



This discussion paper is/has been under review for the journal Atmospheric Measurement Techniques (AMT). Please refer to the corresponding final paper in AMT if available.

Predicting ambient aerosol Thermal Optical Reflectance (TOR) measurements from infrared spectra: elemental carbon

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Received: 18 March 2015 – Accepted: 31 May 2015 – Published: 23 June 2015

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Published by Copernicus Publications on behalf of the European Geosciences Union.

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Abstract

Elemental carbon (EC) is an important constituent of atmospheric particulate matter because it absorbs solar radiation influencing climate and visibility and it adversely affects human health. The EC measured by thermal methods such as Thermal-Optical Reflectance (TOR) is operationally defined as the carbon that volatilizes from quartz filter samples at elevated temperatures in the presence of oxygen. Here, methods are presented to accurately predict TOR EC using Fourier Transform Infrared (FT-IR) absorbance spectra from atmospheric particulate matter collected on polytetrafluoroethylene (PTFE or Teflon) filters. This method is similar to the procedure tested and developed for OC in prior work (Dillner and Takahama, 2015). Transmittance FT-IR analysis is rapid, inexpensive, and non-destructive to the PTFE filter samples which are routinely collected for mass and elemental analysis in monitoring networks. FT-IR absorbance spectra are obtained from 794 filter samples from seven Interagency Monitoring of PROtected Visual Environment (IMPROVE) sites collected during 2011. Partial least squares regression is used to calibrate sample FT-IR absorbance spectra to collocated TOR EC measurements. The FTIR spectra are divided into calibration and test sets. Two calibrations are developed, one which is developed from uniform distribution of samples across the EC mass range (Uniform EC) and one developed from a uniform distribution of low EC mass samples ($EC < 2.4 \mu\text{g}$, Low Uniform EC). A hybrid approach which applies the low EC calibration to low EC samples and the Uniform EC calibration to all other samples is used to produce predictions for low EC samples that have mean error on par with parallel TOR EC samples in the same mass range and an estimate of the minimum detection limit (MDL) that is on par with TOR EC MDL. For all samples, this hybrid approach leads to precise and accurate TOR EC predictions by FT-IR as indicated by high coefficient of variation (R^2 ; 0.96), no bias ($0.00 \mu\text{g m}^{-3}$, concentration value based on the nominal IMPROVE sample volume of 32.8 m^3), low error ($0.03 \mu\text{g m}^{-3}$) and reasonable normalized error (21 %). These performance metrics can be achieved with various degrees of spectral pretreatment (e.g., including or

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excluding substrate contributions to the absorbances) and are comparable in precision and accuracy to collocated TOR measurements. Only the normalized error is higher for the FT-IR EC measurements than for collocated TOR. FT-IR spectra are also divided into calibration and test sets by the ratios OC/EC and ammonium/EC to determine the impact of OC and ammonium on EC prediction. We conclude that FT-IR analysis with partial least squares regression is a robust method for accurately predicting TOR EC in IMPROVE network samples; providing complementary information to TOR OC predictions (Dillner and Takahama, 2015) and the organic functional group composition and organic matter (OM) estimated previously from the same set of sample spectra (Ruthenburg et al., 2014).

1 Introduction

Elemental carbon (EC) in atmospheric aerosols adversely impacts human health (Janssen et al., 2011) and contributes to climate warming (Bond et al., 2013) and decreased visibility (Hand et al., 2014). Elemental carbon is measured by large monitoring networks such as the Interagency Monitoring of Protected Visual Environments (IMPROVE) network, (Hand et al., 2012; Malm et al., 1994) in rural areas of the US, the Speciation Trends Network/Chemical Speciation Network (Flanagan et al., 2006) in urban areas of the U.S, and the European Monitoring and Evaluation Programme (EMEP, Torseth et al. (2012)) throughout Europe. These regional, multi-year data sets are useful for observing trends in particulate concentrations (Hand et al., 2013; Hidy et al., 2014; Torseth et al., 2012) and visibility (Hand et al., 2014), evaluating aerosol transport models (Mao et al., 2011), constraining climate models (Liu et al., 2012) and assessing health impacts (Krall et al., 2013). EC is measured using a thermal optical method (TOR, Chow et al., 2007, NIOSH 5040, Birch and Cary, 1996, European Supersites for Atmospheric Aerosol Research, EUSAAR-2 protocol, Cavalli et al., 2010), or similar, in which particles collected on quartz filters are subjected to a temperature gradient; first in an inert environment and then in an oxidizing environment (Chow

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et al., 2007). Organic carbon (OC) and EC are operationally defined by the temperature and environmental conditions under which the carbon evolves from the aerosol sample. Charred organic material is subtracted from the measured EC based on laser reflectance or transmittance (Cavalli et al., 2010; Chow et al., 2007). These thermal optical methods are not applicable to the PTFE (polytetrafluoroethylene) media used for gravimetric mass, elemental composition and sometimes light absorption in sampling networks because they are the filter material is unstable at high temperatures. TOR methods are also destructive to the sample and expensive.

