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A newly identified calculation discrepancy of the Sunset semi-continuous carbon analyzer

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Sunset Semi-Continuous Carbon Analyzer (SCCA) is an instrument widely used for carbonaceous aerosol measurement. Despite previous validation work, here we identified a new type of SCCA calculation discrepancy caused by the default multi-point baseline correction method. When exceeding a certain threshold carbon load, multipoint correction could cause significant Total Carbon (TC) underestimation. This calculation discrepancy was characterized for both sucrose and ambient samples with three temperature protocols. For ambient samples, 22 %, 36 % and 12 % TC was underestimated by the three protocols, respectively, with corresponding threshold being ~ 0 . 20 and 25 µgC. For sucrose, however, such discrepancy was observed with only one of these protocols, indicating the need of more refractory SCCA calibration substance. The discrepancy was less significant for the NIOSH (National Institute for Occupational Safety and Health)-like protocol compared with the other two protocols based on IMPROVE (Interagency Monitoring of PROtected Visual Environments). Although the calculation discrepancy could be largely reduced by the single-point baseline correction method, the instrumental blanks of single-point method were higher. Proposed correction method was to use multi-point corrected data when below the determined threshold, while use single-point results when beyond that threshold. The effectiveness of this correction method was supported by correlation with optical data.

1 Introduction

Carbonaceous aerosol, usually fractionated into elemental carbon (EC) and organic carbon (OC), is of increasing interest due to its important roles in human health and environmental effects. The most widely used method for aerosol OC/EC measurement is thermal-optical analysis, which classifies OC and EC based on their differences in both thermal and optical characteristics. The IMPROVE (Interagency Monitoring of PROtected Visual Environments) (Chow et al., 1993) and NIOSH (National Institute

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for Occupational Safety and Health) (Birch and Cary, 1996) protocols, as well as their variants (Watson et al., 2005; Chow et al., 2007b), are typically adopted in modern thermal-optical analysis. Compared with low time resolution results (usually 24 h averaged), hourly OC/EC data is more capable of investigating their short-term patterns 5 due to the variations of emission, transportation, gas-particle partitioning, chemical reaction, etc. Therefore, several (semi-)continuous instruments have been developed for carbonaceous aerosol measurement (e.g., Aerosol Mass Spectrometer and Single Particle Soot Photometer). Among these instruments, the Semi-Continuous Carbon (or OC/EC) Analyzer (SCCA) manufactured by Sunset Laboratory Inc. (Beaverton, Oregon, USA) is the only commercial available one that could measure OC and EC simultaneously, thus was widely used in field measurement (Lin et al., 2009; Hu et al., 2012; Docherty et al., 2011).

The initial design of SCCA was described in Birch and Cary (1996). The products have gone through several generations since then, and the newest version now is Model 4 developed in 2000 (Sunset Laboratory, http://www.sunlab.com/about/ about-us.html). Several improvements have been made, such as the change from a flame ionization detector (FID) to a flow-through Non-Dispersive InfraRed (NDIR) system to reduce the requirement for consumable gases (air or hydrogen) at the field site (Bauer et al., 2009). The US EPA has recently launched the Sunset OC/EC Evaluation Project to evaluate the performance of Sunset Model 4 SCCA in various locations and conditions, as part of the preparation work for the use of Sunset SCCA as the routine instrument for long-term monitoring in the Chemical Speciation Network (Sunset OC/EC Evaluation Project, http://www.epa.gov/ttnamti1/spesunset.html).

