace gases such as HNO<sub>3</sub>, HO<sub>2</sub>NO<sub>2</sub>, CIONO<sub>2</sub> and others (see references listed in Table 1). On the other hand, field inter-comparison of a TD-CIMS and a cavity ring-down system showed a high degree of correlation between the two methods

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(Slusher et al., 2004; Huey et al., 2007) indicating the capability of this method in measuring N<sub>2</sub>O<sub>5</sub> in the atmosphere. Although non-negligible and varying background signals at 62 amu have been observed during field studies (Chang et al., 2011), there have been no reports of detailed assessment of these potential interferences under different atmospheric conditions and for different configurations of CIMS. Kercher et al. (2009) developed a method to detect  $N_2O_5$  via the  $I(N_2O_5)^-$  cluster ion at 235 amu with an unheated inlet to address the interference at 62 amu in their CIMS, however this method has its own limitation including a lower sensitivity and larger impact of water vapor than that at 62 amu.

In autumn 2010, a TD-CIMS (thermal dissociation-CIMS), which is the same type used by Slusher et al. (2004) and Huey et al. (2007), was deployed to an urban site in Hong Kong which is characterized by large quantities of NO<sub>v</sub>, ozone and particulate matters. Unexpectedly, concentration peaks of N<sub>2</sub>O<sub>5</sub> + NO<sub>3</sub> were frequently observed in daytime in our study. To investigate this unusual observation, we have conducted a series of laboratory and field tests, including testing interferences individually and in combination from PAN, NO<sub>2</sub>, O<sub>3</sub>, HNO<sub>3</sub>, detection of N<sub>2</sub>O<sub>5</sub> by using a cold inlet via the detection of I(N<sub>2</sub>O<sub>5</sub>)<sup>-</sup> cluster ion at 235 amu, and examination of daytime CINO<sub>2</sub> which is a product of N<sub>2</sub>O<sub>5</sub>. In this paper, we will first present the ambient observations, and then the detailed test results on interferences and other evidence for daytime N<sub>2</sub>O<sub>5</sub>. A surprising result from the tests is that PAN + NO<sub>2</sub> can have large interference to the TD-CIMS at 62 amu, which has not been reported in previous studies. Despite large chemical interference, we show that the observed daytime N<sub>2</sub>O<sub>5</sub> signal may be in part due to real contribution from NO<sub>3</sub> and N<sub>2</sub>O<sub>5</sub>.

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#### 2.1 Measurement site

The field study was conducted in the urban center of Hong Kong (22°18′ N, 114°11′ E, ~ 15 ma.s.l.). The sampling site was located on the rooftop of a seven-story teaching building (about 20 m above the ground) on the campus of the Hong Kong Polytechnic University (HKPU). To the southeast of the campus, there is a cross-harbor tunnel in the north-south direction with a large flow of vehicles, especially during rush hours (see Fig. 1). Victoria Harbor is located about 1 km south of the measurement site. Thus emissions from vehicles and marine vessels are the most important local anthropogenic sources. The sampling site is surrounded by business districts, tourism and residential areas, with no large industrial sources nearby. The field measurements were carried out from 15 October to 4 December 2010, which is the season with the most severe photochemical pollution in Hong Kong (Wang et al., 2009).

#### 2.2 The CIMS apparatus

A TD-CIMS was deployed to measure the sum of N<sub>2</sub>O<sub>5</sub> and NO<sub>3</sub> in this study. The system was developed at the *Georgia Institute of Technology* and is based on a soft and selective ionization process resulting from the reaction between a reagent ion and the target compounds, with the generated ions detected by a mass spectrometer. In the present study, the measurement method and operating parameters of TD-CIMS were the same as those described by Slusher et al. (2004) which was configured to simultaneously measure ambient N<sub>2</sub>O<sub>5</sub> and PANs. The schematic diagram of our TD-CIMS is shown in Fig. 2. I<sup>-</sup>, which was produced from passing a flow of 2 sccm of 0.3 % CH<sub>3</sub>I/N<sub>2</sub> through an alpha ion source (Po-210), served as the reagent ion. Ambient air samples were drawn through a PFA-Teflon tube (I.D., 9.5 mm; O.D., 12.7 mm; length, 2 m) to the CIMS at a flow rate of 1.5 standard liters per minute (SLPM). To reduce the residence time of air samples in the sampling tube, an extra bypass flow of 7.7 SLPM

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was directed to the exhaust. Immediately before the flow tube, the last 14.8 cm of inlet was heated to 180 °C. The temperature of the air on the axis of the heated inlet was estimated at ~117°C (Slusher et al., 2004), under which condition more than 99% of the N<sub>2</sub>O<sub>5</sub> would decompose into NO<sub>3</sub> with an ambient NO<sub>2</sub> level of 40 ppbv. The produced and the original NO<sub>3</sub> then reacted with I to produce NO<sub>3</sub> ions in the flow tube, which were subsequently detected and quantified by a quadruple mass spectrometer at 62 amu with a time resolution of 6 s. The NO<sub>3</sub> signal is proportional to the concentration of total NO<sub>3</sub>, the I<sup>-</sup> signal, the effective reaction rate constant between them, and the effective reaction time (Huey, 2007). Figure 3 depicts the mass spectrums of ambient air in urban Hong Kong obtained both at day and at night, which clearly shows the I<sup>-</sup> signal at 127 amu, NO<sub>3</sub> (N<sub>2</sub>O<sub>5</sub> + NO<sub>3</sub>) signal at 62 amu, and CH<sub>3</sub>C(O)O<sup>-</sup> (PAN) signal at 59 amu, etc.

