

1 **Determining air pollutant emission rates based on mass balance using airborne measurement data**
2 **over the Alberta oil sands operations.**

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13

14 **Abstract**

15
16 Top-down approaches to measure total integrated emissions provide verification of bottom-up,
17 temporally-resolved, inventory-based estimations. Aircraft-based measurements of air pollutants from
18 sources in the Canadian oil sands were made in support of the Joint Canada-Alberta Implementation
19 Plan on Oil Sands Monitoring during a summer intensive field campaign between August 13 and
20 September 7, 2013. The measurements contribute to knowledge needed in support of the Joint Canada-
21 Alberta Implementation Plan on Oil Sands Monitoring. This paper describes a Top-down Emission
22 Rate Retrieval Algorithm (TERRA) to determine facility emissions of pollutants, using SO₂ and CH₄ as
23 examples, based on the aircraft measurements. In this algorithm, the flight path around a facility at
24 multiple heights is mapped to a two-dimensional vertical screen surrounding the facility. The total
25 transport of SO₂ and CH₄ through this screen is calculated using aircraft wind measurements, and
26 facility emissions are then calculated based on the divergence theorem with estimations of box-top
27 losses, horizontal and vertical turbulent fluxes, surface deposition, and apparent losses due to air
28 densification and chemical reaction. Example calculations for two separate flights are presented.
29 During an upset condition of SO₂ emissions on one day, these calculations are within 5% of the
30 industry-reported, bottom-up measurements. During a return to normal operating conditions, the SO₂
31 emissions are within 11% of industry-reported, bottom-up measurements. CH₄ emissions calculated
32 with the algorithm are relatively constant within the range of uncertainties. Uncertainty of the emission
33 rates is estimated as 20%, which is primarily due to the unknown SO₂ and CH₄ mixing ratios near the
34 surface below the lowest flight level.
35

1. Introduction

Aircraft-based measurements have been previously used to derive emission rates from point and area sources of compounds including CO₂, CH₄, CO, NO_x, and SO₂ (See Table 1 for references). This analysis is accomplished by flying downwind and/or around the source, in some cases at multiple heights, and inferring the emissions rate based on a mass-balance analysis. This top-down approach offers an advantage over a bottom-up, inventory-based estimation as it attempts to capture the total integrated emissions, some of which may be missed by inventories or difficult to assess, particularly for large and complex industrial facilities spanning tens to hundreds of square kilometers that are comprised of a large number of activities. Simplifying assumptions may be used in the analysis to reduce the prohibitive cost of aircraft flight time; however, these assumptions result in reduced accuracy of the derived emissions estimates. Flight patterns can be grouped into a) single-height transects, b) upwind and downwind spirals, c) single-screen flights, and d) box flights. In the latter case, the box can refer to a cylinder, a rectangular cuboid, or any other prism shape that is uniform with height.

Table 1. Reported studies of ground source emission estimations from aircraft-based measurements.

Reference	Measurement Technique	Measured Compound(s)
Turnbull et al., 2009	Single Height Transect	CO ₂ , CH ₄ , CO
Peischl et al., 2013	Single Height Transect	CO ₂ , CH ₄ , CO
Karion et al., 2013	Single Height Transect	CH ₄
Wratt et al., 2001	Up and downwind spirals	CH ₄
Gatti et al., 2014	Up and downwind spirals	CO, CO ₂
Mays et al., 2009	Single Screen	CO ₂ , CH ₄
Cambaliza et al., 2013	Single Screen	CO ₂ , CH ₄
Walter et al., 2012	Single Screen (DOAS)	SO ₂
Kalthoff et al., 2002	Box	CO, NO _x
Alfieri et al., 2010	Box	CO ₂

The simplest flight pattern, which we refer to as a single-height transect, is a single flight path at one height perpendicular to the mean wind direction and downwind of the point or area sources (Turnbull et al., 2009; Peischl et al., 2013; Karion et al., 2013). This approach assumes a well-mixed boundary layer, such that the species mixing ratio is constant and equal to the measured value between the surface and the boundary layer height. Upwind of the source, the species mixing ratio is assumed to be equal to a constant background value determined either from the outside edges of the downwind transect (Turnbull et al., 2009), or from a second, upwind transect (Peischl et al., 2013; Karion et al., 2013). Author-derived uncertainties in the calculated emission rate based on this approach are estimated as $\pm 50\%$ (Peischl et al., 2013; Karion et al., 2013). In comparing method uncertainties it is noted that different authors use inconsistent methodologies to estimate total uncertainties, and some estimates are more

1 conservative than others. Hence the relative values of author-derived uncertainties in this section is
2 considered a qualitative comparison only.

3
4 The vertical variation in mixing ratio can be determined by flying in an ascending or descending spiral
5 pattern upwind and downwind of a source (Wratt et al., 2001; Gatti et al., 2014). This gives the total
6 emission rate of a surface line source connecting the two spiral locations. This approach is ideal for
7 large area sources with little variation in emission rate perpendicular to the wind direction.

8 Uncertainties in the calculated emission rate based on this approach are estimated as $\pm 40\%$ (Wratt et al.,
9 2001; Gatti et al., 2014).

10
11 For the single-screen method, horizontal and vertical variation in mixing ratio can be accounted for by
12 flying perpendicular to the mean wind direction and downwind of an area source at multiple heights
13 (Cambaliza et al., 2013; Mays et al., 2009). Each traverse follows the same path above the surface at a
14 different height, which allows the measurements to be interpolated to a two-dimensional screen normal
15 to the mean horizontal wind direction. The upwind, background mixing ratio is estimated from the
16 lateral edges of the screen, which are assumed to be located far enough from the area source to contain
17 no emissions from that source. Uncertainties for this method are conservatively estimated at $\pm 50\%$
18 (Cambaliza et al., 2013). Cambaliza et al. (2013) reanalyzed their results using the single-height
19 transect method and estimated the uncertainty based on that approach as ranging from 23% to 65%. The
20 single screen method can also be approximated by flying at a single height above the boundary-layer
21 and measuring a species profile to the surface using a remote sensing such as a Differential Optical
22 Absorption Spectroscopy (DOAS) instrument (Walter et al., 2012). It is unclear what the uncertainty is
23 based on this approach.

24
25 The box method expands on the screen method by including multiple screens upwind and surrounding
26 the emissions area (Kalthoff et al., 2002; Alfieri et al., 2010). This analysis is accomplished by flying a
27 square (Alfieri et al., 2010) or a polygon (Kalthoff et al., 2002) pattern around the emissions area and
28 repeating the pattern at multiple heights. The box method refers to either cuboid or other prism shapes,
29 although a cylindrical spiral would follow a similar methodology. Species mixing ratios are interpolated
30 between the multiple flight-path heights and extrapolated to the ground to give a two-dimensional
31 screen or wall surrounding the emissions area (the lateral sides of the box). A mass balance approach is
32 then employed to derive the emission rate within the enclosed volume by calculating the total advective
33 fluxes of the emitted material through the surrounding screen. A model comparison (Panitz et al., 2002)
34 of the Kalthoff et al. (2002) study concluded that the advective fluxes account for between 85% and 95%
35 of the total emissions, suggesting a much lower uncertainty compared to the single height transect or
36 single screen methods described above.

37
38 In this paper, we present an algorithm for calculating the emissions from an area source using the box
39 method. We attempt to improve upon the analysis of Kalthoff et al. (2002) and Alfieri et al. (2010) by
40 investigating all possible sources of error and through modified extrapolation of the measurements to

1 the near surface, below the lowest flight path. We have named this improved algorithm the Top-down
2 Emission Rate Retrieval Algorithm (TERRA). Aircraft-based measurements of air pollutants were
3 made during a summer intensive field campaign between August 13 and September 7, 2013 and support
4 the Joint Canada-Alberta Implementation Plan on Oil Sands Monitoring. We use SO₂ and CH₄ to test
5 TERRA from two flights around a facility in the oil sands region on two separate days. Using TERRA
6 with the appropriately designed flight paths allows us to demonstrate improvements on the uncertainties
7 in emission estimates compared to the previously reported aircraft top-down emission estimate methods.
8 SO₂ and CH₄ are chosen as example in these analyses as they represent respectively emissions from a
9 stack source with a low background level and emissions from ground area source with a relatively high
10 background level. Uncertainties due to the method of interpolation and extrapolation are estimated
11 using one of the flight paths with simulated plumes. Sensitivity of the estimation to uncertainties in
12 mixing ratio, wind speed, and various algorithm parameters is analyzed, and SO₂ emission rates are
13 compared to industry reported, bottom-up measurements.

15 **2. Methods**

17 **2.1 Aircraft and Instrumentation**

19 Instrumentation was installed aboard the National Research Council of Canada Flight Research
20 Laboratory (NRC-FRL) Convair-580 research aircraft. The Convair-580 is equipped to measure three-
21 component wind speed (U_x , U_y , w), and temperature (T) with a Rosemount 858 probe (sampled at 32
22 Hz). Aircraft state parameters (latitude y , longitude, x , and ellipsoid height altitude, z) are measured by
23 GPS and Honeywell HG1700 Inertial Measurement Unit (IMU). Kalman filtering of the integrated
24 IMU data is combined with the GPS data to provide state parameters at a rate of 100 Hz. The wind
25 speed measurement uncertainty is estimated as 0.4 m s⁻¹ (Khelif et al., 1999). Dewpoint temperature
26 (T_d) is measured with an Edgetech Hygrometer, and pressure (P) is measured with a DigiQuartz sensor.
27 Additional instrumentation installed specifically for this study comprised a comprehensive suite of fast
28 response instruments to measure gases and aerosols. This paper uses measurements of SO₂ and CH₄ to
29 demonstrate the mass balance approach to estimate emission rates.

31 SO₂ measurements were made with a Thermo Scientific 43iTLE analyzer at a rate of 1 Hz. The SO₂
32 instrument was calibrated three times throughout the project over a range of 0 to 400 ppb. The standard
33 deviation of the three calibration slope measurements, which can be used to quantify long-term drift is
34 0.9%. The average standard deviation of the 1 Hz data during calibrations, which demonstrates short-
35 term variability, is 1.99 ppb.

37 CH₄ measurements were made with a Picarro G2204 cavity-ring-down spectrometer (Picarro, Inc.) at a
38 variable acquisition rate of approximately 0.3 Hz. The CH₄ instrument was calibrated five times
39 throughout the project, over a range of 2 to 3 ppm. The standard deviation of the five calibration slope
40 measurements is 1.3%. The standard deviation of the 0.3 Hz data during calibrations was 0.33 ppb.

1
2 The time delay of the instruments (relative to the wind speed and aircraft state parameter measurements)
3 was measured using automated switching in laboratory experiments with the same inlet systems that
4 were used on the aircraft (including all inlet plumbing configurations). The total delay including inlet
5 flow and instrument response time was 6 seconds for the SO₂ instrument and 8 seconds for the CH₄
6 instrument.

7
8 To consolidate the various measured parameters, high-frequency data (wind and state parameters) were
9 averaged to a frequency of 1 Hz, while low-frequency data (CH₄ were linearly interpolated to a
10 frequency of 1 Hz. The Convair has a cruising speed of 90 m s⁻¹, which gives a spatial resolution of 90
11 m for SO₂ and 270 m for CH₄.

12 13 **2.2 Study Area**

14
15 The aircraft flew a total of 22 flights over the Athabasca oil sands region in northern Alberta between
16 August 13 and September 7, 2013. Thirteen flights included area emissions investigations, comprising a
17 total of 21 box flights around 7 separate oil sands facilities, mostly surface mining operations. Each box
18 flight path was designed to include one facility only and box flights were only done during directionally
19 consistent winds with speeds > 5 m s⁻¹. For this paper, data from two flights surrounding the Canadian
20 Natural Resources Limited (CNRL) Horizon oil sands mining and upgrading facility are analyzed in
21 order to construct an algorithm for the box method and to estimate the uncertainties in the resulting
22 emission rates.

23
24 The CNRL Horizon processing facility is located near 57.34 N, 111.75 W, approximately 4 km west of
25 the Athabasca River and 70 km NNW of Fort McMurray, Alberta, Canada. It is a relatively isolated
26 facility, with only boreal forest to the west and north for hundreds of km. Production at the CNRL
27 Horizon facility in 2013 was approximately 100,000 oil barrels per day (5.8×10^9 L yr⁻¹,
28 www.cnrl.com, 2014). In 2012, the Horizon facility emitted an average of 6.7 t d⁻¹ (metric tonnes per
29 day) of SO₂ (Canadian Natural Resources Ltd, 2014). The flight dates of Aug 20 and Sep 2 were chosen
30 as comparative tests of the emissions algorithm because the SO₂ scrubbing unit was temporarily offline
31 on Aug 20. During this event, CNRL reported an average of 12.9 t h⁻¹ of SO₂ released for the 6-hour
32 period from 12:00 to 18:00 (LT, MDT) on Aug 20, which is compared to a CNRL reported release of
33 0.17 t h⁻¹ of SO₂ for the 6-hour period from 12:00 to 18:00 LT on Sep 2, during normal SO₂ scrubber
34 operation.

35
36 Average meteorological conditions during the flights were determined from a 167 m tall tower and a 75
37 m tall tower, (<http://www.wbea.org>), both located approximately 40 km from the CNRL facility
38 (AMS03: 57.032 N, 111.505 W). During the Aug 20 flight average temperature, wind speed, and
39 direction were 15.2°C, 6.3 m s⁻¹, and 235°. During the Sep 2 flight average temperature, wind speed,
40 and direction were 18.2°C, 8.3 m s⁻¹, and 204°. A more detailed analysis of wind conditions during the

1 flights is given in Section 4.2. The bulk Richardson number (Garrett, 1994) calculated during each
 2 flight demonstrates unstable conditions during both flights ($R_i = -20$ on Aug 20, and $R_i = -6$ on Sep 2).

