

Author comment: Additional corrections

During the discussion stage of the manuscript we realized an issue with the time axes, which meanwhile could be completely resolved. The issue required corrections for those data where NABEL and Xact data were combined, i.e. for the North/South source apportionment. The regressions with the filter (ICP) data are not concerned, as these data were carefully (manually) extracted from the raw data. The changes concern Table 1 (new S1), Fig. 1 (right panel), Figs. 4, 11 (new 9), and S5, the time specifications in the text, and the mass budget estimate in Section 3.3. The changes to the statistics were minor and did not influence the conclusions. As some figures required corrections, we took advantage and also redrew other figures to make them consistent (axes ranges, element groups). Resulting changes to the text were applied where necessary.

The following changes were made:

All times are now reported in local time LT (= CET + 1 = UTC + 2).

Table 1 (new S1): Statistics for non-fireworks north and south sectors updated.

Table 3 (new 2): Column 2 ("all days") removed.

Fig. 1: Right panel updated.

Fig. 4: Xact data shifted by 1 h. The fireworks peaks of the Xact and the TEOM show a 1-h difference.

Fig. 8 (new 7): Elements arranged into groups A, B, C as in the text.

Fig. 11 (new 9): Data updated. Elements arranged into groups A, B, C as in the text.

The mass budget estimate had to be adjusted. Instead of comparing just the two 1-h maxima (TEOM and Xact), we integrated the mass over the full peaks (2 h for Xact and ACSM; 3 h for TEOM).

Response to Referee #1

(Referee's statements in *italics*)

The authors would like to thank the referee for careful reading and critical commenting of the manuscript.

The abstract is too long and should be more focused on what the real novel results are. It reads more like a conclusion than an abstract.

We shortened the abstract from 408 to 262 words, deleting the details referring to source apportionment. We focussed the abstract on the main findings regarding the intercomparison of the Xact with external measurements.

Figure 2 is too small and too many to be useful for interpretations. Needs to be redone with a focus on what the authors want us to see from this figure.

Figure 2 was redrawn, re-scaling the axes with respect to the maximum values of the filter data (ICP). Labels were enlarged. We included all data, spread over four panels, as we find it important to show the complete range of variability.

Data in tables 1 and 2 are excessive and again what is it the authors really want is to observe from these tables. Maybe plots of MDLs versus elements would be a better way to see this and more efficient. Why are regression tables like Table 2 useful - maybe a few sentences in the text with a few selected plots would show these correlations better.

We rearranged Tables 1 and 2. We moved the MDL information from Table 1 to Table 2. Then we moved Table 1, which now only comprises of statistical data characterizing the different periods, to the supplementary material (Table S1). Table 2 (new 1) now contains the regression coefficients and the MDLs. The data of Table 2 is presented in the new Figure 3, which shows the comparison between ICP and Xact MDLs, slopes and intercept-to-average concentration ratios for all studied elements. We consider it as advantageous when the data represented in Figures 2 and 3 can also be looked up quantitatively in a Table (as also supported by Referee #2), showing the full variability of slopes and intercepts.

Response to Referee #2

(Referee's statements in *italics*)

The authors would like to thank the referee for thoughtful reading and critical commenting of the manuscript.

The comparison with ICP data is correctly carried out, showing both elements with good ICP-Xact agreements and elements that are not well quantified by the spectrometer. However, in my opinion, some criticism (on both the use of this instrument and the comparison method) should be more explicitly quoted and discussed.

We add a statement on other possible sampling techniques and their benefit (see also remark further down).

We add a reference to Yatkin et al. (2012) and (2016).

"The NABEL network provided the reference for previous data intercomparisons (Hueglin et al., 2005; Lanz et al., 2010), as well as for the intercomparisons of this study. Comparisons between SR-XRF and filter samples analysed with ICP-OES and ICP-MS have been performed previously (Richard et al., 2010). Comparisons of XRF on samples collected on different substrates were performed, e.g., by Yatkin et al. (2012). A recent interlaboratory comparison of PM₁₀ filter analysis methods is presented in Yatkin et al. (2016), where XRF/PIXE and ICP methods were compared for several metrics. Some of these metrics are also applied in this study."

We added a remark on the regression analyses in Section 3.1:

"The regression intercepts were not forced to be zero to enable examination of potential differences in the measurement accuracy of each of the compared methods, e.g. blank subtraction. The slopes are more relevant and indicate biases between the methods. Orthogonal least squares regressions metrics were calculated which incorporates measurement errors in both quantities being compared. The slopes differed by less than 3.5 % between the two regression methods for the Group A elements. Ba and Pb achieved an almost perfect match with slopes around 1 and negligible intercepts. The other extreme is Zn with a slope of 1.8. Ti is another peculiar case with a slope of 1.13 and the largest intercept/average concentration ratio of 0.37. On average, the Xact 625 yielded approximately 28% higher elemental concentrations than ICP for the Group A elements."

We added the statement in the last paragraph of the conclusions:

"Xact streamlines near-real time monitoring of multi-metals despite not being as cost effective relative to conventional samplers that could be deployed in larger numbers at many sites simultaneously, or that could sample several size fractions at once, although their actual analysis costs (laboratories, accelerator facilities and staffing needs) are not considered here and they may surmount the instrument costs manifold."

It is true that synchrotron-XRF or PIXE require expensive and not-easy-to-obtain accelerator time, but at the same time (if the experiment set up is properly optimized) these techniques allow a very accurate elemental analysis of an high number of samples collected in many sampling sites in very short times, while it is difficult to have many online spectrometer to simultaneously collect the PM in different locations.

True. We have added two statements covering these aspects in the introduction:

"Access restrictions limit the number of collected samples to be analysed, and hence field campaigns are predominantly episodic."

"Instruments of this type can be used for continuous (months, years) monitoring at a site, but their cost restricts the simultaneous deployment of a larger number of devices for different size fractions or at different sites. The benefit of long-term, quasi real time data access, favourable, e.g., for air quality monitoring, contrasts with the possibilities of relatively low cost, multi-site and multi-size samplers used so far in episodic field studies."

For the elements of group A (main PM elements), it is true that the Xact-ICP correlation is very good, but it is not sufficient to say that their concentrations are well reproduced by the spectrometer (as stated in the conclusions and in the abstract). Intercepts arrive up to 40% of the average concentrations (as stated in the paper) and, even if they are not so big, I would not say that they are “small” (pag. 7) or negligible. Deviations of slopes from unit, although, again, not very big, are however significant (Xact/ICP ranging from 1 to 1.8, average 1.28). Possible reasons, like sampling and X-ray absorption, are suggested, but, as the authors themselves state, they are not completely supported/demonstrated by this study. Also they are not always convincing. In particular, X-ray absorption would produce underestimation while Xact concentrations are higher than those obtained by ICP; sampling would produce higher deviations for elements in big particles, while also S slope (1.37) significantly deviates from 1; slope of Zn (1.8) is significantly higher than the others. In this situation, it is not possible to conclude that the spectrometer correctly reproduces the concentrations of all elements of group A and that systematic differences have been attributed to specific reasons (as reported in the conclusion section). Looking at obtained results, I would conclude that correlation is very good (for group A), not big but significant differences are however observed (lower than. . .), possible reasons have been investigated but further studies are needed.

Agreed. We have changed our concluding statement to:

“Excellent correlation between Xact 625 and ICP-OES/ICP-MS was observed for 24-h averages of the elements S, K, Ca, Ti, Mn, Fe, Cu, Zn, Ba, and Pb (“Group A”). The daily averages calculated from hourly measurements by Xact were on average 30 % higher (range -3 % to +80 %, dependent on elements) than 24-h integrated filter measurements by ICP. ... Further research on these issues is needed.”

(I think it is important to keep Table 2 and all the panels of Figure 2, but fonts should be bigger as it is difficult to read them as they now are).

We have rearranged Table 1 by moving the MDL part to Table 2 and the remainder of Table 1 to the supplement which now only contains statistics of the complete Xact dataset. We have also redrawn Figure 2 completely and have replaced Figure 3 with a new one showing the interference free MDLs of Xact and ICP MDLs as well as the slopes (with standard deviations) and the ratios of the intercept with the average concentrations for each element. In this way, complete information on all intercomparison data is presented in graphical and tabular form for quick reference. See response to referee #2.

It is also important noting that the comparison is made on daily averages and the accuracy of hourly concentrations has not been directly tested.

It is true that the hourly concentrations have not been compared to ICP data at their original time resolution, since the ICP data are only available with 24h time resolution, and we included text in the manuscript stating this limitation. Nevertheless, while indeed the emphasis is on the intercomparison of 24-h averages of filters and Xact, hourly S data from Xact have been compared with PM₁ SO₄ collected with the ACSM (Figure 5). On the other hand, the estimated mass budget in Section 3.3 is also based on hourly data (2 h interval for the Xact, 3 h interval for the TEOM, to capture the full peak period with high concentrations). Fig. 4 shows hourly resolved time series of Xact, TEOM and ACSM data.

We insert the following statement at line 29 of Section 2.6:

“Comparisons of hourly Xact data were only possible for S with the ACSM data (in the form of PM₁ sulphate, assuming that all S occurs in PM₁), and between the total Xact element mass and PM₁₀ of the NABEL TEOM instrument, see Sections 3.2 and 3.3.”

And in the conclusions we added:

“The accuracy of hourly values has only been tested for the case of the fireworks peak late on 1 August 2015, where the sum of all elements has been compared to the total mass of the NABEL TEOM. Good agreement between the Xact and TEOM mass was found when corrections derived from the 24-h filter analyses were applied. This was a special case dominated by just three elements, S, K, Cl, and a generalization to all measured elements is not recommended.”

Finally, I have some comment on the description of the spectrometer (section 2.3). There are in my opinion important pieces of information that are not reported and that would be very useful: sampling area, irradiated area and X-ray detector used (including entrance window and collimation size).

True. However, having such proprietary information published is of severe concern for CES due to potential competition with other manufactures. Furthermore, we do believe that removal of such information is consistent with discussion of other methodologies (for example, most journal papers focused on ambient data from say, ICP-MS do not provide details of the liquid impingers or cones being used). We checked to ensure that the data interpretation is indeed not impacted by elimination of such details.

Minimum detection limits reported in Table 1 seem very small for 1-h sampling.

The reported MDLs are interference-free MDLs at the 1-sigma confidence level for unsampled tape (see reference Currie 1977 in the manuscript). These MDLs give an estimate on the minimum detectable mass for pure elements, i.e. for single element standards, hence they represent the achievable minimum of detection under ideal conditions. MDLs in a multi-element sample are expected to be higher, depending on sample composition and concentration, but they would have to be reported individually for each measurement. No such routine is presently implemented in the Xact. MDLs decrease with increasing sampling time ($\sim t^{-1/2}$).

“XRF based MDLs are inversely proportional to the square root of the X-ray analysis time (Currie, 1977), which in the case of Xact is same as the sampling duration. Hence, Xact MDLs are lower for longer sampling durations. Interference free MDLs, while true are idealized lower limits of detection of one single element. As with most analytical methods, matrix effects in ambient samples from interferences between different elements and analyte concentrations could potentially result in MDLs of ambient samples to be higher and vary across samples, which makes them difficult to generalize and report. It is therefore often preferred to report measurement uncertainties to characterize a measurement.”

Uncertainties are surely much higher than 5% for concentrations close to MDLs.

Agreed. We changed the wording from ‘may be higher’ to ‘are higher’ (p 4, line 14):

An uncertainty of 5 % or less due to fitting errors and uncertainties in the standards has been derived from laboratory experiments with NIST standards (benchtop XRF, filter analyses). Uncertainties are expected to be higher for concentrations close to the MDL; for elements with potential for line interferences in multi-element samples; and, from self absorption effects for lightest elements (Si, S, Cl, K, Ca).

Response to Referee #3

(Referee's statements in *italics*)

The authors would like to thank the referee for thoughtful reading and critical commenting of the manuscript.

Issue 1: in the discussion of the level of agreement between the Xact 625 results with the other standard techniques, the Authors consider that some differences could be due the use of different sampling devices placed not exactly in the same position. My question is: why the substrates used in the Xact625 have not been analyzed off-line by other techniques? This would have removed any possible ambiguity related to different amount of sampled material...I have not found in the text any comment on this possibility. I consider a detailed discussion on this point absolutely necessary.

Good point. An offline analysis of the Xact sampling tape was not considered for two reasons: 1) Hourly samples do not collect sufficient mass for ICP analysis. 2) The samples collected are typically not amenable to post-sampling offline analysis due to potential for cross-contamination from sampled filter tape wound upon itself onto the filter wheel.

We added the following statement to the text in Section 2.3:

"While this approach is non-destructive, the samples collected are typically not amenable to offline analysis post-sampling due to potential for cross-contamination from sampled filter tape wound upon itself onto the filter wheel."

Issue2: the Authors discuss quite in deep the differences in the PM composition in the two periods (with and without fireworks) of the campaign. I do not find in such discussion any new or general element which could deserve to stay in the text. I think that this falls outside the main focus of the paper and that most of this discussion should be moved to the on-line supplementary material (or maybe it should find space in some local report). On this point, see the punctual comments below.

Partly agreed. It was originally the goal to demonstrate the capabilities of the Xact in field research in Switzerland, and to show an application of Xact data for source apportionment. From all three reviewers' comments we see that focussing on the first aspect was more desired than emphasizing the source apportionment. We therefore decided to shorten the source apportionment part, and to save part of the material for an upcoming publication. On the other hand, we elaborated the intercomparison part more deeply. This led to some re-writing of various parts of the manuscript, dropping two figures (Figs. 6 and 9), moving Table 1 to the supplementary material, and keeping the source apportionment part at a level demonstrating a few basic capabilities of the Xact data. We do, however, not agree with removing the whole Section 3.3 to the supplementary material, as it demonstrates important findings for operating the Xact. We instead condensed the section to the necessary, documenting the value of 1-h time resolution and the large dynamic range of the Xact (a concentration jump to $48 \mu\text{g m}^{-3}$ within an hour). A comparison of the Xact's and TEOM's peaks could demonstrate how closely the Xact elemental mass represents the total measured PM_{10} mass in this particular case. We consider this to be a significant finding of this study.

Punctual comments:

Abstract, line 17: the wording "Xact PM10 mass" could be misleading since by ED-XRF just a small fraction of the elements presents in PM10 can be detected. I recommend to use "the total concentration of the elements detected by Xact in PM10"

Agreed. However, when condensing the abstract as suggested by Referee #1 this sentence was deleted from the abstract.

Abstract, line 19: Begin the statement with "Ten" instead of "10"

Done.

Introduction, line 38: replace “historically required” with “require”

Done.

Introduction, line 39: This is not true: there are well known methods (e.g.: streaker sampler + PIXE, DRUM impactor + SXRF) which provide hourly or even sub-hourly time resolution with very low MDL. The statement must be changed accordingly.

Agreed. We added the sentences and references at the appropriate place:

“For high time resolution, impactors are used where the sample is collected on a foil (e.g. a rotating drum impactor, Lundgren, 1967), or on a combination of an impactor plate and a filter, such as in a streaker sampler (e.g. Annegarn et al., 1992). These samples are then exposed to X-rays or a particle beam without further treatment, which provide quantitative data with low detection limits.”

Page 2, line 3: replace “similar X ray facility” with “accelerator facilities”

Done.

Page 2, line 5: delete “overwhelming”

Done.

Page 2, line 10: the advantages offered by high time resolution have been discussed in literature well before the “older” reference given in the list. . .to my memory come some papers dating back to the eighties and I think that the Authors should be more precise on this point. Two reference papers are: Annegarn et al., Source profiles by unique ratios (SPUR) analysis: Determination of source profiles from receptor-site streaker samples. Atmos. Env. 26, 1992 D’Alessandro et al., Hourly elemental composition and sources identification of fine and coarse PM10 particulate matter in four Italian towns. JAS 34, 2002

Agreed. We added the two references and quoted them in the appropriate places in the introduction.

Page 3, line 37-38: a list of element detectable by Xact is reported with the explanation that the system sensitivity has been determined for each element by a reference sample. Actually, there are in the list couples of elements which interfere in a X-ray analysis (Fe-Co, Pb-As, Ba-Ti) and I really wonder that, for instance, Co and As can be safely detected in ambient aerosol (usually much richer in Fe and Pb). I have not found in the text any comment of this point. I think that the calibration procedure should be better described including a discussion on these possible interferences. This impacts on the data summarized in Table 1 too.

Very good point. We added some text discussing this point:

“Line interference is well-known for element couples like Fe-Co, Pb-As, Ba-Ti and makes detection of one element difficult if the other is abundant in the sample. The linear least squares reference deconvolution algorithm implemented in the Xact fits the measured sample spectrum with the library of pure element reference spectra to resolve concentrations of each calibrated element.”

The last remark on the impact of line interference on the data in Table 1 concerns only the pair Ba-Ti, as for each other pair one element is below MDL. We found a high correlation between Ti and Ba for the fireworks ($r^2=0.94$), where it is expected, and a different, but still good correlation ($r^2=0.55$) for the non-fireworks cases (see attached Figure), where both elements also show a different behaviour and different enhancement ratios (Fig. 7 in the manuscript). This is a hint towards a good element separation despite the potential Ba-Ti line interference. In addition, from Fig. 2 it can be seen that both Ba and Ti are highly correlated with their corresponding ICP-MS data, with the largest values corresponding to the fireworks cases. One would expect

the Ti values of the fireworks to deviate more from the ICP data when Ba interference were not adequately considered.

We added the following sentences:

“Potential line interference between Ti and Ba can be largely excluded, because the element couple reveals two different regressions for fireworks and non-fireworks cases, as well as distinct diurnal variations in the non-fireworks cases.”

“The element couples of Fe-Co and Pb-As do not show correlations within the couples, because most of their data points are below their respective MDLs, and no conclusion about the deconvolution of interfering lines can be drawn for these elements. Comparing the Xact values with the NABEL annual mean values (Table 2) shows differences smaller than 40 % for Cu, Pb, and Ni, while the differences are much larger for As and Cd. The latter two elements are below their respective MDL (Fig. 2).”

Page 4, first lines: the procedure to determine the MDL takes into account the spectrum collected in a blank portion of the filter. This way the MDL get underestimated since, in each portion of the spectrum, the continuum is not due to the filter only but also to the tails of all the peaks due to the PM elements. Moreover, this method completely not consider the interferences discussed above. More realistic MDL values should be given for each element and a given sampling/analysis time in a dedicated table but should be calculated as an average of PM samples.

Partly agreed. The MDLs reported by the manufacturer are interference-free, 1-sigma confidence level, minimum detection limits determined for single element standards. They are clearly defined and well reproducible, and they are re-determined at regular intervals (1 to 3 months). They indicate the lowest possible detection limits under ideal conditions. While we agree that the table suggested by Referee #3 would be very helpful and probably more meaningful, its determination would require numerous experiments with different sample compositions and concentrations which is beyond the scope of the current study.

“Interference free MDLs, while true are idealized lower limits of detection of one single element. As with most analytical methods, matrix effects in ambient samples from interferences between different elements and analyte concentrations could potentially result in MDLs of ambient samples to be higher and vary across samples, which makes them difficult to generalize and report.”

Page 8, Par 3.3: the whole section with related figures should move to the supplementary material

Partly agreed, see issue 2, above. We condensed the Section 3.3 to the necessary, but want to keep it in place. We consider the Section relevant for the article.

We try to clarify this point with the following text sequence:

“Investigation of the highest peaks reveals the performance of the Xact under high load conditions, when sample thickness may become critical for XRF analysis. A comparison of the two instruments’ peaks could demonstrate how closely the Xact mass represents the total measured PM₁₀ mass. Inspection of the different time series indicates that the TEOM peak is broader (3 h) and higher (59.6 µg m⁻³), and its maximum concentration is reached 1 h later (at 2 Aug 2015 0000 LT), but its increase in concentration starts at the same time as the Xact (at 2200 LT). The delay in the maximum can be attributed to the time constant of the TEOM used for reducing measurement noise and to the averaging procedure. For a comparison of the two peaks their measured masses were integrated over the duration of the peaks, i.e. over 2 h for the Xact data and the ACSM data, and 3 h for the NABEL data..”

Pag. 11, line 9-14: the origin of possible discrepancy have not been clearly identified and the wording “are attributed” is not correct and should be replaced by “could be attributed”

Done.

Pag. 11 , line 25-31: these lines should be removed from the conclusions. . .they are of very local interest and more than a conclusion are just a summary of the findings

Agreed. Done.

Figure 1: the plot in the top panel should be shown in log. Scale: the present picture is not so informative

Not agreed. We want to emphasize the size of the fireworks peak in relation to the background concentrations, and this is best visible with a linear y-axis. Furthermore, on a logarithmic scale with stacked element concentration, the first element is shown with proportionally more area than the subsequent elements, yielding a distorted picture of its importance. Therefore, we prefer to leave the linear scale on the y-axis in the top panel as is.

Figure 2: the number of digits in the value of the a and b coeff. Is too high (i.e. show just significant digit). The values of the correlation coeff. Should be added in each plot.

Agreed. However, we redrew Fig. 2 completely and moved all the regression coefficients to (new) Table 1 (former Table 2), where all coefficients are given to 2 decimal digits. The new figure represents the elements in Group A with colours, while Group B and C elements are represented in black and white, indicating that their data has issues with MDLs and others. Still, we retain all data in the graph to present the complete picture – in accordance with referee #2. We have also reordered that data in Figs. 7 and 11 (new: 7 and 9) according to the groups, hence we consistently show good and bad data throughout the article.

Figure 3: In my opinion the right part of the picture (from V on) does not give any information and should be deleted

Partly agreed. We completely redrew Fig. 3, adding the MDLs for Xact and ICP, the regression slopes and intercept-to-average ratios for all elements. See also comment on Fig. 2, above.

Figure 5: the fit of the red points is more or less meaningless. please add the R2 values in the plot both for blue and red points

Agreed. Done.

Table 2: should be deleted by inserting the R2 values in the plot of fig. 2

The changes made to Figs. 2 and 3 and Table 2 make this suggestion obsolete.