Identified factors that affected accuracy of SCCA results generally fall into two categories, namely artifacts produced during the sampling process and those interfere the thermal-optical analysis. The positive and negative sampling artifacts can be largely reduced with efficient denuders and a backup guartz filter (Lin et al., 2009; Arhami et al., 2006). Identified analysis interferences included: (a) the inorganic carbonate particulate matter. The concentration of carbonate carbon was generally small compared to

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Total Carbon (TC, sum of OC and EC). When needed, it could be quantified manually since SCCA has allowed it to appear as a unique peak (Karanasiou et al., 2011). (b) Refractory metal oxides, which influence thermal-optical analysis results in two ways. The first of which is that they are usually colored, while that does not cause problems except for samples collected near ore-producing industries (Sunset Laboratory, http://www.sunlab.com/about/technology.html). The second mechanism was that they were found to cause premature EC evolution in the inert mode. Suggested measures to avoid this artifact included frequent change of filters, lowered maximum temperatures in the inert mode, and automated laser-temperature correction (Jung et al., 2011). (c) Colored organic materials, the errors caused by which was found to be minimal. (d) Secondary organic aerosol and biomass burning organic aerosol (Cheng et al., 2011). Organics from these two sources tend to be relatively refractory and easy to pyrolyze, thus complicated the OC/EC split.

Previous efforts on artifact identification and correction have improved the accuracy of SCCA results to a large extent. However, as far as we know, no work about SCCA has focused on the possible calculation discrepancy during the processing of raw signals. In this paper we reported a newly identified type of calculation discrepancy produced by the baseline correction of raw signals. This discrepancy was characterized under several conditions. Also, possible influencing factors and reduction measures were discussed.

2 Experimental method

2.1 Sunset SCCA

The Sunset Model 4 semi-continuous OC/EC analyzer measures aerosol OC and EC simultaneously. Detailed configuration has been well documented elsewhere (Birch and Cary, 1996; Lin et al., 2009; Arhami et al., 2006; Jung et al., 2011; US Environmental Protection Agency, http://www.epa.gov/ttnamti1/spesunset.html). Briefly, with a flow

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rate of $8\,L$ per minute, the air was drawn through two back-to-back quartz-fiber filters, and ambient $PM_{2.5}$ was collected onto a sampling spot of $1.31\,cm^2$. The collected sample was subsequently analyzed with thermal-optical method. Aerosol carbon thermally evolved during each heating steps was converted to CO_2 and detected by NDIR sensor. Optical charring correction by light transmittance was based on a tuned diode laser (660 nm). At the end of each analysis, methane gas (5.0 %; UHP He balance) was injected as an internal standard. After deduction of the determined baseline signals, raw NDIR signals in each step was integrated and converted to carbon mass with a calibrated constant. Then the analytical result is normalized to the response of the methane standard, in order to adjust for the slight variations in flow rates.

In addition to the thermal-optical analyzed OE/EC data (referred to as thermal OC/EC below), SCCA also provided optical EC by recording the laser attenuation every minute throughout the sampling period. This data was related to EC mass via a predetermined conversion factor (Arhami et al., 2006; Jeong et al., 2004). Optical OC was then defined as thermal-optical analyzed TC minus optical EC. Since optical EC was determined independently, it's insusceptible to any calculation discrepancy produced during the thermal-optical analysis.

2.2 Temperature protocol

Three temperature protocols were applied in this work (Table 1). RT-IMPROVE (referred to as IMPshort hereinafter) was an instrument-included IMPROVE-like SCCA protocol, with setting in He mode simplified to one-step (550 °C) based on the idea that detailed separation of OC peaks was less meaningful in SCCA. RT-IMPROVE-A (referred to as IMPlong hereinafter) was modified from offline version of IMPROVE-A (Chow et al., 2007a) protocol. It differs from IMPshort in the more temperature steps in He mode, and the prolonged hold time of He mode in total, especially Step OC4. RT-NIOSH5040 (referred to as rtNIOSH hereinafter) was an SCCA protocol comparable to NIOSH5040, and was well validated against off-line samples (Bae et al., 2004).