In the present study, the TD-CIMS instrument was calibrated once a week using the on-line N<sub>2</sub>O<sub>5</sub> synthesis method (Bertram et al., 2009). The calibration source was generated from the reactions of NO<sub>2</sub> with O<sub>3</sub> and subsequently NO<sub>3</sub> with NO<sub>2</sub>. The concentrations of the prepared N<sub>2</sub>O<sub>5</sub> were determined via the change in NO<sub>2</sub> concentrations after adding ozone, and conversely verified by the change in ozone after adding NO2. Zero air that was free from moisture and aerosols served as the diluent so as to prevent the hydrolysis of N<sub>2</sub>O<sub>5</sub> during the calibrations (see Fig. 2). NO<sub>2</sub> was monitored with a chemiluminescence analyzer equipped with a photolytic converter, which ensured measurements of the true NO<sub>2</sub> (Xu et al., 2012). Based on the relative standard deviation of the sample signal, the precision of our TD-CIMS was 3% for 1000 pptv N<sub>2</sub>O<sub>5</sub>. The sensitivity of  $N_2O_5$  during the campaign was  $2.8 \pm 0.2$  (mean  $\pm$  SD) Hz pptv<sup>-1</sup>. The instrument background was automatically measured for 2 min once an hour by adding a small flow (5 mL min<sup>-1</sup>) of NO (1000 ppm) to the sample flow (diluted to 9.2 SLPM), titrating NO<sub>3</sub> and thus N<sub>2</sub>O<sub>5</sub>. The background signal of the NO<sub>3</sub> ion during this field study was  $71.7 \pm 36.0$  (mean  $\pm$  SD) Hz. Figure 4 shows the raw signals of  $N_2O_5 + NO_3$ (NO<sub>3</sub><sup>-</sup>, at 62 amu) with hourly automatic background determination on 30 November, when the maximum concentration of N<sub>2</sub>O<sub>5</sub> + NO<sub>3</sub> was observed. Clearly, the signals

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exhibited relatively low background compared to ambient N<sub>2</sub>O<sub>5</sub> + NO<sub>3</sub> during both day and night. According to three times the standard deviation of the background signal, the typical detection limit of  $N_2O_5$  for 6s average time was estimated to be 13 pptv for our TD-CIMS.

#### Other instruments

In addition to the TD-CIMS, a large number of other instruments were deployed concurrently. Here, we briefly describe those that were used to aid the presentation of the N<sub>2</sub>O<sub>5</sub> and NO<sub>3</sub> data. O<sub>3</sub> was measured by a commercial UV photometric analyzer (Model 49i, Thermo Environmental Instruments (TEI), USA). NO and NO2 were analyzed with a chemiluminescence instrument (Model 42i, TEI) equipped with a photolytic NO<sub>2</sub>-converter (Air Quality Design, USA) (Xu et al., 2012). Solar radiation was measured using a LI-200 Pyranometer Sensor (LI-COR, USA). The ambient RH and temperature were monitored with a RH/temperature probe (Model 41382VC/VF, M.R. YOUNG, USA). During the field measurements, the minute-average data of trace gases and meteorological parameters were collected in real time by a data logger (Model 8816, Environmental Systems Corporation, USA).

#### Observation results

The time series of hourly mixing ratios of N<sub>2</sub>O<sub>5</sub> + NO<sub>3</sub> measured in urban Hong Kong from 15 October to 4 December 2010 is shown in Fig. 3. Similar to the measurement results obtained in other locations, several night-time concentration peaks of N<sub>2</sub>O<sub>5</sub> + NO<sub>3</sub> were noticed in our study (e.g., 31 October, 1 and 10 November). However, very high mixing ratios of  $N_2O_5 + NO_3$  were frequently observed during the daytime (e.g., 23 and 24 October, 19, 22 and 30 November, and 2 December). During the 50 day measurement period (excluding 11 November due to a lack of data), the average hourly concentration of  $N_2O_5 + NO_3$  was 86.9 (± 85.6) pptv with the maximum value of 1033 pptv recorded at 15:00 LT on 30 November. The mean daytime mixing ratio of  $N_2O_5 + NO_3$  was 102.5 pptv (06:00–17:59 LT), even higher than the value of 71.3 pptv at nighttime (18:00–05:59 LT).