Elemental composition of ambient aerosols measured with high temporal resolution using an online XRF spectrometer

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Abstract. An Xact 625 ambient metals monitor was tested during a three-week field campaign at the rural, traffic-influenced site Härkingen in Switzerland during summer of 2015. The field campaign encompassed the Swiss National Day fireworks event, providing increased concentrations and unique chemical signatures compared to non-fireworks (or background) periods. The objective was to ~~characterize the handling and operation of the instrument,~~ evaluate the data quality by intercomparison with other independent measurements, and test its applicability for aerosol source quantification. The Xact was configured to measure 24 elements in PM₁₀ with 1-h time resolution. ~~Hourly element concentrations ranged from a few ng m⁻³ for trace elements in background conditions to tens of µg m⁻³ for major elements during a high emission event (fireworks).~~ The total Xact element mass comprised approximately 20 % of the total PM₁₀ mass. ~~The six major elements Si, S, Cl, K, Ca, and Fe contributed 95 % to the Xact PM₁₀ mass, the remaining 5 % were attributed to the trace elements.~~ Data quality was evaluated for 10 24-h averages of Xact data by intercomparison with 24-h PM₁₀ filter data analysed with ICP-OES for major elements, ICP-MS for trace elements, and gold amalgamation atomic absorption spectrometry for Hg. 40-Ten elements (S, K, Ca, Ti, Mn, Fe, Cu, Zn, Ba, Pb) showed an excellent correlation between the compared methods, with r^2 values ≥ 0.95 . ~~These elements demonstrate the high precision of the Xact instrument, even though~~ However, the slopes of the regressions between Xact 625 and ICP data varied from 0.97 to 1.8 (average 1.28) and thus indicated generally yielded ~~approximately 28% higher Xact elemental concentrations than ICP for these elements.~~ ~~These elements demonstrate the high precision of the Xact instrument.~~ An average 28 percent Possible reasons for these differences are discussed, but further investigations are needed. ~~to ICP analyses might in part be attributed to the differences in the sampling systems (inlets), the geographic distance between the inlets and between the inlets and the freeway, and to uncertainties in the different analysis methods.~~ 40-For the remaining elements no conclusions could be drawn about their quantification for various reasons, ~~mainly detection limit issues.~~ additional elements (Cr, V, Co, Ni, As, Se, Cd, Sn, Hg, Bi) could not be compared to a reference, because their concentrations were close to or below the minimum detection limits of at least one of the analysis methods. ~~Sb revealed a calibration issue with the Xact, which requires correction. Si, Cl and Pt were not analysed with ICP, and thus could not be evaluated.~~ An indirect intercomparison of hourly values was performed for the fireworks peak, which brought good agreement of total masses when the Xact data was corrected with the regressions from the 24-h value ~~intercomparison.~~ The well-quantified elements were further used for a simple investigation of sources. ~~The field campaign encompassed the Swiss National Day fireworks event, providing increased concentrations and unique chemical signatures compared to non fireworks (or background) periods.~~ Fireworks and traffic or rural background emissions could clearly be identified with their element mixture. The results demonstrate that multi-metal characterization at high-time resolution capability of Xact is a valuable and practical tool for ambient monitoring, ~~exhibiting significant advantages compared to~~ traditional elemental analysis methods.

The quantification of trace elements in airborne particulate matter (PM) can be achieved with various techniques, such as inductively-coupled plasma mass spectrometry (ICP-MS), inductively-coupled plasma optical emission spectrometry (ICP-OES), X-ray fluorescence spectrometry (XRF), and proton-induced X-ray emission spectrometry (PIXE). These methods ~~historically required~~ a two-step procedure, i.e. sample collection in the field followed by laboratory analysis. Ambient pollutants are ~~typically conventionally~~ collected on filter substrate for large time duration such as 8-h or 24-h sampling time to ensure that sufficient elemental mass is available for analytical analysis (ICP). For high time resolution, impactors are used where the sample is collected on a foil (e.g. a rotating drum impactor, Lundgren, 1967), or on a combination of an impactor plate and a filter, such as in a streaker sampler (e.g. Annegarn et al., 1992). These samples are then exposed to X-rays or a particle beam without further treatment, which provide quantitative data with low detection limits. In contrast to the non-destructive XRF method, sample preparation for ICP analysis is very laborious, and the samples are destroyed during this process. XRF method has been successfully applied to aerosol characterization in the last decades. Measurement of low sample mass typically requires a high sensitivity (or a low minimum detection limit, MDL) and hence access to a synchrotron or ~~similar X-ray accelerator facilities~~ (Bukowiecki et al., 2005; Bukowiecki et al., 2008; Calzolari et al., 2010; Calzolari et al., 2015; Lucarelli et al., 2011; Richard et al., 2010; Visser et al., 2015b; Yatkin et al., 2016), which is notoriously difficult due to the ~~overwhelming~~ demand for analysis time at such facilities. Access restrictions limit the number of collected samples to be analysed, and hence field campaigns are predominantly episodic. Technical advances in X-ray sources and detectors have recently resulted in the development of commercial systems capable of sampling and analysing ambient PM samples in sub-hourly or hourly resolution in quasi real time. Instruments of this type can be used for continuous (months, years) monitoring at a site, but their cost may restrict the simultaneous deployment of a larger number of devices for different size fractions or at different sites. The benefit of long-term, quasi real time data access, favourable, e.g., for air quality monitoring, contrasts with the possibilities of relatively low cost, multi-site and multi-size samplers used so far in episodic field studies.

Sampling with high time resolution generates large ~~quantity-quantities~~ of data capable of capturing source emission patterns occurring at shorter time duration. For source apportionment of PM components like elements, a high time resolution of the order of 1 hour or less is advantageous, as temporally variable environmental factors such as wind direction and speed or insolation may affect transport and source processes, (e.g. resuspension, ~~(Annegarn et al., 1992; D'Alessandro et al., 2003; Sánchez-Rodas et al., 2007; Sarmiento et al., 2007; Visser et al., 2015b; Yadav and Turner, 2014).~~ One such instrument, Cooper Environmental's Xact® 625 Ambient Metals Monitor, performs in-situ automated measurements of ambient PM₁₀ or PM_{2.5} elemental concentrations for a user-defined set of 24 or more elements with a user-selected sampling time resolution of 15 to 240 minutes. The instrument is transportable and can be deployed in field campaigns where a suitable shelter with electric power and an appropriate sampling line connecting the outdoor with the indoor is available. Remote access to the data is possible during operation, which allows for a continuous, quasi real time (i.e. with a delay of one sampling interval) monitoring of the operation status and the ambient metal content. An in-depth evaluation of the forerunner instrument Xact 620 was previously published by (Park et al., 2014).

An Xact 625 monitor was deployed for a month to test the handling and data production of the instrument. A small field campaign was organized at a monitoring station of the Swiss Air Pollution Monitoring Network (NABEL), where quality-controlled air pollution measurements are performed continuously. The NABEL network provided the reference for ~~various previous~~ data intercomparisons (Hueglin et al., 2005; Lanz et al., 2010), as well as for the intercomparisons of this study.

Comparisons between SR-XRF and filter samples analysed with ICP-OES and ICP-MS have been performed previously (Richard et al., 2010). Comparisons of XRF on samples collected on different substrates were performed, e.g., by Yatkin et al. (2012). A recent interlaboratory comparison of PM₁₀ filter analysis methods is presented in Yatkin et al. (2016), where XRF/PIXE and ICP methods were compared for several metrics. Some of these metrics are also applied in this study.

The goals of this ~~technical note~~article are 1) to characterize the measurements of the test period in Härkingen and compare them with previous studies in Switzerland and elsewhere; 2) to examine the achieved data quality for the selected elements with respect to their minimum detection limits; 3) to quantify the measurement quality based on intercomparisons between the Xact and NABEL PM₁₀ data (1-h TEOM data and 24-h filter samples) for Härkingen; 4) to evaluate the applicability of the instrument at high time resolution in typical summer conditions and concentration ranges at a traffic-influenced rural site in Switzerland; and 5) to gauge the advantages of high time resolution sampling for a preliminary investigation of sources based on enhancement ratios and diurnal variability of elements. A pollution episode captured during the campaign resulted in high ambient concentrations, widening the range of studied concentrations. The selected elements represented a typical mix of elements at the selected site. In addition, a few elements notoriously difficult to measure in Switzerland due to their generally low ambient concentration were included, namely Ni, As, Pt, and Hg.

2 Experimental setup

2.1 Site characteristics

The field campaign was performed at the permanent station Härkingen (47.311877° N, 7.820453° E) of the Swiss Air Pollution Monitoring Network (NABEL, <http://www.bafu.admin.ch/luft/00612/00625/index.html?lang=en>) from 23 July until 13 Aug 2015. This station is located next to the A1 freeway, the main traffic route between eastern and western Switzerland. About 1 km to the west the A2 freeway branches off towards the north. The local terrain is level and the traffic flows freely even during rush hours, limiting incidences of excessive braking or forced acceleration. There are villages to the south and east of the site, and agricultural land immediately to the west and north. Other local activities include industrial buildings approximately 500 m to the northwest (logistics businesses), and a metal processing company to the southeast across the freeway. The site is well documented with respect to gas phase traffic emissions, PM number concentrations and particulate elemental carbon (EC, Hueglin et al., 2006), but an in-depth local source apportionment has not been realised so far except for organic aerosols measured in May 2005 (Lanz et al., 2010).

The first week of the campaign was characterized by ~~lower-moderate summer~~ temperatures (maximum temperature <25 °C, except 23 July) with some occasional rain cleansing the atmosphere. The remaining two weeks were part of a summer heat wave, with temperatures reaching 36.4 °C at the maximum, and values above 30 °C on 7 days of this period. Only one precipitation event occurred during the hot period. The Swiss National Day (1 August) fell on a Saturday, and the weekend weather promoted outdoor barbeques and fireworks. The bulk of the fireworks were burnt on 1 August after 2200 ~~CET~~LT (local time = UTC + 2 h), but some individual fireworks were also burnt on 31 July, and 2 and 3 August.

For this study, the Xact 625 monitor and a Q-ACSM (quadrupole aerosol chemical speciation monitor, Aerodyne Inc.) were installed in an air-conditioned trailer parked next to the NABEL station. This trailer was placed to the north of the freeway at ~23 m away from the centre of the freeway. This placed the trailer on the orthogonal transect between the freeway and NABEL shelter, which is located ~27 m from the centre of the freeway. The trailer's instruments were connected to the NABEL station's power grid and Ethernet.

2.2 NABEL instrumentation

The NABEL station is equipped with a broad range of air quality instrumentation and standard meteorological sensors. The relevant instruments for this field test were the Digital DA-80H HiVol sampler with a DPM 10/30/00 inlet for 24-h PM₁₀ samples collection, and a TEOM FDMS 8500 (Tapered Element Oscillating Microbalance, Filter Dynamics Measurement System, Thermo Scientific) for continuous (10-min) PM₁₀ mass concentration measurements. The time constant used for noise reduction in the TEOM and the averaging procedure caused a significant time delay in peak concentrations. For 1-h values, the random error of a TEOM as derived from parallel operation of two identical instruments is about 2 µg m⁻³.

Standard meteorological measurements such as temperature, wind speed and direction, and precipitation records are also monitored at this station. Furthermore, the station also provided hourly traffic counts for the freeway in the form of total number of vehicles, number of heavy duty vehicles (HDV), and number of light duty vehicles (LDV).

2.3 Xact 625

The Xact 625 ambient metals monitor (Cooper Environmental Services (CES), Beaverton, OR, USA) is a sampling and analysing X-ray fluorescence spectrometer designed for online, semi-continuous measurements of elements in aerosols. In this study, ambient air was sampled with a flow rate of 16.7 actual lpm (~~i.e. temperature and pressure corrected~~) through a PM₁₀ flow separator (Tisch Environmental, TE-PM10-D) and the PM collected onto a Teflon filter tape. The flow is maintained to within about 1 %. After each sampling interval the filter tape is moved into the analysis area of the spectrometer, where it is illuminated with an X-ray tube ~~in with~~ three excitation ~~modes conditions~~ (25 kV and an Al filter, 48 kV and a Pd filter, and 48 kV and a Cu filter), and the excited X-ray fluorescence is measured with an silicon drift detector (SDD) ~~X-ray detector in the energy range from 0 to 40 keV~~. During this XRF analysis, the next sample is collected on a clean spot of the filter tape. This cycle is repeated during each sampling interval, which was configured to be 60 minutes for this study. While this approach is non-destructive, the samples collected are typically not amenable to offline analysis post-sampling due to potential for cross-contamination from sampled filter tape wound upon itself onto the filter wheel. After each analysis interval, raw and calibrated (for the actual volume in units of ng m⁻³) concentration data was stored on the hard disk of the control unit. Daily advanced quality assurance checks (QA energy calibration test, QA upscale test) were performed during 30 min after midnight to monitor shifts in the calibration. Thus, the sampling interval following midnight was limited to 30 min only.

The instrument was configured to quantify 24 elements (Si, S, Cl, K, Ca, Ti, V, Cr, Mn, Fe, Co, Ni, Cu, Zn, As, Se, Cd, Sn, Sb, Ba, Pt, Hg, Pb, Bi, plus Pd for QA). Each of these elements was calibrated individually with a reference sample. ~~An uncertainty of 5 % or less due to fitting errors and uncertainties in the standards has been derived from laboratory experiments with NIST standards (benchtop XRF, filter analyses). The uncertainty may be higher for concentrations close to the minimum detection limit (MDL), and due to self absorption effects for the lightest elements (Si, S, Cl, K, Ca).~~ Minimum detection limits (MDL)s for 1-h sampling for each element are listed in Table 1. CES calculates MDLs using the sensitivity of the element and the counts in the region of interest of a blank unsampled section of tape, from where one sigma interference free detection limits are reported. XRF based MDLs are inversely proportional to the square root of the X-ray analysis time (Currie, 1977), which in the case of Xact is same as the sampling duration. Hence, Xact MDLs are lower for longer sampling durations. Interference free MDLs, while true are idealized lower limits of detection of one single element. As with most analytical methods, matrix effects in ambient samples from interferences between different elements and analyte concentrations could potentially result in MDLs of ambient samples to be higher and vary across samples, which makes them difficult to generalize and report. It is therefore often preferred to report measurement uncertainties to characterize a measurement. XRF based MDLs are inversely proportional to the square root of the X-ray analysis time so that Xact MDLs are lower for longer sampling durations (Currie, 1977). ~~XRF based MDLs are inversely proportional to the square root of the X-ray analysis time so that Xact MDLs are lower for longer sampling durations (Currie, 1977).~~ An uncertainty of 5 % or less due to fitting errors and uncertainties in the standards has been derived from laboratory experiments with NIST standards (benchtop XRF, filter analyses). Uncertainties are expected to be higher for concentrations close to the MDL; for elements with potential for line interferences in multi-element samples; and, from self absorption effects for lightest elements (Si, S, Cl, K, Ca). Line interference is well-known for element couples like Fe-Co, Pb-As, Ba-Ti and makes detection of one element difficult if the other is abundant in the sample. The linear least squares reference deconvolution algorithm implemented in the Xact fits the measured sample spectrum with the library of pure element reference spectra to resolve concentrations of each calibrated element. The Xact reports purely elemental mass

concentrations, which are the focus of discussion and unless otherwise noted ~~(e.g. SO_4^{2-}) we refer to these pure elemental concentrations.~~

2.4 Q-ACSM

A Q-ACSM (Aerodyne Inc., Billerica, MA, USA) was operated in the trailer housing the Xact 625 during the campaign ~~(Crenn et al., 2015; Ng et al., 2011)~~. The Q-ACSM determines quantitative mass spectra of non-refractory particles up to mass to charge ratios (m/z) of 150 ~~(Crenn et al., 2015; Ng et al., 2011)~~. Ion fragments were attributed mainly to organic aerosols, nitrate, sulphate, ammonium, and chloride, which comprise the reported data used in this study. The collection efficiency (CE) was determined for each spectrum according to Middlebrook et al. (2012), and its distribution peaked at the mode of 0.62 (± 0.11). 293 out of 1055 CE values of the full ACSM dataset were equal to 0.45. The Q-ACSM collected sub-micron (PM_{10}) particles and chemically analysed them in 30-min intervals, which were aggregated to 1-h averages for comparison with the Xact 625 data. All concentrations used in this study were CE corrected.

2.5 24-h PM_{10} filter samples

The 24-h PM_{10} samples collected by the HiVol sampler on quartz filters were weighed at Empa laboratory in Dübendorf, Switzerland to determine the gravimetric daily PM_{10} concentrations. These values were then used to correct the TEOM PM_{10} concentrations on a daily basis. Therefore the 24-h TEOM values correspond to 24-h gravimetric PM_{10} . Ten 24-h PM_{10} samples were analysed for their elemental composition at IDAEA-CSIC laboratory in Barcelona. A quarter of each filter was acid digested using a mix of $\text{HF}:\text{HNO}_3$ (2.5:1.25 mL), the solution was kept in a Teflon reactor at 90°C for at least 6 h, and after cooling 2.5 mL of HClO_4 were added. The acid solution was brought to evaporation and the dry residual was re-dissolved with HNO_3 and diluted with milli-Q water for subsequent ICP-OES and ICP-MS analysis. This method has been validated and used in many studies, and is discussed in detail elsewhere (Minguillón et al., 2012; Querol et al., 2001; Querol et al., 2008). A total of 41 elements from Li to U were analysed: the major elements Na, Mg, Al, P, S, K, Ca, Ti, and Fe with ICP-OES; the trace elements with ICP-MS. Si, Cl, and Pt were not analysed on the filters. Analyses of the reference material NIST 1633b (constituent elements in coal fly ash) using the same methodology as that for the samples yielded satisfactory results, with approximately 100% recoveries for the elements under study. Tests of the used methodology with respect to other ICP sample preparation and analysis methods, and applications of the methodology to NIST standards indicated the reliability of the method, exerting a maximum scatter of 10 % for any of the elements, with most uncertainty values clearly below this limit. Relative uncertainties (precision) of the ICP measurements are less than 5 % for the elements with concentrations well above their respective detection limits, whereas the overall uncertainty reflecting the entire sampling procedure, the digestion and the ICP analysis is on the order of 25 %. Minimum detection limits for ICP were determined according to Escrig et al. (2009), and the values for the elements relevant for our intercomparison are listed in Table 1. Hg was analysed with a Hg gold amalgamation atomic absorption analyser (AMA-254, LECO instruments, Botasini et al., 2013; Diez et al., 2007). The three methods are referred to as the offline or ICP methods (ICP-OES, ICP-MS), and the Hg gold amalgamation atomic absorption spectrometry is abbreviated with AuAAA in this paper.

Three of the 10 filters were also analysed with a benchtop XRF system by CES, and with ICP-MS by an independent lab (Eastern Research Group, ERG, Research Triangle Park, NC, USA) to investigate inter-laboratory scatter. ERG used a different digestion method than IDAEA-CSIC. In addition, three filters were prepared with a reference aerosol of known concentration for Fe, Cu, Zn, Sr and Pb, which then were analysed by CES, IDAEA-CSIC, and ERG, again to gain insight into the inter-laboratory scatter. Details on these data and the methods are given in the supplement of this article.

2.6 Data coverage and synchronization

The Xact 625 measurements started on 23 July 2015 1200 CETLT, and ended on 13 August 2015 0600 CETLT. The sampling interval was set to 1 h. Two interruptions occurred during the sampling period: one due to an Xact 625 computer problem (33 h), the other one due to a delayed filter tape change (10 h). The Xact dataset consists of 456 valid 1-hour samples out of 499 possible samples, attaining a coverage of 91.4 %. The NABEL data were tailored to coincide with the Xact data. The 10-min TEOM PM₁₀ values were aggregated to 1-h values to synchronize them with hourly Xact 625 measurements. Additionally, TEOM data were also adjusted to the gravimetrically determined PM₁₀ masses from the HiVol filters to provide an independent reference for intercomparisons. The data used here contained some gaps which were only partly synchronous for the selected parameters. Wind speed and direction missed 12 data points (2.4 %), precipitation 26 (5.21 %), and PM₁₀ 53 (10.6 %) at hourly time resolution. The ACSM data contained a gap of 14.5 h due to an erroneous DAQ value, which caused the data to be very noisy for that short period. These values were rejected from the analyses, and only the remaining 972 data points were averaged to 465 1-h values, which then were resampled to the 456 Xact data points. For the comparisons of the different instruments and sampling intervals, all data were resampled to the corresponding times of the Xact 625, according to the sub-classifications of the data set (e.g. according to wind sectors). For the intercomparisons with the 10 filter samples, the hourly Xact data of the corresponding days were averaged to the 24 h of the filter samples. During each 24-h period, Xact generated 23 1-h values and 1 30-min value were aggregated to 24-h daily averages. This procedure implicitly assumes that the half-hour sample of the first sampling hour is representative for the whole hour. Tests with a 23.5-h weighted average yielded differences of less than 3 % between the two calculation methods. Comparisons of hourly Xact data were only possible for S with the ACSM data (in the form of PM₁ sulphate, assuming that all S occurs in PM₁), and between the total Xact element mass and PM₁₀ of the NABEL TEOM instrument, see Sections 3.2 and 3.3.

3 Results and discussion

3.1 Data validity derived from general statistics and minimum detection limits

The complete Xact dataset is visualized in Figure 1, and general statistics are given in Table S1. The salient feature of the concentration time series is the huge peak late on 1 Aug, caused by the National Day's fireworks episode. Further peaks before and after that day warranted dividing the full data set into a fireworks period and a non-fireworks period. The fireworks period started on 31 July 2015 2200 CETLT and lasted until 4 August 2015 1100 CETLT, as will be discussed in more detail in Sect. 3.3. The remaining non-fireworks period is representative for the typical background concentrations at Härkingen, and can be compared to literature values.

MDLs for the Xact 625 and for ICP-OES/MS and the Hg AuAAA method are listed in Table 1. Note that MDLs of elements measured by Xact 625 are based on 1-hr sampling time while MDLs of filter based elemental concentrations are based on 24-hr. Generally, values below 3*MDL are expected to have much higher uncertainty. Hence, elements with more than 80 % of the data below 3*MDL were rejected from further examination. Xact 625 MDLs have not been determined by the manufacturer for Si, S, and Cl, because self-absorption effects for elements lighter than Ca become more important with decreasing atomic number (Formenti et al., 2010). However, these three elements are abundant, and we assume that they are well above their Xact detection limit. For these elements, an ICP MDL is only given for S, because Si cannot be determined in the filter samples, as it is a main constituent of the quartz filters and is also digested during sample preparation, and Cl cannot be determined by ICP. The table indicates the amount of data points >MDL in percent for the different analysis methods. The elements K, Ca, Ti, Mn, Fe, Cu, Zn, Sn, Sb, Ba, and Pb have most values above the MDL, and their measurement should thus be reliable. Seven Xact elements have >50 % of their data points below MDL and more than 90 % below 3*MDL: V, Co, As, Se, Cd, Pt, and Bi. Cr and Cd show the same behaviour for ICP. Ni revealed variable blank concentrations in the filters and could therefore not be reliably measured with ICP. Hg is also mostly below MDL in the

AuAAA measurement, and Pt was not measured with ICP at all. In summary, 12-11 elements (K, Ca, Ti, Mn, Fe, Cu, Zn, Sn, Sb, Ba, Pb) are above their MDLs for both the XRF and the offline methods, 7 elements (V, Co, As, Se, Cd, Pt, Bi) are below MDL for the XRF, and 4-3 elements (Cr, Cd, Hg) are below MDL for the offline methods (of which only Cd is below MDL for both XRF and ICP).