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Calculation discrepancy identification

The calculation discrepancy was initially observed during the sucrose tests conducted as external calibrations with IMPshort protocol. Sucrose recovery decreased to only 78% (blue markers in Fig. 1a), far beyond the uncertainty range of 5% (Polidori et al., 2006). To examine the possible influence of the calibration constant used to convert NDIR signals to carbon mass, we further conducted that calibration with another two protocols, i.e. IMPlong and rtNIOSH, with identical constants and conditions. For that two protocols, however, sucrose recoveries was satisfactory (blue markers in Fig. 1b and c), indicating the calibration constant was acceptable. For IMPlong protocol, a slightly deviation (~ 2%) seemed to occur when more than 42.07 µg C of sucrose loaded, while that was well within the uncertainty range. For rtNIOSH, sucrose was fully recovered within the tested range, i.e. up to 100.97 µg C of sucrose.

To clarify the underlying mechanism responsible for the underestimated sucrose recovery by IMPshort protocol, we examined the raw signals, and found that the base-

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line correction method appeared to be a likely cause. In Model 4 SCCA, a Multi-Point baseline correction method (MP) was adopted in the instrument-included calculation software (RTcalc, version 522) to improve the linearity of NDIR detectors. This method determined the baseline according to signal at the start point, the center point (the lowest datapoint around the introduction of He/O₂) and the end point (datapoint just before the methane calculation peak). It was such designed to deal with an increasing or rounded NDIR baseline, which, rather than a horizontal one, is more often the case observed in instrumental blank tests (Bauer et al., 2009). MP was the default method and should have been used by all previous SCCA-related work. However, MP corrected baseline was appropriate only if the NDIR signal have truly returned to baseline level at center datapoint, which we suspect not the case especially under heavy carbon loads, due to the incomplete evolution of OC within the limited residence time (Chow et al., 2005). In this case, the baseline would be biased to a higher level, therefore TC being

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underestimated. A schematic diagram of the suspected mechanism was depicted in Fig. 2.

To make sure the relationship of MP and the calculation discrepancy, we reanalyzed the raw signals of sucrose tests with another baseline correction method, namely the Single-Point method (SP). SP equaled the signal at the starting point as the baseline all through the analysis. Results of quantified TC with MP and SP method were referred to below as TC_{MP} and TC_{SP} , respectively. When SP correction was applied, recovery of sucrose spikes were statistically indistinguishable from unity for all protocols (red markers in Fig. 1), confirming the assumption that the "lost" sucrose with IMPshort protocol was caused by MP correction.

3.2 Evidence from ambient samples

Although no difference was observed in sucrose tests of IMPlong and rtNIOSH protocols, their performance in ambient samples was unknown. To examine the possibility of calculation discrepancy occurrence in ambient samples, field sampling was conducted in Tsinghua University (40°19′ N, 116°19′ E), a site in urban area of Beijing without major industrial sources nearby. Three protocols were used sequentially, with IMPshort from mid-July to November 2011, rtNIOSH in January and February 2012, and IMPlong in June, July, and November 2012. Time resolution (sampling time + analysis time) was 1 h (45 min + 15 min) for IMPshort, and 2 h (100 min + 20 min) for IMPlong and rt-NIOSH. Filters changing and sampling cyclone cleaning were done every 4-7 days, and sucrose calibration was conducted monthly. Raw signals were analyzed with both MP and SP. The adopted detection limit here was $2.0 \,\mu \text{g m}^{-3}$ for OC and $0.5 \,\mu \text{g m}^{-3}$ for EC⁵, and data below this value was screened out. TC concentrations (μgCm⁻³) were then converted into mass (µqC) by multiplying the sample volume (m³), to avoid the ambiguity caused by different sampling time in one cycle. For ambient samples, the accurate TC value of which being unknown, TC_{SP} was used as reference of true TC based on results in sucrose tests. The calculation discrepancy was thus defined as -(TC_{MP} - TC_{SP}), the negative sign indicating that a TC underestimation was expected.

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Accordingly, the relative magnitude of the discrepancy was defined as the percentage of underestimated TC to TC_{SP}.