To learn more about the atmospheric conditions associated with the elevated day-time  $N_2O_5$  and  $NO_3$  signals at 62 amu, we present six cases during which the hourly  $N_2O_5+NO_3$  values exceeded 400 pptv. Figure 4 presents the 5 min data of  $N_2O_5+NO_3$ ,  $O_3$ ,  $O_x$  ( $O_3+NO_2$ ),  $O_3$ ,  $O_4$  ( $O_3+NO_2$ ),  $O_4$ ,  $O_5$ , O

#### 4 Chemical interferences

A major drawback of the TD-CIMS technique is that there could be potential interferences to the selected  $NO_3^-$  ion that can give rise to significant background noise at 62 amu (Chang et al., 2011). Previous laboratory studies have proposed many possible ion-molecule reactions yielding the  $NO_3^-$  ion, as summarized in Table 1. As stated earlier, the detailed results of these interferences in the real atmospheres have not been reported to date. To examine the influences of these potential interferences to our  $N_2O_5 + NO_3$  measurements, a series of tests were conducted in various atmospheric conditions and in laboratory during and after the measurement campaign.

The background determination by adding excess NO provided a first examination of the interferences. Such determined background reflects the interferences from

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compounds that are unreactive to NO, including most potential ones other than PAN and  $HO_2NO_2$ . As stated above, the background was periodically measured once an hour throughout the campaign. Considerable amount of background signals at 62 amu in our CIMS was indeed seen during the present study with a mean counts ( $\pm$ SD) of 71.7 ( $\pm$ 36.0) Hz, and they also exhibited a diurnal pattern with higher values in the late afternoon (see Fig. 5). This indicated that the  $N_2O_5 + NO_3$  measurements via the 62 amu channel in our TD-CIMS were subject to some interference during the present study. However, the background signals were much lower than the ambient signals. For the six cases with daytime  $N_2O_5 + NO_3$  concentrations exceeding 400 pptv, the instrument background only accounted for on average 10 % of the ambient values. This background signal due to most of the gases shown in Table 1 and has been accounted for (i.e., subtracted from the total signals at 62 amu) in our final data. The interference from PAN and  $HO_2NO_2$  could not be determined because they also reacted with NO during zeroing. Their interferences are addressed below.

#### 4.1 Interference of PAN

PAN is the most possible compound to interfere the TD-CIMS measurements in this study, not only due to its relatively high ambient abundances but also because it can escape from background determination by adding NO. In the previous studies, higher  $N_2O_5$  signals than the steady-state predictions were usually observed with elevated mixing ratios of PAN (Brown et al., 2005; Osthoff et al., 2006). The interference from PAN to the  $NO_3^-$  signals in our TD-CIMS was evaluated post the field campaign by adding synthetic PAN to zero air and in the ambient air samples. The PAN was generated from a PAN calibrator (*Meteorologie Consult GmbH*), which is based on the reaction sequence of NO and acetone in ultrapure air with a Penray lamp. The concentrations of PAN that were added to the CIMS were quantified simultaneously by a  $NO_y$  analyzer (*TEI* 42CY). We also generated PAN using a conventional wet chemistry method by reacting peracetic acid with HNO3 (Gaffney et al., 1984). The test

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results from the two PAN sources are consistent, thus this study only shows the result which used the PAN generated from the photolytic source.

For tests of PAN in zero air and in relatively clean ambient air at a coastal site (Hok Tsui), measurable interference from PAN was observed at 62 amu. The result in zero air indicates  $17 \sim 25$  ( $22.8 \pm 4.1$ ) pptv from  $NO_3 + N_2O_5$  for every ppbv of PAN (see intercepts in Fig. 7). Adding PAN to ambient air at the coastal site yielded similar results. Table 2 shows the results from four tests conducted with varying ambient pollution levels and meteorological conditions. Introducing 5.6-6.3 ppb of PAN to the ambient air resulted in an increase in the  $NO_3^-$  signal of 42-75 pptv of  $NO_3 + N_2O_5$  in our TD-CIMS, again indicating interference from PAN to the field  $NO_3/N_2O_5$  measurements via the  $NO_3^-$  by TD-CIMS. The mechanism how PAN interferes the  $NO_3^-$  detection is unclear. Some researchers proposed that the  $CH_3C(O)O^-$  ion (produced from the reaction of  $I^-$  with  $CH_3C(O)O_2$  — the product of PAN thermo-dissociation) likely reacts with HNO $_3$  to produce  $NO_3^-$  (Veres et al., 2008; Roberts et al., 2010). This appeared not to be the case in our CIMS as additions of both PAN and HNO $_3$  didn't lead to any increase at 62 amu compared to the addition of PAN alone (see Fig. 6).