Thermal optical methods are one type of method that seeks to measure carbon in atmospheric aerosols that is structurally similar to graphite in that it is composed of sp^2 -bonds and is strongly light absorbing (Bond and Bergstrom, 2006). Thermal optical methods refer to this material as EC. Other methods use light absorption to characterize sp^2 -bonded carbon in aerosol (Andreae and Gelencser, 2006) and typically refer to the constituent being measured as black carbon (BC). Continuous light absorption measurements are made by such instruments as the PSAP (Bond et al., 1999), aethalometer (Coen et al., 2010), and multi-angle absorption photometer (MAAP) (Petzold et al., 2005). Time-integrated absorption measurements are made from filters samples using instruments such as the hybrid-integrating plate method used by the IMPROVE network (White, 2015). Fourier Transform Infrared (FT-IR) spectroscopy has also been used to characterize sp^2 -bonded carbon in aerosols and other environmental samples. FT-IR spectra of ground graphite and ground activated carbon have prominent peaks at 1580 cm^{-1} which were assigned to the graphic structure of the material (Friedel and Carlson, 1971). FT-IR spectra of ground charcoal and synthetic and marine sediments had similar peaks at 1580 cm^{-1} and additional peaks at 1720 and 1240 cm^{-1} , which were assigned to carbonyl and single carbon-oxygen bonds, respectively (Smith et al., 1975). FT-IR and partial least squares regression (PLSR) have been used to quantify BC in soil samples by measuring benzene polycarboxylic acids which are aromatic markers for black carbon (Bornemann et al., 2007). In a companion paper (Takahama and Dillner, 2015), we identify wavenumbers in the spectra of ambient

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aerosol samples that correlate strongly to TOR EC. The wavenumbers that correlate to TOR EC are not the same as those in sediments, charcoal and graphite indicating that the EC in aerosols is functionally different than other sample types. However, the presence of these wavenumbers indicates the potential for FT-IR spectra calibrated to TOR EC using PLSR to predict TOR EC in aerosol samples.

In this work, FT-IR spectra are calibrated to TOR EC using PLSR, similar to our previous method for predicting TOR OC (Dillner and Takahama, 2015). PLSR is a general method that has been used to calibrate FT-IR spectra of environmental samples such as dust (Weakley et al., 2014), food (Polshin et al., 2011) and soil (Bornemann et al., 2007) to constituents of interest. As described above, thermal optical methods currently provide EC measurements in air monitoring network ambient particle matter samples. Such networks simultaneously sample particles on PTFE filters which are used for gravimetric mass and elemental composition analysis and can also be used to obtain FT-IR spectra. In this work, EC is predicted from infrared spectra of PTFE filter samples of aerosols using PLSR. The methods are developed and tested using EC from TOR analysis and FT-IR spectra from parallel PTFE filters from one year of samples from seven IMPROVE sites.

The objective of this work is to demonstrate the feasibility of predicting TOR EC from infrared spectra as a second step in developing an inexpensive, fast and non-destructive method for carbon measurements in PM monitoring networks. Prediction of TOR OC was the first step in developing this method and was established in Dillner and Takahama (2015). Sampling networks which only use PTFE filters, such as the Federal Reference Method network used for compliance with National Ambient Air Quality Standards for PM mass concentrations in the United States, can also use the FT-IR and PLSR method presented here and in Dillner and Takahama (2015) to obtain information about the carbonaceous aerosol provided that the samples have similar aerosol composition as the calibration samples. In this work, we will show that the prediction of TOR EC can be accomplished with accuracy on par with TOR EC measurement precision. Furthermore, we will mechanistically explain important differences in sam-

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ple composition between calibration and test sets that can lead to increased prediction errors; for this we use additional IMPROVE measurements to aid in our interpretation. And, finally, we will demonstrate how sensitivity to sample composition is manifested in predictions for sites that are not included in the calibration set.

2 Methods

2.1 IMPROVE Network samples

This study uses 794 IMPROVE particulate matter samples collected on PTFE filters and 54 blank PTFE filters. The IMPROVE samples were collected at seven sites during 2011 (Fig. S1 in the Supplement). These are the same samples, blanks and consequent FT-IR spectra used for developing the OC method in Dillner and Takahama (2015) and the OM/OC method in Ruthenburg et al. (2013). See these papers for additional details. In the IMPROVE network, filter samples of particles less than $2.5\text{ }\mu\text{m}$ ($\text{PM}_{2.5}$) are collected every third day from midnight to midnight local time at a nominal flow rate of 22.8 liters per minute which yields a nominal volume of 32.8 m^3 .

The FT-IR analysis is applied to 25 mm PTFE filters (Teflo, Pall Gelman, 3.53 cm^2 sample area) that are analyzed for gravimetric mass, elements and light absorption in the IMPROVE network. Quartz filters collected in parallel to the PTFE filters are analyzed by TOR and adjusted to account for charring of organic material prior to reporting EC mass in the IMPROVE network (Chow et al., 2007). For this work, the EC values are also adjusted to account for flow differences between the quartz and PTFE filters.

In order to provide reference performance metrics for the evaluation of the FT-IR to TOR comparisons (see Sect. 2.4 for a description of the metrics), measurements from seven IMPROVE sites with collocated TOR measurements (Everglades, Florida; Her- cules Glade, Missouri; Hoover, California; Medicine Lake, Montana; Phoenix, Arizona; Saguaro West, Arizona; Seney, Michigan) are used.

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2.2 FT-IR analysis

2.2.1 Spectra acquisition

PTFE filters are analyzed using a Tensor 27 Fourier transform infrared (FT-IR) spectrometer (Bruker Optics, Billerica, MA) equipped with a liquid nitrogen-cooled wide-band mercury cadmium telluride detector. The samples are analyzed using transmission FT-IR over the mid infrared wavenumber region of 4000 to 420 cm^{-1} (see Ruthenburg et al., 2014 for more details). Absorbance spectra are calculated using a recent spectrum of the empty sample compartment as a zero reference. Air free of water vapor and carbon dioxide (delivered by purge-gas generator; PureGas LLC, Broomfield, CO) is used to continuously purge the optical compartments of the instrument and to purge the sample compartment for 4 min before each sample or reference spectrum is acquired.