For ambient samples, this calculation discrepancy was observed with all protocols (Fig. 3) in a similar pattern with that of sucrose, namely being present only when above 5 a certain threshold carbon load, and being positively correlated with carbon load once above that threshold. However, detailed characteristics could differ much among protocols. For IMPshort protocol, ambient samples and sucrose performed much alike. The threshold carbon levels were both less than the detection limit, with magnitude of underestimation almost the same (22%). For IMPlong protocol, however, performance of ambient samples was in sharp contrast with that of sucrose. Threshold carbon for ambient samples was determined to be about 20 µgC. Although the difference was present even before that threshold, it was thought to be caused by the systematic error resulted from the significantly higher instrumental blanks associated with SP, as discussed in Sect. 3.3. This was further supported by the sharp increase in discrepancy magnitude up into 36% once after that threshold. In comparison, for sucrose an insignificant underestimation of 2 % was seen only when over 42 µg C was loaded. As to rtNIOSH protocol, although the threshold was not detected up to 100 µg C of sucrose, it was seen at only 25 µg C of ambient samples. Nevertheless, the magnitude of its discrepancy (12%) ranked the lowest among three protocols. The distinct SCCA performance on sucrose and ambient samples suggested the need of a more refractory calibration substance to better represent ambient samples.

The characteristics of this discrepancy, i.e. the discrepancy-emerging threshold carbon load and the magnitude of TC underestimation, varied with the type of samples as well as temperature protocols. The underlying reason of these variations could come down to thermal stability distribution of carbonaceous samples. Compared with sucrose, TC underestimation in Beijing ambient samples was generally seen at significantly lower threshold and larger magnitude. In addition, the magnitude of TC underestimation varied with seasons as well. rtNIOSH was applied only in winter, thus its seasonal variation was not available. While for IMPshort and IMPlong, a bifurcation

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in magnitude was both seen among seasons (Fig. 3a and b). Generally, the discrepancy tend to be larger in summer (June to August) samples than winter (November to February next year) ones. These differences could be largely explained by variation of thermal properties among sample types. The significant lower threshold for sucrose than ambient samples suggested a more refractory nature of ambient OC. More refractory OC needed more time or higher temperature to evolve completely, thus increased the possibility that the signal not returning to baseline level at center point. Compared with sucrose of same TC, the increased portion of refractory OC in ambient sample made it more sensitive to this calculation discrepancy. The seasonal variation of this discrepancy could be explained similarly. It was well proved that summer samples contained a larger portion of water soluble organic carbon (Lin et al., 2009; Cheng et al., 2011), which was an important source of organic compounds that was relatively refractory and easy to pyrolyze during the inert mode. The increased proportion of water soluble organic aerosol could push the thermal properties of the whole ambient sample towards a more refractory direction, thus increasing the possibility of discrepancy occurrence. For a fixed sample type, the threshold was always in the increasing order of IMPshort, IMPlong and rtNIOSH, indicating a prolonged time or an increased temperature in the step just before center point (i.e. the highest temperature step of He mode) could reduce the possibility of discrepancy occurrence.

Although frequently exceeded in the heavily polluted Beijing samples, the observed threshold carbon load was a rather high value for American samples. That could largely explain why previous validation work failed to identify this type of calculation discrepancy. For example, Bae et al. (2004) examined the rtNIOSH protocol against off-line instrument and found good agreement. However, the maximum TC concentration in their work was $\sim 20 \,\mu g \, m^{-3}$ (corresponding to 9.6 $\mu g \, C$), well below the rtNIOSH threshold of 25 µgC. In addition, the semi-continuous carbon analyzer used in their work was an older version equipped with the FID detector. Also, the validation work of Bauer et al. (2009) adopted a protocol similar to rtNIOSH, the performance of which was also expected to be similar. Maximum TC in that work was even lower, corresponding to only

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3.24 µg C. In another inter-protocol comparison work (Arhami et al., 2006), a modified-IMPROVE protocol was shown to agree well with a modified-NIOSH protocol. In that modified-IMPROVE protocol, the total hold time in He mode was 315 s, with a 125 s residence time in the temperature plateau of 550 °C. The threshold of that protocol was thus expected to be a little lower than the IMPlong threshold of 20 µg C. In comparison, the maximum TC concentration in that study was only $\sim 30 \,\mu \mathrm{g\,m^{-3}}$ (corresponding to 10.8 µgC). Moreover, the difference of thermal properties between Beijing and American aerosols may also contribute to the different picture.