A surprising finding was much large interferences at  $62\,\mathrm{amu}$  when the same spike tests were conducted at the PolyU site, which has very high  $\mathrm{NO_x}$  concentrations, and the interference appeared to increase with ambient  $\mathrm{NO_2}$  suggesting that reaction between PAN and  $\mathrm{NO_2}$  leads to significant interference at  $62\,\mathrm{amu}$ . To confirm this, a series of tests were conducted. Figure 7 shows the signal at  $62\,\mathrm{amu}$  as a function of PAN and  $\mathrm{NO_2}$  concentrations in zero, which clearly shows that the interference increases with both PAN and  $\mathrm{NO_2}$ . For example, at 5 ppbv of PAN, adding  $60\,\mathrm{ppbv}$  of  $\mathrm{NO_2}$  produces  $400\,\mathrm{ppt}$  equivalent  $\mathrm{NO_3}$  signal, compared to 150 pptv without  $\mathrm{NO_2}$ , indicating an amplifying effect of  $\mathrm{NO_2}$  on the previously reported PAN interference. The exact chemical reaction that leads to the interference is not clear. We believe that interference of  $\mathrm{PAN} + \mathrm{NO_2}$  is related to the thermal dissociation of PAN followed by radical reactions with  $\mathrm{NO_2}$  in the heated inlet. The radical reactions might serve as a source of  $\mathrm{NO_3}$  and  $\mathrm{N_2O_5}$ , or the products could react with  $\mathrm{I}^-$  to produce  $\mathrm{NO_3}^-$  ions which is detected at

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62 amu. We will evaluate the possible contribution of PAN + NO<sub>2</sub> to the observed high

#### 4.2 Interference of HNO<sub>3</sub>

daytime  $N_2O_5 + NO_3$  for the 2010 autumn study later.

Another possible source of interference to the NO<sub>3</sub> ion is the HNO<sub>3</sub> since it has a NO<sub>3</sub> group. Several studies have proposed the production of NO<sub>3</sub> from ion-molecule reactions involving HNO3, with somewhat inconsistent results. Huey and co-workers showed that the reaction of I with HNO<sub>3</sub> (producing NO<sub>3</sub>) was rather inefficient. Thornton and coworkers on the other hand found a significant background signal (10-50 Hz) at the NO<sub>3</sub> mass in a cold iodide CIMS under conditions of long ion-molecule reaction time, and attributed this to the reaction of HNO<sub>3</sub> with I<sup>-</sup> (Thornton et al., 2003; Thornton and Abbatt, 2005). Roberts and coworkers reported that the HNO<sub>3</sub> was sensitive to the acetate ions and react to produce NO<sub>3</sub> ions at the 62 amu (Veres et al., 2008; Roberts et al., 2010).

The relatively low background signals by adding NO in our study suggest insignificant interference from HNO<sub>3</sub> to our CIMS, which has been corrected during data reduction. This was further confirmed by the addition of HNO<sub>3</sub>. The test was carried out three times for varying ambient conditions to check the repeatability of the results, which are listed in Table 2 with an example being shown in Fig. 6. It can be seen that after introducing a trace amount of HNO<sub>3</sub> to the inlet tube of our TD-CIMS, there was no significant increase in the NO<sub>3</sub> signal compared to the ambient air. Additionally, adding HNO<sub>3</sub> to a trace level of PAN which resulted in an acetate ion signal of  $\sim 0.9 \times 10^4$  Hz, the NO<sub>3</sub> signal showed no apparent change compared to that for only adding PAN. These results suggest that the HNO3 itself and its mixture with acetate ions have no significant interference to the detection of  $N_2O_5 + NO_3$  via the  $NO_3^-$  ion in our TD-CIMS. It should be noted that the result of non-reactivity of HNO<sub>3</sub> to the acetate ion in our TD-CIMS is different from that obtained by Roberts and coworkers using a NI-PT-CIMS

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(negative-ion proton-transfer CIMS) (Veres et al., 2008; Roberts et al., 2010), possibly due to different configurations and operation conditions.

#### 4.3 Interference of other possible compounds

CIONO<sub>2</sub>, BrONO<sub>2</sub> and HO<sub>2</sub>NO<sub>2</sub> also react efficiently with I<sup>-</sup> to produce NO<sub>3</sub><sup>-</sup> (Huey et al., 1995; Hanson et al., 1996; Zhang et al., 1997; Amelynck et al., 2001). Again, the relatively low instrument background indicated small interference from CIONO<sub>2</sub> and BrONO<sub>2</sub> and any interference from them would have been corrected in our measurements because they cannot be removed by adding NO. The zeroing would not work for HO<sub>2</sub>NO<sub>2</sub>. However, it is impossible for HO<sub>2</sub>NO<sub>2</sub> to pass through the heated inlet tube in our TD-CIMS considering its thermally unstable nature.