The comparisons between online Xact 625 and offline 24-h PM₁₀ elemental concentrations for 21 elements are shown in Fig. 2, Fig. 3, and Table 21. Only the 21 elements analysed by both methods are compared by dividing them into two groups based on data characteristics.

Group A shows excellent correlations between the two measurement methods (r^2 values >0.95) and ~~only small~~ intercepts (<40% of mean concentration), and consists of the elements *S, K, Ca, Ti, Fe, Mn, Cu, Zn, Ba, and Pb* (elements in *italics* were analysed with ICP-OES; the others by ICP-MS). The regression intercepts were not forced to be zero to enable examination of potential differences in the measurement accuracy of each of the compared methods, e.g. blank subtraction. The slopes are more relevant and indicate biases between the methods. Orthogonal least squares regressions metrics were calculated which incorporates measurement errors in both quantities being compared. The slopes differed by less than 3.5 % between the two regression methods for the Group A elements. On average, the Xact 625 yielded approximately 28% higher elemental concentrations than ICP for the Group A elements. Ba and Pb achieved an almost perfect match with slopes around 1 and negligible intercepts. The other extreme is Zn with a slope of 1.8. Ti is another peculiar case with a slope of 1.13 and the largest intercept/average concentration ratio of 0.37. On average, the Xact 625 yielded approximately 28% higher elemental concentrations than ICP for the Group A elements.

The high linearity and little scatter in the regressions testify for the precision of both the Xact and the ICP methods, but the differences in the slopes (range 0.97 to 1.8) for different elements require further investigation. No systematic deviations based on elemental molecular weight or X-ray ~~energy excitation~~ conditions were observed in these slopes. The deviations of the slopes from unity may be partially attributed to the different inlets for the Xact and the HiVol samplers (Panteliadis et al., 2012), which may produce a difference in collected mass on the order of 10 %. The inlets were not explicitly tested for their cut-off characteristics in this study. A slightly different cut-off value for the particle size may lead to differences in the collected mass, especially for the largest and heaviest particles in PM₁₀, and hence to an underestimation or overestimation of the total mass collected with a particular inlet. This may be of special relevance in a near-road setting with lots of re-suspended dust (ACES, 2012). Potential line interference between Ti and Ba can be excluded, because the element couple reveals two different regressions for fireworks and non-fireworks cases, as well as distinct diurnal variations in the non-fireworks cases.

The results of additional investigations of a few selected filters by independent labs and analytical methods for understanding these differences are discussed in the supplementary material. Examination of reference samples indicated a high precision in XRF measurements despite consistently ~~underestimating-underestimated the~~ absolute concentration (~~underestimated by range of underestimation varying between~~ 6 to 14% depending on the element). In contrast, ICP measurements indicated greater variability (range of 30% underestimation to 60% overestimation, depending on the element~~underestimation by 30% to overestimation by 60%, depending on the element~~) and hence higher uncertainty in estimated ambient concentrations. Examination of three filter samples collected during the campaign by an offline XRF instrument (by CES) and by ICP at an external laboratory (ERG) indicated a variability of about 30% for most elements, and an almost perfect match (-0.1 %) between ICP labs for the average concentrations when Ca and Se are excluded, and deviations of less than 5 % between benchtop XRF and ICP (Table S2). These results are comparable to Gerboles et al. (2011). Comparisons between the Xact and benchtop XRF in previous tests by the manufacturer with better control of the sampler inlet conditions almost always were within 5 % of each other. The relative mean difference of 28 % between Xact and filter data (analysed with ICP) for samples collected during the field campaign appears to be systematic. Such differences may result from a difference in location of the Xact and filter sampling inlets (~5 m) and their relative distance

from the freeway. Ultrafine particle number concentrations from dust resuspension due to vehicle traffic are known to decrease with increasing distance from the road, with the sharpest decline observed within the first 50 m (Hagler et al., 2009). Crustal elements, which dominate in the PM₁₀ size fraction, are expected to settle faster due to larger aerosol size. Hence the difference in Xact and ICP reported PM₁₀ elemental concentrations may be indicative of a gradient in PM occurring for some elements in close proximity to roadways. To quantify ~~such near-source PM gradient~~~~the different effects~~, a field campaign with a different design would be needed, e.g. an array of samplers along a line perpendicular to the freeway. However, since the difference is also observed for S, which is typically found in the fine mode, does not have a major traffic related source ~~apart from coarse re-suspended dust~~, and is not expected to suffer from incomplete digestion we assign part of the differences also to calibration issues ~~in the~~~~with~~ Xact.

Group B, the remainder, consists of the elements V, Cr, Co, As, Se, Cd, and Bi, i.e. of those elements that are close to or below their Xact MDL, plus Ni, Cd, Sn, Sb and Hg. Ni, Cd and Hg were below MDL for the offline methods (Cd for Xact and offline methods). Although an intercomparison of these elements may not be justified, we observed some features in the regressions of the Group B elements in Fig. 2 that are worth commenting. Cr is below the ICP-MS MDL for 60 % and below 3*MDL for all filter values, but 75% are above the Xact MDL. The ICP measurement is below MDL because of the high and variable blank concentrations, which make a meaningful blank subtraction difficult, and which increases the Cr MDL in these samples. Although the slope of Cr is 1.23 and thus comparable to the other Group A slopes, a comparison with ICP values is statistically not robust. However, Cr seems to be quantifiable with the Xact. The regression plot of Bi shows two extreme values on 31 Jul and 1 Aug corresponding to the fireworks days. These two points are above MDL for both methods and suggest good quantitative agreement between XRF and ICP for these two high-concentration cases. Sb shows a moderate correlation ($r^2 = 0.47$), and a large intercept. Sn behaves similarly as Sb, with an r^2 value of 0.15. The large intercepts hint toward a problem in processing the Xact blanks. In addition, when Ca is abundant, as in our case in Härkingen, the Sb La line interferes with the Ca K α line, producing low signal-to-noise ratios for Sb, and similarly for the pair K – Sn. Hence, the reported Xact concentrations of these two elements reflect mainly spectral noise. 60 % of the filter Hg data were below MDL and thus cannot be well compared with the Xact Hg data. Inspection of the Xact raw Hg spectra showed a possible interference from Br causing the fitting routine to attribute some Br mass to non-existent Hg peaks in the spectra. Br was not calibrated in the fitting routine. Thus the Hg concentrations reported by the Xact seem to be due to this interference and are not realistic, even though 86 % of the measured values are above the Xact MDL. Values $< 1.5 \text{ ng m}^{-3}$ are in the same order of magnitude as the fortnight values of Chiaradia and Cupelin (2000) for the city of Geneva (Table 32). ~~The element couples of Fe-Co and Pb-As do not show correlations within the couples, because most of their data points are below their respective MDLs, and no conclusion about the deconvolution of interfering lines can be drawn for these elements. Comparing the Xact values with the NABEL annual mean values (Table 2) shows differences smaller than 40 % for Cu, Pb, and Ni, while the differences are much larger for As and Cd. The latter two elements are below their respective MDL (Fig. 2).~~ To summarize, the Group B elements show issues with the minimum detection limits of at least one of the analysis methods, which ~~restricts thorough interpretation~~~~made a comparison meaningless~~. Individual data points above MDL reveal nevertheless a usable quantification by the Xact in these particular cases. ~~Determination of Sb, Sn and Hg by Xact are potentially impacted by XRF line interferences and warrant further improvements for better quantification by Xact. Sb, Sn and Hg showed instrumental problems (line interferences) for the Xact technique that need to be improved.~~

The Group C elements Si, Cl, and Pt were not measured on the filters. An Xact MDL for S has not yet been determined. ~~Data quality of S measurements was inferred by. We can obtain information on the accuracy of the S measurement by~~ comparing its Xact concentrations with the concentrations of another element originating from the same source, or belonging to the same chemical compound. ~~Hence, Xact reported~~ S (with unknown MDL) and K (with known MDL) concentrations were ~~used to this end. K and S were~~ highly correlated ($r^2 = 0.99$) during the fireworks period, with a slope of 2.30 ± 0.05 ~~for K vs. S concentrations~~, which agrees with the stoichiometric relation between K and S when forming K₂SO₄. For the non-

fireworks periods the correlation was weak ($r^2 = 0.16$), which hints towards a completely different, more random relationship between the two elements, as expected. ~~Hence, high concentration K reported by the Xact provided reliable S during the fireworks episode. Based on these results, it can be concluded that S can be measured as reliably as K by the Xact,~~ despite the lack of an established MDL for S. Xact Pt measurements were always below MDL, and no conclusion about the Pt accuracy can be drawn.

3.2 Comparisons with other data

Figure 1 shows that roughly 95% of the total analysed elemental mass by Xact is comprised of 6 elements: Si, S, Cl, K, Ca, and Fe. These major elements all show average concentrations $>100 \text{ ng m}^{-3}$. Si, S, Ca, and Fe are observed throughout the study, although with high variability. Cl and K are abundant only episodically: Cl is strongly tied to westerly winds during the last week of July, and is practically absent after 2 Aug. K is prominent during the fireworks period. Ti, Cu, and Zn show daily average concentrations between 11 and 34 ng m^{-3} . The other analysed elements were found in daily average concentrations $<10 \text{ ng m}^{-3}$. The concentrations are of the same order of magnitude as those recently measured at other places in Switzerland, e.g. at an urban background site in Zurich (Minguillón et al., 2012; Richard et al., 2011, Table 32), but are generally lower than the measurements in older studies (Chiaradia and Cupelin, 2000; Gälli Purghart et al., 1990; Rösli et al., 2001, Table 32). The decreasing trends in PM and trace element concentrations have been documented in numerous NABEL reports on the air quality in Switzerland (e.g., BAFU and Empa, 2015; Gianini et al., 2012). These trends make it preferable to use modern studies for comparisons. Furthermore, the episodic nature of the 2015 campaign also demands for some generosity when comparing the measured values with annual or seasonal mean values.

A time series of the Xact 625 total element concentrations together with the NABEL TEOM PM_{10} data and the total ACSM non-refractory (NR)- PM_1 concentrations with 1-h resolution is presented in Figure 4. The total ACSM NR- PM_1 concentration is the sum of sulphates, nitrates, ammonia, chlorides and organic aerosols. Total PM_{10} shows a generally increasing trend over the whole campaign, with a strong peak superposed on 1 Aug 2015, ~~which-whose increase~~ coincides with the peak in the Xact data, ~~whereas the maximum is delayed by one hour relative to the Xact data.~~ The peak is due to the fireworks burnt on that evening. On average the Xact 625 elements make up about 20 % of the total PM_{10} mass (regression slope 0.2, $r^2 = 0.63$). A complete mass closure cannot be achieved, because the NABEL station only reports the total gravimetric PM_{10} mass and $\text{PM}_{2.5}$ elemental carbon (EC) concentrations with diurnal or better time resolution.

~~The Xact's~~ measurement accuracy for S was tested by comparison with ACSM sulphate measurements (Figure 5) performed at 1-h resolution. The S concentrations of ~~the~~ Xact 625 were multiplied with a factor of 3, assuming that all S occurred in the form of SO_4^{2-} . The slope of the regression line for the non-fireworks case is 1.34, with $r^2 = 0.85$, in line with the Group A elements, and in agreement with the comparison of S from Xact and from ICP, hence the slope between ACSM SO_4 and ICP SO_4 would be ~ 1 . The high linear correlation suggests a high precision of the Xact 625 data, but does not allow a definitive answer on the accuracy because of expected self-absorption effects, ~~which would increase the slope further if corrected. The comparison for~~ during the fireworks period, ~~looks different. T~~he scatter is large, and the correlation coefficient is only 0.1. We hypothesize that fireworks produce larger and non-refractory particles (e.g. K_2SO_4) not measured by the ACSM.

In summary, the on average 25 to 30 % difference between the Xact and ICP data can probably be explained by differences in the sampling inlets, the distance between the inlets, and uncertainties of the different analysis methods. The correlation coefficients close to 1 for many elements demonstrate the high precision of the Xact and ICP methods. The obtained time series of those elements can thus reliably be used for source apportionment. The subsequent analyses (e.g. elemental concentration ratios, enhancement ratios) were done with the unmodified Xact data. The only exception is estimation of a mass budget in the discussion of the extreme concentrations in section 3.3.

3.3 Extreme concentrations: the fireworks period

As mentioned in Sect. 3.1 the measurement campaign can be divided into a fireworks and a non-fireworks period. A K concentration $> 220 \text{ ng m}^{-3}$ served as the criterion to distinguish between these periods, and we required the fireworks period to be contiguous from the first increase in K on 31 July 2015 2200 CET-LT to the final decay to background values on 4 August 2015 1100 CET-LT. The average K concentration during the fireworks period was $2 \text{ } \mu\text{g m}^{-3}$, but this period showed extremely high hourly PM_{10} concentrations and an element mix different from the remainder of our test campaign.

Investigation of the highest peaks reveals the performance of the Xact under high load conditions, when sample thickness may become critical for XRF analysis. A comparison of the two instruments' peaks could demonstrate how closely the Xact mass represents the total measured PM_{10} mass. Inspection of the different time series indicates that the TEOM peak is broader (3 h) and higher ($59.6 \text{ } \mu\text{g m}^{-3}$), and its maximum concentration is reached 1 h later (at 2 Aug 2015 0000 LT), but its increase in concentration starts at the same time as the Xact (at 2200 LT). The delay in the maximum can be attributed to the time constant of the TEOM used for reducing measurement noise and to the averaging procedure. For a comparison of the two peaks their measured masses were integrated over the duration of the peaks, i.e. over 2 h for the Xact data and the ACSM data, and 3 h for the NABEL data.

~~Figure 4 shows an extreme peak on 1 Aug 2015 2300 CET, when the integrated NABEL PM_{10} mass reached a 1-h concentration of $59.6 \text{ } \mu\text{g m}^{-3}$. The Xact 625 monitor reported a total of $48.4 \text{ } \mu\text{g m}^{-3}$ for the sum of all analysed elements (except Pd, which was used only as an internal standard).~~

The bulk of this concentration ($84.3 \text{ } \mu\text{g m}^{-3}$) was made up of a few elements (in brackets: concentrations in $\text{ } \mu\text{g m}^{-3}$, and abundances relative to total analysed element mass, $\text{PM}_{10-\text{element}}$): K (2748.36 , 56.56%), S ($12.021.5$, $24.925.1\%$), Cl ($4.57.4$, $9.28.6\%$), Fe ($1.53.2$, $3.2.7\%$), Ba (12.1 , $2.3.4\%$), Si (1.70 , $2.21.9\%$). Absolute K and S concentrations are in good agreement with the values in Drewnick et al. (2006). K likely originated from KNO_3 , a basic constituent of black powder (Drewnick et al., 2006; Kong et al., 2015). ~~The period was characterized by four strong peaks (with decaying intensity) in the K/S ratio (Fig. 6). The expected K/S ratio for black powder is 2.76 (Drewnick et al., 2006), and was nearly attained on the first two evenings.~~

~~The ACSM NH_4 did not show an effect of the fireworks in the time series (Figure 6); NO_3 and organic aerosol showed a quick drop immediately before due to a wind shift, and only SO_4 and chloride showed a five fold and 2.6 fold increase, respectively, at the time of maximum fireworks activity, relative to the pre fireworks period, and a subsequent decay. The absence of fireworks NO_3 has been observed previously (Drewnick et al., 2006) and indicates that all nitrogen of the black powder is converted into N_2 or NO_x . The NO_x option is not supported by a strong increase in NO_x in the NABEL data, nor is there a strong correlation between NO_x and K, and hence our measurements indicate the N_2 pathway. SO_4 reached a peak concentration of $5.9 \text{ } \mu\text{g m}^{-3}$, and chloride reached $0.5 \text{ } \mu\text{g m}^{-3}$. The organic aerosol concentrations showed a value of $3.2 \text{ } \mu\text{g m}^{-3}$, and were further increasing after the fireworks. The other species showed minima during the hours before the fireworks, which coincides with northwesterly flow, followed by a slow increase and south westerly flow over the next few hours. In total, the ACSM PM_1 contributed $11.920.5 \text{ } \mu\text{g m}^{-3}$ (25.17%) to the aerosol concentration in the fireworks hour peak (2300 CET). The SO_4 peaks coincide with a strong drop of the PM_1 NO_3 concentration, which indicates that NH_4NO_3 reacted with H_2SO_4 , and HNO_3 escaped into the gas phase. The slight increase in ACSM chloride may be due to reaction of NH_3 with HCl .~~

The comparison of the Xact concentration and the TEOM PM_{10} concentration for the 1 August peak in Fig. 4 requires taking into account the systematic difference between the Xact and TEOM measurements discussed above, and the fact that the elements are typically not present in elemental but rather in their oxidized form, such that the mass of the latter needs to be included for a quantitative comparison. We therefore estimate a mass budget for the fireworks peak hour at 2300 h CET, when the six elements K, S, Cl, Fe, Ba, and Si comprise the bulk of the total mass. We calculate the ion balance for the positive ions K^+ , Fe^{2+} , Ba^{2+} , Si^{2+} , submicron NH_4^+ and negative ions Cl^- , SO_4^{2-} , and submicron NO_3^- . We further add all

available components from the NABEL station (PM_{2.5} ~~ECeBC~~) and the ACSM (NO₃, NH₄, organic aerosols, but not SO₄, as this is already considered with the S in the Xact data) to the mass balance. Using the measured values of the Xact 625 yields an excess of negative ions of 2.7 %, and a total mass of ~~77.6141.0~~ 59.6121.9 µg m⁻³, which overestimates the TEOM value of ~~59.6121.9~~ µg m⁻³ by ~~30-16~~ %. If ~~we scale the~~ Xact concentrations were scaled towards the ICP concentrations with the corresponding regression slopes from Table ~~21~~, using for Cl and S the average slope of 1.22 from the other four elements, then the calculation yields an excess of 11% in positive ions which are then assumed to be neutralized by oxygen. This yields a total mass of ~~63.7116.3~~ µg m⁻³, which is ~~only 74.6 % higher-lower~~ than the TEOM value. Our values are lower limits of the total mass, because the balance is incomplete with respect to relevant elements in the fireworks (e.g. Sr) and other chemical species like carbonaceous and nitrogen containing molecules in the coarse fraction. The result shows that the bulk of fireworks PM₁₀ aerosols are a few metals oxidised to sulphates, chlorides, and oxides. Overall, these results further demonstrate the advantage of Xact's high time resolution sampling for associating high metal concentration episodes with source emission activities. ~~The result further demonstrates the applicability of the Xact in conditions with high concentrations, and the advantages of high time resolution sampling.~~

3.4 Investigation of sources

Trace elements can be excellent tracers for specific aerosol sources (e.g. Hopke, 2016; Park et al., 2014; Querol et al., 2007; Visser et al., 2015a). A simple approach for characterizing a common source for a group of elements is to study the temporal covariation of the elements in this group. For our Härkingen data, the time series indicate the strong influence of the fireworks on the concentrations of K, S, Ti, Cu, and Ba (Figures ~~7-6~~ and ~~87~~), which are important constituents of fireworks along with Sr as fireworks tracer (Kong et al., 2015; Moreno et al., 2007). ~~In addition, we would expect Sr as a fireworks tracer (Kong et al., 2015; Moreno et al., 2007). We checked~~ Examination of a few raw spectra from the fireworks and non-fireworks periods indeed identified and could clearly identify an enhancement of Sr during the fireworks duration, while the peak was definitively absent during the non-fireworks period. Sr was, however, not quantified in our configuration, as we put emphasis on crustal elements and some special trace elements difficult to detect in Switzerland (Hg, Pt). ~~The gradual decay of the K/S ratio to ~0.5 over the fireworks period (Figure 6) hints towards a depletion of K relative to S, which may indicate the increasing presence of secondary sulphate from SO₂ oxidation, or to the influence of a source other than fireworks. Barbecues are a typical summer evening activity, especially on weekends. The charcoal K/S ratio ranges from 0.3 to 2 (Humphreys and Ironside, 1980), which brackets our measurements.~~

The 1-hour sampling interval further enabled examination ~~allows for the resolution~~ of diurnal variations of the elements. Ca and Ba are presented in Figure ~~7-6~~ and the other elements in the supplement (Fig. ~~S3S4~~), which shows the classification of the data according to fireworks and non-fireworks periods. It can be seen for the time from 2300 h to 0600 h that the elements Ba, Bi, Cu, K, S, and Ti show a clear distinction between the two periods. The fireworks elements show a maximum concentration at 2300 ~~h-LT~~ and a gradual decay over the next 6 to 12 h into the morning hours of the (following) day. Mn, Fe and Zn also show an increased and then decaying concentration after midnight, but the difference compared to non-fireworks days is within the data variability. Si is depleted during the fireworks period relative to the non-fireworks period. This is probably a weekday vs. weekend effect, when fewer heavy duty vehicles (HDV) circulate (Switzerland does not permit HDV use on Sundays), and less road dust is re-suspended (Bukowiecki et al., 2009; 2010). For the non-fireworks cases the transition elements Mn, Fe, Zn, and the element Pb are characterized by a broad morning peak with a maximum around 1000 ~~hLT~~, correlating well with the increasing traffic in the morning hours, and the breakup of the temperature inversion. ~~The non fireworks Si increase in the morning hours runs parallel to the increasing traffic and NO_x (Fig. 9). However, Si and traffic deviate in the afternoon, when traffic still increases until the evening rush hour, while both Si and total PM concentrations decrease. The Si curve thus resembles more the number of HDV, which remains constant throughout the working hours, than the total number of vehicles (Bukowiecki et al., 2010).~~ It should be pointed out that

this observation is consistent with the differences between the Xact and ICP-MS being due to the sampler's closer proximity to the road.

To identify the fireworks tracers, an enhancement ratio (ER) was defined as the ratio between the mean concentrations of an element for the fireworks period to the concentrations in the non-fireworks period (Figure 87). For K, Cu, and Ba the ER is larger than 2 (Cu), and goes up to 10.6 (Ba). S, Cl, Ti, Zn and Pb show an intermediate ER between 1 and 2. Cr, Mn, and Fe ER are close to 1. Si and Ca are depleted with an ER around 0.5, both of which are probably related to the above weekend effect. The elements with the high ER are clearly identified as elements of fireworks: S, K, Ti, Cu, Zn, Ba, Bi.