Instrumental blanks

To examine the influence of true baseline signals, instrumental blanks measured daily at 00:00 LT were respectively calculated with MP and SP correction. The distribution was shown in Fig. 4. Based on MP corrected results, the upper limit of instrumental blanks was suggested to be 0.3 µgC in standard operating procedure. This criteria was met for all protocols tested. By contrast, SP-corrected instrumental blanks were distributed with much higher means and larger variances than MP results. The higher means indicated the incapability of SP to exclude true baseline signals, while the large variances implied that it was not appropriate to correct SP results simply by deducting a constant instrumental blank value. In addition, both the variances and means of SPcorrected results were in the increasing order of IMPshort, rtNIOSH, and IMPlong. TC_{SP} blank level of IMPlong was the highest, reaching $0.99 \pm 0.86 \,\mu g \,C$ (average \pm one standard variation). In comparison, the blank level of rtNIOSH as $0.48 \pm 0.78 \,\mu$ gC was more acceptable. As IMPshort could not provide accurate TC estimation, rtNIOSH was recommended over IMPlong in terms of a more stable blank level.

3.4 Reduction and correction method

When ambient carbon load exceeded a certain threshold value, the default MP baseline correction could cause significant TC underestimation. Although this TC

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underestimation could be corrected with SP method, the SP method could not exclude the influence of fluctuating true baseline signals, thus would introduce a TC overestimation which can be significant under low carbon load. Since no single correction method is valid under all circumstances, a thorough solution should include measures to avoid this calculation discrepancy, as well as a systematic procedure for discrepancy detection and correction.

One possible measure to avoid this calculation discrepancy was to shorten the sampling time within a cycle. As carbon load equals ambient concentration multiplies sampling time in a cycle, this could effectively reduce the carbon load with the same ambient concentration, thus increasing the threshold TC concentration. Take rtNIOSH for example, when the sampling time was 2 h (100 min sampling plus 20 min analysis), the threshold carbon load of 25 μ gC corresponded ambient TC concentration of 31.25 μ gCm⁻³. When changed to 1 h cycle (40 min sampling plus 20 min analysis), the threshold-corresponding concentration was increased to 78.13 μ gCm⁻³, which was close to the maximum TC concentration observed in this work. However, this was at the cost of reduction of proportion of effective sampling time, thus the representativeness of reported concentration. Moreover, this measure could solve only part of the problem for IMPlong and IMPshort protocol, the threshold carbon load of which was much lower than rtNIOSH.

Another choice was to increase the temperature plateau and/or prolong the hold time in the step just before center point (i.e. the highest temperature step of He mode), so as to increase the possibility that $\rm CO_2$ signal having returned to baseline at center point. The effectiveness of increasing temperature was validated by the better performance of rtNIOSH over IMPshort and IMPlong. The effectiveness of prolonging hold time in He mode was examined with an additional series of sucrose experiment. Filters loaded with a certain sucrose mass (42.07 μ g) was analyzed with a series of IMPlong-like protocols. This series of protocols differs only in OC4 hold time, which varied from 60 s to 240 s with a 30 s interval. Checks were repeated 3 times, and averaged results were shown in Fig. 5. The differences of $\rm TC_{MP}$ to target mass was found to decrease

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with prolonged hold time, and down into the uncertainty range when hold time was longer than 210 s. This measure was an effective solution for application of IMPROVE protocols on SCCA, while it also suffers from the low efficient sampling time ratio.