In addition, the  $NO_3^-$  ion may also come from the ion-molecule reactions involving  $CI^-$  (35 amu),  $C_2H_2N^-$  (40 amu),  $NO_2^-$  (46 amu),  $O_3^-$  (48 amu),  $CO_3^-$  (60 amu), and  $CO_4^-$  (76 amu) (listed in Table 1). These reagent ions were observed in quite low levels (i.e., < 20 Hz for 40 and 48 amu, and 10–500 Hz for 35, 46, 60 and 76 amu) in our TD-CIMS during the field measurements (see Fig. 3). Therefore, the ion-molecule reactions induced by  $CI^-$ ,  $C_2H_2N^-$ ,  $NO_2^-$ ,  $O_3^-$ ,  $CO_3^-$  and  $CO_4^-$  are also believed to have no significant influence to the  $NO_3^-$  detection based on the relationship between the product ion and reactants.

In summary, according to the above tests and discussions, reactions between PAN and  $NO_2$  in the heated inlet are found to have significant interference to signal at 62 amu in our TD-CIMS, while  $HNO_3$  and other compounds are not believed to have contributed to signal at 62 amu, although our tests are not exhaustive for including all other chemicals.

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As indicated above, laboratory and field tests revealed significant interference from PAN and NO2 to 62 amu in our TD-CIMS. The contribution from this interference to the observed daytime N<sub>2</sub>O<sub>5</sub> + NO<sub>3</sub> in this study was examined and corrected based the tests results in Fig. 7. During the field campaign at PolyU, PAN was concurrently measured by the same TD-CIMS and a GC-ECD instrument (gas chromatography with an electron capture detector). For the six episodes shown in Fig. 4, the interference would contribute to 41.3-67.0%, 29.0-39.8%, 32.2-73.2%, 25.6-49.2%, 19.4-43.4%, 20.4-77.8%, respectively, to the daytime signals in the 6 episodes. (Two examples on 23 October and 30 November were shown in Fig. 8.) The remaining daytime signal may be a real contribution from N<sub>2</sub>O<sub>5</sub> and NO<sub>3</sub>, but interferences from other untested chemicals cannot be ruled out. As interferences at the 62 ion channel are large and variable, it is difficult to determine the fraction of real signal from  $N_2O_5 + NO_3$  in our study at 62 amu.

#### Other evidence of daytime $NO_3 + N_2O_5$ in Hong Kong

To check the measurement results by at 62 amu, we attempted to measure ambient N<sub>2</sub>O<sub>5</sub> with a cold version of CIMS immediately after the present campaign at the same site. The CIMS was configured similar to that described by Kercher et al. (2009). When an unheated inlet tube is used, the reactions of N<sub>2</sub>O<sub>5</sub> with I<sup>-</sup> produce both the NO<sub>3</sub> and  $I(N_2O_5)^-$  cluster ions. The  $I(N_2O_5)^-$  ion (at 235 amu) is thought to be free from the chemical interferences that can perturb the NO<sub>3</sub> ion, and thus provides a better measure of N<sub>2</sub>O<sub>5</sub> (Kercher et al., 2009). In our study, the sensitivity of N<sub>2</sub>O<sub>5</sub> from the  $I(N_2O_5)^-$  ion was determined at  $0.55 \pm 0.003 \, \text{Hzpptv}^{-1}$ , which is smaller than the value of 0.93 Hzpptv<sup>-1</sup> obtained by Kercher et al. (2009) possibly due to a smaller sample flow rate and/or stronger electric field in the collisional dissociation chamber in our CIMS. It's also noteworthy that the  $N_2O_5$  data from  $I(N_2O_5)^-$  in our study were

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not re-scaled according to the  $I(H_2O)^-/I^-$  ratio as described by Kercher et al. (2009), because the  $I^-$  ion at 127 amu was not detected in real-time in our study. The rescaled  $N_2O_5$  concentrations during the daytime would be even higher, because the daytime  $I(H_2O)^-$  signals were generally lower than (by a factor of 0.4–1.0) those during the calibrations. Despite the above uncertainty in determining the absolute value, the general variation pattern of  $N_2O_5$  should be trustworthy. Consistent with the setup with the heated inlet and the detection at 62 amu, signal at 235 amu with the cold CIMS also showed a daytime peak. Figure 9 gives an example of the ambient results taken on 20 December 2010.