Further refinement of sources can be obtained when classifying the non-fireworks data by wind direction into a north (270° – 0° – 90°) and a south (90° – 180° – 270°) sector (Figure 108), with the south sector more strongly influenced by highway emissions (Hueglin et al., 2006). ~~The freeway runs from 120° to 270°, but a shift of the applied wind sectors by 20° showed no significant difference.~~ The north sector characterizes the (rural) background concentrations of the central Swiss plateau. Table S1 summarizes the mean element concentrations for the campaign divided into the different periods and wind sectors. Ba, Cr, Cu, Fe, and Mn show the signature of continuous freeway traffic emissions during the day. Pb and Zn show a morning peak only and are well correlated in both sectors. Si, K, Ca, and Ti show another pattern that could reflect the local and regional transport of crustal material partly re-suspended by traffic (south sector), partly originating from the agricultural area north of the freeway. ~~S shows a high variability and no clear difference between the sectors, and an interpretation is difficult. This could be the result of a specific, perhaps episodic wind pattern advecting higher concentrations from south during the night and from north during the day. Cl also shows a unique behaviour. An increase in Cl was seen in the ACSM data only during the fireworks. We therefore conclude that Cl during the non fireworks period was of regional (probably maritime) origin and hence rather an indicator of long range transport. A full understanding of the Cl behaviour requires a more detailed study of the wind field and a more sophisticated source apportionment which is beyond the scope of this study.~~ Figure 119 shows the enhancement ratios south/north. Apart from Cl, all south – north differences are positive, and Si, S, Ca, and Fe concentration differences are larger than 80 ng m⁻³. These are mainly crustal elements (although Fe is also emitted from vehicles). The enhancement ratios of the transition elements Cr, Mn, Fe, Cu, Zn, Ba, and Pb are larger than 1.2 and related to traffic emissions (engine abrasion, tyre wear, brake wear).

4 Conclusions

A three-week test of a Cooper Environmental Xact 625 Ambient Metals Monitor was performed at the Swiss NABEL station Härkingen. The instrument was configured to measure 24 elements continuously with 1-h time resolution. The selection of elements ranged from Si to Bi, thus covering a range of environmentally relevant elements. Besides the 'standard' elements from K to Pb, which have been well characterized by the manufacturer with respect to their accuracies and detection limits, we included several abundant light elements (Si, S, Cl) and – more for curiosity - some low-concentration elements (As, Pt, Hg) in our selection to test the behaviour of the instrument in a typical Swiss environment. We tested the measurement quality of the Xact 625 by intercomparison with well-established methodologies (ICP-OES and ICP-MS analyses on 24-h PM₁₀ samples for major and trace elements, and AuAAA for Hg), ACSM, and TEOM, and used additional meteorological data for the interpretation of the results.

The general findings are:

- The total of elements analysed with the Xact comprised of approximately 20 % of the PM₁₀ mass.
- The Xact 625 produced element concentration time series that were highly correlated with the ICP analyses of 24-h filter samples ($r^2 \geq 0.95$), even though the slopes deviated from 1.
- Element concentrations ranged from ng m⁻³ (in background conditions) to tens of µg m⁻³ (during the fireworks), and no instability in operation due to sample overload or else could be observed.

- Measured concentrations agreed reasonably well with other recent field measurements in Switzerland.

The results for measurement accuracy, precision and data quality are:

- ~~We found an excellent~~ Excellent correlation between Xact 625 and ICP-OES/ICP-MS was observed for 24-h averages of the elements S, K, Ca, Ti, Mn, Fe, Cu, Zn, Ba, and Pb (“Group A”). The daily averages calculated from hourly measurements by Xact were on average 30 % higher (range -3 % to +80 %, dependent on elements) than 24-h integrated filter measurements by ICP, indicating that all methods reproduce these concentrations well. Systematic differences of on average 25 to 30 % ~~are~~ could be attributed to physical reasons in the experiment settings, such as the different characteristics of the two inlet systems, the distance between the inlets and to the main source (freeway), and uncertainties in the various analysis methods. For XRF this includes particle size dependent self-absorption effects for the lighter elements and line interferences between different elements. For ICP this includes the entire sampling, digestion and the analysis procedure, as indicated by limited inter-laboratory and inter-method comparisons, as well as the impurities in blank filters). Further research on these issues is needed.
- The accuracy of hourly values has only been tested for the case of the fireworks peak late on 1 August 2015, where the sum of all elements has been compared to the total mass of the NABEL TEOM. Good agreement between the Xact and TEOM mass was found when corrections derived from the 24-h filter analyses were applied. This was a special case dominated by just three elements, S, K, Cl, and a generalization to all measured elements is not recommended.
- The remaining elements (“Group B”) of the filter intercomparison, V, Cr, Co, Ni, As, Se, Cd, Sn, Sb, Hg, and Bi (11 elements) were mostly below detection limit of at least one method, or showed issues with the analysis procedures (Sn, Sb, Hg). A general quantitative statement on their quality could not be achieved. Notice here that a longer sampling time, e.g. 2 or 4 hours, would have lowered the Xact MDLs and therefore increased the number of good measurements, but ~~on at~~ the cost of a reduced time resolution.
- Si and Cl were not analysed on the filters, and their Xact detection limits have not yet been determined. Hence their accuracies could not be quantified directly. Some indirect approaches were calculated.
- The Pt concentrations reported by the Xact 625 were below MDL, and Pt was not analysed on the filters. No conclusion about the accuracy of this difficult to measure element can be drawn.

~~The results from the investigation of sources indicate:~~

- ~~The period influenced by fireworks was clearly distinguished from the normal conditions. It showed extremely high or strongly enhanced concentrations of elements S, K, Cl, Cu, Zn, Ba, and Bi.~~
- ~~The normal, non-fireworks conditions could be split according to wind direction into a traffic-influenced sector and a rural background sector. The enhancement of traffic-related elements relative to the background mix could clearly be shown.~~
- ~~Average diurnal variations of element concentrations could be calculated. They further demonstrated the capabilities of the Xact 625 instrument to refine the investigation of PM sources.~~

Compared to rotating drum impactor sampling with synchrotron radiation induced XRF or streaker sampling with PIXE analysis, the Xact 625 measures ambient concentrations of the most relevant elements in near real time, and provides data with a delay of only one sampling/analysis cycle. Continuous operation capability of Xact also circumvents the sample number limitations due to restricted beamtime assignments at synchrotrons. This is a major advantage compared to the usual time delay of a couple months caused by the restricted access to synchrotron or accelerator facilities. Of course, the high time resolution of the Xact 625 comes at the cost of sensitivity, visible in the minimum detection limits, which are higher than the MDLs for the offline methods. Xact can be set for longer sampling intervals to extend the number of samples with analysed elements above their MDLs depending on the objective of monitoring campaigns. Xact streamlines near-real time monitoring of multi-metals despite not being as cost effective ~~In our short-time test study, we fixed the sampling interval to 1~~

h, but longer sampling intervals and therefore lower MDLs could be set at the instrument, thus extending the number of successfully analysed elements. Another advantage is the continuous operation capability that circumvents sample number limitations due to restricted beamtime assignments at synchrotrons. This enables long term sampling and routine monitoring relative to conventional samplers that could be deployed in larger numbers at many sites simultaneously, or that could sample several size fractions at once, although their actual analysis costs (laboratories, accelerator facilities and staffing needs) are not considered here and they may surmount the instrument costs manifold. Overall, high time resolution sampling of metals provides a rich dataset for associating high metal concentration episodes with source emission activities. Useful extensions of the present capabilities of the Xact could be the addition of more elements to be analysed (especially under the circumstance that the full mix of observed elements cannot always be known in advance), improved quantification of the lightest elements (especially their MDLs), a vacuum or helium device for analysing light elements like Na and Mg, and an inlet switch to alternately measure PM₁₀ and PM_{2.5} with one single instrument.

5 Supplementary Material

6 Competing interests

Krag Petterson and Varun Yadav are employed by Cooper Environmental Services, the manufacturer of the Xact® 625.

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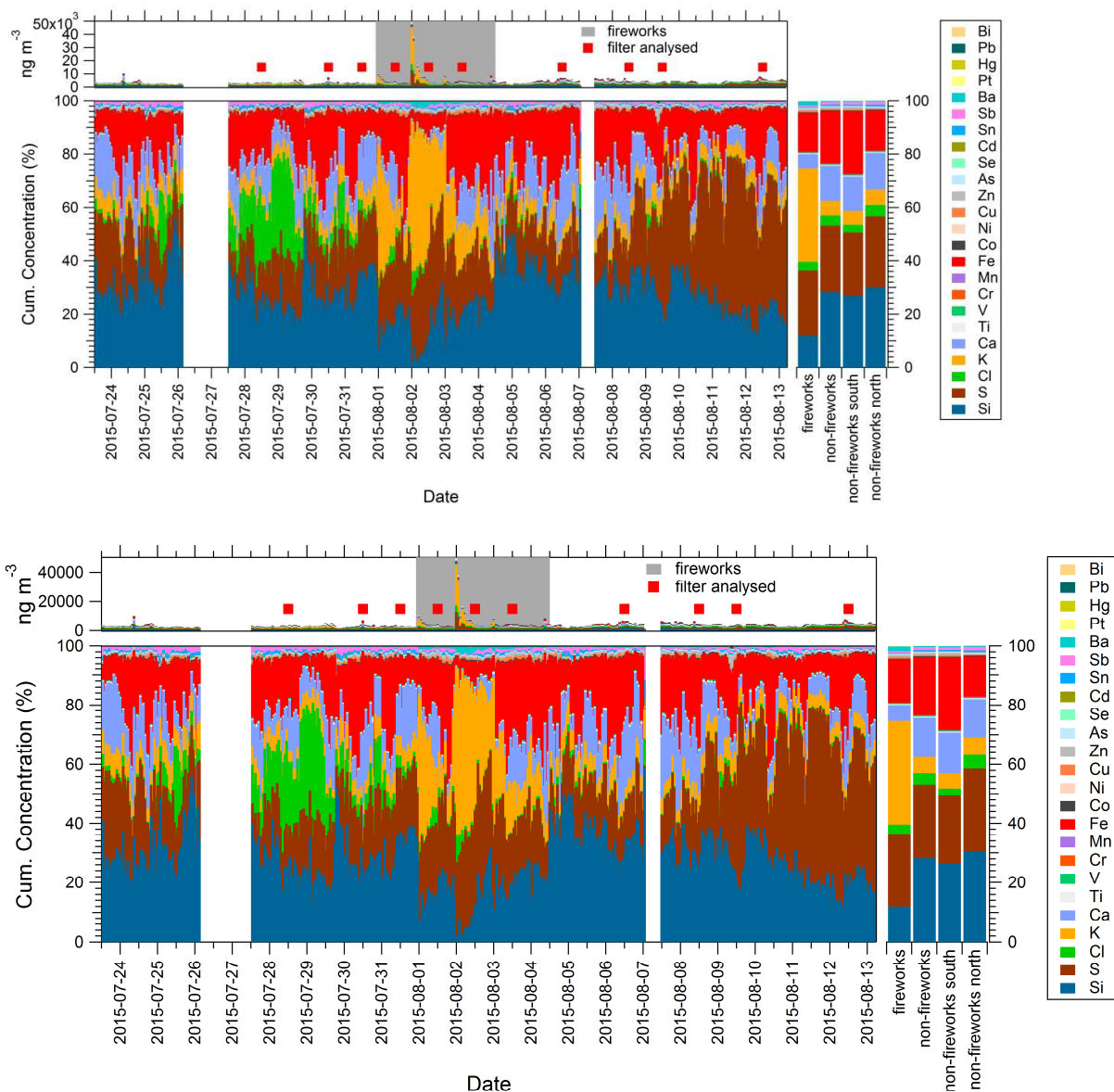
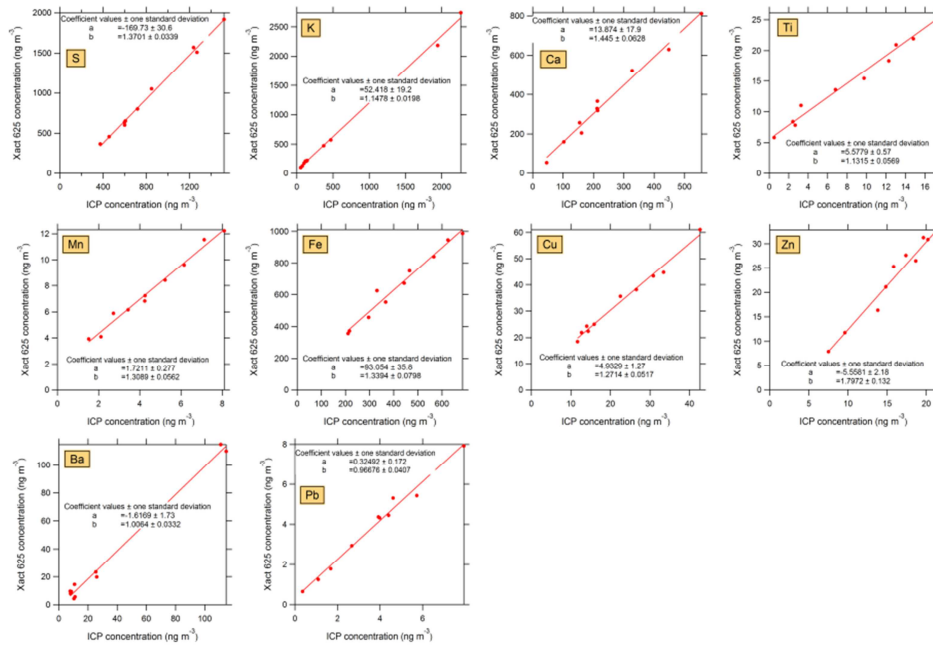
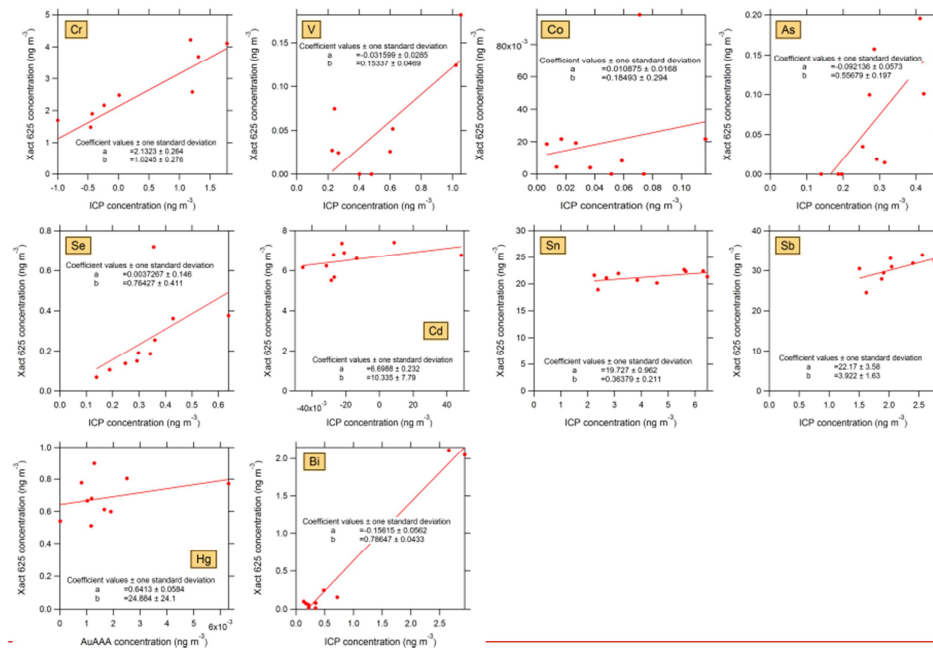


Figure 1: Main panel: Relative amount of analysed elements by Xact 625 during the field campaign. Top panel: Absolute concentrations, stacked. The grey shaded area denotes the fireworks period. The red squares mark the days when 24-h filters were analysed and used for comparisons in this study. Bottom panel: relative cumulative elemental concentrations, stacked. Right panel: Average relative contributions (in %) of elements for the fireworks period, the non-fireworks period, and for the south and north sectors during the non-fireworks period.

Group A



Group B



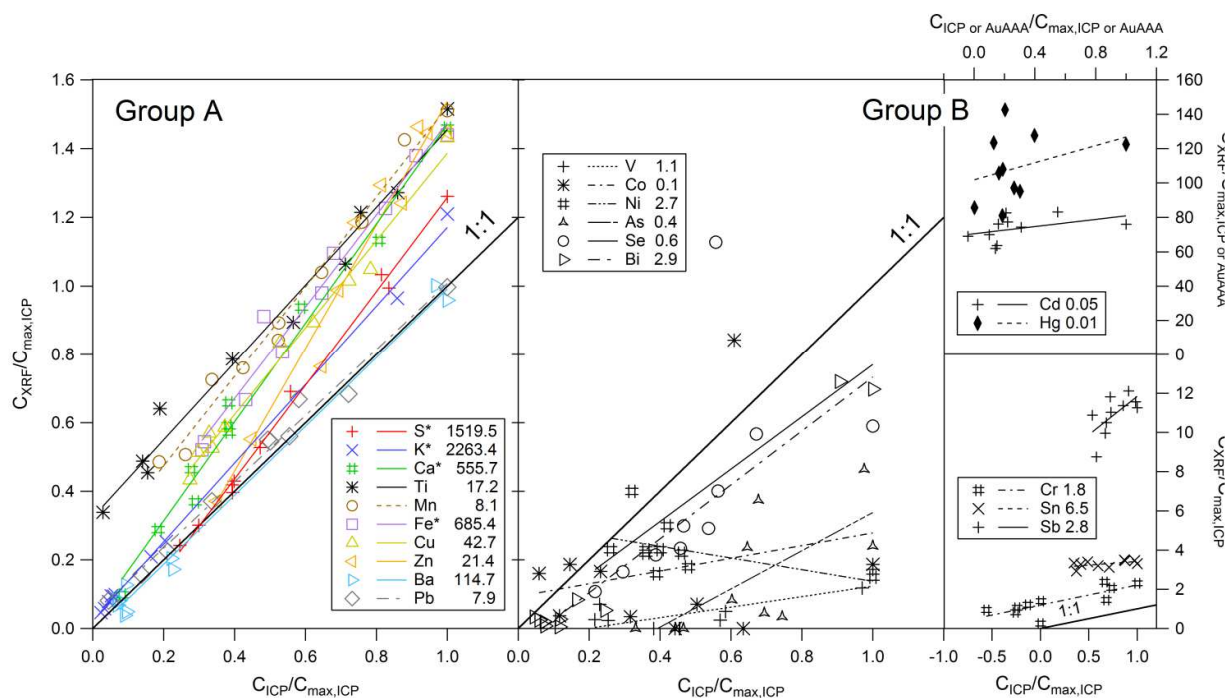


Figure 2: Scatterplots and regression lines of Xact 625 (ordinate) vs. ICP-OES/MS (abscissa) data for Groups A and B. The axes have been scaled by the maximum concentration $C_{\max,ICP}$ for each element ($C_{\max,AuAAA}$ for Hg). The Levenberg-Marquardt linear least squares fitting method was applied, taking the ICP measurements as the independent data. Regression equation is $y = a + bx$. See Table 1 for data.

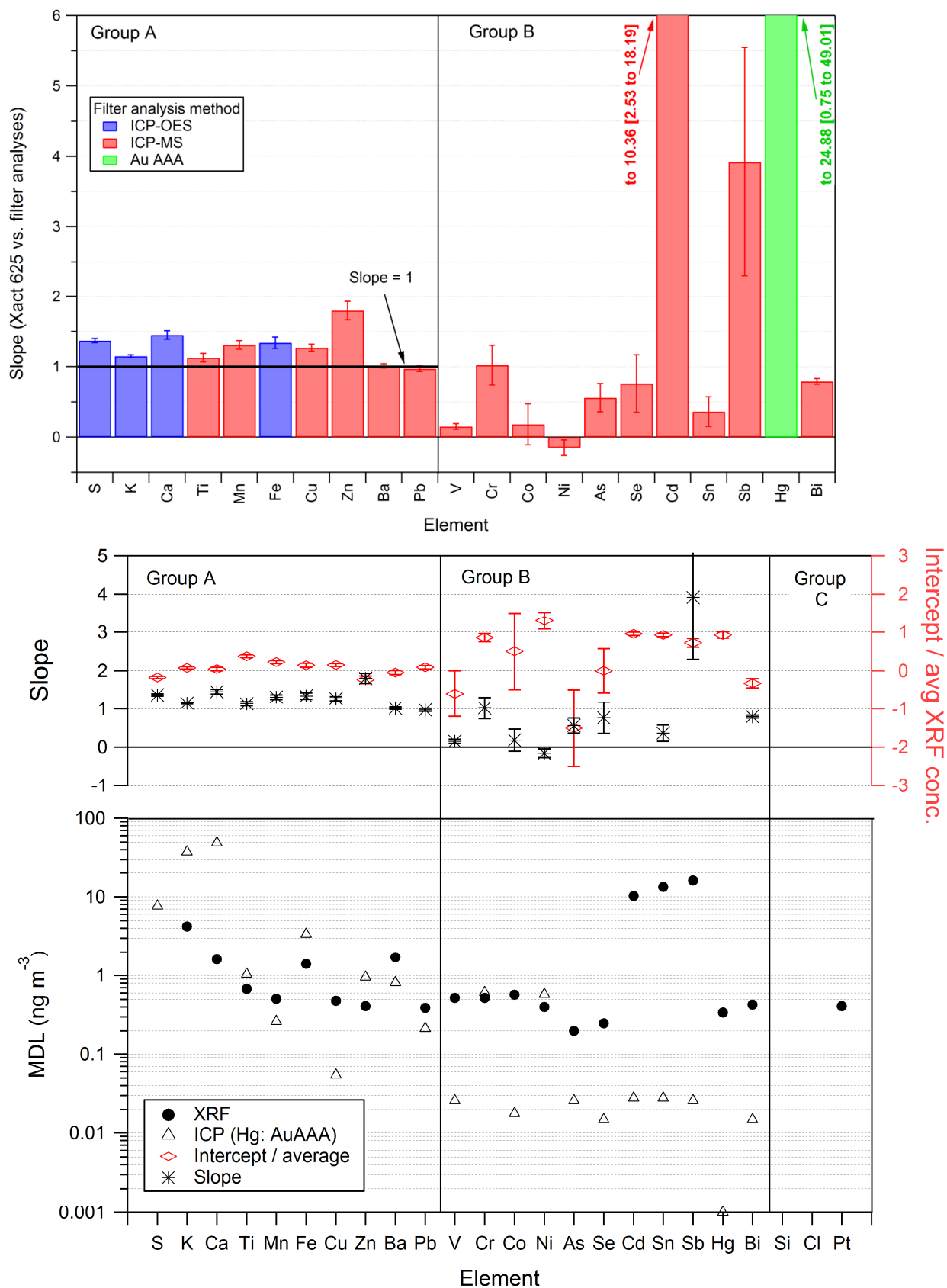


Figure 3: Bottom: Minimum detection limits (MDL), interference free for the Xact 625 (XRF), and for the ICP analyses. Hg was analysed with AuAAA spectrometry. Slopes of the Levenberg-Marquardt least-squares fit regression analyses of Xact 625 vs. filter analyses with ICP-OES, ICP-MS and Hg AuAAA analysis. Error bars indicate the computed uncertainties of the fits. Top: Slopes and intercept ratios (intercepts divided by the average concentrations measured with XRF) with standard deviations for all elements measured with the Xact.