Since this calculation discrepancy is inevitable under high ambient concentrations, we propose a systematic procedure for check and correction of this discrepancy, as described below. (i) Analyze raw data with both SP and MP correction. (ii) Plot the difference between TC_{SP} and TC_{MP} against TC_{SP}, as illustrated in Fig. 3. (iii) If the difference is distributed randomly around a certain value (true baseline signals), then the threshold is not reached. Under this circumstances, MP-corrected results are suggested. (iv) If the difference increased positively with TC_{SP} when above a certain threshold, the data should be treated separately. Use MP-corrected results for data below that threshold, and SP results when beyond that threshold.

As the correlation between thermal and optical EC was observed to be rather good in previous SCCA validation work (Arhami et al., 2006; Yu, 2011), it provided another criteria to testify the validity of suggested correction method. Since the TC underestimation existed nearly all the time for IMPshort, this protocol was not considered here. For the other two protocols, the correlation was tested between optical EC and thermal EC including MP-corrected EC (referred to as EC_{MP} hereinafter), and SP-corrected EC (referred to as EC_{SP} hereinafter) (Fig. 6). General pattern was the same for both protocols. Although EC_{SP} correlated better with optical EC compared to EC_{MP} in terms of overall performance (Fig. 6c and g), EC_{MP} correlated better with Optical EC when the threshold was not reached (Fig. 6a and e). We also checked the performance of EC data corrected with the method we proposed above (Fig. 6d and h), i.e. using EC_{MP} when below the threshold and EC_{SP} when above that threshold. Stronger correlation between optical and thermal EC was found, with R^2 reaching 0.87 and 0.85 for IMPlong and rtNIOSH protocol, respectively. However, an obvious gap in corrected EC data was observed for IMPlong protocol, caused by the significantly higher EC_{SP} data than EC_{MP} data. In comparison, corrected EC showed good continuity for rtNIOSH protocol. In addition, the absolute values of corrected EC data given by rtNIOSH protocol was

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Conclusion and implications

The widely use of Sunset SCCA in carbonaceous aerosol studies required high data accuracy. Previous validation work has focused on the artifacts associated with sampling and analytical methods. The results of this study suggested the significance to take calculation discrepancy produced during data processing into consideration. Under high ambient aerosol concentrations, the default multi-point baseline correction method could generally result in a TC underestimation of 12% to 36%, depending on the temperature protocol used. This indicated the potential need of re-analysis of previous reported data, especially those with IMPROVE-like protocols (e.g., Pan et al., 2012), and/or under heavily polluted period or areas (e.g., Lin et al., 2009; Andreae and Gelencser, 2006). Single-point correction combined rtNIOSH protocol could give the most reliable results under high carbon load. A framework to identify and avoid this calculation discrepancy was suggested. The effectiveness of this framework was supported by correlation with optical data.

Supplementary material related to this article is available online at http://www.atmos-meas-tech-discuss.net/7/377/2014/ amtd-7-377-2014-supplement.pdf.

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Table 1. Temperature protocols used for sunset semi-continuous OC/EC analyzer.

Step	Gas	RT-IMPROVE (IMPshort)		RT-IMPROVE-A (IMPlong)		RT-NIOSH5040 (rtNIOSH)	
		Temperature (°C)	Hold Time (s)	Temperature (°C)	Hold Time (s)	Temperature (°C)	Hold Time (s)
OC1	He	550	110	140	60	340	60
OC2	He	_	_	280	60	500	60
OC3	He	_	_	480	60	615	60
OC4	He	_	_	580	240	870	90
Cooling Oven	He	_	_	_	_	Oven off	60
He Time			110		420		330
EC1	He/O ₂	550	75	580	90	550	45
EC2	He/O ₂	700	75	740	90	625	45
EC3	He/O ₂	850	90	840	240	700	45
EC4	He/O ₂	_	_	_	_	775	45
EC5	He/O ₂	_	_	_	_	850	45
EC6	He/O ₂	_	_	_	_	900	120
He/O ₂ Time	-		240		420		345
Total Time			350		840		675

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Interactive Discussion



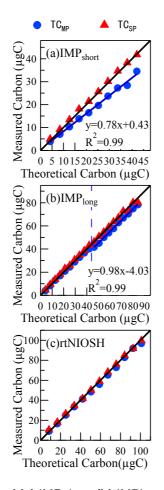


Fig. 1. Sucrose calibration results of **(a)** IMPshort **(b)** IMPlong and **(c)** rtNIOSH protocol, baseline corrected with MP and SP, respectively. The 1:1 line was shown for reference. Linear regression of TC_{MP} was indicated in blue line for reference.