Another independent piece of evidence for the daytime N<sub>2</sub>O<sub>5</sub> + NO<sub>3</sub> in Hong Kong is concurrent increase in the mixing ratios of CINO<sub>2</sub> (a product of N<sub>2</sub>O<sub>5</sub> hydrolysis) observed in a follow-up study in western Hong Kong (Tung Chung; see SI for the experiment information). At this site, elevated N<sub>2</sub>O<sub>5</sub> + NO<sub>3</sub> concentrations were also found at daytime during photochemical episodes, with the CINO2 signals (208 amu) showing concurrent increases. Figure 10 shows an example for 28 August 2011. On that day, N<sub>2</sub>O<sub>5</sub> + NO<sub>3</sub> from 62 amu exhibited an afternoon peak of 670 pptv (5 min average, at 14:00 LT), and CINO<sub>2</sub> had a concurrent enhancement to 120 pptv. To estimate the levels of N<sub>2</sub>O<sub>5</sub> that would be needed to sustain such amount of CINO<sub>2</sub>, we assumed a photostationary steady state for CINO2 in the afternoon with an uptake coefficient of 0.03 for N<sub>2</sub>O<sub>5</sub> hydrolysis on aerosol surfaces and a CINO<sub>2</sub> yield of 10 %. The photolysis rate of CINO<sub>2</sub> was estimated as  $7.8 \times 10^{-4} \, \text{s}^{-1}$  using the method by Simon et al. (2009), and the aerosol surface area was 979 mm<sup>2</sup> m<sup>-3</sup> from concurrent measurements of aerosol size distribution. The calculation showed that to produce the observed 120 pptv of CINO<sub>2</sub>, at least 518 pptv of N<sub>2</sub>O<sub>5</sub> would be required. This result provides additional corroboration of the possible elevated daytime concentrations of  $N_2O_5+NO_3$  in Hong Kong.

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A TD-CIMS technique, which has been previously applied to field measurements in the US, was deployed to measure ambient NO<sub>3</sub> and N<sub>2</sub>O<sub>5</sub> in urban Hong Kong in a photochemical season. Surprisingly, concentration peaks of NO<sub>3</sub> + N<sub>2</sub>O<sub>5</sub> were frequently observed in daytime at 62 amu channel in the TD-CIMS, which is in contrast to our current understanding of reactive nitrogen chemistry. Our subsequent laboratory and field tests provide new insights into chemical interferences in the TD-CIMS. In particular, we have discovered that reaction between NO<sub>2</sub> and PAN can amplify the interference from PAN at 62 amu. This interference could have contributed 30-50 % to the average daytime NO<sub>3</sub> and N<sub>2</sub>O<sub>5</sub> at our site. Despite the large interference, evidence exists to suggest that the elevated N<sub>2</sub>O<sub>5</sub> in daytime may be in part due to real contribution from NO<sub>3</sub> or N<sub>2</sub>O<sub>5</sub>. In viewing of the large and variable interferences at 62 amu and difficulty in correcting them, we conclude that it is not suitable to use the TD-CIMS to measure NO<sub>3</sub> or N<sub>2</sub>O<sub>5</sub> at 62 amu in a high NO<sub>y</sub> environment like the present study site. Adoption of either 235 amu with a cold inlet or an optics-based technique is recommended. We also suggest more studies to examine the abundance of daytime NO<sub>3</sub> and N<sub>2</sub>O<sub>5</sub> in similar urban areas with co-existence of high ozone and NO<sub>2</sub>.

Supplementary material related to this article is available online at http://www.atmos-meas-tech-discuss.net/6/7473/2013/amtd-6-7473-2013-supplement.pdf.

Acknowledgements. We thank Steven Poon and K.S. Lam for their help in the field study and setting up the instruments. This research was supported by the Environment and Conservation Fund of Hong Kong (Project No. 2009-07), Niche Area Development Scheme of the Hong Kong Polytechnic University (1-BB94), and the Research Grants Council of Hong Kong (PolyU 5015/12P).

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**Table 1.** Ion-molecule reactions with product of  $NO_3^-$  and the rate constants.

Reaction	Rate constant (molecule <sup>-1</sup> cm <sup>3</sup> s <sup>-1</sup> )	Comments	Reference		
$I^- + HNO_3 \rightarrow NO_3^- + HI$	< 5 × 10 <sup>-11</sup>		Fehsenfeld et al. (1975)		
•	No reaction		Huey et al. (1995)		
$I^- + CIONO_2 \rightarrow NO_3^- + ICI$	$9 \times 10^{-10}$	±40 %	Huey et al. (1995)		
$I^- + BrONO_2^- \rightarrow NO_3^- + IBr$	Unknown		Hanson et al. (1996)		
$I^- + HO_2NO_2 \rightarrow NO_3^- + HIO$	Unknown		Zhang et al. (1997)		
$CO_3^- + HNO_3 \rightarrow NO_3^- + products$	$8 \times 10^{-10}$		Fehsenfeld et al. (1975)		
$CO_3^- + NO_2 \rightarrow NO_3^- + CO_2$	$2 \times 10^{-10}$	±50%	Ikezoe and Viggiano (1987)		
$CO_4^- + NO \rightarrow NO_3^- + CO_2$	$4.8 \times 10^{-11}$	±30 %	Ikezoe and Viggiano (1987)		
$C_2H_2N^- + HNO_3 \rightarrow NO_3^- + CH_3CN$	$1.4 \times 10^{-9}$	297 K	Ikezoe and Viggiano (1987)		
$C_2H_3O_2^- + HNO_3 \rightarrow NO_3^- + C_2H_4O_2$	Unknown		Roberts et al. (2010)		
$Cl^- + HNO_3 \rightarrow NO_3^- + HCl$	$1.6 \times 10^{-9}$	±40 %	Ikezoe and Viggiano (1987)		
$NO_2^- + HNO_3 \rightarrow NO_3^- + HONO$	$1.6 \times 10^{-9}$	±40 %	Ikezoe and Viggiano (1987)		
$NO_2^- + NO_2 \rightarrow NO_3^- + NO$	$< 2 \times 10^{-13}$		Ikezoe and Viggiano (1987)		
$NO_{2}^{-} + N_{2}O \rightarrow NO_{3}^{-} + N_{2}$	$< 1 \times 10^{-12}$		Ikezoe and Viggiano (1987)		
$NO_{2}^{-} + O_{3} \rightarrow NO_{3}^{-} + O_{2}^{-}$	$1.2 \times 10^{-10}$	±40 %	Ikezoe and Viggiano (1987)		
$O_3^- + NO_2^- \to NO_3^- + O_2^-$	$2.8 \times 10^{-10}$	$\pm$ 30 %, 280 K	Ikezoe and Viggiano (1987)		