4 2.3 Mass Conservation Equations

5
 6 Emissions are determined by flying in a pattern that approximates a rectangular box shape surrounding
 7 the facility area. Some other flights during the airborne study used five-sided polygon shapes. The
 8 number and orientation of the box walls has no effect on the analysis discussed herein. For simplicity,
 9 the walls of the box for the Aug 20 and Sep 2 flights were aligned with the north, south, east, and west
 10 directions, regardless of wind direction. Figure 1 illustrates the path of the Aug 20 and Sep 2 flights
 11 from Fort McMurray to the facility and the box surrounding the facility. The box walls are
 12 approximately 5 to 10 km from the edges of the facility boundaries. The Aug 20 flight also included
 13 two profiles from spirals in the north-east (downwind) and south-west (upwind) corners of the box as
 14 well as three north-south transects over the facility. The south-west spiral was flown first, then the box
 15 flight around the facility, followed by the north-east spiral, and finally the north-south transects. The
 16 Sep 2 flight includes two spiral profiles south-east (downwind) of the box, a second profile near the east
 17 wall (upwind) of the box, and two north-south transects over the facility. One south-east spiral was
 18 flown first, then the box flight around the facility, followed by the east wall spiral, the north-south
 19 transects, and finally the second south-east spiral.

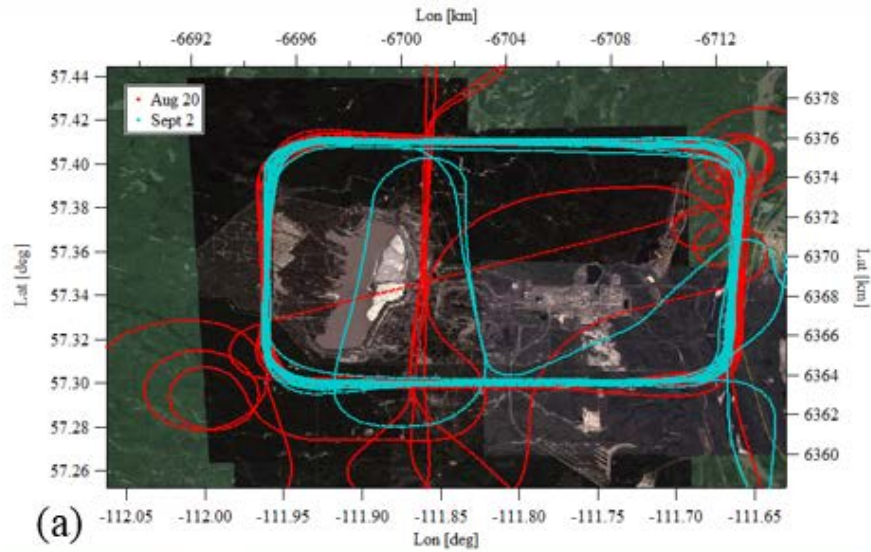
20
 21 The following Sections describe the Top-down Emissions Rate Retrieval Algorithm (TERRA)
 22 developed herein and used to calculate emission rates from these flight data. Area emission rates are
 23 estimated using the Divergence Theorem, which equates the change in mass within a control volume
 24 with the integrated mass flux through the walls of the control volume. This gives a mass balance in the
 25 control volume for a given compound (C) of

$$26 \quad E_C = E_{C,H} + E_{C,HT} + E_{C,V} + E_{C,VT} + E_{C,VD} - E_{C,M} - E_{C,X}, \quad (1)$$

27 where E_C is the total emissions rate integrated over all activities within the facility, $E_{C,H}$ is the horizontal
 28 advective flux through the box walls, $E_{C,HT}$ is the horizontal turbulent flux through the box walls, $E_{C,V}$ is
 29 the advective flux through the box top, $E_{C,VT}$ is the turbulent flux through the box top, $E_{C,VD}$ is the
 30 deposition to the surface, $E_{C,M}$ is the increase in mass within the volume due to a change in air density,
 31 and $E_{C,X}$ is the increase in mass due to chemical changes of the compound within the box volume. For
 32 comparative purposes $E_{C,H}$ can be separated into inwards and outwards fluxes such that $E_{C,H} = E_{C,H,out} -$
 33 $E_{C,H,in}$, where subscript out denotes horizontal advective flux leaving the box and in denotes horizontal
 34 advective flux entering the box. Similarly, the mass balance for air in the control volume is

$$35 \quad 0 = E_{air,H} + E_{air,V} - E_{air,M}, \quad (2)$$

36 where $E_{air,H}$ is the horizontal advective flux of air, $E_{air,V}$ is the box-top advective flux of air, and $E_{air,M}$ is
 37 the change in air mass within the volume. The horizontal advective flux can also be separated as $E_{air,H} =$
 38 $E_{air,H,out} - E_{air,H,in}$.

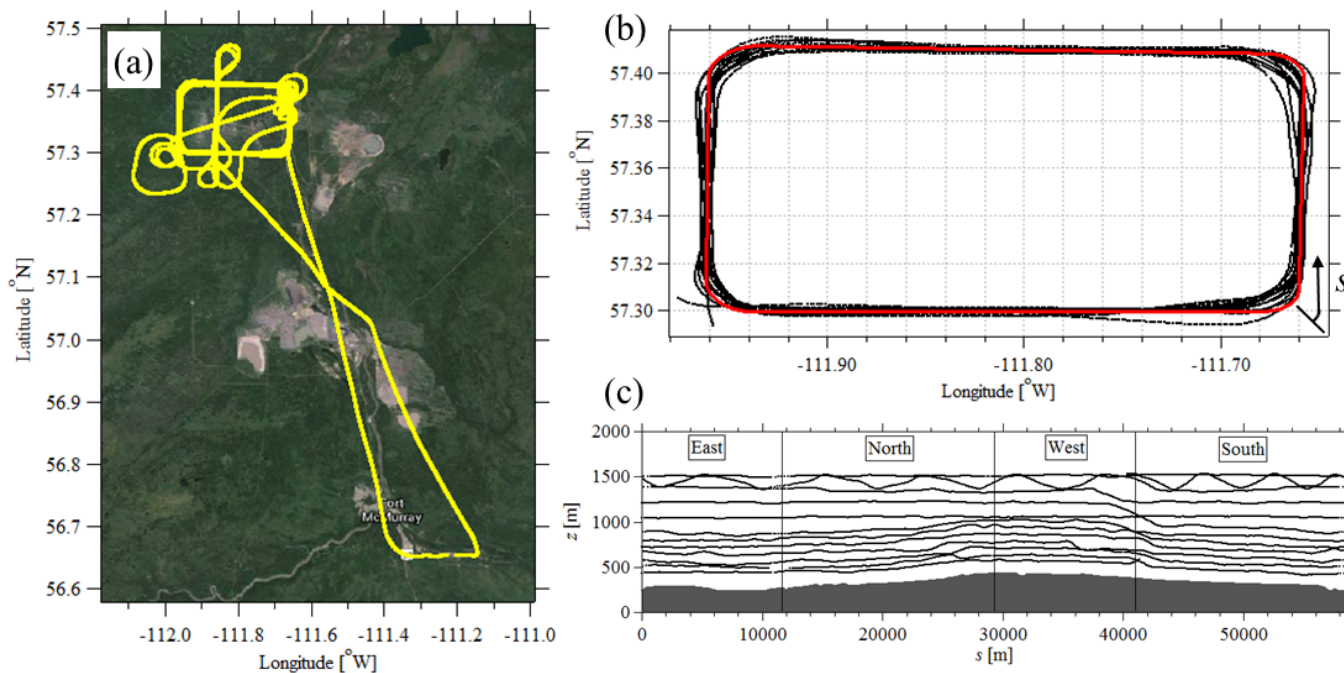


1
 2 Figure 1. A composite Google image (a) shows the path of the Aug 20 (red) and Sep 2 (cyan) flights. Google
 3 Earth images demonstrate the path of (b) the Aug 20 flight and (c) the Sep 2 flight. The yellow arrow shows wind
 4 direction, the blue arrow shows north. Map image data provided by CNES/SPOT, Digital Globe, and Google.

5 **2.4 Position Mapping and Interpolation**

6
 7 Figure 2 demonstrates the process by which the 1-s flight position data for the Aug 20 flight are mapped
 8 to the 2-dimensional screen, which comprises the lateral box walls. First the box flight position data are

1 separated from the other flight sections (e.g. to and from the airport, spirals, transects) by visual
 2 inspection. The flight time for the Aug 20 box was 2 hour and 10 min and the flight time for the Sep 2
 3 box was 1 hour and 48 min. In each case, a single horizontal path with four linear components
 4 (corresponding to the east, north, west, and south walls) is determined using a least-squares fitting as a
 5 function of latitude and longitude (Fig. 2b). The corners of the box path are rounded with a turning
 6 radius to produce a smooth path without discontinuities, which further allows a proper calculation of
 7 wind speed normal to the curved path at the corners (see Section 2.5). The start of the horizontal path is
 8 arbitrarily defined as the south-east corner and the horizontal path distance (s) increases in a counter
 9 clockwise direction. The selection of the starting position for the horizontal path has no effect on the
 10 overall calculation. Each 1-s aircraft position datum during the box flight is then mapped to the closest
 11 position on the least-squares path fit. This procedure results in a translation of each flight position point
 12 from a 3-dimensional position of latitude (y), longitude (x), and altitude (z , above mean sea-level) to a 2-
 13 dimensional screen position of horizontal path distance $s=f(x,y)$, and altitude, z , as shown in Fig. 2c.
 14 Herein the term screen is used to refer to the full unwrapped composite of the four walls (with
 15 dimensions $s \times z$), whereas wall refers to each of the four box sides.
 16



17
 18 Figure 2. The mapping of aircraft position during the box flight to the box walls for the Aug 20 flight showing (a)
 19 the complete flight path, (b) the box flight aircraft position data (black dots) and least squares fit (red line), and
 20 (c) the unwrapped screen in the horizontal path length (s) and height (z) dimensions. The ground elevation ($z_g(s)$)
 21 beneath the flight path fit is shown in grey shading (c).

22
 23 The wind speeds are separated into northerly and easterly components ($U_N(s,z)$, $U_E(s,z)$). The air density
 24 ($\rho_{air}(s,z)$) is calculated at each aircraft position as (Rogers and Yau, 1996)

$$\rho_{air} = \frac{P}{RT(1+0.6\chi_{H2O})}, \text{ with } \chi_{H2O} = \frac{A_d \varepsilon}{P} \exp\left(\frac{T_d}{B_d}\right), \quad (3)$$

where $R = 287.1 \text{ J kg}^{-1} \text{ K}^{-1}$, χ_{H2O} is the water vapour mixing ratio, $A_d = 3.41 \times 10^9 \text{ kPa}$, $B_d = 5420 \text{ K}$, $\varepsilon = 0.622$, and T , p , and T_d are the measured temperature, pressure, and dew-point temperature, respectively. Five interpolated s - z screens are created for each flight: U_N , U_E , ρ_{air} , SO_2 mixing ratio (χ_{SO_2}), and CH_4 mixing ratio (χ_{CH_4}).

Interpolation of the screens can be done with a variety of methods. Three techniques are compared using simulate plumes in Section 3.2: inverse distance weighting (IDW), natural neighbour (Sibson, 1981), and kriging (Isaaks and Srivastava, 1989). Each technique calculates the interpolated values ($\chi(s,z)$) as a weighted average of surrounding points (χ_i) giving $\chi(s,z) = \Sigma(\lambda_i \chi_i)$, with weights $\Sigma\lambda_i = 1$. For IDW, each point in the interpolated image is weighted as the inverse distance to a given power. Initial trials determined that a fourth power gives the best results, which gives $\chi(s,z) = \Sigma d_i^4 \Sigma(\chi_i d_i^{-4})$, where d_i is the distance between the interpolation location and each surrounding data point.

Natural neighbour interpolation creates a Voronoi diagram from the discrete data points. Each point in the interpolated image is used to create an overlapping Voronoi pattern with the surrounding measured data points. The value of the interpolated point is then calculated as a weighted sum average of the surrounding points with weighting equal to the amount of overlap between Voronoi patterns. For this analysis we use the Voronoi image interpolation function from Igor Pro data analysis software (Wavemetrics Inc.).

Kriging requires an approximation of the semivariance, $\gamma(d)$ (half the calculated variance), which is a measure of the variation in measured data points as a function of distance (d) between the points. Here we use what is termed ‘‘simple kriging’’. Each weight is calculated as $\lambda_i = \mathbf{K}^{-1} \mathbf{k}$, where \mathbf{K} is a 2-dimensional matrix with values $\gamma(\infty) - \gamma(d_{i,j})$, and \mathbf{k} is a 1-dimensional matrix with values $\gamma(\infty) - \gamma(d_i)$. Here $d_{i,j}$ is the distance between measured points i and j , and d_i is the distance between the measured point i and the interpolation location.