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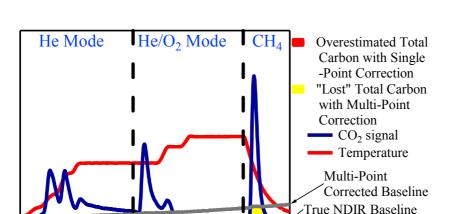


Fig. 2. Schematic diagram of the suspected mechanism of the calculation discrepancy produced by Multi-Point baseline correction method.

End Point

Start Point Center Point

Single-Point

Corrected Baseline

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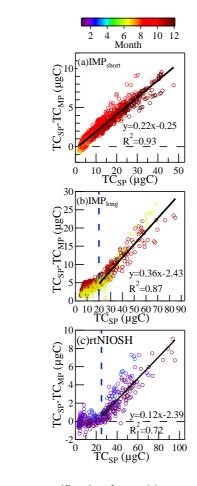


Fig. 3. Calculation discrepancy quantification for ambient samples analyzed with (a) IMPshort, (b) IMPlong and (c) rtNIOSH protocol.

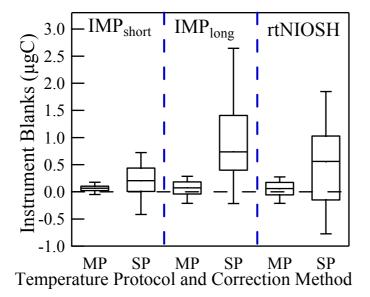


Fig. 4. Instrumental blank distributions for the three protocols respectively corrected with two baseline correction methods. The box-and-stem plots depict the 5th, 25th, 50th (median), 75th, and 95th percentile for instrumental blanks. See text for detailed explanation of the abbreviations.

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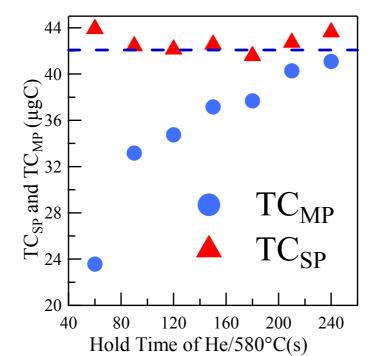


Fig. 5. TC recoveries of sucrose changed with hold time of Step OC4 in IMPlong protocol. Target Sucrose mass (42.07 μg C) was shown in dash line for reference.

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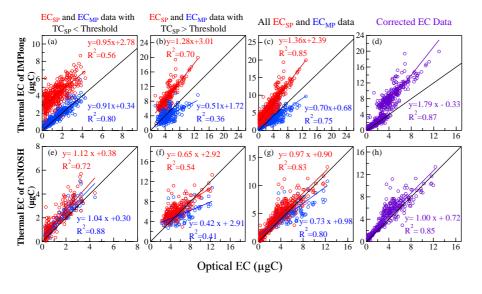


Fig. 6. Thermal EC data, including ECSP, ECMP and corrected EC, plotted against Optical EC for (a-d) IMPlong protocol, and (e-h) rtNIOSH protocol. Linear regression of both ECSP and ECMP with Optical EC was done for (a and e) data with corresponding TCSP < Threshold. (b and f) data with corresponding TCSP > Threshold carbon, and (c and q) all data, respectively. Correlation of corrected EC with Optical EC (d and h) was also shown for comparison. The 1:1 line was shown for reference.