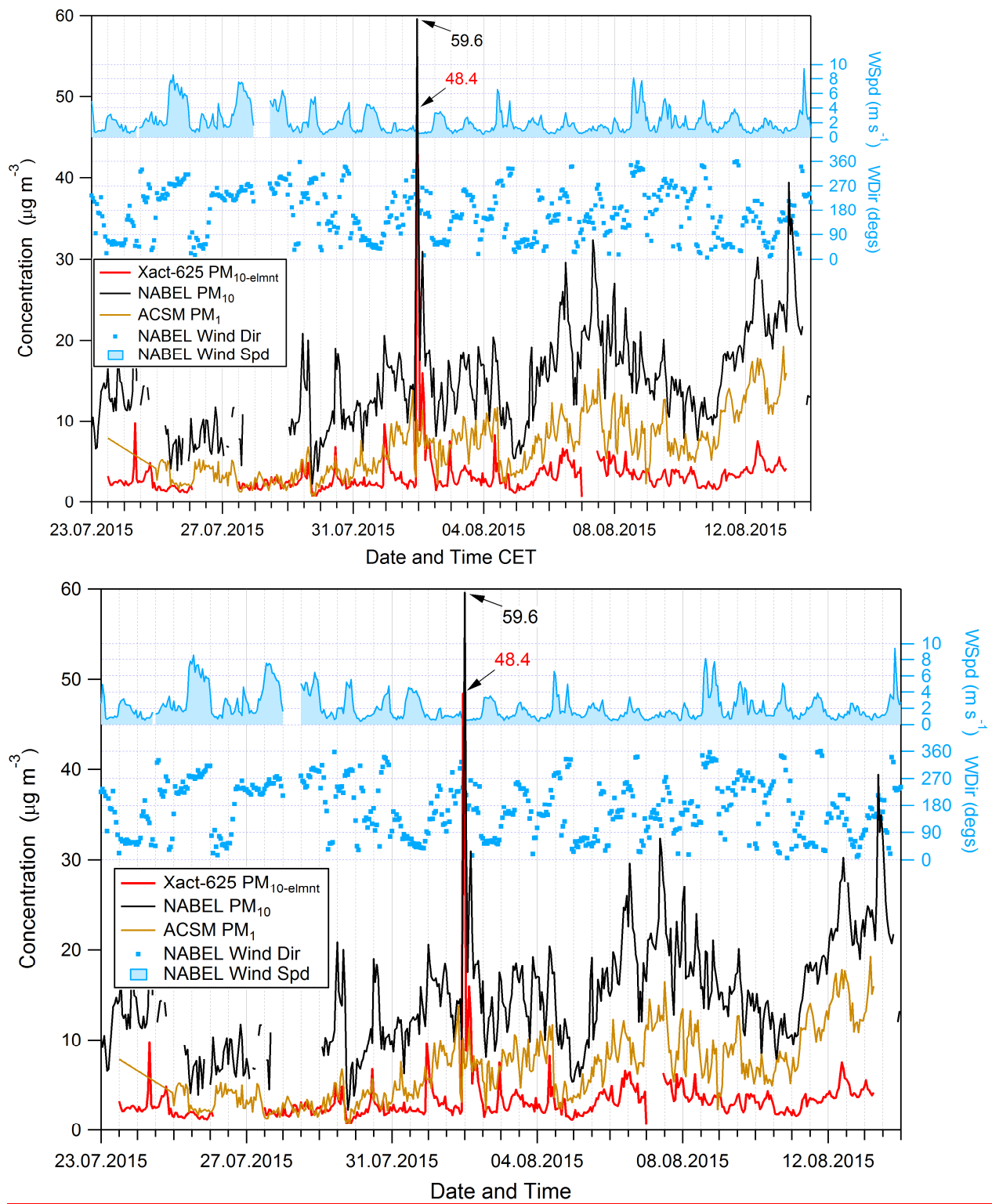


Figure 4: Time series of Xact625 total elemental concentration (**red**), ACSM PM_1 data, NABEL TEOM PM_{10} data (**black**), and wind speed (WSpd) and direction (WDir) measurements (**blue**) in Härkingen. Numbers at the peaks indicate 1-h concentration maxima.

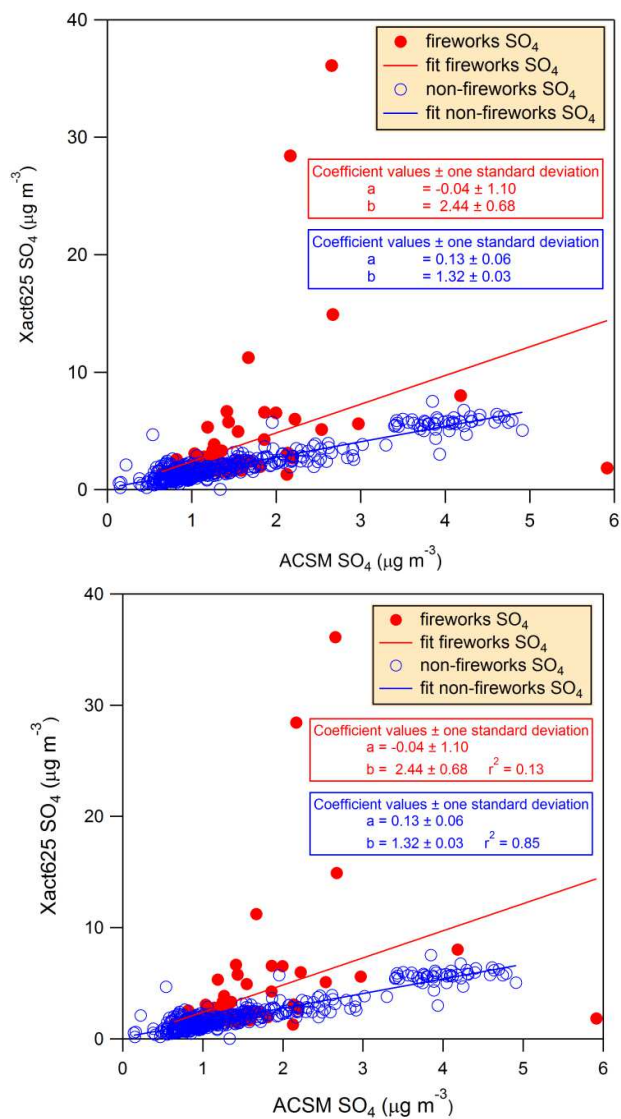


Figure 5: Comparison of Xact PM₁₀ SO₄ vs. ACSM PM₁ SO₄. Data were split into fireworks (red) and non-fireworks (blue) periods.

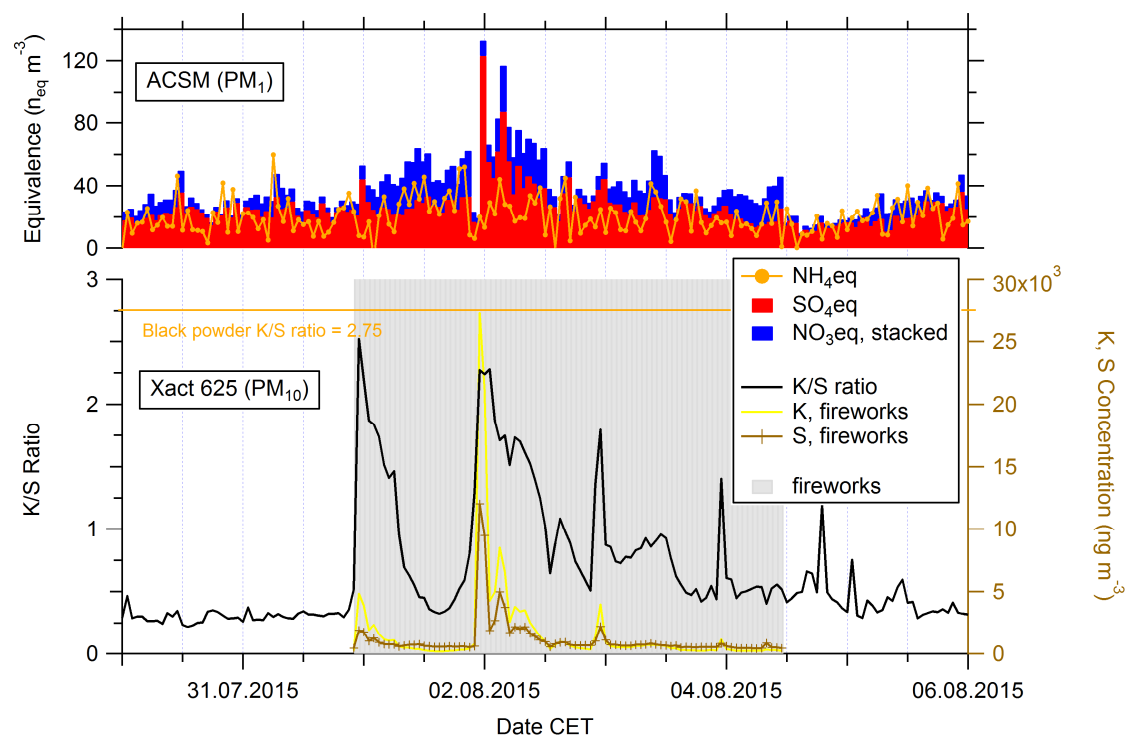
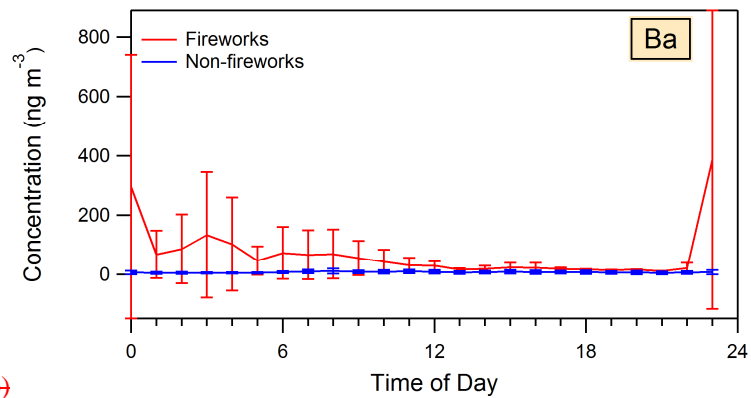
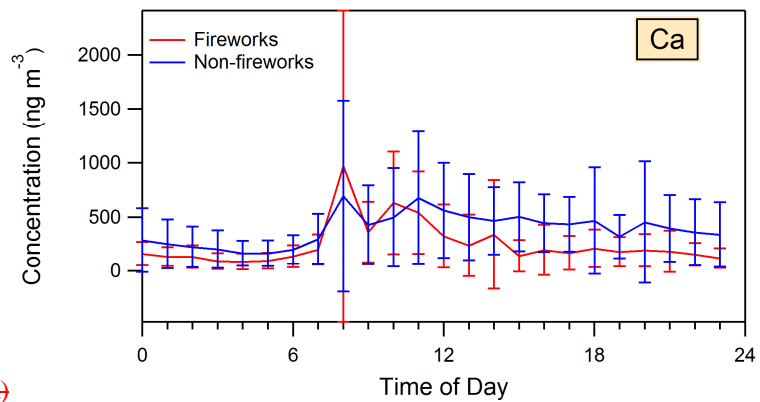


Figure 6: Top: Time series of equivalent concentrations of NH_4 , NO_3 and SO_4 in PM_{10} measured with the ACSM. The NO_3 eq is stacked on SO_4 eq. Bottom: Time series of the K/S mass ratio (black, left axis) of the Xact 625. The K and S concentrations for the fireworks period (31 July 2200 CET to 4 Aug 1100 CET, shaded) are given in yellow and brown (right axis). The orange line indicates the K/S mass ratio of 2.75 for black powder.

5



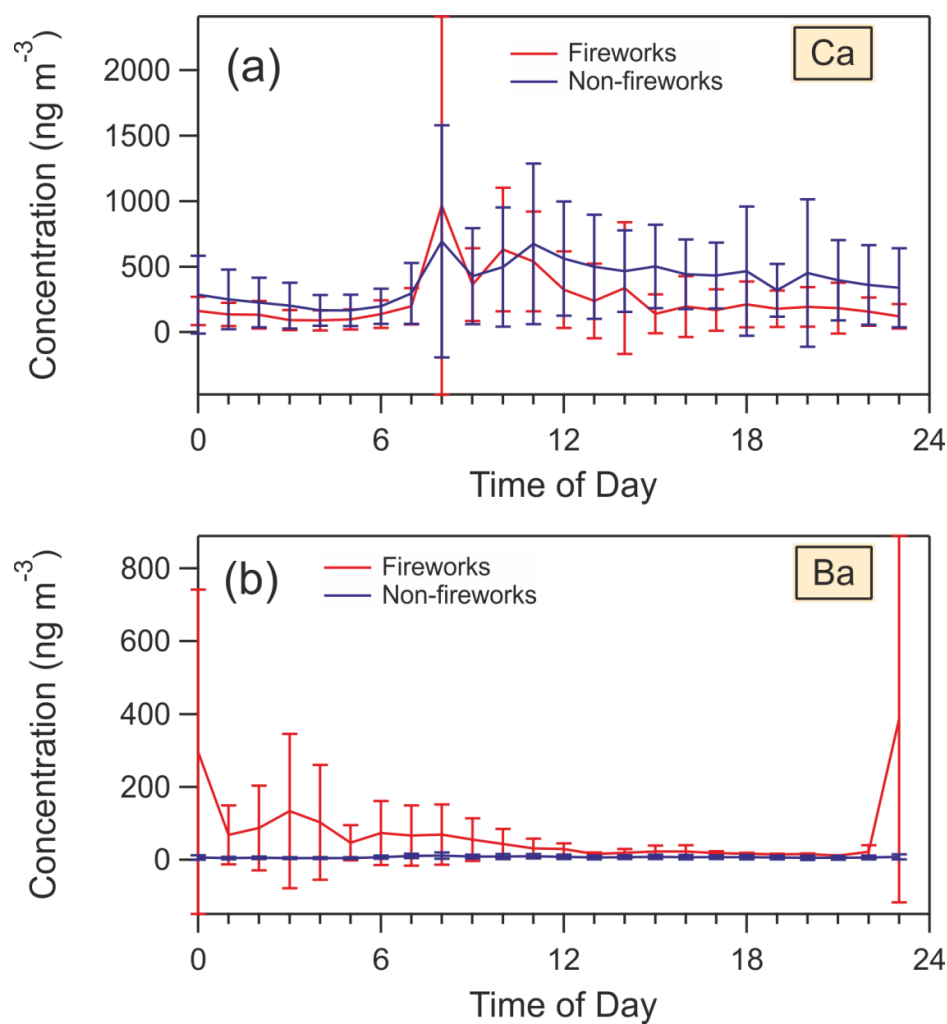


Figure 76: Mean diurnal variations of a) Ca and b) Ba, stratified for fireworks (red) and non-fireworks (blue) periods. Error bars denote ± 1 standard deviation of the averaging period. Diurnal variations for the other elements are shown in the supplement S4 (Fig. S3S4).

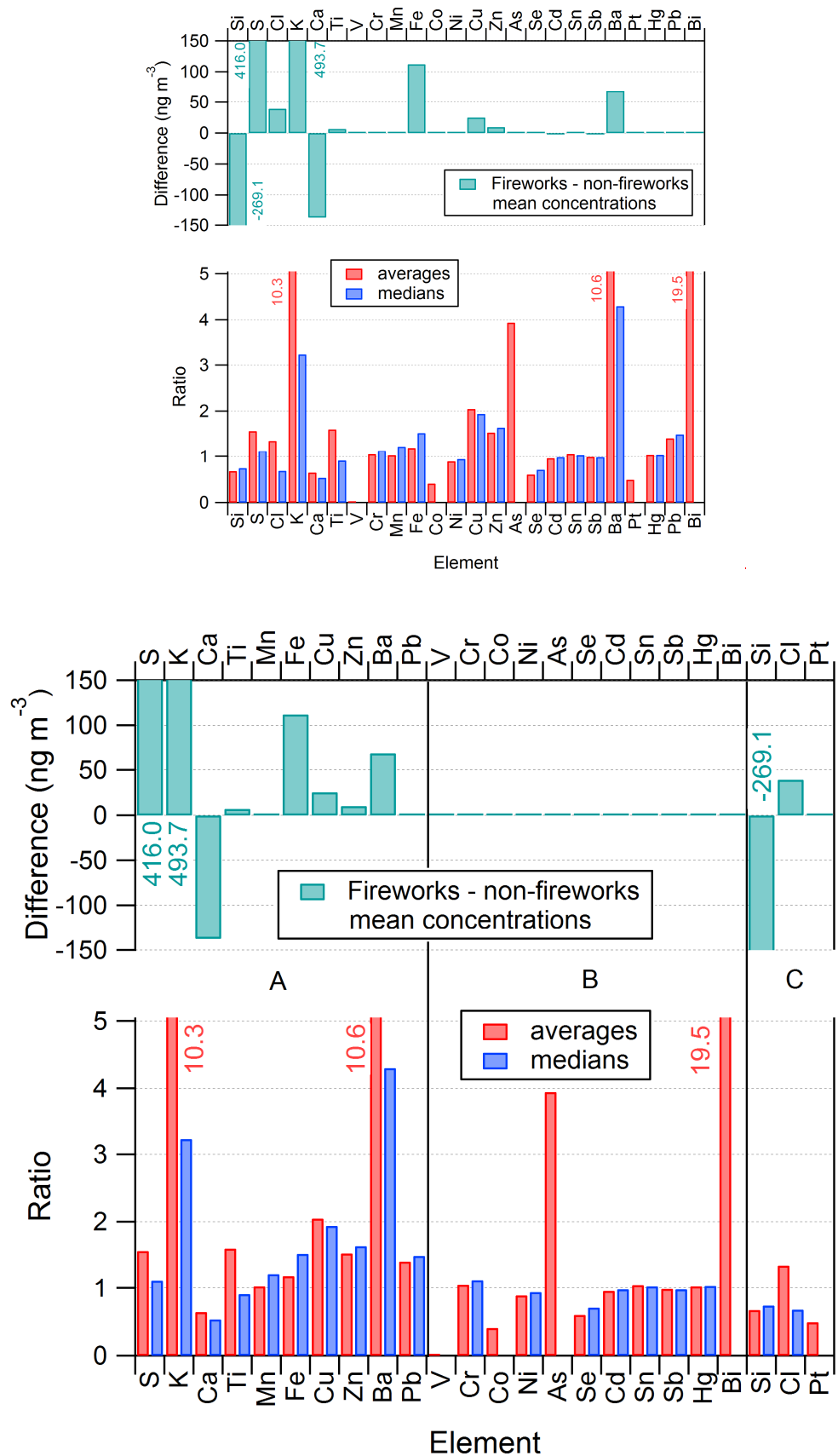


Figure 87: Bottom: Enhancement ratios for all analysed elements for fireworks/non-fireworks classification, sorted by group. Numbers indicate values outside the axis range for S, K, and Si. Top: 'Background'-subtracted mean concentrations of the south

~~sector for the non-fireworks period~~Difference between average concentrations of fireworks and non-fireworks periods. Numbers indicate values outside the axis range for ~~Si (negative), S, and K, and Si (negative).~~