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**Table 2.** Summary of the interference tests for the TD-CIMS.

Test	t Ambient conditions			Standard addition*		CIMS r	CIMS response	
	NO <sub>y</sub> (ppb)	O <sub>3</sub>	<i>T</i> (°)	RH (%)	ΔPAN (ppb)	$\Delta$ HNO <sub>3</sub> (ppb)	$\Delta N_2 O_5$ (ppt)	$\Delta S_{62}/\Delta S_{59}$ (cps/cps)
1	3.5	69	25	83	6.3	0	75	0.014
2	4.8	71	25	87	5.8	0	42	0.009
3	4.5	60	23	83	5.9	0	69	0.014
4	14.0	35	22	69	5.6	0	62	0.013
5	7.2	70	24	81	0	1.4	1	n.a.
6	6.2	76	24	64	0	9.1	1	n.a.
7	15.2	38	22	69	0	8.9	2	n.a.
8	12.8	41	21	68	5.3	8.9	62	n.a.

<sup>\*</sup> The concentrations of PAN and HNO<sub>3</sub> that were added to the ambient air were determined by a NO<sub>y</sub> analyzer.

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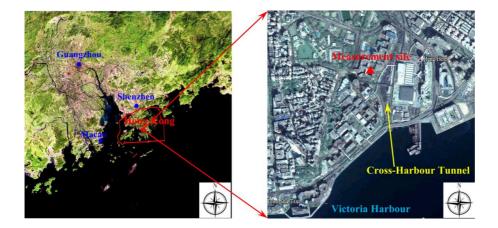


Fig. 1. Location of the measurement site and surrounding layout.

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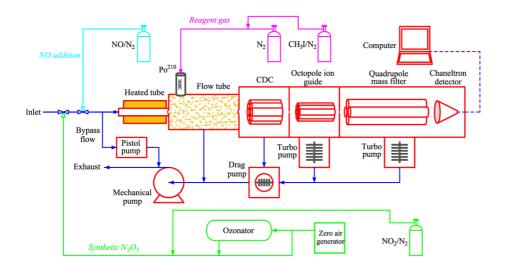


Fig. 2. Schematic diagram of the TD-CIMS deployed in this study.

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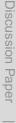
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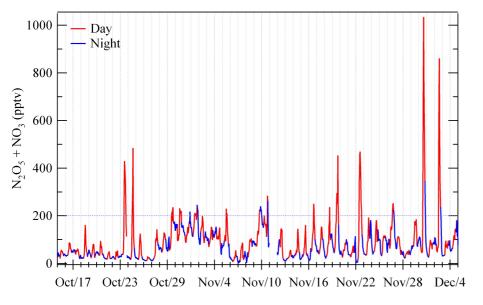
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**Fig. 3.** Time series of hourly average concentrations of  $N_2O_5 + NO_3$  measured in Hong Kong.



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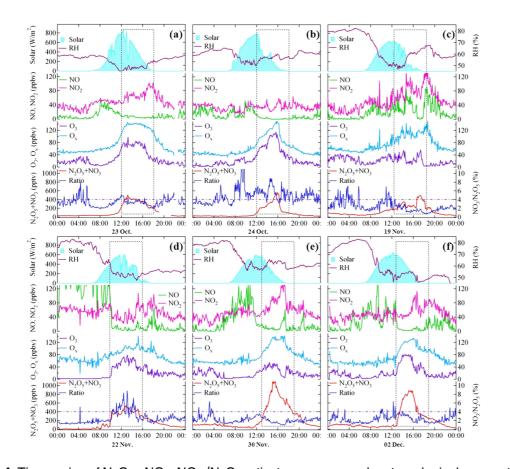


Fig. 4. Time series of N<sub>2</sub>O<sub>5</sub>+NO<sub>3</sub>, NO<sub>3</sub>/N<sub>2</sub>O<sub>5</sub> ratio, trace gases and meteorological parameters for six pollution episodes on (a) 23 October, (b) 24 October, (c) 19 November, (d) 22 November, (e) 30 November, and (f) 2 December 2010. The pollution episodes with daytime concentration peaks of  $N_2O_5$  and  $NO_3$  are shown in the panes.