Interpolation is done to a resolution of $\Delta s = 40 \text{ m}$ and $\Delta z = 20 \text{ m}$. All extrapolation between the lowest flight path and the surface is removed as the lack of known boundary conditions near the surface leads to erroneous results, including potentially negative mixing ratios. These removed data (typically a vertical gap of approximately 150 m) are filled using a method dependent on the measured variable. The methods used to fill this gap are discussed in Section 3.1

2.5 Emissions Algorithm

1 The terms of Eqs. 1 and 2 are listed in Table 2 in order of necessary operation to calculate the total
 2 emission rate, E_C (where C represents SO_2 or CH_4). The terms are expanded to their integral solutions.
 3 M_R is the ratio of the compound molar mass to the molar mass of air, which is $64.07/28.97$ for SO_2 and
 4 $16.04/28.97$ for CH_4 . Other variables specific to individual terms are discussed below.

5
 6 The first term ($E_{air,H}$) is the integrated horizontal advective flux of air mass through the screen. This
 7 term is evaluated using the interpolated and surface-gap filled screens of $U_N(s,z)$, $U_E(s,z)$, and $\rho_{air}(s,z)$.
 8 Since the screen path length, s , is a function of longitude, x , and latitude, y , the normal wind vector (U_{\perp}
 9 (s,z), positive outwards) is calculated through cross-multiplication as

$$10 \quad U_{\perp} = \frac{U_N ds/dx + U_E ds/dy}{\sqrt{(ds/dx)^2 + (ds/dy)^2}}. \quad (4)$$

11 The use of a smooth path length with rounded corners (Fig 2b) allows the lateral flux to be calculated
 12 continuously, including the corner locations. The sign of U_{\perp} is used to separate $E_{air,H}$ into $E_{air,H,in}$ and
 13 $E_{air,H,out}$.

14
 15 The change in air mass within the volume ($E_{air,M}$) is the rate of air mass added to or subtracted from the
 16 total box volume due to change in air density with time. The change in air density is dependent on the
 17 rate of change of temperature and pressure. This term can be estimated by taking the time derivative of
 18 the ideal gas law (see appendix) and integrating the density term with height to give

$$19 \quad E_{air,M} = \iiint \frac{d\rho_{air}}{dt} dx dy dz = \frac{A}{\Delta t} \left(\frac{\Delta p}{p} - \frac{\Delta T}{T} \right) \int \rho_{air} dz, \quad (5)$$

20 where A is the area enclosed by the box, p and T are the average pressure and temperature, and Δp and
 21 ΔT are change in pressure and temperature over the duration of the box flight (Δt). The average pressure
 22 and temperature are approximated as independent of height for this preliminary estimation.

23
 24 From the estimations of $E_{air,H}$ and $E_{air,M}$, the remaining term of Eq. 2 ($E_{air,V}$), which represents the
 25 integrated air mass flux through the top of the box, can be calculated and substituted into $E_{C,V}$ of Eq. 1.
 26 This calculation gives the vertical wind speed at the box-top (w , positive upwards). If it can be
 27 demonstrated that the compound mixing ratio at the top of the box ($\chi_{C,Top}$) is nearly constant, the $E_{air,V}$
 28 term gives the integrated compound mass flux through the box top.

29
 30 The $E_{C,H}$ term in Eq. 1, which represents the integrated lateral mass flux of a compound, can then be
 31 solved using the interpolated and surface-gap filled screens of $U_N(s,z)$, $U_E(s,z)$, $\rho_{air}(s,z)$, and the mixing
 32 ratios of $\chi_{\text{SO}_2}(s,z)$ or $\chi_{\text{CH}_4}(s,z)$. As with the air mass flux, the normal wind vector ($U_{\perp}(s,z)$) is calculated
 33 from Eq. 4.

34
 35 Table 2. Terms from Eqs. 1 and 2 used to solve for the total emission rate, E_C . The necessary input variables are
 36 listed with their functional dependence. See text for explanation of variables.

Term	Integral	Description	Input variables
$E_{air,H}$	$\iint_{Sides} \rho_{air} U_{\perp} ds dz$	Integrated horizontal advection of air mass	$U_N(s,z), U_E(s,z), \rho_{air}(s,z), s(x,y)$
$E_{air,M}$	$\iiint_{Volume} \frac{d\rho_{air}}{dt} dx dy dz$	Change in air mass within volume	$T(t), p(t), \rho_{air}(z)$
$E_{air,V}$	$\iint_{Top} \rho_{air} w dx dy$	Integrated advection of air mass through the box top	$E_{air,H}, E_{air,M}$
$E_{C,V}$	$M_R \chi_{C,Top} \iint_{Top} \rho_{air} w dx dz$	Integrated advection of SO ₂ or CH ₄ mass through the box top	$\chi_{C,Top}, E_{air,V}$
$E_{C,H}$	$M_R \iint_{Sides} \chi_C \rho_{air} U_{\perp} ds dz$	Integrated horizontal advection of SO ₂ or CH ₄ mass	$U_N(s,z), U_E(s,z), \rho_{air}(s,z), s(x,y)$ $\chi_{SO_2}(s,z)$ or $\chi_{CH_4}(s,z)$
$E_{C,HT}$	$-M_R \iint_{Sides} K_x \frac{d\chi_C}{dx_{\perp}} \rho_{air} ds dz$	Integrated horizontal turbulent flux of SO ₂ or CH ₄ mass	Estimated value with $\chi_{SO_2}(s,z)$ or $\chi_{CH_4}(s,z)$
$E_{C,VT}$	$M_R \iint_{Top} \Delta \chi \rho_{air} w_e dx dy$	Integrated turbulent flux of SO ₂ or CH ₄ mass through the box top	Estimated value with $\chi_{SO_2}(s,z_{Top})$ or $\chi_{CH_4}(s,z_{Top})$
$E_{C,VD}$	$M_R \iint_{Bottom} \chi_{sur} \rho_{air} V_D dx dy$	Deposition rate of SO ₂ or CH ₄ mass to the surface	Estimated value with $\chi_{SO_2}(s,z_g)$ or $\chi_{CH_4}(s,z_g)$
$E_{C,M}$	$M_R \iiint_{Volume} \chi_C \frac{d\rho_{air}}{dt} dx dy dz$	Change in SO ₂ or CH ₄ mass with time within volume	Estimated value with $T(t), p(t)$, and $\rho_{air}(z)$
$E_{C,X}$	$M_R \iiint_{Volume} \frac{d\chi_C}{dt} \rho_{air} dx dy dz$	Change in SO ₂ or CH ₄ mixing ratio with time within volume	Estimated with wind speed, chemistry, and source location

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The remaining terms ($E_{C,HT}$, $E_{C,VT}$, $E_{C,VD}$, and $E_{C,M}$) require varying degrees of estimation and their solution is dependent on knowledge of the emissions behaviour and distribution of concentration within the box volume. The results of Panitz et al. (2002) demonstrated the potential relative importance of these terms. Panitz et al. (2002) used a 3-dimensional KAMM/DRAIS model to evaluate the box method for CO and NO_x emissions derived from two flights over a city. The model predicted horizontal turbulent fluxes ($E_{C,HT}$) no greater than 0.3% of the total emission rate (E_C) for either CO or NO_x. The vertical turbulent fluxes through the box top ($E_{C,VT}$) were predicted to be 0.3% E_C for CO and 0 for NO_x on one flight and 13% E_C for CO and 6.3% E_C for NO_x on the other flight, with the high ratios of the second flight likely due to a strong modelled inversion near the box top. Deposition was more consistent with $E_{C,VD}$ between 2.6% and 3% E_C for CO and between 5.0% and 6.7% E_C for NO_x. The change in mass due to temperature and pressure changes was not explicitly stated; however the total change (final box-volume concentration – initial box concentration) was 11.5% and 8.8% E_C for CO and NO_x on first flight and 3.5% and 3.8% E_C for CO and NO_x on the second flight.

1 The horizontal turbulent flux ($E_{C,HT}$) is proportional to the horizontal diffusion constant (K_x) and the
 2 negative change in concentration with downwind distance normal to the screen (x_\perp). A Gaussian plume
 3 from an elevated source can be assumed to expand approximately linearly with downwind distance
 4 (Seinfeld and Pandis, 1998). If it can be assumed that the plume is the only source of horizontal
 5 advective flux ($E_{C,H}$), and the wind direction is perpendicular to the screen wall, the ratio of the
 6 horizontal turbulent flux to horizontal advection simplifies to

$$7 \quad E_{C,HT} / E_{C,H} = 2 K_x / x_\perp U, \quad (6)$$

8 where x_\perp is distance downwind of the source, and U is the mean wind speed. For unstable conditions,
 9 which is typical for the summer afternoon flight times, the diffusion constant can be estimated as $K_x =$
 10 $0.1 h^{3/4} (-\kappa L)^{-1/3} u^*$, where h is the boundary-layer height, $\kappa = 0.4$, L is the Mono-Obukhov length, and
 11 u^* is the friction velocity (Seinfeld and Pandis, 1998).

12
 13 Vertical turbulent fluxes ($E_{C,VT}$) will occur at the box-top if there is an inversion near the box-top height.
 14 Following Alfieri et al. (2010), the integrated vertical flux due to an inversion step change $\Delta\chi$ can be
 15 approximated as $E_{C,VT} = A \rho_{air} w_e \Delta\chi$, where A is the box-top area and w_e is an entrainment rate, which
 16 Alfieri et al. (2010) estimated as 0.01 to 0.03 m s^{-1} . The determination of $\Delta\chi$ requires investigation of
 17 flight spirals that traverse the boundary-layer height.

18
 19 The calculation of deposition to the surface ($E_{C,VD}$) requires an estimation of the deposition velocity (V_D)
 20 and knowledge of the mixing ratio at the surface (χ_{sur}) throughout the box. The estimation of surface
 21 mixing ratio is discussed in Section 3.1. As a rough estimate, the deposition rate can be approximated
 22 as $E_{C,VD} = A \rho_{air} V_D \chi_{sur}$, where χ_{sur} is the average mixing ratio at the surface (i.e. $\chi_{sur} = \int \chi(s, z_g) ds / \int ds$).

23
 24 The change in species (SO_2 or CH_4) mass within the volume ($E_{C,M}$) is the rate of species mass added to
 25 or subtracted from the total box volume due to change in air density with time. Previous mass-balance
 26 approaches (see Table 1 for references) have ignored this and the $E_{air,V}$ term, typically with the
 27 justification that meteorological conditions are nearly constant during early afternoon hours when the
 28 flights were done. The term cannot be estimated directly from measurement as the distribution of
 29 mixing ratio within the box volume is unknown. It can be approximated, following Eq.4, as

$$30 \quad E_{C,M} = M_R \iiint \chi_C \frac{d\rho_{air}}{dt} dx dy dz = \frac{AM_R}{\Delta t} \left(\frac{\Delta p}{p} - \frac{\Delta T}{T} \right) \int \chi_C(z) \rho_{air} dz, \quad (7)$$

31 where $\chi_C(z)$ is approximated as the average screen mixing ratio around the box walls (i.e. $\chi_C(z) =$
 32 $\int \chi_C(s, z) ds / \int ds$).

33
 34 The change in species mass within the volume ($E_{C,X}$) is the rate of species mass created or lost due to
 35 chemical reaction (assuming the emissions are at steady state). If an exponential decay of concentration
 36 due to a chemical reaction is assumed, the magnitude of $E_{C,X}$ can be estimated as $E_{C,X} / E_{C,H} =$
 37 $\exp(-t_0/\tau) - 1$ (the negative result indicates a loss of concentration). Here t_0 is the time the species
 38 spends within the box and τ is the lifetime of the species.

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3. Results

3.1 Near-Surface Extrapolation

Because the lowest flight path ($z_L(s)$) was typically near 150 m above ground level ($z_g(s)$), and there were no ground level measurements along the flight paths, there is a gap in measurement data between the surface and the lowest flight altitude. For many of the studies listed in Table 1, a well-mixed layer below the lowest flight altitude is assumed. Because surface values are unknown, this can lead to unquantified uncertainties. For both surface based and stack emission sources, without constraints of surface measurements along the box walls, this lack of near-surface measurements may lead to large uncertainties in the emission rate estimations based on the interpolation schemes. To reduce these uncertainties, we estimate variables near the surface region with an extrapolation scheme based on known boundary layer meteorological empirical approximations.

3.1.1 Wind Speeds

From flux-gradient relations, it can be shown that wind speeds follow a stability dependent log profile (Garrett, 1996) which can be compared to a least squares fit of U to $\ln(z)$ as

$$U(z) = \frac{u_*}{\kappa} \left(\ln \left(\frac{z - z_g - d}{z_0} \right) - \Psi \right) = \frac{u_*}{\kappa} \ln(z - z_g - d) + f(u_*, z_0, \Psi). \quad (8)$$

Here u_* is the friction velocity, $\kappa = 0.4$, z is the flight altitude, z_g is the ground height beneath the flight path, d is a displacement height and z_0 is the roughness length, which are both characteristic of the terrain and surface characteristics, and Ψ is a stability correction, which depends on atmospheric conditions. The terms of the equation which are independent of height are grouped into a least-squares fit parameter f .