2.2.2 Spectra preparation

Three different versions of the absorption spectra are used in our analysis (Fig. S2), corresponding to different pretreatments and wavelength selection: (1) “Raw” spectra are unmodified spectra except that values interpolated during the zero-filling process are removed. These spectra contain all 2784 wavenumbers. (2) “Baseline corrected” spectra include absorbances above 1500 cm^{-1} and the substrate contribution is removed by subtracting an average blank filter spectrum and then using linear or polynomial baselines by spectral region as described by Takahama et al. (2013). These spectra are standardized to a 2 cm^{-1} resolution, so contain 1563 wavenumbers. (3) “Truncated” spectra are the raw spectra interpolated to match the wavenumbers in the baseline corrected spectra, which excludes the PTFE peaks (the region below 1500 cm^{-1}), so also contain 1563 wavenumbers.

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2.3 Calibration

The FT-IR spectra are calibrated to TOR EC measurements using Partial Least Squares Regression (PLSR) using the kernel pls algorithm, implemented by the pls library (Mevik and Wehrens, 2007) for the R statistical package (R Core Team, 2014).

Conceptual description of PLSR can be found in Dillner and Takahama (2015), Takahama and Dillner (2015) and Ruthenburg et al. (2013) and references therein. Briefly, in PLSR, the matrix of spectra is decomposed into factors and their respective weights. Candidate models are generated by varying the number of factors used to reconstruct the EC mass in the calibration filters.

Using the common approach for model selection and assessment (Hastie et al., 2009; Bishop, 2011; Witten, 2011), two-thirds of the 794 samples and two-thirds of the blanks filters (which are assumed to have zero EC mass) are used for developing the calibration and one-third of the samples and blanks (called the test set) are used to evaluate the model. Ambient samples with EC below TOR EC MDL are excluded from the model so as not to train the calibration with samples that have a low signal to noise ratio. K -fold cross validation with $k = 10$ is used to estimate the accuracy of each candidate model. The minimum root mean square error of prediction (RMSEP; Mevik and Cederkvist (2004)) is used with cross-validation to select the model with least prediction error. This procedure permits development and selection of PLSR models using only the samples in the calibration set, and guards against over-fitting to a single set of samples. The test set, which has not been used in model development or selection, is used for model evaluation.

The above procedures are the same as those used to develop the OC calibration (Dillner and Takahama, 2015). For EC, possibly on account of lower mass concentrations, additional methods for including blanks into the calibration are needed. First, a calibration is developed without blank filters and used to predict the mass of EC on each blank filter. The blanks are then ordered by EC mass and every third is included in the test set and the remaining are put into the calibration set so that there

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are similar distributions (both positive and negative) of blanks in the two sets. Blanks are interspersed in the calibration at roughly equal intervals so that each iteration of the cross-validation process has a roughly consistent number of blanks used for calibration. Calibrations developed without ordering the blanks and without interspersing the blanks in the calibration filter set lead to inconsistent MDL values.

EC is predicted for a sample (i) by taking the product of the absorbance at wavenumber j ($x_{i,j}$) and the calibration vector (b_j) as shown in Eq. (1).

$$EC_i = \sum_j x_{i,j} b_j \quad (1)$$

Blank samples in the test set are used to calculate the MDL. Performance metrics used to assess the quality of the calibration model and MDL determination are described in Sect. 2.4.

Multiple calibrations are developed by varying the spectra type used and by selecting filters for the calibration and test sets using different ordering regimes. Initially, we develop a set of calibrations based on uniform and non-uniform distributions of EC in the calibration and test sets. For the uniform case, the samples are ordered by EC mass and every third sample is put into the test set so the distribution of EC is the same in the calibration and test sets. Three non-uniform cases are developed to assess the impact of EC distribution on the quality of the calibration. For the non-uniform cases, the samples are ordered by EC mass and then selected for calibration and test set based on where each sample falls within the range of EC masses. The three non-uniform cases are: (1) samples with mass in the highest two-thirds of the EC mass range are used for calibration and samples with EC in the lowest one-third of the EC mass range are used as the test set (non-uniform A), (2) Samples in the highest and lowest one-third of the EC mass range are used to predict samples in the middle one-third of the EC mass range (non-uniform B), and (3) samples with EC mass in the lowest two-thirds of the EC range are used to predict samples in the highest one-third of the EC mass range (non-uniform C).

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Figure 1 shows the uniform and three non-uniform calibrations developed for EC. A description of the performance metrics shown in Fig. 1 are given in Sect. 2.4. The top row gives the EC distribution of the test and calibration set for each case. The EC distributions reflect the algorithm used to select the filters for that case. The median and 25th to 75th percentiles (interquartile range) of the bias and normalized error are shown in the lower two rows of Fig. 1 for each of the three spectra types (indicated by symbol shape). Small open symbols are used for sets with low median EC masses. Larger closed symbols have higher median EC masses. The EC distributions for the Uniform case shows that the distribution of samples in the calibration set and test set are very similar. This leads to predictions with small bias (middle plot) and median normalized errors (bottom plot) ranging from 15 to 30 % for the test and calibration sets for the three spectral treatments. However, when low EC mass samples are used to predict high EC samples (non-uniform A) and high samples are used to predict low EC samples (non-uniform C), the test set is biased leading to high error, especially when the EC values are low (non-uniform C).