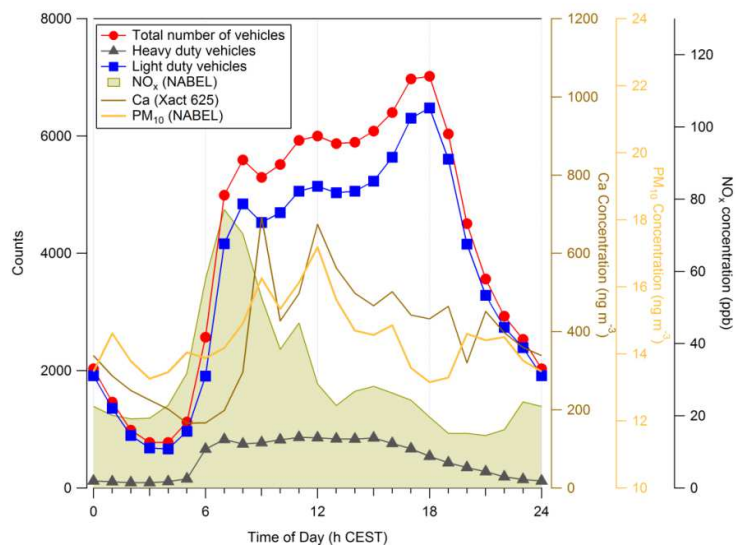
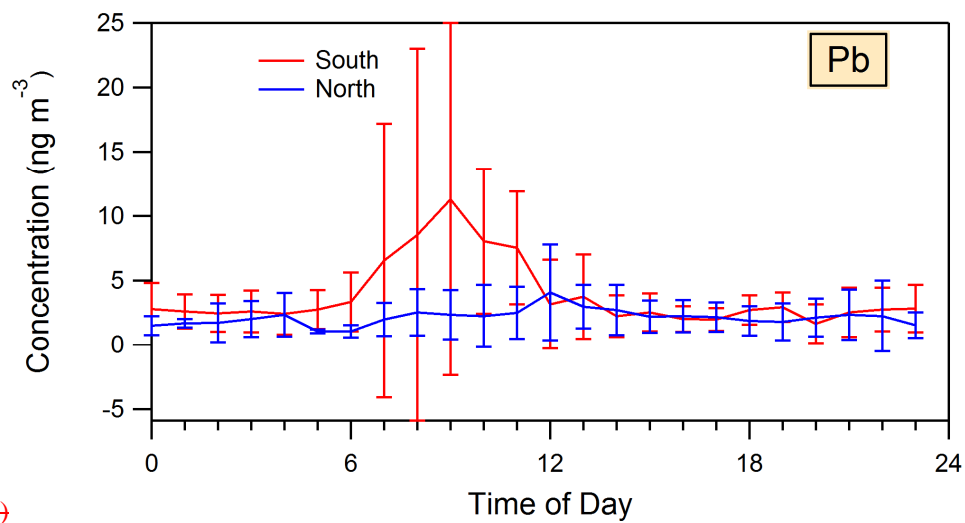
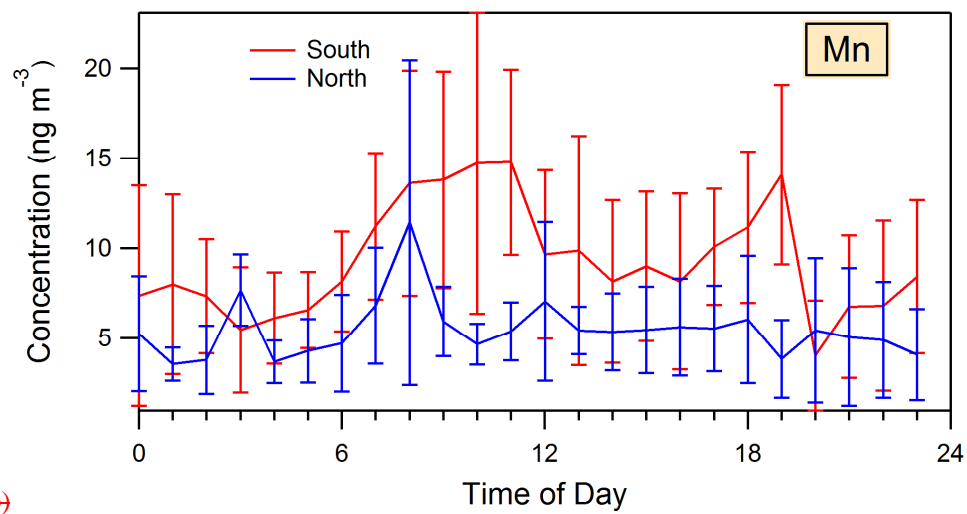
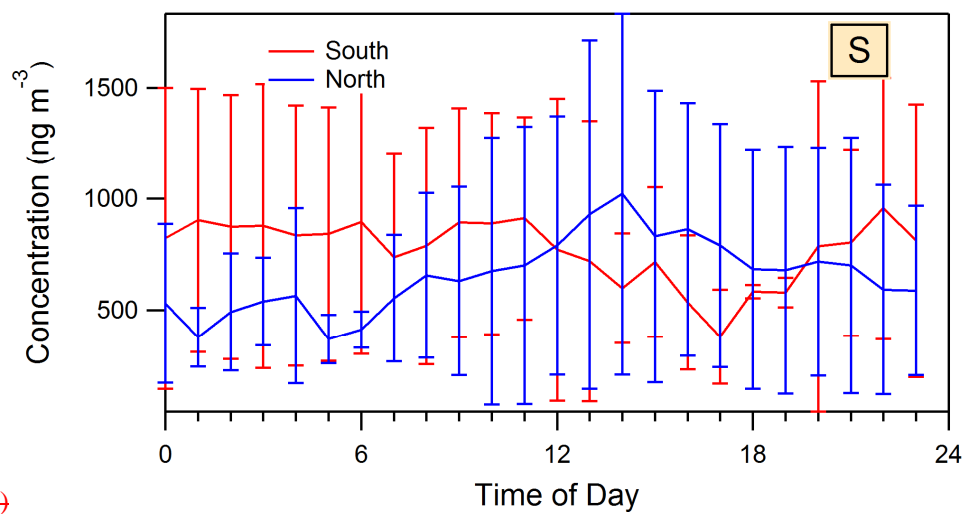
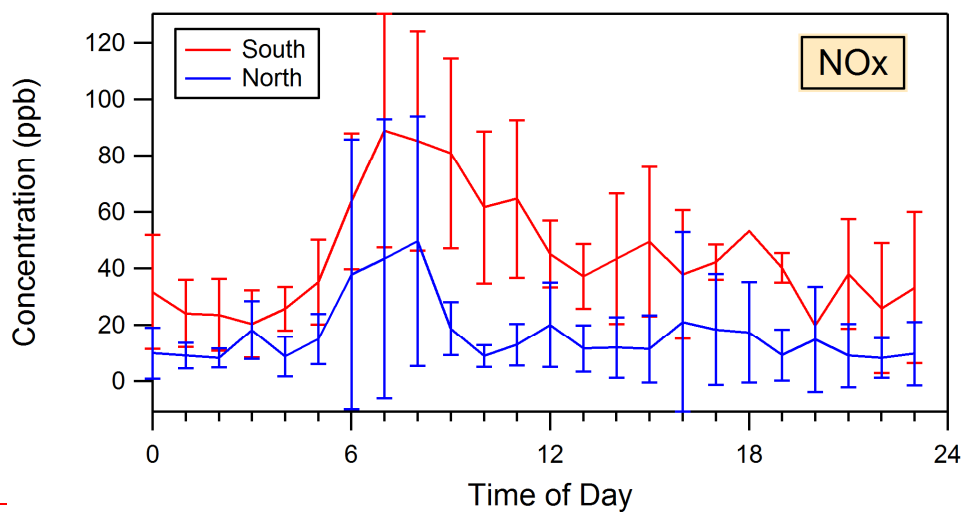


Figure 9: Diurnal variation of traffic counts (average diurnal variations) and NO_x concentration in Härkingen for the non-fireworks period. Note that the time axis is in CEST (Central European Summer Time = UTC + 2h = local time).





(d)

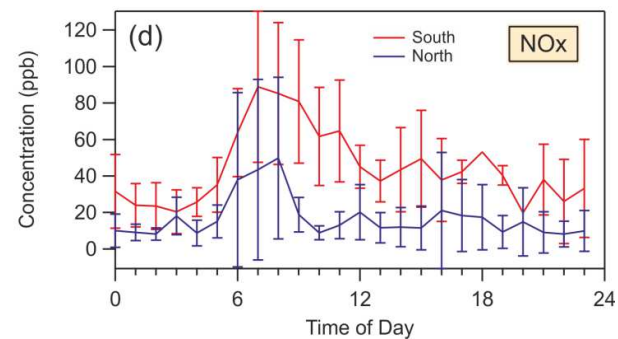
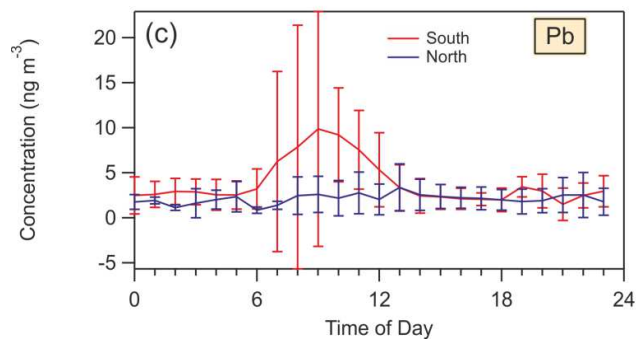
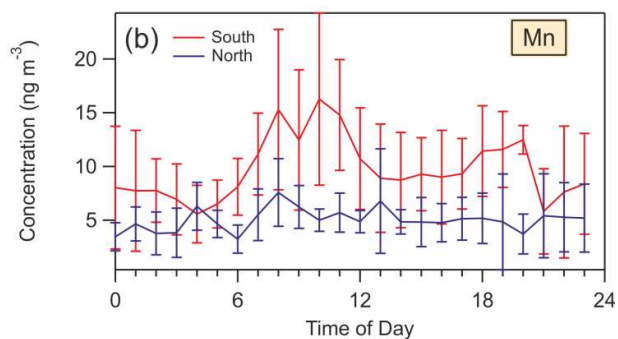
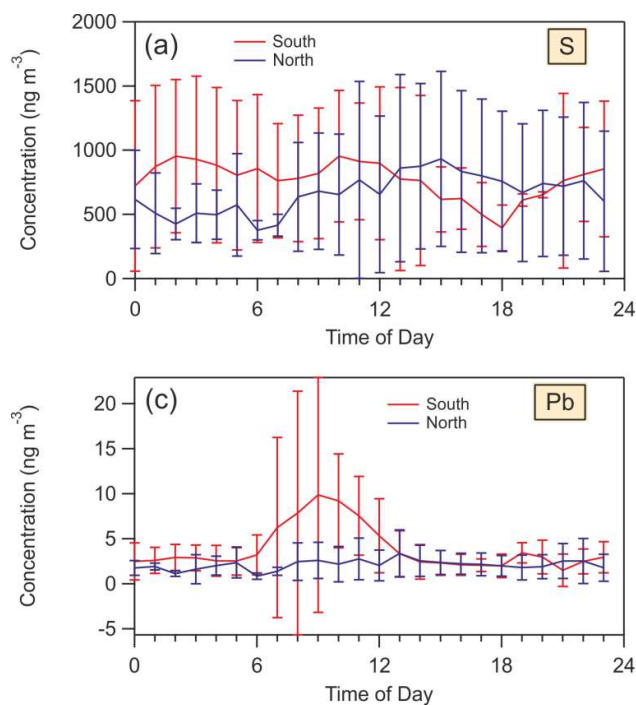


Figure 498: Mean diurnal variations of a) S, b) Mn, c) Pb, and d) NO_x for the non-fireworks period stratified for south (~~red~~) and north (~~blue~~) wind directions. Error bars denote ± 1 standard deviation of the averaging period. Diurnal variations for the other elements are shown in the supplement ~~S4-S5~~ (Fig. ~~S4S5~~).

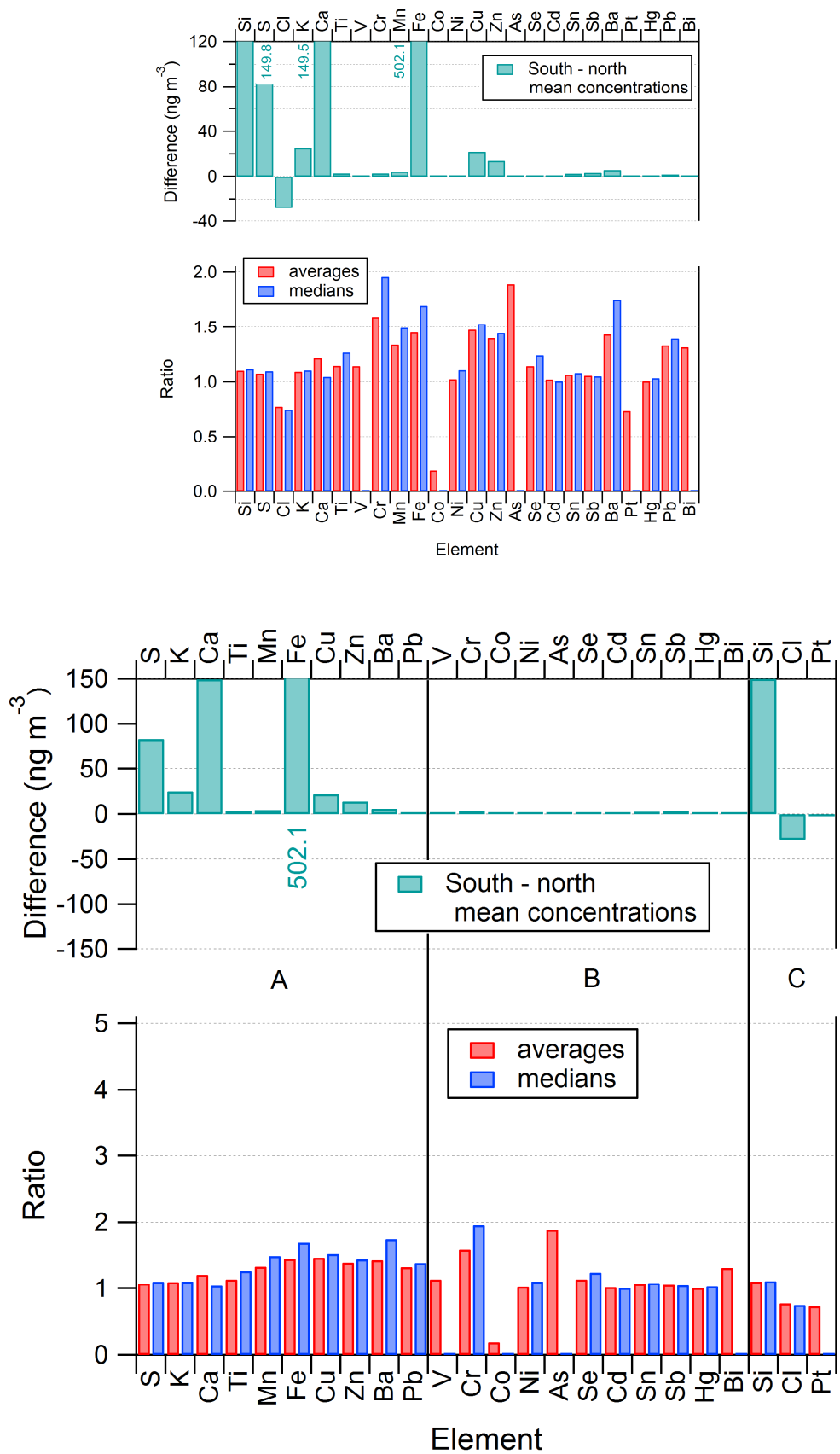


Figure 149: Bottom: Enhancement ratios for all analysed elements for south/north sector classification for the non-fireworks period, sorted by group. Top: 'Background'-subtracted mean concentrations of the south sector for the non-fireworks period. Numbers indicate values outside the axis range for Si, Ca, and Fe.

Table 1: Data characteristics of Xact 625 measurements in Härkingen, and minimum detection limits (MDL) for Xact and ICP/Hg AuAAA. Elements are sorted according to the groups in Table 2). Data were classified into fireworks and non-fireworks periods. The non-fireworks period was further classified into north (rural) and south (freeway) sectors according to the wind direction. Numbers in italics indicate cases where the daily averages were <MDL. The cases for the two wind sectors do not add up to the non-fireworks cases as wind data are missing for a total of 12 h (cf. Fig. 1).

Element nt	Non-Fireworks					Fireworks					South sector (non-fireworks)					North sector (non-fireworks)					Xact		ICP		
	avg ng m ⁻³	sdev ng m ⁻³	max ng m ⁻³	median ng m ⁻³	# cases	avg ng m ⁻³	sdev ng m ⁻³	max ng m ⁻³	median ng m ⁻³	# cases	avg ng m ⁻³	sdev ng m ⁻³	max ng m ⁻³	median ng m ⁻³	# cases	avg ng m ⁻³	sdev ng m ⁻³	max ng m ⁻³	median ng m ⁻³	# cases	MDL (60 min) ng m ⁻³	Pts > MDL %	MDL (24 h) ng m ⁻³	Pts > MDL %	
S	779.28	524.59	2508.00	601.85	155.32	1666.73	12044.00	677.15	799.62	520.55	2141.00	650.32	707.39	522.53	2508.00	529.50	384	184	100	4.20	100.00	7.662	100		
K	161.00	56.58	484.09	152.81	1661.10	3854.66	27349.00	493.69	173.72	57.35	395.34	168.27	152.42	54.59	484.09	144.79	100	100	100	100	1.60	100.00	37.808	100	
Ca	390.66	384.69	3211.00	262.79	253.21	389.74	3109.00	140.82	434.34	393.86	2166.00	271.75	362.94	381.77	3211.00	255.70	10.61	7.52	43.38	8.30	0.68	100.00	1.043	90*	
Ti	11.44	8.12	43.38	8.79	18.29	36.16	282.23	8.04	12.75	8.69	39.56	10.63	10.61	7.52	43.38	8.30	0.68	100.00	100	0.51	100.00	0.264	100		
Mn	7.10	4.62	26.98	5.72	7.30	3.88	22.21	6.59	9.02	5.18	26.98	8.16	5.47	3.30	26.87	4.85	0.30	0.30	0.30	0.30	0.51	100.00	0.264	100	
Fe	587.41	428.85	2338.00	460.08	699.95	385.97	1909.00	699.78	792.89	463.19	2338.00	708.17	404.84	301.70	1828.00	310.16	1.40	100.00	100	0.48	100.00	0.055	100		
Cu	24.07	17.69	109.34	20.07	49.28	48.72	371.81	38.93	32.27	18.92	109.34	27.48	16.48	12.74	71.41	18.70	0.48	100.00	100	0.41	100.00	0.959	100		
Zn	18.07	16.64	143.37	14.31	28.56	18.94	104.12	23.42	24.95	21.08	143.37	18.81	13.31	8.66	67.33	11.62	0.41	100.00	100	0.39	98.20	0.216	100		
Ba	7.12	5.49	25.33	5.25	75.95	165.25	1127.00	22.58	9.31	6.03	25.33	8.02	5.22	4.18	22.94	5.81	1.70	94.30	0.819	100	0.52	2.00	0.216	100	
Pb	2.96	3.89	41.07	1.99	4.17	3.02	15.30	2.96	3.86	5.22	41.07	2.63	2.26	1.77	12.50	1.73	0.39	98.20	0.216	100	0.52	2.00	0.216	100	
V	0.06	0.15	1.22	0.00	0.00	0.00	0.00	0.00	0.06	0.15	1.22	0.00	0.00	0.06	0.16	0.92	0.00	0.52	2.00	0.026	100	0.026	100		
Cr	2.40	2.30	12.96	1.75	2.51	2.22	9.23	1.98	3.44	2.52	12.96	2.75	1.44	1.61	8.47	0.79	0.52	74.60	0.614	40	0.52	74.60	0.614	40	
Co	0.02	0.08	0.70	0.00	0.01	0.03	0.23	0.07	0.03	0.07	0.01	0.05	0.55	0.00	0.03	0.10	0.70	0.00	0.57	0.44	0.018	0.70	0.44	0.018	70
Ni	0.02	0.65	10.32	0.54	0.56	0.35	1.95	0.51	0.64	0.37	2.14	0.59	0.60	0.85	10.32	0.48	0.40	67.80	0.591	0	0.036	100	0.036	100	
As	0.02	0.14	1.31	0.00	0.09	0.21	1.91	0.00	0.04	0.19	1.21	0.00	0.01	0.08	0.64	0.00	0.20	4.00	0.026	100	0.25	38.00	0.015	100	
Se	0.27	0.32	4.39	0.20	0.16	0.13	0.44	0.14	0.31	0.41	4.39	0.24	0.23	0.21	0.90	0.16	0.25	38.00	0.015	100	0.25	38.00	0.015	100	
Cd	6.75	3.61	23.62	6.25	6.49	3.14	21.33	6.14	6.94	3.79	23.62	6.27	6.68	3.49	20.63	6.26	10.30	12.70	0.028	10	13.30	85.00	0.028	10	
Sn	20.79	7.82	55.34	19.57	21.78	8.27	54.28	20.10	21.66	7.80	55.28	20.85	20.08	7.74	55.34	18.98	13.30	85.00	0.028	10	13.30	85.00	0.028	10	
Sb	31.31	11.22	111.88	29.80	30.97	10.42	67.96	29.38	32.98	11.49	77.63	31.08	29.79	10.79	111.88	28.76	16.00	94.00	0.026	100	0.34	86.80	0.001	40	
Hg	0.63	0.25	1.49	0.61	0.64	0.18	1.31	0.63	0.62	0.24	1.23	0.62	0.64	0.25	1.49	0.62	0.34	86.80	0.001	40	0.34	86.80	0.001	40	
Bi	0.07	0.12	0.70	0.00	1.27	3.82	23.47	0.15	0.08	0.12	0.49	0.00	0.05	0.12	0.70	0.00	0.43	7.20	0.015	100	0.43	7.20	0.015	100	
Si	839.20	386.20	3415.00	713.75	570.13	223.75	1758.00	532.38	902.57	433.21	2258.00	792.43	796.09	360.44	3415.00	694.15	30.02	30.02	0.00	0.41	1.75				
Cl	113.70	200.20	969.80	26.44	153.07	578.02	4455.00	18.11	94.66	181.53	969.80	19.72	111.73	198.09	957.00	30.02	0.06	0.11	0.64	0.00	0.41	1.75			
Pr	0.05	0.11	0.60	0.00	0.03	0.07	0.34	0.00	0.04	0.10	0.60	0.00	0.06	0.11	0.64	0.00	0.06	0.11	0.64	0.00	0.41	1.75			

Numbers in italics indicate Xact daily averages < MDL.

both for ICP-OES, ICP-MS

MDL for Hg Gold Analaim Atomic Absorption

Numbers in italics indicate Xact daily averages < MDL.

*both for ICP-OES, ICP-MS

MDL for Hg Gold Anilam Atomic Absorption

Table 21. Regression coefficients for the comparison of Xact 625 and offline data. The 1-h values of the Xact 625 were averaged to 24-h values. Primed quantities are uncertainties.