The displacement height, d , and the fit parameter, f (which incorporates friction velocity, roughness length, and stability), is approximated using measurements from a nearby WindRASS (Scintec) acoustic profiler. The profiler was located in Fort Mckay during the project at 57.19 N, 111.64 W, approximately 18 km SSE of the flight tracks. The profiler measures winds from a height above ground of 40 m to as high as 800 m (in ideal conditions) in 15-min averages. During the Aug 20 and Sep 2 flight times (09:58 to 13:34 and 11:18 to 14:43 LT, respectively) the maximum profiler measurement height ranged from 220 m to 450 m above ground level. For consistency, we limit the data to a height of 220 m, since we are interpolating only the lowest 150 m of the wind screens. The wind measurements during the Aug 20 and Sep 2 flight times were averaged and a least squares fit to Eq. 8 was determined. This fitting gives values of $d = 6.0$ m, $u_* = 0.60$ m s⁻¹, and $f = -2.64$ m s⁻¹ for the Aug 20 data and $d = 3.1$ m, $u_* = 0.68$ m s⁻¹, and $f = -1.87$ m s⁻¹ for the Sep 2 data. Although displacement height, d , should be constant, the difference is small relative to the vertical resolution of the interpolated

1 wind screens ($\Delta z = 20$ m). Comparing these averaged fits with the 15-min measurements (over the 40 m
2 to 220 m height range) for the same time periods gives root-mean squared errors of 0.78 m s^{-1} and 1.16
3 m s^{-1} for wind speeds for the Aug 20 and Sep 2 flight times respectively.

4
5 To interpolate the wind speeds between the surface and the lowest flight height, the friction velocity is
6 determined from each interpolated s - z wind screen. At each s location (with resolution $\Delta s = 40$ m), Eq.
7 8 is solved for u^* with wind data from the lowest flight path ($U(z_L)$, where typically $z_L - z_g \cong 150$ m), and
8 d and f values as determined above. Wind speed data are then filled in at each s location for $z_g < z < z_L$
9 from Eq. 8.

11 3.1.2 Air Density

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13 Although air density varies exponentially with height (amsl), at low altitudes (less than several km), it
14 can be approximated with a linear dependence on altitude ($\rho_{air}(z) = a + b z$). The measured air density
15 from the Aug 20 flight varies linearly with z and correlates as $r^2 = 0.993$ ($a = 1.184 \text{ kg m}^{-3}$ and $b = -$
16 $1.0 \times 10^{-4} \text{ kg m}^{-4}$), and the measured air density from the Aug 20 flight varies linearly with z with $r^2 =$
17 0.990 ($a = 1.185 \text{ kg m}^{-3}$ and $b = -9.2 \times 10^{-5} \text{ kg m}^{-4}$). The gap of $z_g(s) < z < z_L(s)$ is filled for each flight
18 using this linear dependence.

20 3.1.3 Pollutant Mixing Ratios

21
22 Five methods are compared to extrapolate mixing ratio values to the surface, which are termed: 1) zero,
23 2) constant, 3) zero-to-constant, 4) linear-fit, and 5) exponential-fit. The zero method assumes an
24 elevated plume that is completely above the lowest measurement height and a zero background
25 concentration, which gives $\chi(s,z) = 0$ for $z_g(s) < z < z_L(s)$. The constant method assumes an elevated
26 plume with a constant background level. The background level is derived from the lowest flight
27 measurement to give $\chi(s,z) = \chi(s,z_L)$ for $z_g(s) < z < z_L(s)$. The zero-to-constant method assumes non-
28 zero concentrations at the lowest flight level, a zero concentration at the surface, and a linear
29 interpolation between the surface and the lowest flight level. This interpolation gives $\chi(s,z) =$
30 $\chi(s,z_L) \times (z - z_g(s)) / (z_L(s) - z_g(s))$ for $z_g(s) < z < z_L(s)$.

31
32 For a surface based emission or a low plume in which the maximum value is near the surface, the choice
33 of extrapolation method is much more important. For example, emissions of CH_4 from the facility can
34 be from ground sources such as tailings ponds, fugitive emissions from pipe lines, or fresh mine face
35 exposed during continuing mining operations. Hence, the bulk of the emitted CH_4 mass may be below
36 the lowest measurement altitude. The linear-fit method assumes a maximum value mixing ratio at the
37 ground and a linear decrease in mixing ratio with height (z). The rate of change and the surface mixing
38 ratio are determined from a least-squares fit at each s location (with resolution $\Delta s = 40$ m) up to a height

(from ground) of $z(s) - z_g(s) = 300$ m. The exponential-fit method also uses the same data range for a least-squares fit, but assumes an exponential decay of

$$\chi(s, z) = \chi_{Top}(s) + (\chi_{sur}(s) - \chi_{Top}(s)) \exp\left(-\left(\frac{z - z_g(s)}{z_R(s)}\right)^2\right), \quad (9)$$

where $\chi_{sur}(s)$ is the surface mixing ratio and $z_R(s)$ is the scaling distance of the exponential function, both determined by least-squares fitting up to a height (from ground) of $z - z_g(s) = 1000$ m. This method assumes that the surface sourced plume dispersion has a half-Gaussian distribution vertically at locations close to the sources, such as along the box walls. The constant, linear-fit, and exponential fit are compared in a later Section (Fig. 6).

3.2 Interpolation Schemes

To determine the accuracy of the interpolation methods, three simulated emissions scenarios were generated based on: a single elevated source (smoke stack); two nearby sources with overlapping plumes (one tall smoke stack and one smaller stack); and a vertically mixed ground source. All scenarios assume a southerly wind at the location of this facility. A slant factor (β) is added to the equations to simulate a wind shear with height, resulting in Gaussian distributions of mixing ratio at the north wall of

$$\chi(s, z) = \sum_i \exp\left[-\frac{1}{2}\left(\left(\frac{s - s_{o,i} - \beta z}{\sigma_{s,i}}\right)^2 + \left(\frac{z - z_{o,i}}{\sigma_{z,i}}\right)^2\right)\right]. \quad (10)$$

The values for each scenario are listed in Table 3. The flight path for the Aug 20 flight is then used to sample the simulated values. Figure 3 shows the mixing ratios for each simulation along the north wall with the flight path locations superimposed. Values on the east, west, and south walls are near zero in all scenarios. Image interpolation (IDW, Nearest Neighbour, and Kriging) is then used with the sampled flight path positions to recreate the original image at a resolution of $\Delta s = 40$ m and $\Delta z = 20$ m. The interpolation analysis is limited to the north wall ($12 \text{ km} < s < 29 \text{ km}$), with values outside this range assumed to be zero

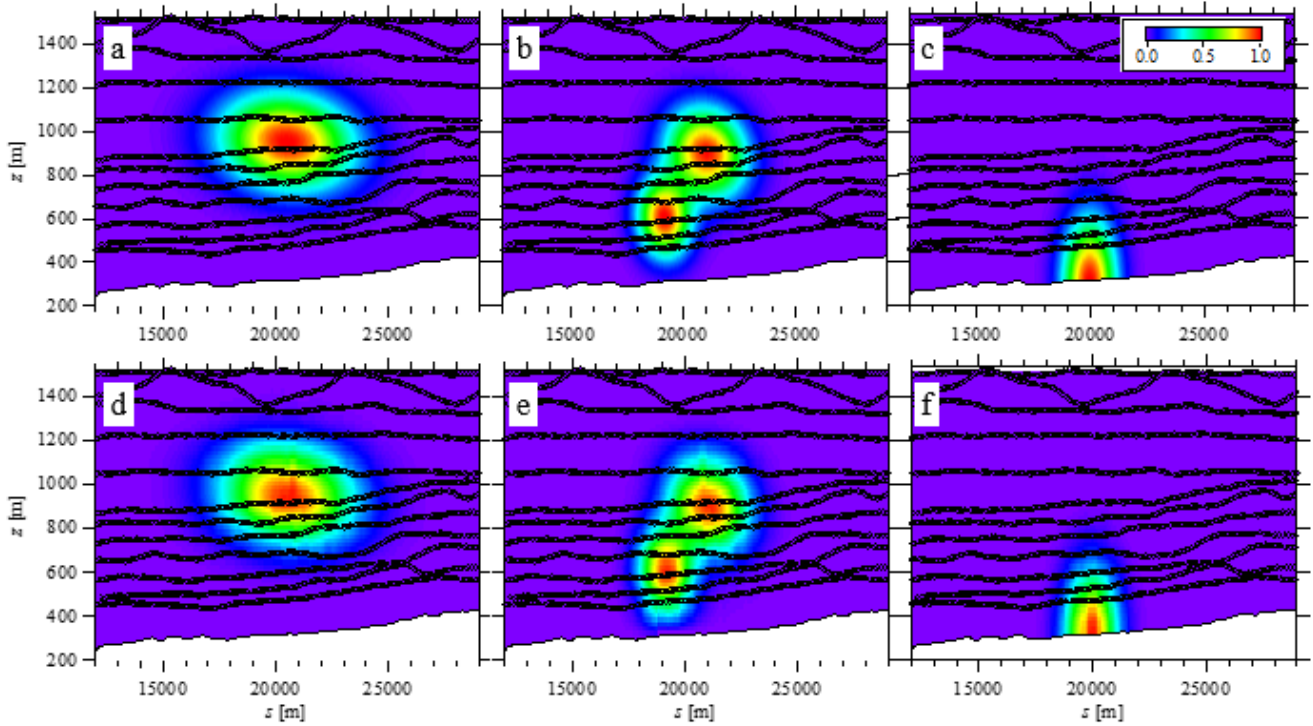
Interpolated data below the lowest flight path are removed and replaced with near-surface extrapolation as discussed in Section 3.1. For the single elevated source, there is clearly no simulated plume in the lowest 150 m (Fig. 3a), and the zero mixing ratio extrapolation method is used. For the two overlapping plumes scenario (Fig. 3b), there are significant mixing ratio values at the lowest flight level, approaching zero near the surface. Here, the zero, constant, and zero-to-constant mixing ratio extrapolation methods are compared. For the ground source scenario (Fig. 3c), there is an increase in mixing ratio towards the surface, and the linear-fit and exponential-fit methods are chosen for comparison. Examples of the resulting interpolated and extrapolated screens are shown for the three scenarios in Fig. 3d-f for the kriged interpolation with zero extrapolation (Fig. 3d), zero-constant extrapolation (Fig. 3e), and exponential fit extrapolation (Fig. 3f).

1 A statistical comparison of the three interpolation routines is shown in Table 4. The interpolated
2 average (μ) is calculated as the average value of $\chi_i(s, z)$ over the range of s and z shown in Fig. 3. The
3 mixing ratio, χ_i , is determined by interpolation and extrapolation of the mixing ratio values along the
4 flight path. The simulated average (μ_{sim}) is calculated as the average of $\chi(s, z)$ as determined by Eq.
5 10. The root-mean-square error and correlation coefficients compare χ_i and χ over the range of s and z
6 shown in Fig. 3.

7
8 For all scenarios and extrapolation methods, IDW demonstrates the highest r.m.s. error and lowest
9 correlation coefficient, while kriging consistently demonstrates the lowest r.m.s. error and highest
10 correlation coefficient. The best results are obtained for the single elevated plume, with an r.m.s. error
11 of 8.6% of the average and $r^2 = 0.998$. For the two overlapping plumes with a significant concentration
12 below the lowest flight path, the linear extrapolation to the surface gives the best results, with an r.m.s.
13 error of 13.4% of the average and $r^2 = 0.997$. For the ground source scenario, the exponential fit
14 extrapolation gives the best results, with an r.m.s. error of 19.2% of the average and $r^2 = 0.998$.

15
16 More complex kriging schemes are available that may further improve the accuracy, but these results
17 demonstrate that a far greater source of uncertainty is the extrapolation of the data between the lowest
18 flight level and the surface. In cases where extrapolation is not necessary (e.g. scenario 1), the average
19 interpolated value is within 0.2% of the simulation average. The other cases require a proper choice of
20 extrapolation technique based on knowledge of the mixing ratio behaviour in this region. For example,
21 the case of an elevated plume with part of the plume beneath the lowest flight is best suited to a zero-to-
22 constant extrapolation of mixing ratio to the surface, while a ground source concentration which
23 decreases with height above the surface is best suited to an exponential-fit extrapolation. Without
24 knowledge of this behaviour, uncertainties due to extrapolation are on the order of $\delta_{Ex} \approx 20\%$, based on
25 a comparison of the r.m.s. errors.

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2 Figure 3.(a-c) The simulated emissions scenarios of Table 3 and Eq. 10 with the flight positions of the Aug 20
 3 flight (open circles) for (a) an elevated source, (b) two elevated sources, and (c) a ground source. The values at
 4 the flight positions are then used with IDW, natural neighbour, and kriging interpolation to attempt to recreate the
 5 original plume image. Resulting krig-interpolated images (d-f) are shown for each scenario with (d) zero
 6 extrapolation, (e) zero-constant extrapolation, and (f) exponential fitting extrapolation below the lowest flight
 7 path.

8

9 Table 3. Parameters used in Eq. 8 for each simulation scenario and plume number (i). The corresponding panel (a,
 10 b, or c) in Figure 3 is also given.

Fig. 3	Interpolation	i	$s_{o,i}$ [m]	$z_{o,i}$ [m]	$\sigma_{s,i}$ [m]	$\sigma_{z,i}$ [m]	β
(a)	Single elevated source	1	22,000	950	2000	150	30
(b)	Two elevated sources	1	22,000	900	1200	140	20
		2	19,500	600	800	120	10
(c)	Ground source	1	20,000	300	800	400	0

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1 Table 4. Statistical comparison of interpolation techniques: inverse distance weighting (IDW), nearest-neighbour
2 (NN) and kriging. The Aug. 20 flight positions are used with three simulated emissions scenarios: 1) a single
3 elevated plume; 2) two overlapping elevated plumes; and 3) a plume from a ground source. Different surface
4 extrapolation methods are used based on the plume location (see Section 3.1). The statistics are determined using
5 the interpolated mixing ratio χ_i and the simulated mixing ration χ (from Eq. 10). The ratio of interpolated
6 average (μ) to the simulate average (μ_{sim}), the ratio of root-mean squared error (E_{rms}) to the simulated average
7 (μ_{sim}), and the coefficient of correlation (r^2) are compared for each scenario.