In order to eliminate bias in the calibration and improve prediction capability for low EC samples, we develop a localized calibration for samples in the lowest one-third of the EC mass range ($EC < 2.4 \mu g$). The Low EC calibration uses samples in the calibration set that are in the lowest one-third of the EC mass range to predict samples in the test set that are also in the lowest one-third of the EC mass range. Localization of the calibration is a commonly used method to improve the performance of the calibration, often at the more difficult to measure low end of the range. The Uniform (which contains samples across the whole range of EC, Fig. 1) and Low EC calibrations are used to predict low EC test set samples and the resulting error and MDL are compared in Fig. 2. The mean error or precision of collocated TOR EC samples below $2.4 \mu g$ and the reported TOR EC MDL are also shown in Fig. 2. The Low EC calibration reduces the error for all three spectral types to a value similar to that of the precision of collocated TOR EC samples in the same mass range. In addition, the MDL (which is based on the prediction of blank filters) is reduced to approximately the TOR MDL for all three

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spectral types. This comparison indicates that a localized calibration greatly improves the prediction quality for low EC samples and that the error in the FT-IR prediction is due primarily to TOR EC measurement uncertainty.

A hybrid calibration method, which includes the Uniform calibration for the whole EC range and the Low EC calibration for the low EC range is used for all results presented in Sect. 3.0, regardless of the ordering scheme used to select filters for the calibration and test sets. In the hybrid calibration method, the Uniform calibration is used to predict all filters in the test set. Those filters in the test set which are predicted to be in the lowest one-third of the EC mass are then analyzed by the Low EC calibration. The prediction from the full calibration is used for samples above the lowest one-third and predictions from the low EC calibration are used for the low EC samples.

A “Base case” hybrid calibration, where the samples are chronologically stratified per site (i.e., ordered by date for each site) is developed as a reference scenario. Every third samples in the ordered list is put into the test set and the rest are put into the calibration set. The base case has the same samples in the test set as the base case for the OC predictions reported earlier (Dillner and Takahama, 2015). Description of a minor error in the OC calibration set for the base case is documented in Sect. S3, and this has been corrected for the EC calculations in this manuscript. The blanks are calculated and distributed as previously stated. This ordered set of samples based on sites and date provides fairly uniform distribution of EC and aerosol composition in the calibration and test sets. In addition to the base case, uniform and non-uniform OC/EC and ammonium/EC and site specific calibrations are developed using the hybrid approach and discussed in more detail in the Sect. 3.

2.4 Methods for evaluating of the quality of calibration

The quality of each calibration is evaluated by calculating four performance metrics: bias, error, normalized error and the coefficient of variation (R^2) of the linear regression fit of the predicted FT-IR EC to measured TOR EC. FT-IR EC is the EC predicted from the FT-IR spectra and the PLSR calibration model. The bias is the median dif-

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and Takahama, 2015) and the R^2 values are the same (0.96, Dillner and Takahama, 2015). The normalized error for the test set (21 %) is higher than the collocated TOR EC precision (13 %) and higher than the TOR OC normalized error (11 %). The hybrid calibration also produces high quality predictions of EC baseline corrected and truncated spectra as indicated by the similarity in performance metrics for all three spectral types (Fig. S4b). The distribution of normalized errors for the calibration and test set for the baseline corrected and truncated spectral pretreatments are similar to raw spectra and the collocated precision for the TOR samples (Fig. S4a). Section S5 demonstrates that the number of samples in the calibration set can be reduced and still provide high quality predictions. The analysis suggests that the accuracy of FT-IR EC predictions is comparable to the precision of collocated TOR EC measurements.

Table 1 gives the MDL and collocated Phoenix sample precision for the hybrid FT-IR EC predictions for each spectral type and TOR EC. The MDL for all hybrid FT-IR EC spectral types are approximately the same as TOR EC with 3 % or less the samples below MDL. Section S5 shows that the MDL is independent of the number of blanks included in the calibration and the number of samples included in the calibration set (from 2/3 to 1/3 of the samples). The collocated Phoenix precision is better for the FT-IR methods than for TOR by about half. The mean predicted value for the blanks filters (last row of Table 1) is less than half of the 1st percentile of predicted EC values for the raw and baseline corrected spectra and equivalent to the 2nd percentile for the truncated case. The precision is on the same order for all three spectral types.

An alternate method for estimating EC is to predict total carbon (TC = OC + EC) and subtract the OC prediction. This method can lead to higher errors on average than direct prediction of EC because the errors of both TC and OC are included in this method of estimation and these errors are larger than EC errors because EC mass is usually much less than OC mass. More discussion of this topic is included in Sect. S6.

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3.2 Evaluating causes of bias and error by selecting the calibration and test sets based on measured parameters

In this section, we consider the role of aerosol composition on the quantification of EC. Aerosol composition is described by the distribution of OC/EC and ammonium/EC. In other work, we show that OC and EC measurements by FT-IR and PLSR, rely on similar wavebands for predictions (Takahama and Dillner, 2015). Therefore, OC could be considered an interferant to the measurement of EC. We evaluate this possible interferant using the ratio of OC to EC because the impact of OC is likely dependent on its mass relative to EC mass. Ammonium absorbs FT-IR radiation in the same wavenumber region as OC and EC and so can also be considered an interferant. We use the ratio of ammonium to EC mass loadings to isolate the effect of ammonium. OC/EC and ammonium/EC are not correlated to 1/EC ($R^2 < 0.2$), which indicates that impacts observed from EC, OC/EC and ammonium/EC are at least somewhat independent of each other. Therefore, separate calibrations developed by ordering the samples by OC/EC and ammonium/EC.