Regression coefficients						
	fitX = a + bx					
Element	a	± a'	b	± b'	r ²	Group
S	-169.73	30.57	1.37	0.03	1.00	A
K	52.42	19.15	1.15	0.02	1.00	
Ca	13.87	17.91	1.45	0.06	0.99	
Ti	5.58	0.57	1.13	0.06	0.98	
Mn	1.72	0.28	1.31	0.06	0.99	
Fe	93.05	35.80	1.34	0.08	0.97	
Cu	4.93	1.27	1.27	0.05	0.99	
Zn	-5.56	2.18	1.80	0.13	0.96	
Ba	-1.62	1.73	1.01	0.03	0.99	
Pb	0.32	0.17	0.97	0.04	0.99	
V	-0.03	0.03	0.15	0.05	0.57	B
Cr	2.13	0.26	1.02	0.28	0.63	
Co	0.01	0.02	0.18	0.29	0.05	
Ni	-0.55	1.02	-0.15	0.13	0.14	
As	-0.09	0.06	0.56	0.20	0.50	
Se	0.00	0.15	0.76	0.41	0.30	
Cd	6.49	0.20	10.36	7.83	0.18	
Sn	19.73	0.96	0.36	0.21	0.27	
Sb	22.17	3.58	3.92	1.63	0.42	
Hg	0.64	0.06	24.88	24.13	0.12	
Bi	-0.16	0.06	0.79	0.04	0.98	C
Si						
Cl						
Pt						
Group A	average slope		1.28			
Group A	standard deviation		0.24			
	Elements analysed with ICP-OES					
	Elements analysed with ICP-MS					
	Element analysed with Au AAA					

			Regression coefficients							Xact		ICP	
Group	Element	Analysis method	fitX = a + bx					Average XRF conc.	Ratio intercept/average	MDL (60 min)	Pts > MDL	MDL (24 h)	Pts > MDL
			a	± a'	b	± b'	r ²	ng m ⁻³		ng m ⁻³	%	ng m ⁻³	%
A	S	ICP-OES	-169.73	30.57	1.37	0.03	1.00	956.48	-0.177			7.662	100
	K	ICP-OES	52.42	19.15	1.15	0.02	1.00	703.47	0.075	4.20	100.00	37.808	100
	Ca	ICP-OES	13.87	17.91	1.45	0.06	0.99	365.41	0.038	1.60	100.00	49.195	90
	Ti	ICP-MS	5.58	0.57	1.13	0.06	0.98	14.94	0.373	0.68	100.00	1.043	90
	Mn	ICP-MS	1.72	0.28	1.31	0.06	0.99	7.59	0.227	0.51	100.00	0.264	100
	Fe	ICP-OES	93.05	35.80	1.34	0.08	0.97	656.22	0.142	1.40	100.00	3.398	100
	Cu	ICP-MS	4.93	1.27	1.27	0.05	0.99	33.49	0.147	0.48	100.00	0.055	100
	Zn	ICP-MS	-5.56	2.18	1.80	0.13	0.96	22.97	-0.242	0.41	100.00	0.959	100
	Ba	ICP-MS	-1.62	1.73	1.01	0.03	0.99	31.92	-0.051	1.70	94.30	0.819	100
	Pb	ICP-MS	0.32	0.17	0.97	0.04	0.99	3.84	0.085	0.39	98.20	0.216	100
B	V	ICP-MS	-0.03	0.03	0.15	0.05	0.57	0.05	-0.619	0.52	2.00	0.026	100
	Cr	ICP-MS	2.13	0.26	1.02	0.28	0.63	2.48	0.860	0.52	74.60	0.614	40
	Co	ICP-MS	0.01	0.02	0.18	0.29	0.05	0.02	0.555	0.57	0.44	0.018	70
	Ni	ICP-MS	0.82	0.14	-0.16	0.11	0.22	0.63	1.310	0.40	67.80	0.581	100
	As	ICP-MS	-0.09	0.06	0.56	0.20	0.50	0.06	-1.481	0.20	4.00	0.026	100
	Se	ICP-MS	0.00	0.15	0.76	0.41	0.30	0.26	0.015	0.25	38.00	0.015	100
	Cd	ICP-MS	6.28	0.28	10.37	7.86	0.18	6.54	0.960	10.30	12.70	0.028	20
	Sn	ICP-MS	19.73	0.96	0.36	0.21	0.27	21.29	0.927	13.30	85.00	0.028	100
	Sb	ICP-MS	22.17	3.58	3.92	1.63	0.42	30.60	0.724	16.00	94.00	0.026	100
	Hg	Au AAA	0.64	0.06	24.88	24.13	0.12	0.69	0.935	0.34	86.80	0.001	40
C	Bi	ICP-MS	-0.16	0.06	0.79	0.04	0.98	0.49	-0.318	0.43	7.20	0.015	100
	Si												
	Cl												
	Pt									0.41	1.75		
	Group A	average slope					1.28						
	Group A	standard deviation					0.24						

Table 32. Comparison of Xact data with published ICP data of other campaigns.

	Xact 625 averages		Switzerland								
	All days	Non-fireworks days	Belp	Geneva	Basel, summer	Payerne	Zürich	Zürich, summer	Payerne, summer	Payerne, summer	Härkingen NABEL
Reference			1)	2)	3)	4)	5)	6)	6)	7)	8)
# cases, size	22	17	PM8	PM10	PM10	PM10	PM10	PM10	PM10	PM10	PM10
Sampling period	2015	2015	1985/86	1996/97	1997/98	1998/99	2008/09	2009	2009	2012	2015
Unit	ng/m ³	ng/m ³	ng/m ³	ng/m ³	ng/m ³	ng/m ³	ng/m ³	ng/m ³	ng/m ³	ng/m ³	ng/m ³
Si	783.14	829.45					210.90	571.00	634.00	370.00	
S	822.74	790.19					2394.30	625.00	637.00	360.00	
Cl	113.17	109.40			41.00		656.50	66.00	190.00	30.00	
K	408.00	166.56			630.00	98.00	1318.20	187.00	188.00	120.00	
Ca	367.99	397.57			720.00	100.00	137.40	451.00	355.00	180.00	
Ti	12.26	11.30			38.00		6.50	13.90	14.30	9.90	
V	0.05	0.06	3.90			0.70	0.40	1.00	1.20	0.70	
Cr	2.39	2.43			8.00		0.90	2.30	1.60	1.00	
Mn	7.11	7.21	31.70		16.00	2.80	5.80	7.20	5.00	2.80	
Fe	600.48	593.92			760.00	89.00	389.70	455.00	202.00	130.00	
Co	0.01	0.02					2.60	0.10	0.10		
Ni	0.60	0.63			8.00	1.20	0.60	0.90	1.00	0.50	0.90
Cu	28.04	24.45	7.90	35.00	75.00	6.00	28.10	17.40	4.30	2.80	19.70
Zn	20.23	19.04	65.00	120.00	73.00		20.30	16.10	9.40	7.50	
As	0.03	0.02	2.20	2.00	1.00	0.53		0.30	0.50	0.20	0.31
Se	0.27	0.30		6.00		0.16		0.30	0.30	0.20	
Cd	6.57	6.67	0.88	0.40	0.00	0.32		0.10	0.10		0.07
Sn	21.00	20.93					2.60	2.60	1.00		
Sb	31.33	31.59			29.00	0.26	2.50	2.40	0.50		
Ba	18.09	7.12			110.00		6.70	6.50	3.90	1.80	
Pt	0.04	0.05									
Hg	0.64	0.64		0.50							
Pb	3.13	3.04	134.00	95.00	51.00	10.00	14.20	3.60	3.10	1.20	4.90
Bi	0.26	0.07						0.20	0.10		
1)	Galli et al. 1990										
2)	Chiaradia and Cupelin 2000 - fortnight averages										
3)	Röösli et al. 2001										
4)	Hueglin et al. 2005										
5)	Richard et al. 2011										
6)	Minguillón et al. 2012										
7)	Alastuey et al. 2016										
8)	BAFU/Empa 2015 - annual mean values										

S1. General statistics of Xact 625 data

Table S1: Data characteristics of Xact 625 measurements in Härkingen. Elements are sorted according to the groups in Table 2). Data were classified into fireworks and non-fireworks periods. The non-fireworks period was further classified into north (rural) and south (freeway) sectors according to the wind direction. Numbers in italics indicate cases where the daily averages were <MDL. The cases for the two wind sectors do not add up to the non-fireworks cases as wind data are missing for a total of 12 h (cf. Fig. 1).

	Non-Fireworks				Fireworks				South sector (non-fireworks)				North sector (non-fireworks)			
Element	avg	sdev	max	median	avg	sdev	max	median	avg	sdev	max	median	avg	sdev	max	median
	ng m ⁻³	ng m ⁻³	ng m ⁻³	ng m ⁻³	ng m ⁻³	ng m ⁻³	ng m ⁻³	ng m ⁻³	ng m ⁻³	ng m ⁻³	ng m ⁻³	ng m ⁻³	ng m ⁻³	ng m ⁻³	ng m ⁻³	ng m ⁻³
# cases	370				86				173				185			
S	739.28	524.59	2508.00	601.85	1155.32	1666.73	12034.00	677.15	795.31	516.19	2254.00	659.85	711.53	537.04	2508.00	499.31
K	161.00	56.98	484.09	152.81	1661.10	3854.66	27349.00	493.69	175.80	60.89	395.34	168.66	150.75	50.00	484.09	144.98
Ca	390.66	384.69	3211.00	262.79	253.21	389.74	3109.00	140.82	474.39	476.21	3211.00	274.58	324.91	266.10	2254.00	252.85
Ti	11.44	8.12	43.38	8.79	18.29	36.16	282.23	8.04	13.08	8.90	39.56	11.14	10.33	7.17	43.38	8.23
Mn	7.10	4.62	26.98	5.72	7.30	3.88	22.21	6.99	9.51	5.14	26.98	8.56	5.03	2.79	20.99	4.71
Fe	587.41	428.85	2338.00	460.08	699.95	385.97	1909.00	699.78	852.36	452.72	2338.00	779.67	350.25	227.50	1309.00	303.58
Cu	24.07	17.69	109.34	20.07	49.28	48.72	371.81	38.91	35.46	18.29	109.34	30.66	13.52	8.55	49.01	9.94
Zn	18.67	16.84	143.37	14.31	28.56	18.94	104.12	23.42	26.13	20.82	143.37	20.67	12.27	7.97	66.14	10.70
Ba	7.12	5.49	25.33	5.25	75.39	169.25	1127.00	22.58	10.19	5.96	25.33	9.19	4.42	3.21	17.22	3.46
Pb	2.96	3.89	41.07	1.99	4.17	3.02	15.30	2.96	3.95	5.26	41.07	2.77	2.18	1.62	9.38	1.72
V	0.06	0.15	1.22	0.00	0.00	0.00	0.00	0.00	0.07	0.18	1.22	0.00	0.05	0.13	0.89	0.00
Cr	2.40	2.30	12.96	1.75	2.51	2.22	9.23	1.98	3.81	2.47	12.96	3.43	1.09	1.10	6.17	0.76
Co	0.02	0.08	0.70	0.00	0.01	0.03	0.23	0.00	0.00	0.02	0.21	0.00	0.03	0.11	0.70	0.00
Ni	0.62	0.65	10.32	0.54	0.56	0.35	1.95	0.51	0.64	0.45	3.99	0.59	0.61	0.81	10.32	0.50
As	0.02	0.14	1.31	0.00	0.09	0.31	1.91	0.00	0.04	0.20	1.31	0.00	0.01	0.06	0.84	0.00
Se	0.27	0.32	4.39	0.20	0.16	0.13	0.44	0.14	0.30	0.41	4.39	0.24	0.24	0.22	0.90	0.17
Cd	6.75	3.61	23.62	6.25	6.49	3.14	21.33	6.14	6.89	3.74	19.10	6.28	6.70	3.55	23.62	6.20
Sn	20.79	7.82	55.34	19.57	21.78	8.27	54.28	20.10	22.15	8.38	55.34	21.15	19.64	7.01	45.41	18.68
Sb	31.31	11.22	111.88	29.80	30.97	10.42	67.96	29.38	33.03	11.93	77.63	31.30	29.90	10.43	111.88	28.90
Hg	0.63	0.25	1.49	0.61	0.64	0.18	1.31	0.63	0.63	0.23	1.18	0.63	0.64	0.26	1.49	0.61
Bi	0.07	0.12	0.70	0.00	1.27	3.82	23.47	0.15	0.09	0.13	0.50	0.00	0.04	0.10	0.70	0.00
Si	839.20	398.20	3415.00	713.75	570.13	223.75	1758.00	532.28	924.98	468.79	3415.00	795.77	775.17	308.86	2052.00	682.42
Cl	113.70	200.20	969.80	26.44	153.07	578.02	4455.00	18.11	87.95	161.12	871.19	19.79	116.38	210.81	969.80	30.50
Pt	0.05	0.11	0.66	0.00	0.03	0.07	0.34	0.00	0.04	0.10	0.66	0.00	0.06	0.11	0.64	0.00

S1S2. Ambient filter samples for method intercomparisons

A quarter each of three NABEL filters of the series analysed at IDAEA were also analysed with XRF at CES and with ICP-MS at ERG. This allows for an intercomparison between benchtop XRF and ICP-MS, between ICP-MS of two different laboratories, and between Xact XRF and benchtop XRF. The elements Zn, Sr, Cu, Pb, Fe, K, Ca, Mn, Se, and Ba were selected for this comparison. Benchtop XRF required no further sample preparation except punching a 47-mm piece of the original filter. IDAEA's digestion protocol is described in the main paper. Of three different filter blanks, blank 1 appeared contaminated and was not further considered, blanks 2 and 3 were averaged and subtracted from the analysis, but the values are not reported. ERG followed a protocol of the USEPA describing the multi-elemental determination of total metals by ICP-MS in ambient air samples collected on 47mm Teflon® filters following guidelines in EPA method IO-3.5 and EPA FEM Method "Standard Operating Procedure for the Determination of Lead in PM10 by Inductively Coupled Plasma Mass Spectrometry (ICPMS) with Hot Block Dilute Acid and Hydrogen Peroxide Filter Extraction" (EQL-0512-202). The filters were digested in a HotBlock™ for 2.5 hours using an extraction fluid containing 1.85% nitric acid (HNO₃), 0.5% hydrochloric acid (HCl), and 0.17% hydrofluoric acid (HF) with 0.33 mg/L of gold added for mercury stabilization. One aliquot of hydrogen peroxide (H₂O₂) was added after 1.5 hours of extraction and was allowed to effervesce. The extract was analyzed by ICP-MS and the data were collected using the manufacturer's software. The results are given in Table S1.

Table S1S2. Analyses of three ambient (NABEL) samples (1, 6 and 12 August 2015) from Härkingen. Comparisons of Xact and benchtop XRF (CES), benchtop XRF (CES) and ICP-MS (IDAEA and ERG), and ICP-MS at two laboratories (IDAEA and ERG). NR = not reported; BD = below limit of detection.

Sample	Element	Xact Daily Average (ng m ⁻³)	CES XRF Results (ng m ⁻³)	Background (ng cm ⁻²)	IDAEA Results (ng m ⁻³)	ERG Values (ng m ⁻³)	Background (ng m ⁻³)	% Difference						
								CES vs. IDAEA (CES - IDAEA) / IDAEA	Average CES vs. IDAEA	CES vs. ERG (CES - ERG) / ERG	Average CES vs. ERG	ERG vs. IDAEA (ERG - IDAEA) / IDAEA	Average ERG vs. IDAEA	Average Xact vs. CES (Xact - CES) / CES
Field_PSI_213	Zn	31.2	19.2	1.0	19.6	22.4	64.0	-1.9	-1.5	-14.2	-0.8	14.4	0.4	62.1
Field_PSI_218		25.4	16.5	1.0	15.8	15.9	64.0	4.2		3.5		0.7		53.9
Field_PSI_224		30.4	18.8	1.0	20.2	17.4	64.0	-6.7		8.2		-13.7		61.4
Field_PSI_213	Sr	NR	60.0	0.0	61.1	58.6	0.9	-1.9	-36.6	2.4	-32.5	-4.1	-8.3	NR
Field_PSI_218		NR	1.5	0.0	2.0	1.9	0.9	-23.9		-18.4		-6.8		NR
Field_PSI_224		NR	0.2	0.0	1.1	0.9	0.9	-84.0		-81.4		-14.0		NR
Field_PSI_213	Cu	58.2	44.6	0.0	42.2	49.9	1.9	5.5	-0.1	-10.6	-15.2	18.0	17.8	30.5
Field_PSI_218		39.1	25.8	0.0	26.1	31.3	1.9	-1.3		-17.6		19.8		51.5
Field_PSI_224		35.8	21.1	0.0	22.1	25.5	1.9	-4.5		-17.3		15.5		69.7
Field_PSI_213	Pb	4.4	NR	0.0	4.4	4.5	8.7	NR	NR	NR	NR	2.2	-14.8	NR
Field_PSI_218		4.3	NR	0.0	4.0	3.4	8.7	NR		NR		-15.5		NR
Field_PSI_224		5.3	NR	0.0	4.6	3.2	8.7	NR		NR		-31.1		NR
Field_PSI_213	Fe	757.1	529.7	33.6	465.5	479.4	74.9	13.8	12.4	10.5	9.7	3.0	2.5	42.9
Field_PSI_218		1021.1	805.4	32.8	685.4	732.4	74.9	17.5		10.0		6.9		26.8
Field_PSI_224		906.6	661.6	32.9	624.7	609.7	74.9	5.9		8.5		-2.4		37.0
Field_PSI_213	K	2640.7	2046.1	0.0	2263.0	2663.0	41.2	-9.6	9.7	-23.2	0.3	17.7	9.9	29.1
Field_PSI_218		225.8	194.2	0.0	143.0	156.4	41.2	35.8		24.2		9.4		16.3
Field_PSI_224		209.5	136.7	0.0	133.0	136.7	41.2	2.8		0.0		2.8		53.2
Field_PSI_213	Ca	213.5	172.1	12.9	160.6	795.5	190.0	7.2	4.7	-78.4	-27.6	395.4	135.6	24.1
Field_PSI_218		791.9	597.8	12.6	555.7	602.9	190.0	7.6		-0.9		8.5		32.5
Field_PSI_224		517.9	324.9	12.7	327.2	336.9	190.0	-0.7		-3.6		3.0		59.4
Field_PSI_213	Mn	7.3	8.3	0.1	4.3	5.8	1.4	94.7	57.5	42.3	25.5	36.9	24.6	-12.2
Field_PSI_218		12.3	11.9	0.1	8.1	9.7	1.4	46.9		22.8		19.6		3.1
Field_PSI_224		11.1	9.3	0.1	7.1	8.4	1.4	30.9		11.5		17.3		19.3
Field_PSI_213	Se	BD	BD	0.0	0.3	0.5	0.1	NR	168.8	NR	-13.3	64.0	105.1	NR
Field_PSI_218		0.3	BD	0.0	0.4	0.5	0.1	NR		NR		41.3		NR
Field_PSI_224		0.7	1.0	0.0	0.4	1.1	0.1	168.8		-13.3		210.0		-25.5
Field_PSI_213	Ba	109.1	88.9	13.6	110.9	111.0	81.1	-19.8	-19.8	-19.9	-19.9	0.1	-33.2	22.7
Field_PSI_218		14.3	BD	13.3	10.7	8.5	81.1	NR		NR		-20.5		NR
Field_PSI_224		9.2	BD	13.3	8.6	1.8	81.1	NR		NR		-79.0		NR

- 5 The data shows a somewhat better comparison between offline XRF and ICP than between online XRF (Xact) and ICP, though the scatter in the relative differences (XRF-ICP/ICP) varied from -37 % (Sr) to +57 % (Mn), if Se and Ba, for which only one filter shows concentrations above the XRF detection limit, are not considered. Comparing the ICP-MS results between the two labs (ERG-IDAEA/IDAEA) shows a range from -33 % (Ba) to +25 % (Mn), when Se and Ca are not considered. Se concentrations are close to their ICP MDL and hence rather uncertain, while Ca shows a problem with one
- 10 ERG measurement. If Ca and Se are excluded, the average relative difference between the two labs is -0.1 %, with a standard deviation of 19 %. Similarly the agreement between each of the labs and bench top XRF is good as well. If Se and Ca are excluded the average percent difference between XRF and IDEA is 5.4% while the difference between XRF and ERG is -3.1%. The comparison of the daily averaged Xact values with the benchtop XRF values shows an average difference of 37 % (Xact-CES)/CES) for the elements Zn, Cu, Fe, K, Ca, and Mn, which is close to the observed mean difference to ICP. It is
- 15 also consistent in the sense that all average differences Xact – CES for these elements are positive. The benchtop XRF and the Xact are typically within 5% when analysing the same standard. Further both benchtop XRF and Xact use the same type of fitting routine (with minor differences in the determination of spectral background), hence the most likely explanation for the difference between the Xact and the labs is differences due to sampling or sampling location.

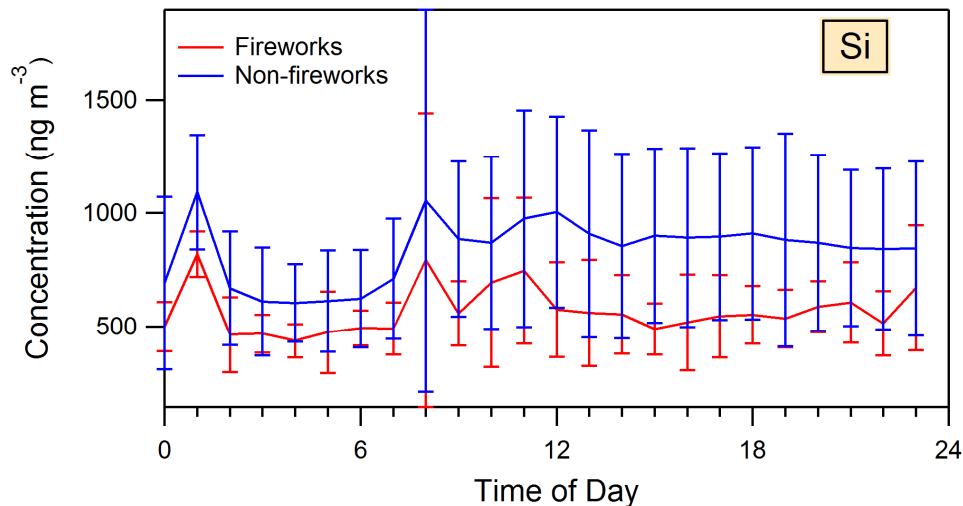
S2S3. Spiked filter samples for method intercomparisons