Scenario	Extrapolation Method	Statistic	IDW	NN	Kriging
1	Zero	μ / μ_{sim}	0.988	0.996	0.998
		E_{rms} / μ_{sim}	0.259	0.151	0.086
		r^2	0.985	0.996	0.998
2	Zero	μ / μ_{sim}	0.955	0.955	0.954
		E_{rms} / μ_{sim}	0.434	0.400	0.374
		r^2	0.967	0.972	0.976
2	Constant	μ / μ_{sim}	1.077	1.083	1.081
		E_{rms} / μ_{sim}	0.544	0.537	0.529
		r^2	0.950	0.950	0.953
2	Zero-Constant	μ / μ_{sim}	1.013	1.016	1.014
		E_{rms} / μ_{sim}	0.254	0.194	0.134
		r^2	0.989	0.994	0.997
3	Linear-Fit	μ / μ_{sim}	1.043	1.065	1.028
		E_{rms} / μ_{sim}	0.409	0.470	0.317
		r^2	0.994	0.991	0.996
3	Exponential-Fit	μ / μ_{sim}	1.000	1.004	0.999
		E_{rms} / μ_{sim}	0.293	0.259	0.177
		r^2	0.996	0.997	0.998

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10 3.3 Interpolated Mixing Ratio Screens

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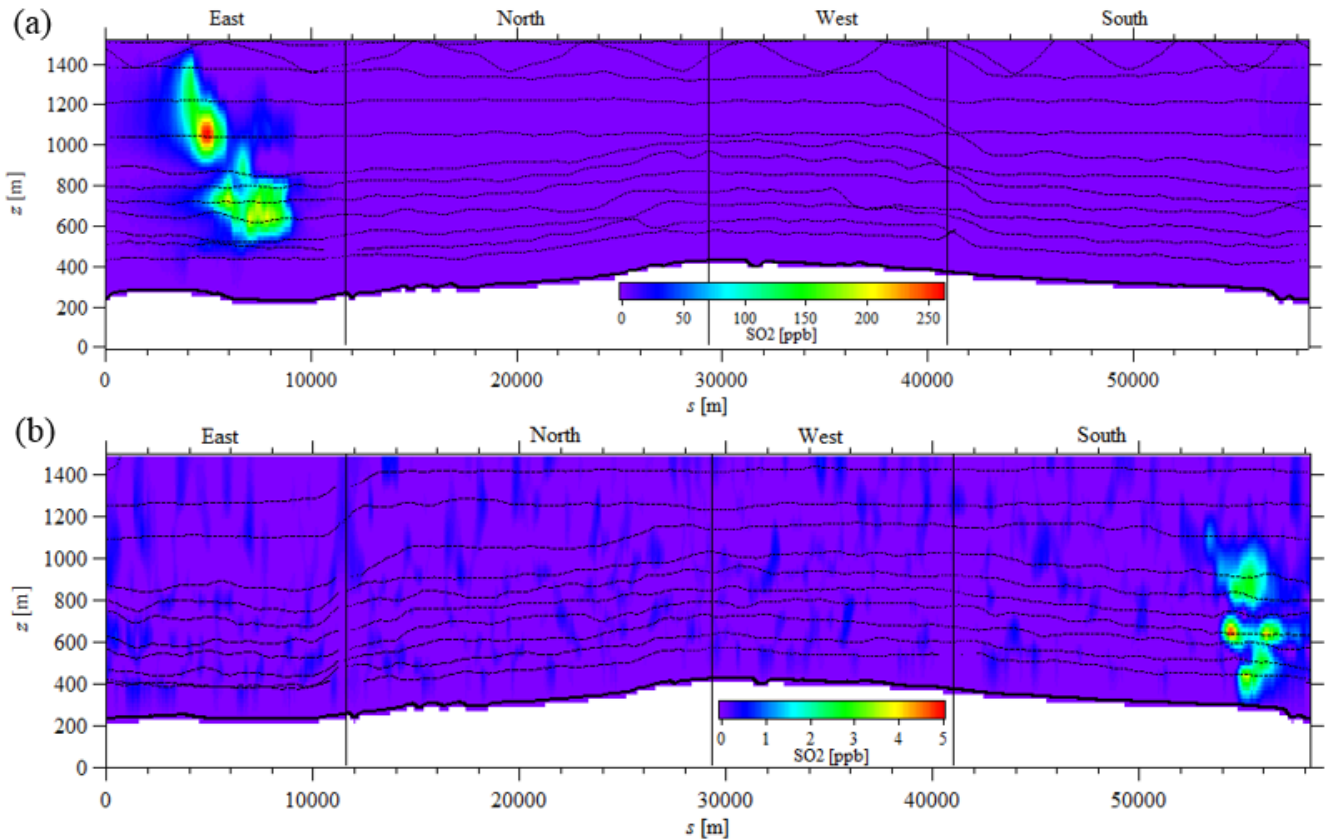
12 The kriging method discussed in Section 3.2 is used for these and all subsequent interpolations. Figure
13 4 shows the mixing ratio screens for SO₂ for the Aug 20 and Sep 2 flights. For the Aug 20 flight (Fig.
14 4a), the primary source of SO₂ appears to be two separate and elevated smokestack plumes. Due to the
15 elevation of the sources, mixing ratios are generally low at the lowest flight altitudes ($z_L(s)$) and the
16 extrapolated mixing ratios within the gap of $z_g(s) < z < z_L(s)$ are expected to be even lower. For the Sep
17 2 flight (Fig 4b), the apparent plumes are generally lower and extrapolation below the $z_L(s)$ level is
18 required. For an initial base-case we use a zero-to-constant extrapolation of mixing ratio to the surface
19 for both flights, and compare the zero and constant extrapolation techniques in Section 4.1. The zero-
20 constant surface extrapolations are shown below the lowest flight paths in Figs. 4a and 4b.

21

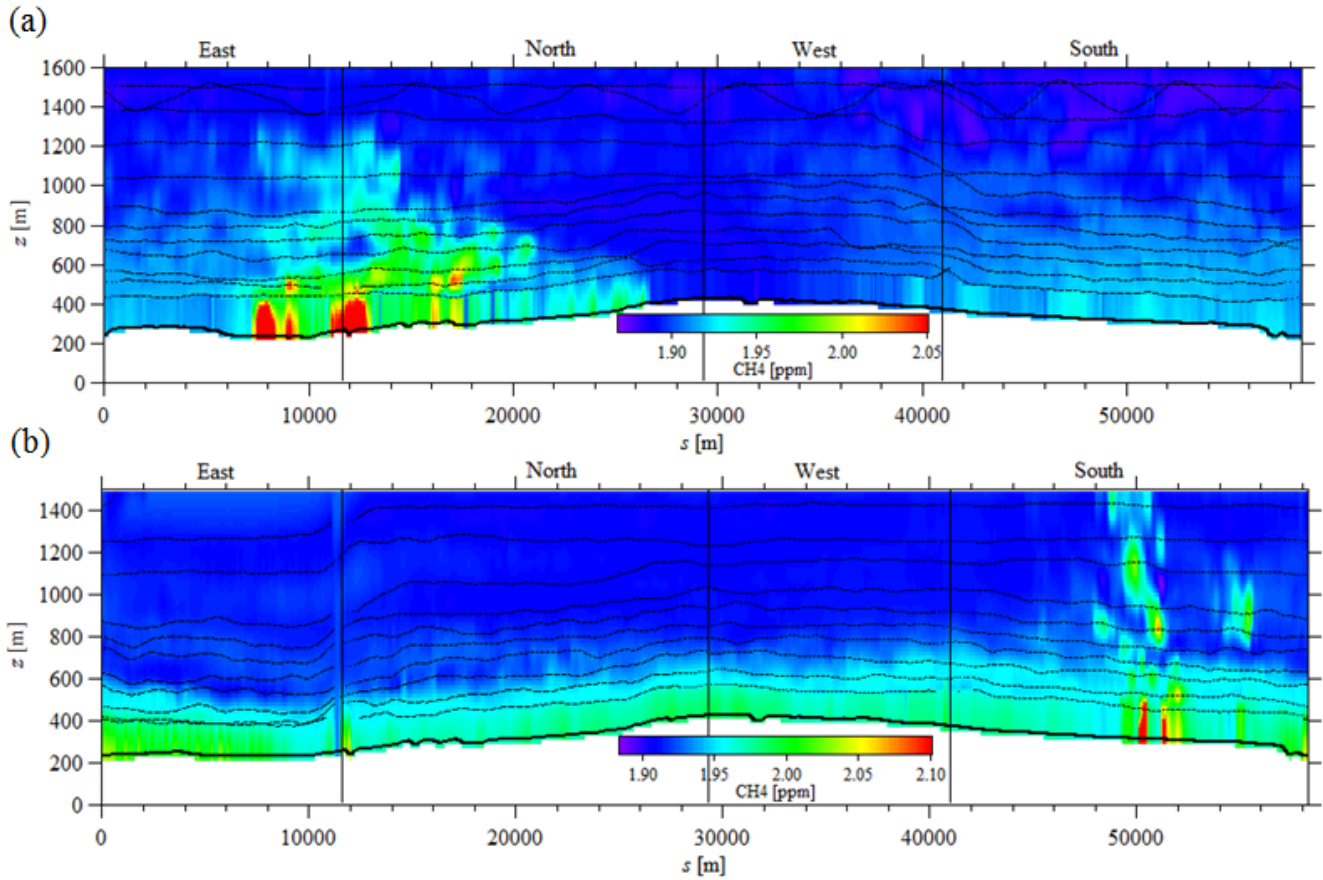
1 The interpolated CH₄ screens are shown in Fig. 5. The near-surface behaviour of CH₄ is more complex
 2 and more significant than the near surface behaviour of SO₂ as the CH₄ emissions appear to be surface-
 3 based. As is shown in Fig 5, the highest measured values of χ_{CH_4} are near the lowest flight path ($z_L(s)$),
 4 clearly indicating surface sources. Hence, the bulk of the emitted CH₄ mass may be below the lowest
 5 measurement locations. For a base case analysis we use an exponential fit to extrapolate mixing ratios
 6 to the surface and compare other extrapolation methods in Section 4.1.

7
 8 Linear and exponential fits with poor fitting statistics (defined here as $r^2 < 0.1$) for each s location are
 9 removed and replaced using the constant extrapolation method. Linear fits which indicate an increase in
 10 concentration with height are replaced with constant value extrapolation ($\chi(s,z) = \chi(s,z_L)$ for $z_g(s) < z <$
 11 $z_L(s)$). After the removal of failed fits ($r^2 < 0.1$), the exponential fitting results in average correlation
 12 coefficients of $r^2 = 0.79$ for the Aug 20 flight and $r^2 = 0.92$ for the Sep 2 flight. The linear fitting of CH₄
 13 (presented in Section 4.1) results in an average r^2 value of 0.82 for the Aug 20 flight and an average r^2
 14 value of 0.84 for the Sep 2 flight. Examples of the extrapolations for the highest recorded values at $z =$
 15 $z_L(s)$ (for each flight) are shown in Fig. 6. These figures compare the measured values (within ± 40 m of
 16 the s location), interpolated values, and extrapolations using constant value, linear-fit, and exponential-
 17 fit.

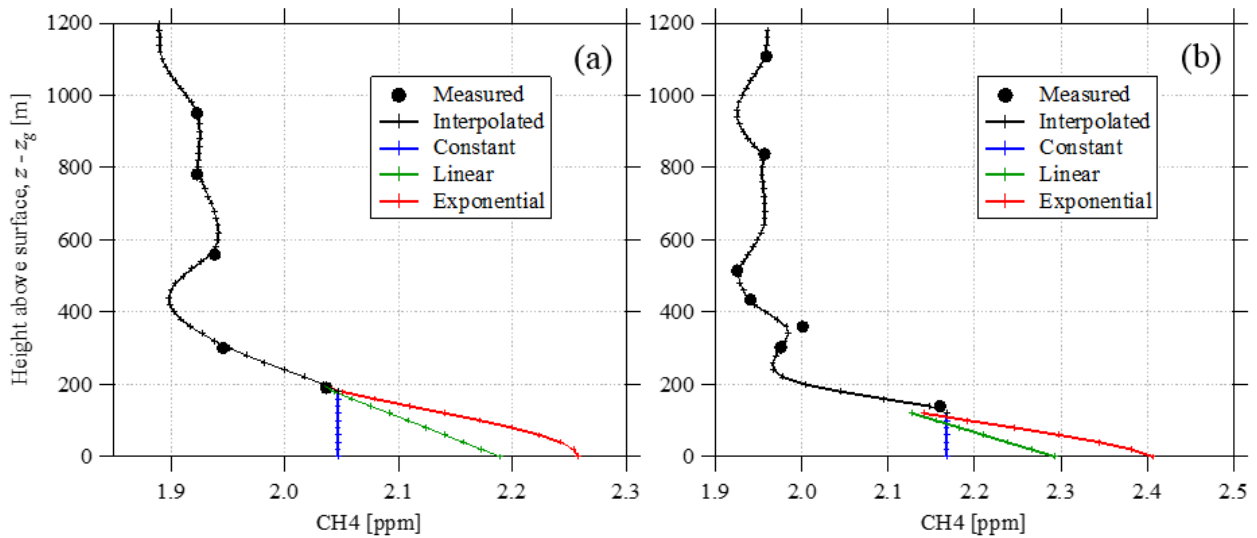
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19
 20 Figure 4. Krig interpolated SO₂ mixing ratios for the Aug 20 (a) and Sep 2 (b) flights. Note that the colour scales
 21 are different. The flight path is superimposed (black dots). The near-surface values are estimated using a zero-
 22 constant extrapolation which varies linearly from the lowest measurement to a zero value at the surface.