As was done for uniform EC case, samples are arranged in ascending order by the parameter of interest prior to selection of filters for the calibration and test sets. Every third sample in the ordered list is put into the test set and the remaining samples are put into the calibration set. These cases are called the Uniform OC/EC case and Uniform Ammonium/EC case. Three non-uniform cases are also considered and the OC/EC non-uniform cases are detailed here as an example: samples in the lowest 2/3 of the OC/EC range are used to predict samples in the highest 1/3 of the OC/EC range (non-uniform A), the highest and lowest 1/3 of the OC/EC range are used to predict the middle 1/3 OC/EC range (non-uniform B), and the highest 2/3 of the OC/EC range are used to predict samples in the lowest 1/3 of the OC/EC range (non-uniform C). All predictions are based on the hybrid calibration model (Sect. 3.1) such that low EC samples in any of the test sets are predicted using a low EC calibration developed for that case.

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The top row of subplots in Fig. 4 shows that the distribution of OC/EC in the test and calibration sets for the base case and the Uniform OC/EC case are similar. The three non-uniform OC/EC cases have distributions that are different for the test and calibrations sets. The Base and Uniform case have zero bias and similar normalized error in the test and calibration sets indicating good predictions for these cases. When there is a large difference in OC/EC distribution between the test and calibration sets (non-uniform cases), the normalized error is higher and sometimes a bias is induced. For non-uniform C, there is a significant negative bias and the normalized error is about 35 % (10 % higher than the calibration set). Non-uniform B, in which the medians are similar but the distribution of OC/EC is quite different between the sets, there is less than 10 % higher error in the test set than the calibration set. For non-uniform A, in which the median EC value is in the low EC range, the median normalized error is at least 60 % and the range of error is high which are due to both the difference in the chemical composition of the aerosol in the test and calibration sets and the low EC values.

The impact of ammonium is evaluated using uniform and non-uniform calibrations of ammonium/EC (Fig. 5). Like EC, the Base case, Uniform case and non-uniform B case have near zero bias and low normalized error. The results for the non-uniform A and C case are very similar to the non-uniform A case for OC/EC. In the non-uniform C case, the calibration set contains a higher amount of ammonium and EC is under-predicted (some of the EC is assumed to be ammonium) in low ammonium/EC test samples (bias, -0.04 to $-0.06 \mu\text{g m}^{-3}$) and the range in bias of individual samples increases. The normalized error increases by about 15 % from the calibration set to the test set but the error bars are about the same for the two sets. When low ammonium/EC samples are used to predict samples with high ammonium/EC (non-uniform A), the normalized error becomes very large. This is due in part to the median EC value being below $2.4 \mu\text{g}$. It is likely that additional error is induced because the calibration set is not trained to disregard ammonium in the prediction of EC, so some of the ammonium is reported to be EC leading to a positive bias and increased error. The distribution of

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EC, OC/EC and ammonium/EC for the test and calibration sets for the Base, Uniform and non-uniform cases are shown in Sect. S7. When the chemical composition, as indicated by OC/EC and ammonium/EC are different between the calibration and test sets, the predictions have higher error than when the chemical compositions are similar.

3.3 Prediction of specific sites

Calibrations are developed using ambient samples for all but one site in the calibration set. The one site omitted from the calibration is predicted. Figure 6 shows the distributions of EC in the test and calibration set, and the median and interquartile range of bias and normalized error for all sites. Three sites, Mesa Verde, Olympic and Trapper Creek, have median EC mass below $2.4 \mu\text{g}$ and although the bias is near zero, they have the highest median (40 to 60 %) and range of normalized error. As shown with the Low EC calibration and comparison to collocated TOR samples (Sect. 3.1), these errors are similar to collocated precision of TOR samples in the same EC range indicating that the error is due primarily to TOR analytical error. All other sites have higher EC mass and are expected to be predicted well, based on EC mass alone. However, only Proctor Maple Research Facility is predicted well (Fig. 6). EC, OC/EC and ammonium/EC distributions at Proctor Maple Research Facility are similar to the calibration set. The increased errors for the remaining three sites, Phoenix, Sac and Fox and St. Marks are due to differences in aerosol composition between the calibration set and these sites. Distributions of EC, OC/EC and Ammonium/EC for the test and calibration sets for all sites are shown in Sect. S7.

Figure 7 shows the OC/EC and Ammonium/EC distributions at Phoenix and the ammonium/EC distributions for Sac and Fox and St. Marks. Phoenix, an urban site, has higher EC, than the rural sites (Fig. 6). Phoenix also has lower OC/EC than the other sites (OC/EC non-uniform C), and has lower ammonium/EC than the rural sites (ammonium/EC non-uniform C). All of these compositional differences lead to a negative bias and increased error as seen in the Phoenix samples (bias up to $0.03 \mu\text{g m}^{-3}$ and error up to 50 %). All six rural sites have similar OC/EC in the test set and the

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calibration set indicating that OC/EC does not impact the errors for these sites. Sac and Fox and St. Marks have higher ammonium/EC than the other sites (similar to ammonium/EC non-uniform A except that the EC median is higher). As shown in Fig. 7, there is an increase range of bias values and higher error (22 to 40 %) in these sites predictions compared to the calibration set, likely due to ammonium/EC differences.

We can therefore estimate how well a site not included in the calibration will be predicted, based on the EC, OC/EC and ammonium/EC for the site. For low EC sites, the error ranges from 40–60 %. Phoenix, which has higher EC, lower OC/EC and lower ammonium/EC than the rural sites, has biased predictions and error between 20 and 50 %. Differences in ammonium/EC alone produce higher errors up to 40 % for Sac and Fox and St. Marks, especially for the truncated spectral type.