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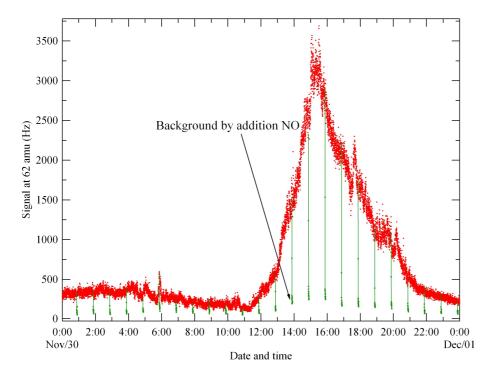
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**Fig. 5.** Raw  $NO_3^-$  (62 amu) signal with hourly automatic background detection by adding excess NO, 30 November 2010.

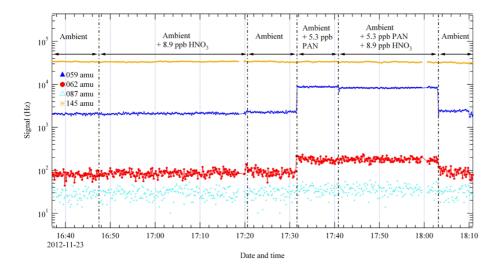
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**Fig. 6.** Changes of raw signals of  $CH_3C(O)O$ -,  $NO_3^-$ ,  $C_3H_7C(O)O$ -,  $I(H_2O)^-$  detected by  $I^-$  TD-CIMS when adding trace amounts of PAN, HNO<sub>3</sub>, or their mix to the ambient air.

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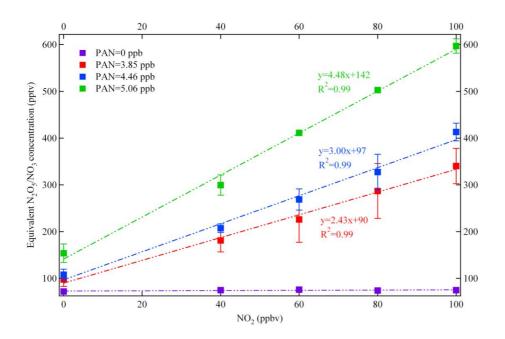


Fig. 7. Interference of  $NO_2$  to equivalent  $NO_3^-$  concentration under different levels of PAN. Vertical bars are standard deviations of tests results.

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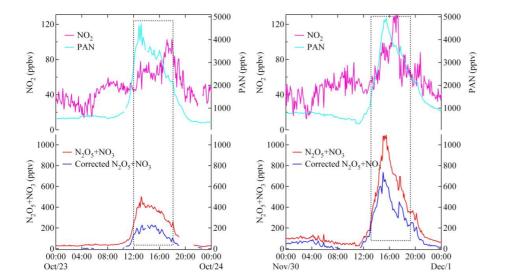
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**Fig. 8.** Time series of corrected and uncorrected  $N_2O_5 + NO_3$ , PAN,  $NO_2$  and NO on 23 October and 30 November 2010.

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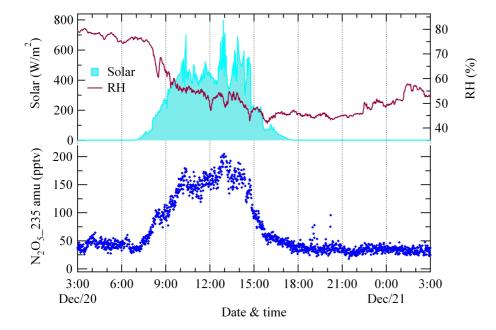
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**Fig. 9.** Time series of of  $N_2O_5$  from the  $I(N_2O_5)^-$  ion, solar radiation and relative humidity in urban Hong Kong on 20 December 2010.

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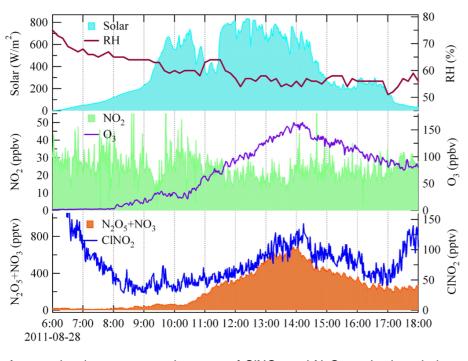


Fig. 10. A case showing concurrent increase of CINO<sub>2</sub> and N<sub>2</sub>O<sub>5</sub> at daytime during a photochemical episode at a suburban site (Tung Chung) in western Hong Kong.