1
 2 Figure 5. Krig interpolated CH_4 mixing ratios for the Aug 20 (a) and Sep 2 (b) flights. The flight path is
 3 superimposed (black dots). Values below the lowest flight path are extrapolated with an exponential fit (Eq. 9).
 4



5
 6 Figure 6. CH_4 mixing ratios with height above ground for the Aug 20 flight (a) at $s = 12.4$ km and for the Sep 2
 7 flight (b) at $s = 50.4$ km. Measured values within ± 40 m (Δs) are shown as dots, krig interpolated values ($z >$
 8 $z_L(s)$, black lines) are compared to extrapolated values ($z < z_L(s)$) for constant (blue), linear-fit (green), and
 9 exponential-fit (red) values. Background values on each day were approximately 1.9 ppm.

10

3.4. Emission Rate Calculation

The change in air mass within the volume, $E_{air,M}$, is estimated based on temperature and pressure changes (from Eq. 5) over the duration of the flights as measured at two locations at Fort McMurray Airport (56.650 N, 111.213 W, <http://climate.weather.gc.ca>) and two meteorological towers (<http://www.wbea.org>). One tower is 167 m tall (AMS03: 57.032 N, 111.505 W) and the other is 75 m tall (AMS05: 56.969 N, 111.482 W). Both towers are located approximately 40 km from the CNRL facility (nearly half-way between the airport and the facility). The average pressure ratio from both airport locations is $\Delta p/p = 0.02\%$ for Aug 20 and $\Delta p/p = 0.13\%$ for Sep 2. The average temperature ratio from the lowest (20 m) and highest (75 m or 167 m) tower heights and two airport measurement locations is $\Delta T/T = 0.99\%$ for Aug 20 and $\Delta T/T = -0.69\%$ for Sep 2. The sensitivity of the final results to the pressure and temperature ratios is discussed in Section 4.1. The change in air mass ($E_{air,M}$) is compared to the horizontal advective flux of air mass ($E_{air,H,in}$ and $E_{air,H,out}$) in Table 5.

The horizontal advective flux ($E_{air,H}$) and the change in air mass ($E_{air,M}$) are then used to determine the vertical advective air mass flux ($E_{air,V}$) from Eq. 2. Dividing the vertical advective air mass flux by the box-top area and an average air density gives an average exit velocity through the box-top of $w = 0.10 \text{ m s}^{-1}$ for Aug 20 and $w = 0.08 \text{ m s}^{-1}$ for Sep 2. This exit velocity is a result of a divergence of the air flow since flow streamlines over a varying terrain are unlikely to follow flat, horizontal trajectories.

Table 5. Mass balance terms for air (Eq. 2) in units of 10^9 kg h^{-1} . Horizontal advective flux is separated into inward and outward as $E_{air,H} = E_{air,H,out} - E_{air,H,in}$.

Term	Aug 20	Sep 02
$E_{air,H,in}$	479.6	722.6
$E_{air,H,out}$	394.6	657.9
$E_{air,M}$	-1.4	1.3
$E_{air,V}$	83.6	66.0

Table 6 lists the integrated lateral flux terms ($E_{C,H,in}$ and $E_{C,H,out}$) for SO_2 and CH_4 . The value of $E_{air,V}$ shown in Table 5 is used to calculate the compound mass flow through the box top as $E_{C,V} = M_R \chi_{C,Top} E_{air,V}$. At the highest level of the interpolated screen ($z = 1540 \text{ m}$ for Aug 20, $z = 1500 \text{ m}$ for Sep 2), the average mixing ratio of SO_2 is near zero for both flights ($<0.02 \text{ ppb}$). The resulting mass flow of SO_2 through the box-top is negligible compared to the lateral advection through the box walls. The average mixing ratio of CH_4 at the top of the screen is $\chi_{C,Top} = 1.89 \text{ ppm}$ for the Aug 20 flight and $\chi_{C,Top} = 1.91 \text{ ppm}$ for the Sep 2 flight. This non-zero mixing ratio results in a significant loss of CH_4 through the box top, which is larger than the net gain of CH_4 through the box walls. Although the horizontal and vertical advection terms are a factor of 20 higher than the estimated emission flux, the scale of these terms is deceiving, as they are calculated from the same integral of air density and wind speed, as is demonstrated in the Appendix.

1 The horizontal turbulent flux for a Gaussian plume can be estimated for SO₂ by assuming a linear
 2 expansion of the plume width with distance downwind. Based on measured wind profiles, estimated
 3 plume height and source location, input variables of Eq. 6 are estimated as $u^* = 0.3 \text{ m s}^{-1}$, $h = 1.6 \text{ km}$, $L =$
 4 50 m , $x = 4 \text{ km}$, and $U = 6 \text{ m s}^{-1}$. These estimates give a ratio of $E_{C,HT} / E_{C,H} \sim 0.03 \%$. For CH₄ the
 5 plume location near the ground would suggest a much smaller diffusion constant and wind speed,
 6 resulting in negligible horizontal turbulent flux for these surface based emissions.

7

8 Table 6. Mass balance terms for SO₂ and CH₄ (Eq. 1). Horizontal advective flux is separated into inward and
 9 outward as $E_{C,H} = E_{C,H,out} - E_{C,H,in}$.

Term	SO ₂ [t h ⁻¹]		CH ₄ [t h ⁻¹]	
	Aug 20	Sep 02	Aug 20	Sep 02
$E_{C,H,in}$	0.226	0.148	501.1	766.2
$E_{C,H,out}$	12.890	0.395	416.5	701.3
$E_{C,V}$	0.003	<0.001	87.3	69.9
$E_{C,HT}$	0.004	<0.001	0.00	0.00
$E_{C,VT}$	0	0	0.07	0.24
$E_{C,VD}$	0	0	0.00	0.00
$E_{C,M}$	-0.015	<0.001	-1.49	1.53
$E_{C,X}$	-0.097	-0.002	0	0
E_C	12.79	0.249	4.21	3.79

10

11

12 For the calculation of the vertical turbulent fluxes ($E_{C,VT}$), an inversion step change of concentration
 13 (determined from the flight spirals) is used as a proxy for the boundary layer depth. A comparison of
 14 flight altitude to SO₂ for the entire flight duration (including vertical spirals to 2 km, as shown in Fig. 1)
 15 demonstrates that SO₂, which has background levels near 0, shows no inversion step change ($\Delta\chi$) for
 16 either flight. In comparison, during the Aug 20 flight, there is an inversion step change for CH₄ of
 17 approximately 8.0 ppb. The height of this step change varies between 1.0 km and 1.8 km amsl depending
 18 on location. During the Sep 2 flight, there is a much stronger inversion step change for CH₄ of
 19 approximately 28 ppb near 1.5 km amsl. Using these values with $w_e = 0.03 \text{ m s}^{-1}$ gives $E_{C,VT} = 0.07 \text{ t/h}$
 20 and 0.24 t/h for Aug 20 and Sep 2, respectively, representing 2% and 6% of the CH₄ emission rate (E_C)
 21 estimated for both days. However, there is a large uncertainty in this $E_{C,VT}$ estimation and it is unclear
 22 from these measurements if the inversion step change occurs near enough to the box top to necessitate
 23 inclusion in the calculated emissions.

24

1 The deposition term is calculated with a surface mixing ratio (χ_{Sur}) estimated with the same near-surface
 2 interpolation schemes used to calculate $E_{C,H}$. For SO_2 , a linear decrease to a zero surface mixing ratio
 3 was used, which would give zero deposition. Hence SO_2 deposition is zero for this base case, but will be
 4 non-zero for other near surface extrapolation techniques (compared in Section 4.1). For example, using
 5 the constant value extrapolation with a deposition velocity for SO_2 of $V_D = 10 \text{ mm s}^{-1}$ (Zhang et al., 2003)
 6 gives depositions of $<2\%$ of $E_{C,H}$. For CH_4 , generally deposition is not considered in mass balance
 7 calculations, although some microbial uptake of CH_4 in soils has been documented (e.g. Whalen and
 8 Reeburgh, 2000). Here we assume that the CH_4 deposition rate ($E_{C,VD}$) is zero.

9
 10 The change in compound mass within the box volume due to change in air density, $E_{C,M}$, is estimated
 11 from Eq. 7 with the average temperature and pressure ratios used to calculate the $E_{air,M}$ term using Eq. 5.
 12 As discussed in Section 2.5, the unknown concentrations within the volume are estimated by averaging
 13 the surrounding box at each height level. The resulting values of $E_{C,M}$ are $<0.2\%$ of the horizontal flux
 14 term $E_{C,H}$ for SO_2 for both days. For CH_4 , the resulting values of $E_{C,M}$ are small relative to the horizontal
 15 flux term $E_{C,H}$ ($\pm 2\% E_C$), but are large compared to the final calculated emission rate ($-35\% E_C$ and 40%
 16 E_C for Aug 20 and Sep 2 respectively). However, it will be demonstrated in Section 4.1 that the final
 17 emission rate is not strongly dependent on the change in air density due to temperature and pressure
 18 changes. This is because the change in density influences both $E_{air,M}$ (Eq. 5) and $E_{C,M}$ (Eq. 7) as is
 19 shown in the appendix.

20
 21 The change in compound mass within the volume due to oxidation of SO_2 ($E_{C,X}$) is estimated for a
 22 source to box-wall distance of 4 km, an average wind speed of $U = 6 \text{ m s}^{-1}$, and a chemical lifetime of τ
 23 = 24 hours (Walter et al., 2012). These estimates give $t_0 = 11 \text{ min}$ and $E_{C,X}/E_{C,H} = -0.8\%$. The
 24 chemical reaction of CH_4 is assumed to be insignificant.

25 26 **4. Discussion**

27 28 **4.1 Calculation of Uncertainties**

29
 30 To calculate uncertainties in the final emission rate, we attempt to identify and estimate each source of
 31 uncertainty. Most of the uncertainties in the calculated emission rates are due to five subcomponents:
 32 measurement error in wind speed and mixing ratio (δ_M), the near-surface mixing ratio extrapolation
 33 technique (δ_{Ex}), the near-surface wind extrapolation (δ_{Wind}), the box-top mixing ratios (δ_{Top}), and the
 34 temperature and pressure ratios (δ_{dens}). We also investigate the uncertainty due to the location of the step
 35 inversion (δ_{VT}) and the box-top height (δ_{BH}). In some cases, such as wind speed measurement error, wind
 36 speed extrapolation, and box-top height, the uncertainty affects multiple variables simultaneously. For
 37 this reason all uncertainties are expressed as a fraction of the base case emission rate (E_C) and also in
 38 units of t h^{-1} . Each uncertainty is assumed to be independent and they are added in quadrature to give the
 39 total estimated emission rate uncertainty as

$$40 \quad \delta^2 = \delta_M^2 + \delta_{Ex}^2 + \delta_{Wind}^2 + \delta_{Top}^2 + \delta_{dens}^2 + \delta_{VT}^2 + \delta_{BH}^2. \quad (11)$$

1 All uncertainties are listed for each flight and each species in Table 7.

2 3 **4.1.1. Wind Speed and Mixing Ratio Measurement** 4

5 Uncertainties in wind speed and mixing ratio measurements are incorporated into the algorithm through a
6 Monte Carlo simulation in which the wind and mixing ratio time series are modified from the base case,
7 the wind and mixing ratio screens are re-interpolated, and the horizontal advective fluxes of air ($E_{air,H}$)
8 and compound ($E_{C,H}$) are recalculated to determine the uncertainty in the final emission rate (E_C). Wind
9 speed and mixing ratio measurement uncertainties are given in Section 2.1. For SO₂, this gives a total
10 uncertainty in the E_C term of 0.076 t h⁻¹ (0.6% of E_C) for the Aug 20 flight and 0.002 t h⁻¹ (0.9%) for the
11 Sep 2 flight. For CH₄, this gives a total uncertainty in the E_C term of 0.025 t h⁻¹ (0.6%) for the Aug 20
12 flight and 0.023 t h⁻¹ (0.6%) for the Sep 2 flight. Hence we conservatively estimate the uncertainty due
13 to wind speed and mixing ratio measurement error as $\delta_M \approx 1\%$.

14 15 **4.1.2. Extrapolation Method** 16

17 For SO₂, the base case was calculated by assuming a linear decrease from the mixing ratio measured at
18 the lowest flight altitude to zero at the surface. Assuming a constant value of $\chi(s,z) = \chi(s,z_L)$ below $z_L(s)$
19 results in an increase of 0.038 t h⁻¹ (0.3% of E_C) for the Aug 20 flight and an increase of 0.021 t h⁻¹
20 (8.2%) for the Sep 2 flight. Assuming a constant value of $\chi = 0$ below $z_L(s)$ results in a decrease of 0.041
21 t h⁻¹ (-0.3%) for the Aug 20 flight and a decrease of 0.029 t h⁻¹ (-12%) for the Sep 2 flight. As
22 demonstrated by Fig. 4, the plume is well above the lowest flight path on Aug 20, but much closer to the
23 surface on Sep 2, which increases the uncertainty due to surface extrapolation. A constant value of zero
24 is an extreme assumption, and it is likely that the true profile is somewhere between the constant ($\chi(s,z) =$
25 $\chi(s,z_L)$) and linear extrapolation techniques. Based on this sensitivity analysis, we estimate an
26 uncertainty due to extrapolation of $\delta_{Ex} \approx 0.3\%$ on Aug 20, when the emissions are higher, and $\delta_{Ex} \approx 12\%$
27 on Sep 2, when there is less SO₂ emission. These values are within the uncertainty estimates of $\delta_{Ex} \approx$
28 20% determined with simulated plumes in Section 3.2.