4 Conclusions

Non-destructive, fast and inexpensive FT-IR analysis of routinely collected PTFE filters in the IMPROVE network can be used to predict TOR EC mass in IMPROVE aerosol samples. The FT-IR spectra and parallel TOR EC measurements are used in partial least squares regression to develop calibrations to predict TOR EC. A hybrid approach is used for calibration in which samples with low EC are calibrated with a low EC calibration and all other samples are calibrated with a calibration that spans the range of EC samples. The low EC calibration produces predictions with mean error similar to that of collocated TOR samples in the same mass range and similar MDLs, indicating that the errors in the FT-IR method are primarily due to TOR measurement uncertainty. All three spectral types produce high quality predictions. The hybrid calibrations developed from samples ordered by site-date (base case), EC, OC/EC, and ammonium/EC produce nearly bias-free predictions with low error ($\sim 0.02 \mu\text{g m}^{-3}$ or 20–25 %). Blank filters in the test set are used to calculate MDL, but the number of blanks in the calibration set does not impact the value of the MDL. Using a calibration set that does not have similar composition to the test set (i.e., using samples in the calibration set that

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do not span the full range of EC, OC/EC or ammonium/OC) leads to higher bias and errors and is not recommended for obtaining high quality predictions. In the proposed method, error and bias are kept to a minimum by including samples in the calibration set that have a similar composition as the test set, as indicated by EC, OC/EC and ammonium/EC. Therefore, we conclude that FT-IR spectra calibrated to TOR EC using partial least squares regression is a robust method for predicting TOR elemental carbon from particulate matter samples. Future work includes establishing that the calibration developed here can be used to predict TOR EC for samples collected during other years and developing a calibration that includes samples with a broader range of aerosol composition.

**The Supplement related to this article is available online at
doi:10.5194/amtd-8-6325-2015-supplement.**

Acknowledgements. The authors acknowledge funding from the IMPROVE program (National Park Service cooperative agreement P11AC91045) and EPFL startup funding.

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Table 1. MDL and precision for Hybrid FT-IR EC and TOR EC.

	TOR-EC	Hybrid FT-IR EC raw spectra	Hybrid FT-IR EC baseline corrected spectra	Hybrid FT-IR EC truncated spectra
MDL ($\mu\text{g m}^{-3}$) ^a	0.01 ^b	0.02	0.01	0.02
% below MDL	3	1	2	3
precision ($\mu\text{g m}^{-3}$) ^a	0.11	0.04	0.06	0.06
Mean blank (μg)	NR ^c	0.06 ± 0.17	0.08 ± 0.15	0.12 ± 0.19

^a Concentration units of $\mu\text{g m}^{-3}$ for MDL and precision are based on the IMPROVE volume of 32.8 m^3 .

^b Value reported for network ($0.44 \mu\text{g}$) in concentration units.

^c Not reported.

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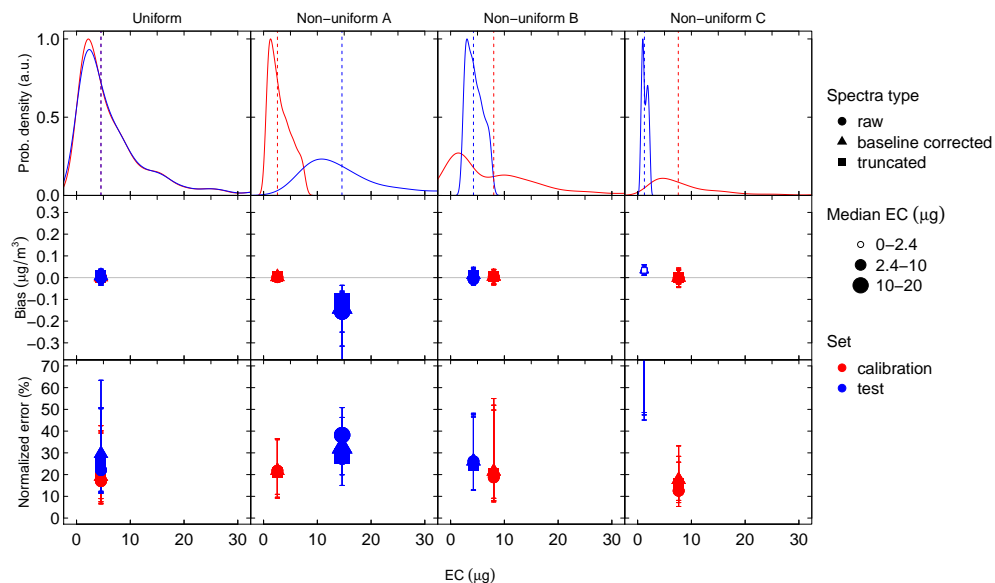


Figure 1. Uniform and non-uniform EC calibrations. The probability density distribution of EC and bias and normalized error (with the interquartile range shown by error bars) in the calibration (red) and test (blue) sets for the Uniform EC case and three non-uniform EC cases. Vertical lines are the median of the EC mass distributions color-coded for calibration and test sets.

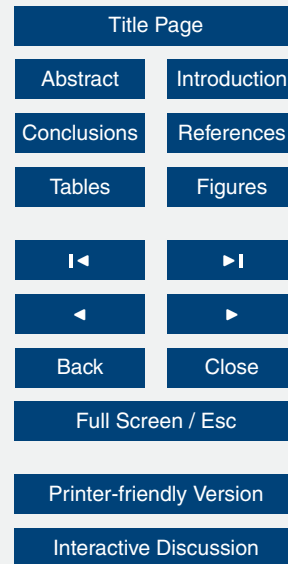


Figure 2. Mean error of a low EC test set (EC < 2.4 µg) and MDL from the Uniform EC and Low Uniform EC calibrations. The same low EC test set is predicted by the Uniform and Low Uniform calibrations. The mean error of collocated EC samples less than 2.4 µg and the reported EC MDL for the TOR method are shown for comparison.

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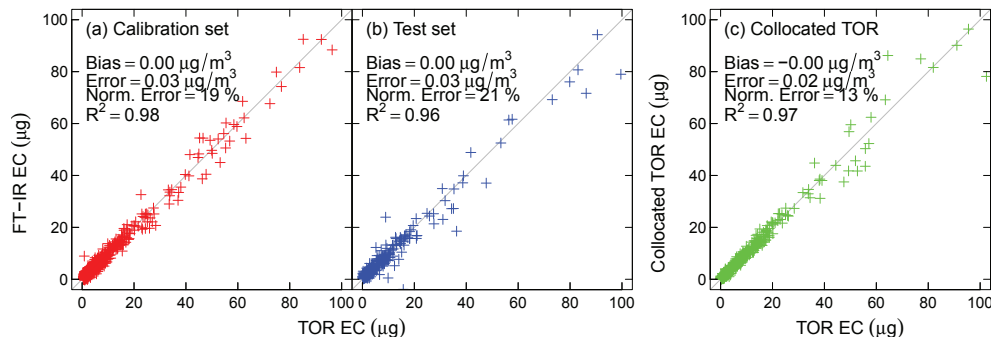


Figure 3. Predicted EC for calibration set **(a)** and test set **(b)** for the base case using a hybrid calibration model. The collocated TOR samples **(c)** are from sites with parallel quartz filters that are both analyzed by TOR. Only the Phoenix site has samples in the calibration, test, and collocated data sets. There are 521 samples in the calibration set **(a)**, 265 samples in the test set **(b)** and 431 samples in the collocated TOR data set **(c)**. Concentration units of $\mu\text{g m}^{-3}$ for bias and error are based on the IMPROVE nominal volume of 32.8 m^3 .

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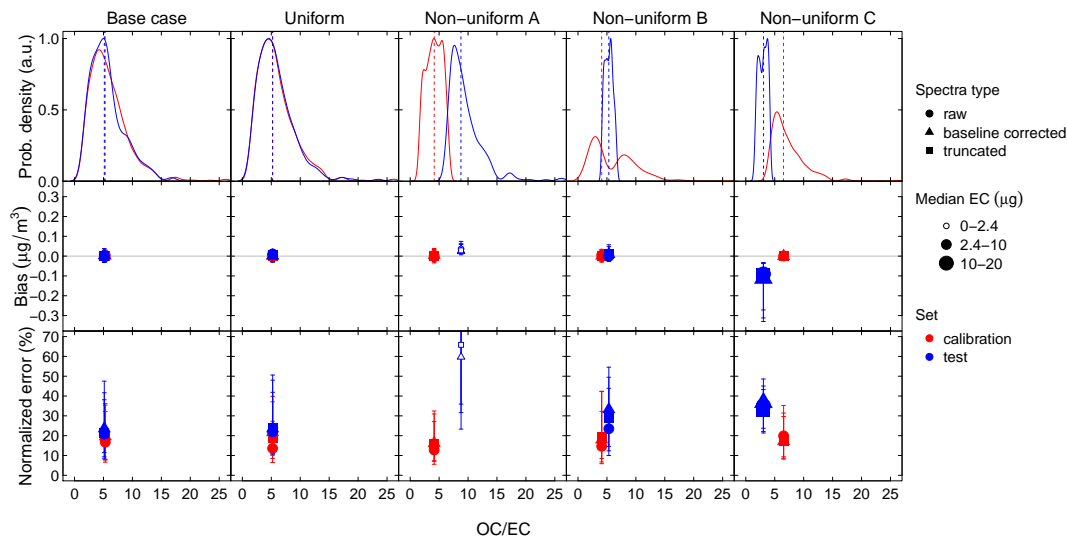


Figure 4. The probability distribution of OC/EC and bias and normalized error (with the interquartile range shown by error bars) in the calibration (red) and test (blue) sets for five hybrid calibration cases: the base case, the uniform OC/EC case and three non-uniform OC/EC cases. Vertical lines on the probability distributions are the color-coded median of the OC/EC distributions.