29
30 For CH₄, the base case is an exponential extrapolation below the lowest flight path, which results in
31 downwind surface mixing ratios as high as 1 ppm above background levels on Aug 20 and 0.5 ppm
32 above background levels on Sep 2 (typical background levels on both days are near 1.9 ppm). Use of a
33 constant extrapolation results in a 1.1 t h⁻¹ reduction of the emission rate (-27% of E_C) for Aug 20 and
34 negligible change (-1.1%) for Sep 2. Use of a linear extrapolation results in an emission rate reduction
35 of 0.8 t h⁻¹ (19%) for Aug 20 and a reduction of 0.5 t h⁻¹ (13%) for Sep 2.

1 Table 7. Emission rate uncertainties as percent of E_C , rounded up to the nearest integer. Uncertainties are
 2 combined to give the total uncertainty (δ) from Eq. 11.

		SO ₂		CH ₄	
		Aug 20	Sep 2	Aug 20	Sep 2
Measurement Error	δ_M	1	1	1	1
Mixing Ratio Extrapolation	δ_{Ex}	1	12	27	15
Wind Extrapolation	δ_{Wind}	1	1	1	1
Box-top Mixing Ratio	δ_{Top}	1	6	3	5
Density Change	δ_{dens}	0	0	2	6
Vertical Turbulence	δ_{VT}	0	0	2	7
Box-top Height	δ_{BH}	0	0	0	16
Total Uncertainty	δ	2	14	28	25

3

4

5 This dependency on extrapolation method is consistent with the uncertainty of a surface-based emission
 6 source and a low altitude plume (as shown in Fig. 5). Large-eddy simulation modeling by Vinuesa and
 7 Galmarini (2009) demonstrate that a ground source with a mean wind speed of 5 m s⁻¹ develops from an
 8 exponential profile to a constant value near the surface within 2 to 6 km distance from the source,
 9 suggesting that the constant value extrapolation to the surface may be a better physical representation of
 10 the plume. However, under ideal conditions, a constant emission of CH₄ from the upwind boreal forest
 11 would result in an exponential vertical profile of mixing ratio going into the box. Hence we propose a
 12 fourth method of extrapolation to the surface which uses a combination of the three extrapolation
 13 techniques, depending on the flux direction. In situations where air masses pass over the boreal forest
 14 and then enter the box, an exponential extrapolation to the surface is used, and when the air mass leaves
 15 the box, a constant extrapolation to the surface is used, thereby considering the results of the large eddy
 16 simulation modeling (Vinuesa and Galmartini, 2009). This extrapolation method results in an emission
 17 rate reduction of 1.1 t h⁻¹ (-27% of E_C) for Aug 20 and a reduction of 0.6 t h⁻¹ (-15%) for Sep 2. Hence
 18 we estimate the uncertainty due to extrapolation for CH₄ as $\delta_{Ex} \approx 27\%$ for Aug 20 and $\delta_{Ex} \approx 15\%$ for Sep
 19 2.

20

21 4.1.3. Wind Speed Extrapolation

22

23 To estimate sensitivity to the extrapolation of wind speed, the base case scenario was rerun assuming a
 24 constant wind speed of $U = U(z_L)$ for $z_g < z < z_L$. Because the horizontal advection of air is also affected
 25 by the wind speed, the calculation of vertical advection (both of air and chemical species) will also
 26 change with an assumed wind speed (see Table 2). The resulting change in calculated emission rate is
 27 less than 1% for both flights for both species, suggesting that the correct parameterization of wind speed
 28 near the surface is not significant. The uncertainty for wind speed extrapolation is therefore
 29 conservatively estimated as $\delta_{wind} \approx 1\%$ for all cases.

30

4.1.4. Box-Top Mixing Ratio

The assumed constant mixing ratio at the top of the box is estimated using the average measured value at the top of the interpolated screen walls (i.e. $\chi_{C,Top} = \int \chi(s, z_{Top}) ds / \int ds$). This value is then used in the calculation of the vertical advection term, $E_{C,v}$. For a normal distribution of error there is 95% confidence that the true mean is within approximately $2\sigma/\sqrt{n}$ of the measured mean, where σ is the standard deviation of the measurements and n is the number of independent samples. To estimate n , the auto-correlation of each series of mixing ratio, $\chi(s, z_{Top})$, is calculated and a length scale is approximated at a distance where the auto-correlation approaches zero. The results demonstrate an independent length scale between 1.5 km and 3 km. For the lap distance of nearly 59 km, this gives n between 20 and 40. For our error estimation we conservatively use the lower value of $n = 20$ sample points. For SO₂, this gives real mean values of $\chi_{C,Top}$ are 0 ± 0.14 ppb on Aug 20 and 0 ± 0.092 ppb on Sep 2. For CH₄, the real mean values are $1.89 \text{ ppm} \pm 2.0$ ppb on Aug 20 and $1.91 \text{ ppm} \pm 4.5$ ppb on Sep 2. This range results in an emission rate uncertainty for SO₂ of 0.026 t h^{-1} (0.2% of E_C) for Aug 20 and 0.013 t h^{-1} (5.4%) for Sep 2. For CH₄, the emission rate uncertainty is 0.09 t h^{-1} (2.2%) for Aug 20 and 0.16 t h^{-1} (4.3%) for Sep 2. These uncertainties are listed (δ_{Top}) in Table 7.

4.1.5. Change in Air Density

The temperature and pressure ratio difference from Eqs. 5 and 7 ($\Delta p/p - \Delta T/T$) used in the base case is an average of four meteorological stations in the surround area. These ratios are used to determine the magnitude of the change in air density within the box, which defines the $E_{air,M}$ and $E_{C,M}$ terms. The average value is -0.97% on Aug 20, indicating a reduction in air density with time, and 0.82% on Sep 2, indicating an increase in air density with time. Using the minimum and maximum ratios derived from the stations with the highest (or lowest) temperature (or pressure) ratios gives an indication of the uncertainty due to the air density change in the region. The minimum ratios are -1.07% and -0.45% , and the maximum ratios are -0.85% and 1.30% on Aug 20 and Sep 2, respectively. The derived SO₂ emission rate is not sensitive to variation in air density due to the low background mixing ratios, with differences less than 2 kg h^{-1} . The CH₄ emission rate of Aug 20 is also not sensitive to density change; however the emission rate of Sep 2 does show some dependence, with changes in emission rates between -0.07 t h^{-1} (-1.9% of E_C) and 0.19 t h^{-1} (5.1%) from the base case for the given range of temperature and pressure ratios. Hence, for SO₂ with near-zero background, we estimate $\delta_{dens} \approx 0$. For CH₄, with high backgrounds, we estimate an uncertainty due to density changes within the box of $\delta_{dens} \approx 2\%$ on Aug 20 and $\delta_{dens} \approx 6\%$ on Sep 2.

4.1.6. Boundary-Layer Height

Boundary-layer height is estimated using the variation of dew point temperature (T_d) with height. Visual inspection of the data demonstrates a boundary-layer height of approximately 600 m near the start of the Aug 20 flight (~10:30 LT) and a height of 1100 m after the box portion of the flight (~12:30 LT). During the Sep 2 flight, the boundary layer height was invariant with time, and ranged from 1100 m to 1500 m. These heights are consistent with the step inversions described in Section 4.3. The calculation of the emission rate with the TERRA method does not depend on boundary-layer height, so long as the plume extent is below the box height at the downwind screen location. Below we investigate the uncertainty in the calculated emission rate due to the assumed vertical turbulent mixing across the step inversion and due to the choice of box height.

For CH₄, There is a large unknown uncertainty in the application of the vertical turbulent flux term ($E_{C,VT}$), due to the unknown exact location of the step inversion. The $E_{C,VT}$ used in the base case (Section 3.4) is 1.7% of the emission rate term (E_C) for Aug 20 and 6.3% E_C for Sep 2. For SO₂ this is not an issue due to the near-zero background levels. Hence we estimate the magnitude of the uncertainty of $\delta_{VT} \approx 0\%$ for SO₂ and $\delta_{VT} \approx 2\%$ for CH₄ on Aug 20 and $\delta_{VT} \approx 7\%$ for CH₄ on Sep 2.

To investigate the sensitivity of the model to the choice of box height, we recalculate the emission rate with an arbitrarily reduced box height of 100 m. This has the effect of modifying three terms: the horizontal advective flux term ($E_{C,H}$), due to reduced screen size; the vertical advective flux term ($E_{C,V}$), due to the both the recalculation of box-top mixing ratio, $\chi_{C,Top}$, and the change in vertical advection of air, $E_{air,V}$ (see Table 2); and the mass change term ($E_{C,M}$), due to the change in box volume. Reducing the box-top height by 100 m, results in a change in the calculated emission rate (E_C) of less than 1%, except in the case of the Sep 2 flight for CH₄, in which the modified emission rate is reduced by 0.61 t h⁻¹ (16% of E_C). In the Sep 2 case for CH₄, there is a highly elevated plume (Fig. 5b) which is not fully captured by the reduced box. Since there is potential that this plume could further extend above the unmodified box-top height, we estimate the uncertainty in CH₄ emission rate for the Sep 2 flight as $\delta_{BH} \approx 16\%$.

4.1.7. Total Estimated Uncertainties

The uncertainties listed above and in Table 7 are summed with Eq. 11 to give the total emission rate uncertainties in each case. The largest uncertainties are due to the mixing ratio extrapolation below the lowest flight level. The height and location of the plume (Figs. 4 and 5) clearly relates to the mixing ratio extrapolation. Generally, uncertainties are lower for the SO₂ emission rate (2% on Aug 20 and 14% on Sep 2) compared with the CH₄ emission rate (28% on Aug 20 and 25% on Sep 2). This is consistent with the lower surface emissions of CH₄ compared to the elevated stack emissions of SO₂. When the plume is clearly elevated and fully contained in the flight range (as with SO₂ on the Aug 20 flight), the uncertainty is insignificant. When the bulk of the mixing ratio is closer to the surface (as with CH₄ on the Aug 20 flight) the uncertainty is very large (approaching 30%).

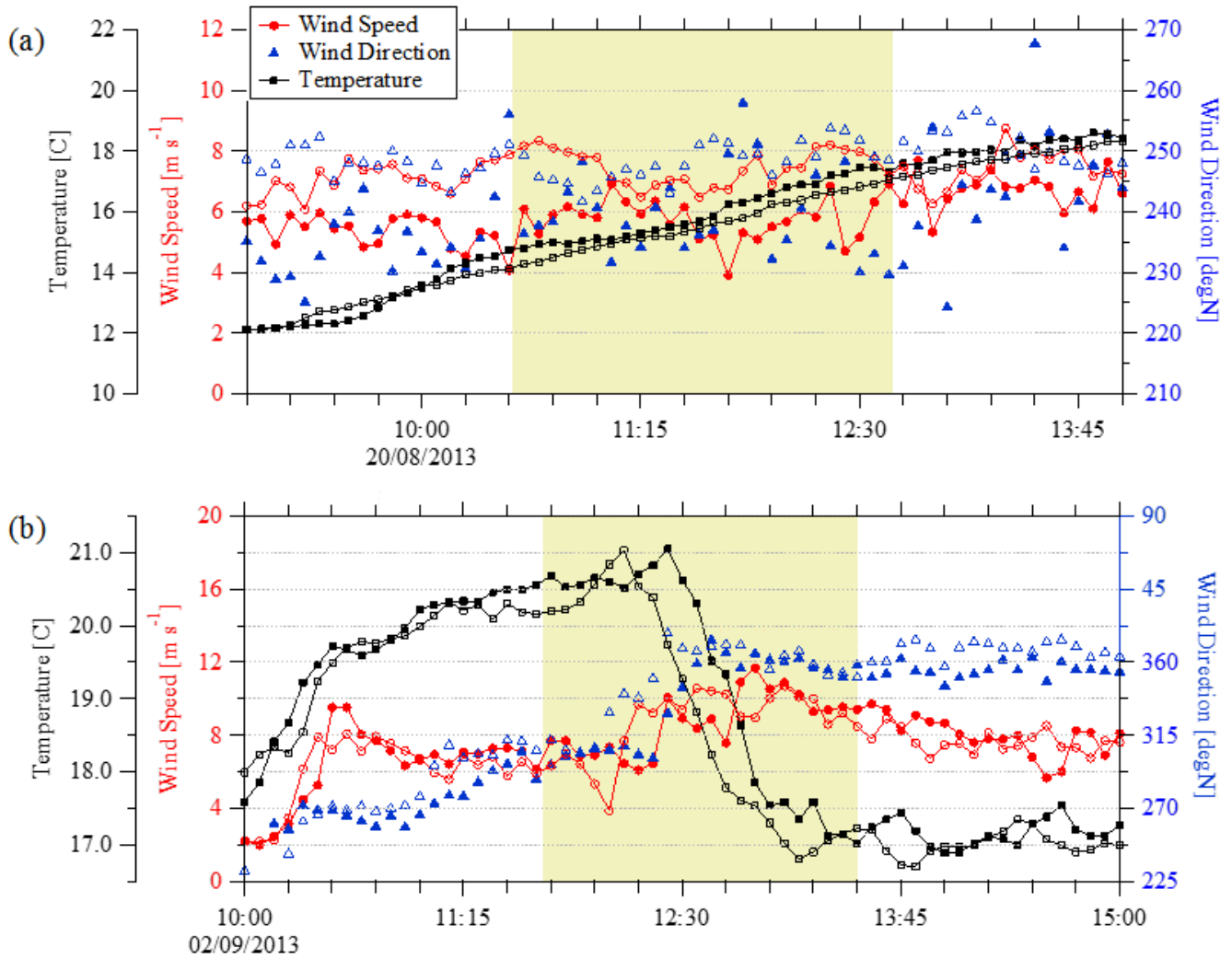
1
2 All other uncertainties are relatively small (<7%), with the exception of CH₄ on the Sep 2 flight, where
3 the proximity of a plume segment to the top of the box leads to suspicion that the entire plume may not
4 have been captured by the flight screen.