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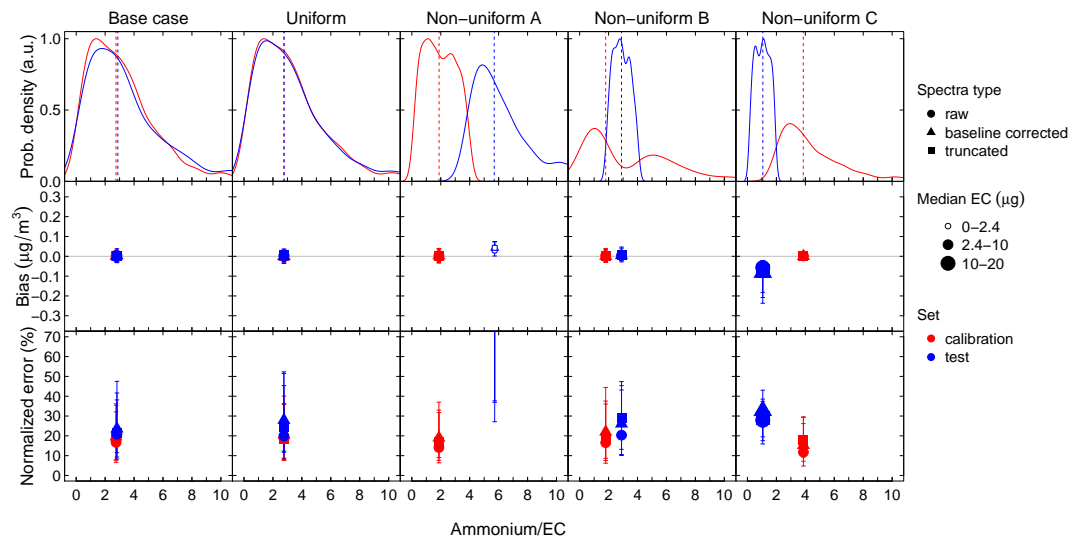


Figure 5. The probability distribution of ammonium/EC and bias and normalized error (with the interquartile range shown by error bars) in the calibration (red) and test (blue) sets for five hybrid calibration cases: the base case, the uniform ammonium/EC case and three non-uniform ammonium/EC cases. Vertical lines on the probability distributions are the color-coded median of the ammonium/EC distributions.

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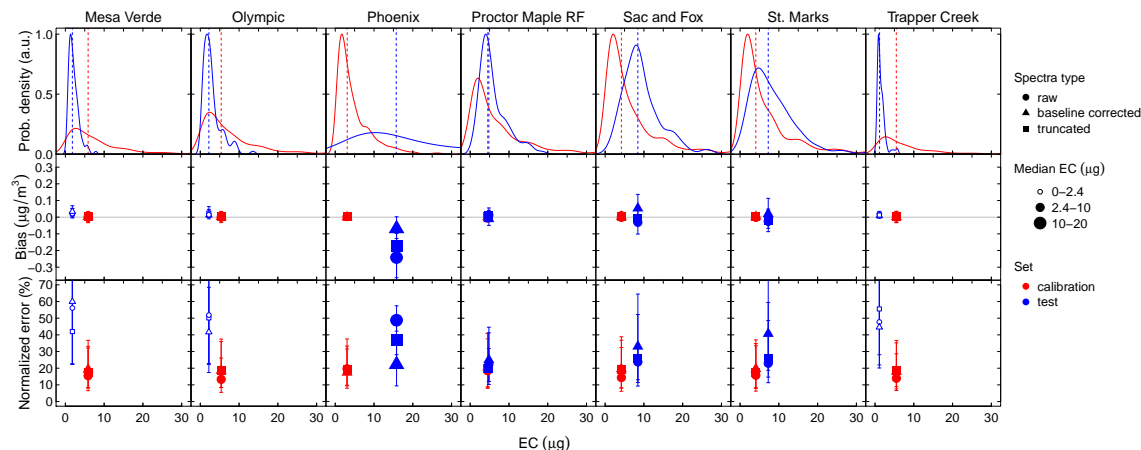


Figure 6. The distribution of EC and bias and normalized error (with the interquartile range shown by error bars) in the calibration (red) and test (blue) sets for calibrations developed for each site. Vertical lines are the median of the EC mass distributions color-coded for calibration and test sets.

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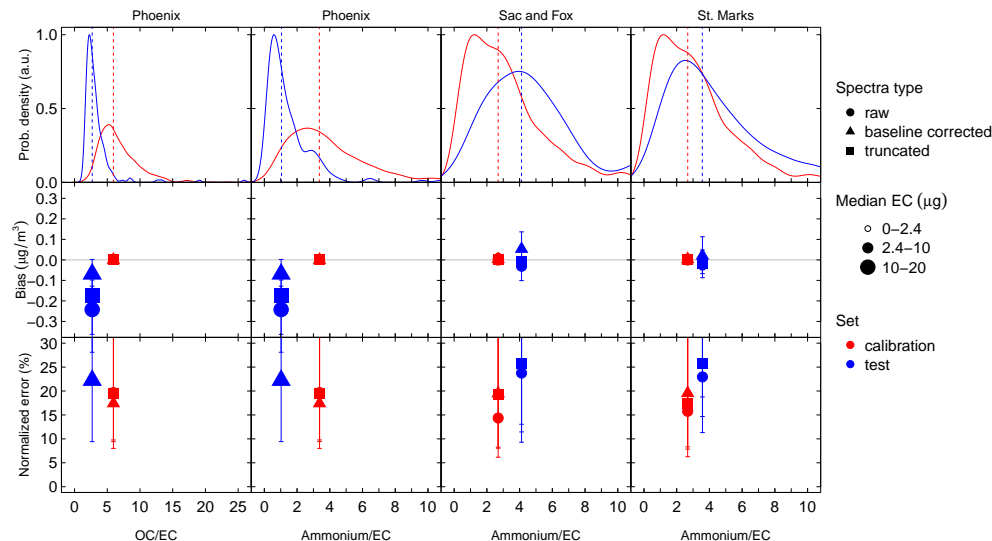


Figure 7. The distribution of OC/EC and bias and normalized error (with the interquartile range shown by error bars) in the calibration (red) and test (blue) sets for calibrations developed for Phoenix and the distribution of ammonium/EC, bias and normalized error for Phoenix, Sac and Fox and St. Marks. Vertical lines are the median of the distributions color-coded for calibration and test sets.

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