5
6 Hence the surest method to reduce uncertainties during flights using the TERRA method is to ensure
7 that the plume is fully contained between the lowest and highest flight paths, although it is recognized
8 that this is not always possible due to low-level plumes and minimum flight altitude restrictions. In
9 these cases, mixing ratio constraints based on downwind surface measurements would reduce
10 uncertainty significantly. Beyond this approach, only a reduction in wind and mixing ratio uncertainty
11 would have a significant impact on improving the accuracy of the emission rate estimation.

12 13 **4.5 Wind Consistency**

14
15 A potential source of error with the SO₂ plume is the assumption of constant wind speed during the box
16 flight. This assumption is less of an issue with CH₄ as it is a surface source and the bulk of the ground-
17 source plume is sampled in the lowest two flight tracks where winds are generally lower. The aircraft
18 took approximately 11 to 12 min to complete one level track around the facility. If there is a shift in
19 mean wind direction or plume buoyancy during that time, the plume could potentially be over- or under-
20 sampled. Fig. 4 demonstrates two or three separated plume maxima on both Aug 20 (Fig. 4a) and Sep 2
21 (Fig. 4b). On both flights the progression of the flight path increased in altitude level upward from near
22 the surface to ~1.5 km. Hence the separation of plume maxima could be due to turbulent fluctuation and
23 large scale eddies, multiple plumes, or a sudden change in plume position, which would result in
24 oversampling of the same plume. In the case of the Aug 20 flight, this drift would need to be
25 approximately 1 km to the south (decreasing *s* on the east wall in Fig. 4a) and 400 m upward in the
26 duration of one level track completion. In the case of the Sep 2 flight, the plume would need to move
27 300 m upward within 11 to 12 minutes.

28
29 Figure 7 shows the wind speeds, wind direction, and temperatures measured at the two towers
30 (<http://www.wbea.org>) located between the facility and Fort McMurray. During the Aug 20 flight (Fig.
31 7a), wind speeds and direction appear consistent at these locations. Temperature rises consistently
32 throughout the two hour duration by approximately 3°C. Based on these measurements, it is unlikely that
33 a major shift would occur in plume position, as there is no apparent shift to a more northerly flow and an
34 increase in air temperature would cause a relative decrease in plume buoyancy, resulting in
35 undersampling instead of oversampling as the aircraft altitude increases.



1
 2 Figure 7. Wind speed (circles), direction (triangles), and temperature (squares) as recorded by two WBEA towers.
 3 Open symbols are at AMS03 (57.032 N, 111.505 W) at a height of 167 m above ground level and closed symbols
 4 are at AMS05 (56.969 N, 111.482 W) at a height of 75 m above ground level. The shaded area shows the duration
 5 of the flight box on Aug 20 (a) and Sep 2 (b). Times are LT (MDT).

6
 7
 8 During the Sep 2 flight (Fig. 7b) a shift in winds and temperature is apparent resulting in higher wind
 9 speeds, a 3.5 °C drop in air temperature, and a shift from WNW to N winds. A decrease in air
 10 temperature could result in a sudden increase in plume buoyancy, which would move the plume upward
 11 during the box flight. A shift from westerly to northerly wind speeds would result in a shift in the plume
 12 position to the west (decreasing s on the south wall in Fig. 4b), so that the higher plume (presumably
 13 sampled later) would be west relative to the lower plume (presumably sampled before). This shift in
 14 lateral plume position is not seen in Fig. 4b, suggesting that the small shift in wind speed and direction
 15 has no apparent effect on the measurements.

16

1 These results suggest that wind shifts do not affect our measurements for these two flights; however, this
2 demonstrates the necessity of stationarity testing such as this for mass-balance flight emissions
3 calculations. This is especially the case when weather conditions change during the box flight duration.
4

5 **4.6 Comparison to Industry-Reported SO₂ Emissions**

6

7 Individual stack emission rates of SO₂ on a minute-by-minute basis were provided by CNRL for the two
8 flight dates. Using average wind speed and direction and approximate distance from the stack to the
9 box-wall, we estimate a delay of approximately 30 min. between plume emission and interception by the
10 aircraft on Aug 20, and 20 min. between emission and interception by the aircraft on Sep 2.
11

12 During the aircraft box flight time period of 10:31 to 12:41 LT on Aug 20 and adjusting for the 30 min.
13 between plume emission and interception by the aircraft, the minute-by-minute SO₂ emissions give an
14 average emission rate of 12.20 t/h. Our TERRA derived SO₂ emission rate is 12.79 t/h, with an
15 estimated uncertainty of 2%. This difference of 4.8% is larger than the uncertainty associated with the
16 emission calculation on this day (primarily due to wind speed measurement uncertainty); however, it is
17 possible that the difference of 4.8% is real and due to non-stack sources of SO₂, such as gas combustion
18 or oxidation of the sulphur stockpile in the facility.
19

20 For Sep 2, the minute-by-minute SO₂ emission rates from CNRL give an average of 0.224 t/h during the
21 aircraft box flight duration of 11:18 to 14:43 LT and accounting for a plume travel time of 20 min.
22 between emission and interception by the aircraft. This is compared to our TERRA derived SO₂
23 emission rate of 0.249 t/h, with an uncertainty of 14%. The aircraft derived SO₂ emission rate is 11%
24 higher (within the estimated uncertainty) but the absolute difference is small at 0.025 t/h. This
25 difference may be either due to non-stack SO₂ emissions at the CNRL facility, or due to the
26 uncertainties in the aircraft derived emission rates.
27
28

29 **5. Conclusions and Recommendations**

30

31 The results presented above demonstrate the relative importance of terms in the mass balance equation
32 and potential sources of uncertainty in the emission rate calculations. For a low-background compound,
33 such as SO₂, the horizontal advection is by far the most significant term. In contrast, CH₄ has large
34 background levels and a small incremental concentration due to facility emissions. For CH₄, the
35 significance of the vertical advection and mass density change terms ($E_{C,V}$ and $E_{C,M}$) highlight the
36 potential disadvantages in using simplified techniques to estimate emission rate, such as single height
37 transects or single screens. For example, in the case of a surface-based emitted compound such as CH₄,
38 failure to include advective fluxes through the box top (with proper background subtraction) can result in
39 apparent negative emission rates.
40

1 The comparison of interpolation techniques demonstrates that kriging is superior to IDW or natural
2 neighbour interpolations. Although our example simulations demonstrate that kriging can overestimate
3 the average real values slightly (~1%), the uncertainty is small compared to other unknowns.
4

5 The conditions of Aug 20 are clearly reproduced by the TERRA SO₂ emissions calculation. For this
6 day the SO₂ emissions show weak dependence (<1% difference) on the method of extrapolation to the
7 surface. The emissions calculated over this 2 hour and 10 min period are 4.8% higher than minute-by-
8 minute emissions (12.20 t/h) reported by CNRL. This difference could be due to non-stack sources of
9 SO₂ not included in the CNRL reported values. During normal SO₂ capture operations on Sep 2, the
10 CNRL reported value (0.224 t/h) is within the range of uncertainty (0.215 to 0.282 t/h), and is within
11 11% of the base case TERRA calculated value (0.249 t/h).
12

13 The TERRA calculated CH₄ emissions show a stronger dependence on the choice of near-surface
14 interpolation methods, as would be expected for a compound emitted from the surface. Although there is
15 some uncertainty near the surface, validation of this emission calculation is demonstrated by the similar
16 values in CH₄ emission rate estimates for the two days. Values based on varying inputs within a range of
17 uncertainty give emission rates of 3.43 ± 0.9 t/h on Aug 20 and 3.28 ± 0.8 t/h on Sep 2. Within the range
18 of uncertainty, there is no significant variation in emission rates between the dates measured.
19

20 The results of this study demonstrate that uncertainty in the emission rate calculation is very low (~2%)
21 for plumes which are entirely captured within the sampling region. For plumes with high near-surface
22 concentrations uncertainties can be as high as 28%. These uncertainties could be improved significantly
23 with simultaneous ground level measurements, especially directly downwind of the emissions source.
24

25 It is preferable to fly on days when winds and atmospheric stability are consistent to avoid plume drift
26 during the flight. If this is not possible, it is necessary to compare the changing conditions to plume
27 location along each flight path around the box, as changing conditions will not always produce
28 significant error.
29

30 Generally, the technique is most suitable for an elevated plume from an easily identifiable source, such
31 as stack emissions. The box flight pattern should be organized such that the entire plume is fully
32 captured between the minimum and maximum flight altitudes downwind of the source. We have
33 demonstrated that the technique can also be applied to ground-source surface-area emissions of
34 unidentified origin, albeit with much higher uncertainties. In the case of surface-area emissions the
35 flight path must be low enough (or the emissions source high enough) to ensure that there is adequate
36 mixing ratio above background in order to provide goodness of fit for the extrapolation between the
37 lowest flight level and the surface. The use of high-resolution atmospheric and chemical models could
38 also prove useful in determining the appropriate shape of the extrapolated fit. In these non-ideal
39 situations, surface measurements would be very valuable in constraining the extrapolated fit to surface
40 values.

1
2 Although we have demonstrated the applicability of the TERRA model to box sizes on the scale of 10 to
3 15 km, the algorithm would apply to either much smaller or much larger locations. A larger size would
4 imply greater mixing of the emitted compound, making a constant surface extrapolation most
5 appropriate; however, this would work best with a conserved compound, as the chemical reaction term
6 ($E_{C,X}$) becomes more important with greater downwind distance.

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10
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11

1 Appendix

2
3 Equations 5 and 7 are derived by differentiating $\rho_{air} = p/RT$ with respect to time to give

$$4 \frac{d\rho_{air}}{dt} = \frac{dp}{dt} \frac{1}{RT} - \frac{dT}{dt} \frac{p}{RT^2} = \rho_{air} \left(\frac{dp}{dt} \frac{1}{p} - \frac{dT}{dt} \frac{1}{T} \right) \approx \frac{\rho_{air}}{\Delta t} \left(\frac{\Delta p}{p} - \frac{\Delta T}{T} \right) \quad (A1)$$

5 where ΔT and Δp are the change in temperature and pressure over time duration Δt , and p and T are the
6 average pressure and temperature for the time duration.

7
8 The following is a demonstration of the weak dependence of the estimated emissions rate on
9 temperature and pressure changes. The advective flux through the box top is determined by solving Eq.
10 2 for $E_{air,V}$ and substituting into the $E_{C,V}$ term as

$$11 E_{C,V} = M_R \chi_{C,Top} (E_{air,M} - E_{air,H}). \quad (A2)$$

12 Equation 1 can then be rewritten as

$$13 E_C = E_{C,H} + E_{C,HT} + [M_R \chi_{C,Top} (E_{air,M} - E_{air,H})] + E_{C,VT} + E_{C,VD} - E_{C,M}. \quad (A3)$$

14 Substituting the volume integrals (Table 2) for $E_{air,M}$ and $E_{C,M}$ and collecting terms gives

$$15 E_C = E_{C,H} + E_{C,HT} + M_R \chi_{C,Top} E_{air,H} + E_{C,VT} + E_{C,VD} - M_R \iiint_{Volume} (\chi_C - \chi_{C,Top}) \frac{d\rho_{air}}{dt} dx dy dz. \quad (A4)$$

16 Hence the effect of the pressure and temperature change on the emission rate is proportional to the
17 difference between the mixing ratio within the box (χ_C) and the box-top mixing ratio ($\chi_{C,Top}$). This
18 result implies that any change in density within the box will modify the estimated emission rate
19 proportionally to the excess mixing ratio, above the ‘background’ level. Hence changes in $d\rho_{air}/dt$ of
20 more than $\pm 50\%$ result in small changes in the estimated emission rate, ranging from -2% and 5% as
21 shown in Tables 7 and 8.

22
23 The final calculated emission rate of CH_4 is dominated by the difference between the horizontal and
24 vertical advective flux terms ($E_{C,H}$) and ($E_{C,V}$) both of which are larger than the emission rate by a factor
25 of 20. Equation A4 can be expanded further using the surface integrals of air and mixing ratio advective
26 flux ($E_{C,H}$ and $E_{air,H}$ in Table 2) to give

$$27 E_C = M_R \iint_{Sides} (\chi_C - \chi_{C,Top}) \rho_{air} U_{\perp} ds dz + E_{C,HT} + E_{C,VT} + E_{C,VD} - M_R \iiint_{Volume} (\chi_C - \chi_{C,Top}) \frac{d\rho_{air}}{dt} dx dy dz. \quad (A5)$$

28 Hence, the total horizontal advection is effectively a background-subtracted mixing ratio ($\chi_C - \chi_{C,Top}$).
29 This implies that uncertainties in the wind speed and concentrations do not scale with the larger $E_{C,H}$ and
30 $E_{C,V}$ terms, but instead scale with the smaller background-subtracted mixing ratio.