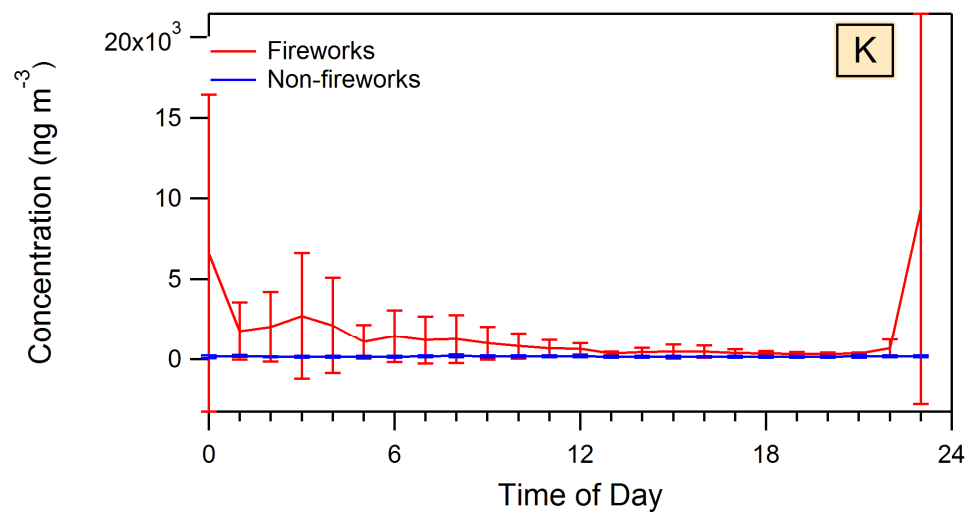
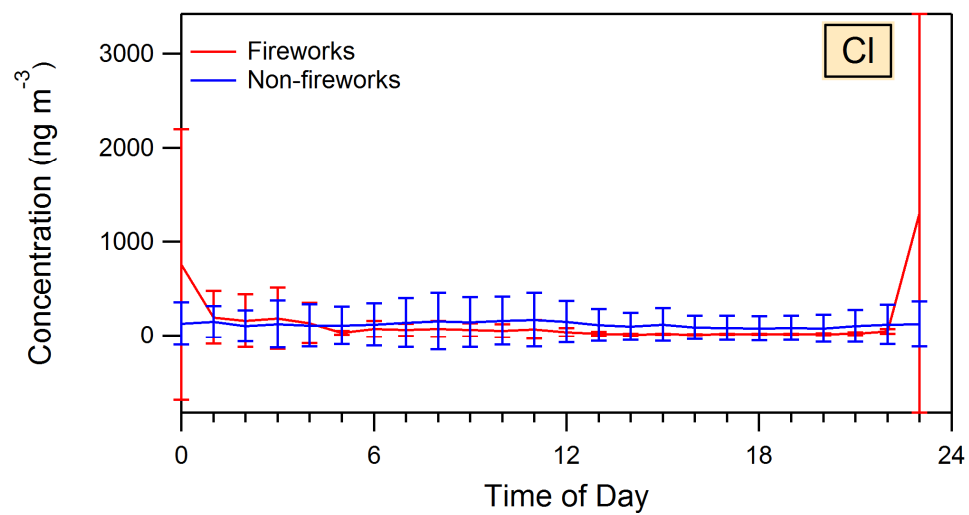
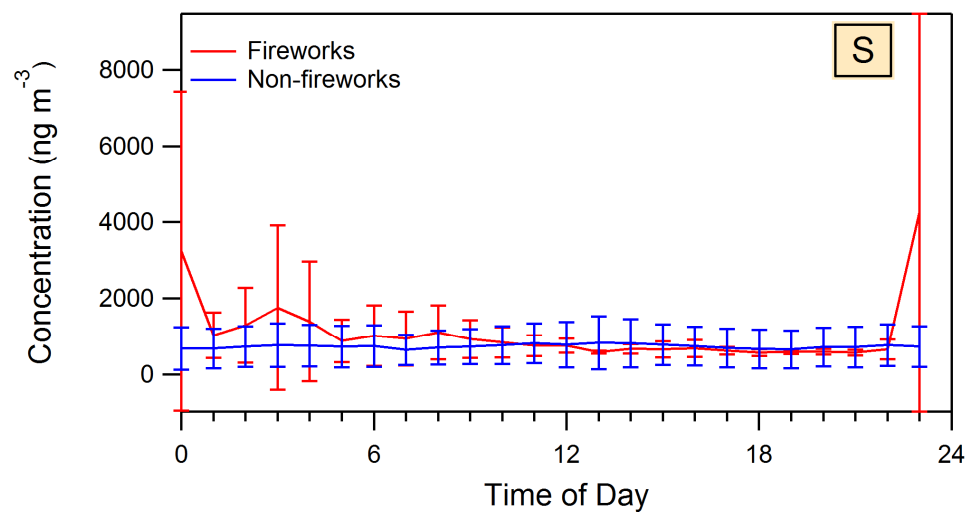
- 20 CES produced a set of six quartz filters coated with known amounts of the elements Zn, Sr, Cu, Pb, and Fe. These filters were analysed with a benchtop XRF instrument by CES, and three each of them were sent to IDAEA-CSIC, and ERG for analysis with ICP-MS. The results are presented in Table S2. Notice that Pb is not reported for XRF, because of large variations of the measured values for quartz filters. This indicates a problem with the XRF fitting routine for quartz filters, as the issue is not seen with Teflon filters.
- 25 **Table S2S3. Spiked filter analyses for five elements. Comparison between XRF and ICP-MS analyses performed at three independent laboratories.**

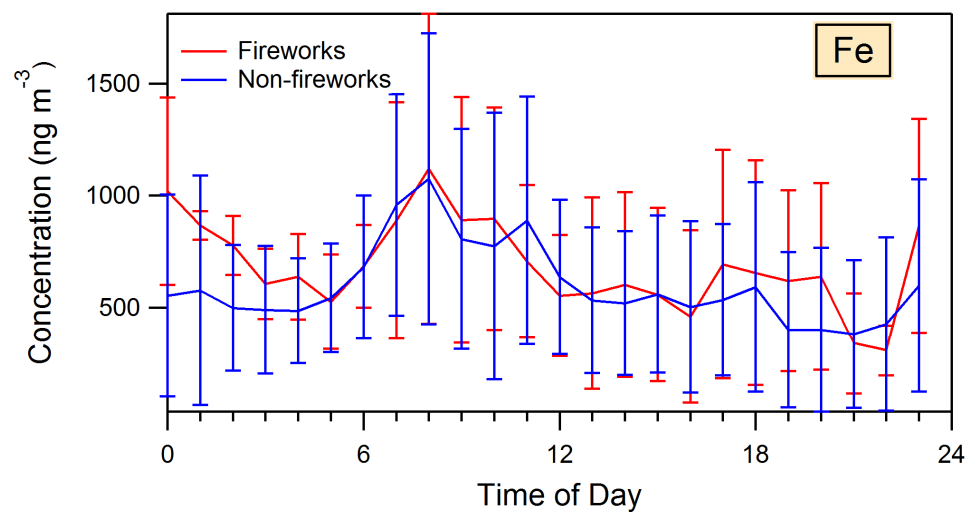
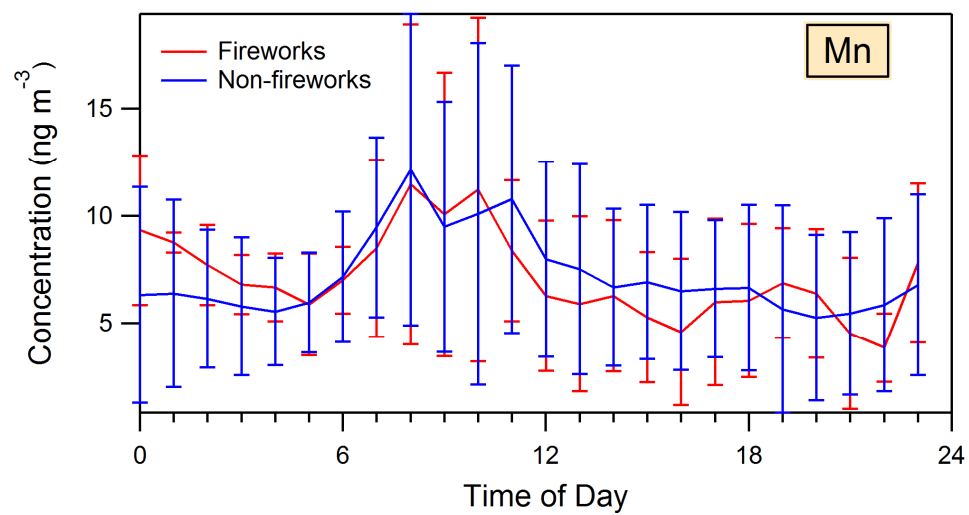
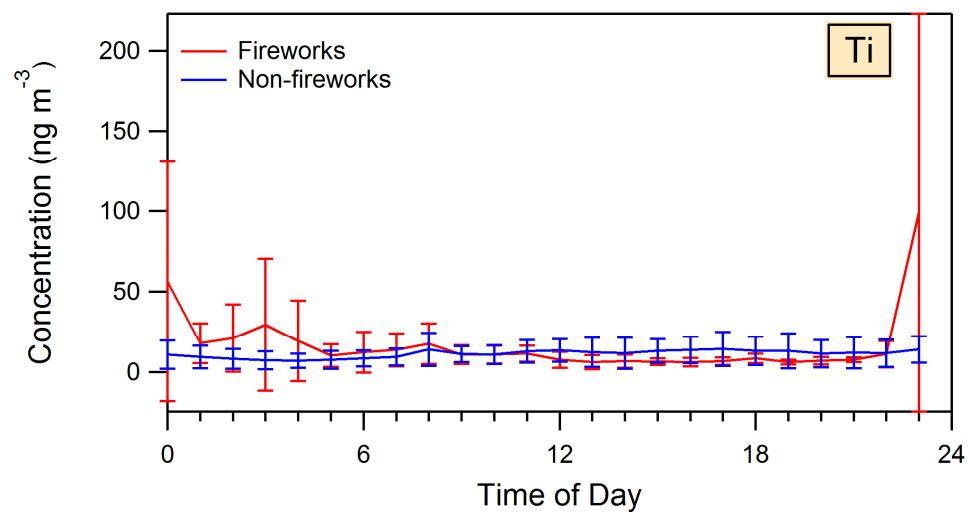
Sample Start Time	Sample	Element	CES			IDAEA	ERG			% Difference					Average Percent Difference				
			Spiked Conc. (ng cm ³)	XRF Conc. (ng cm ²)	Blank (ng cm ²)	Conc. (ng cm ³)	ERG Values (ng cm ²)	Blank (ng cm ²)	Spiked vs. CES (CES - spiked) / spiked	Spiked vs. IDAEA (IDAEA - spiked) / spiked	Spiked vs. ERG (ERG - spiked) / spiked	IDAEA vs. CES (CES - IDAEA) / IDAEA	ERG vs. CES (CES - ERG) / ERG	Spiked vs. CES (CES - spiked) / spiked	Spiked vs. IDAEA (IDAEA - spiked) / spiked	Spiked vs. ERG (ERG - spiked) / spiked	IDAEA vs. CES (CES - IDAEA) / IDAEA	ERG vs. CES (CES - ERG) / ERG	
21.04.2016 11:50	PQ042116A	Zn	97.4	88.1	10.2	133.1			-9.6	36.6		-33.8		-9.7	32.0	22.4	-30.3	-24.6	
21.04.2016 12:25	PQ042116B		97.4	89.2	10.2	155.0			-8.4	59.1		-42.4							
21.04.2016 13:00	PQ042116C		97.4	83.5	10.2	97.7			-14.3	0.3		-14.5							
21.04.2016 15:13	PQ042116D		97.4	88.3	10.2		104.1	30.7	-9.3		6.8		-15.1						
21.04.2016 15:46	PQ012116E		97.4	92.4	10.2		123.2	30.7	-5.1		26.4		-25.0						
21.04.2016 16:19	PQ042116F		101.5	90.2	10.2		136.0	30.7	-11.2		34.0		-33.7						
21.04.2016 11:50	PQ042116A	Sr	206.0	191.6		178.6			-7.0	-13.3		7.3		-6.2	-15.7	-7.0	12.7	0.8	
21.04.2016 12:25	PQ042116B		206.0	194.0		194.5			-5.8	-5.6		-0.2							
21.04.2016 13:00	PQ042116C		206.0	193.9		147.8			-5.9	-28.2		31.1							
21.04.2016 15:13	PQ042116D		206.0	191.6			190.5	1.5	-7.0		-7.5		0.5						
21.04.2016 15:46	PQ012116E		206.0	194.0			189.8	1.5	-5.8		-7.9		2.2						
21.04.2016 16:19	PQ042116F		206.0	193.9			194.6	1.5	-5.9		-5.5		-0.3						
21.04.2016 11:50	PQ042116A	Cu	127.6	108.1		111.9			-15.3	-12.3		-3.4		-13.6	-17.0	5.8	5.9	-18.2	
21.04.2016 12:25	PQ042116B		127.6	110.6		117.8			-13.3	-7.7		-6.1							
21.04.2016 13:00	PQ042116C		127.6	111.9		87.9			-12.3	-31.1		27.2							
21.04.2016 15:13	PQ042116D		127.6	108.1	0.8		131.2	1.3	-15.3		2.8		-17.6						
21.04.2016 15:46	PQ012116E		127.6	110.6	0.8		129.4	1.3	-13.3		1.4		-14.5						
21.04.2016 16:19	PQ042116F		127.6	111.9	0.8		144.3	1.3	-12.3		13.1		-22.5						
21.04.2016 11:50	PQ042116A	Pb	20.5	NR		22.6			NR	10.3		NR		NR	32.4	57.7	NR	NR	
21.04.2016 12:25	PQ042116B		20.5	NR		37.9			NR	84.7		NR							
21.04.2016 13:00	PQ042116C		20.5	NR		20.9			NR	2.1		NR							
21.04.2016 15:13	PQ042116D		20.5	NR			27.1	1.1	NR		32.0		NR						
21.04.2016 15:46	PQ012116E		20.5	NR			29.1	1.1	NR		41.8		NR						
21.04.2016 16:19	PQ042116F		20.5	NR			40.9	1.1	NR		99.4		NR						
21.04.2016 11:50	PQ042116A	Fe	3024.6	2759.6		2827.6			-8.8	-6.5		-2.4		-8.1	-3.6	-0.8	-2.1	-7.2	
21.04.2016 12:25	PQ042116B		3024.6	2795.5		3543.9			-7.6	17.2		-21.1							
21.04.2016 13:00	PQ042116C		3024.6	2786.4		2377.9			-7.9	-21.4		17.2							
21.04.2016 15:13	PQ042116D		3024.6	2759.6	154.5		2901.2	140.5	-8.8		-4.1		-4.9						
21.04.2016 15:46	PQ012116E		3024.6	2795.5	154.5		2901.2	140.5	-7.6		-4.1		-3.6						
21.04.2016 16:19	PQ042116F		3024.6	2786.4	154.5		3202.9	140.5	-7.9		5.9		-13.0						

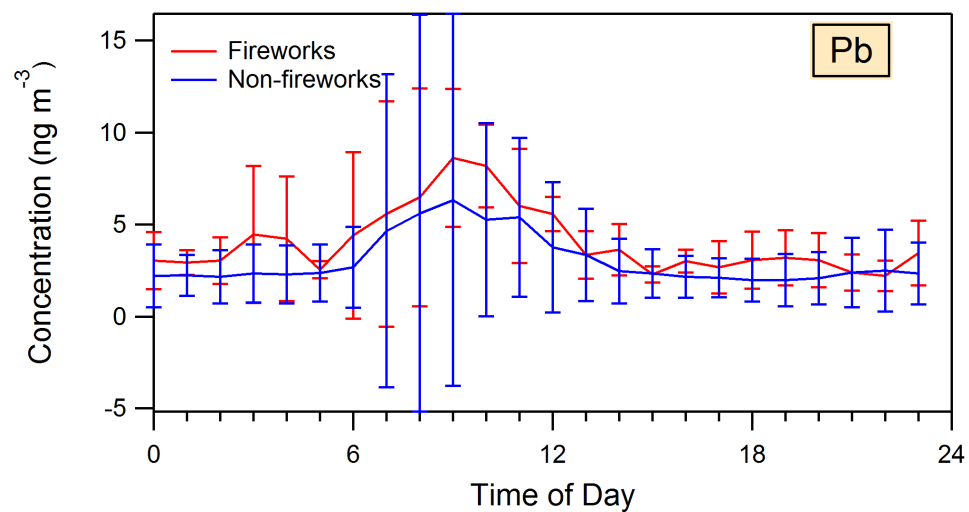
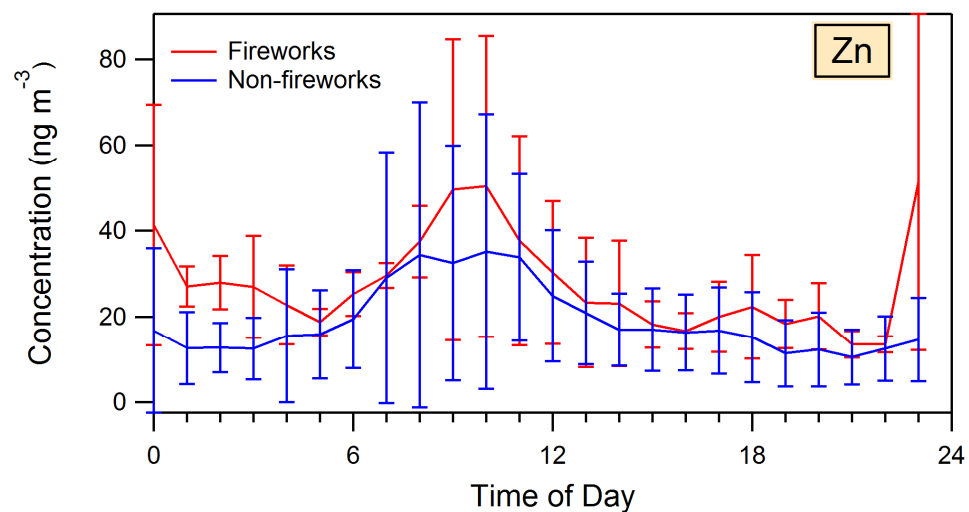
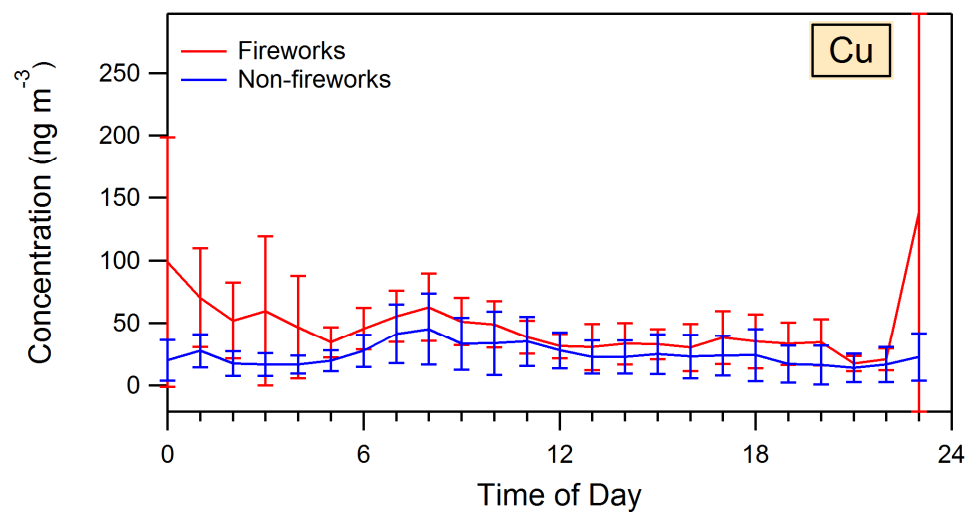
Tests with specifically produced reference samples of Fe, Cu, Zn, Sr, and Pb (Table S2) showed relative differences between the measured concentrations and the theoretically expected concentrations ranging from -6.2 % (Sr) to -13.6 % (Cu) for benchtop XRF, on average -9.4 % (without Pb). For all these elements, XRF underestimated the expected value, -as expected for absorption of fluorescence radiation by the quartz fiber material (Tanner et al. 1974). Similarly spiked teflon filters (not shown) also showed underestimation of the expected concentrations, though not as much as for the quartz filters. A statistical analysis revealed that at the 99 % confidence level only Cu showed a significant difference between the two filter types. ICP showed differences between -17 % and +32 % (average 5.6 %) for IDAEA-CSIC, and -7 % and +58 % (average 15.6 %) for ERG for quartz filters. The scatter is much larger than for the field samples, and differences can be positive or negative.

§3.4. Diurnal variations of elements for fireworks and non-fireworks periods









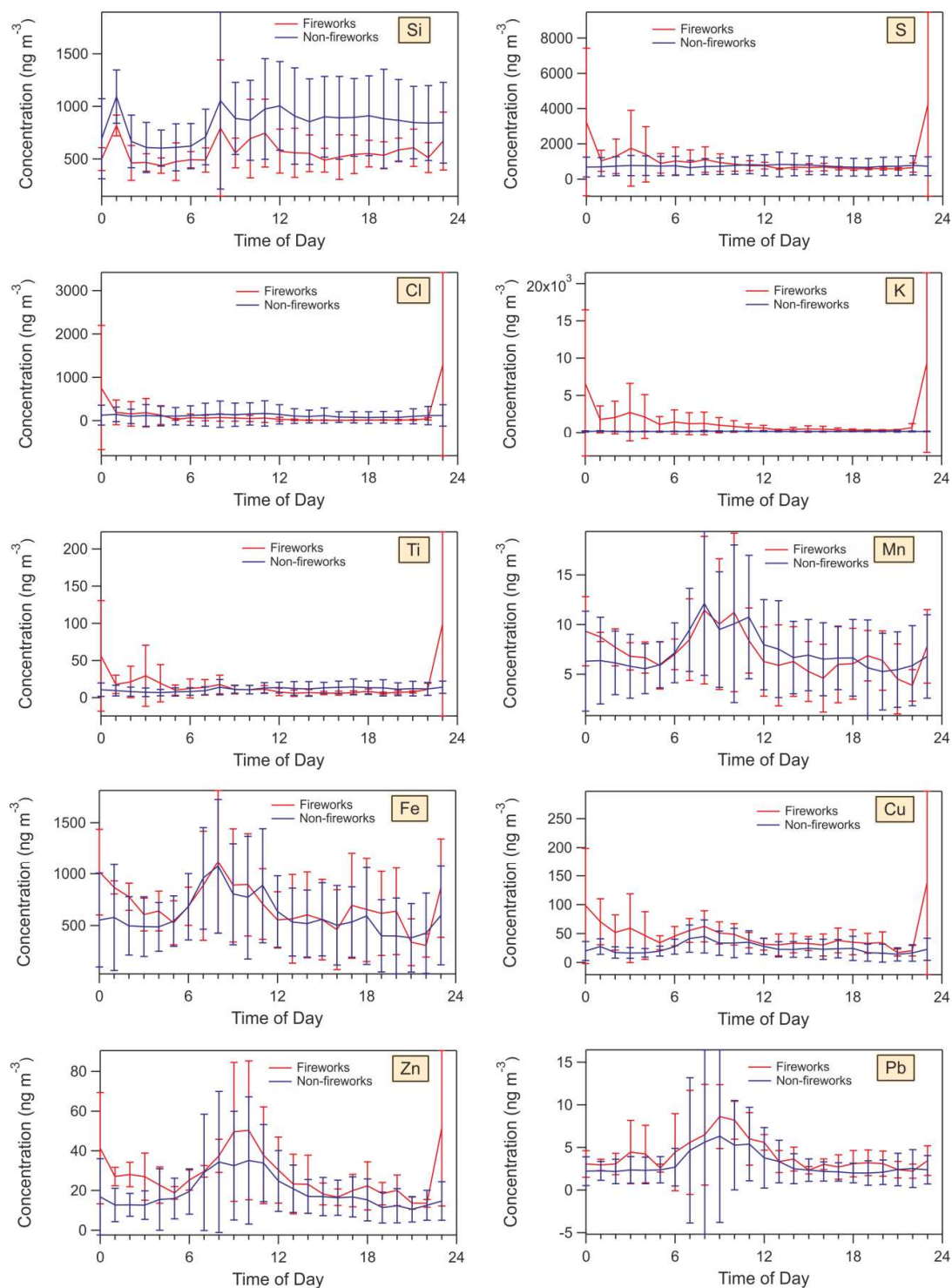
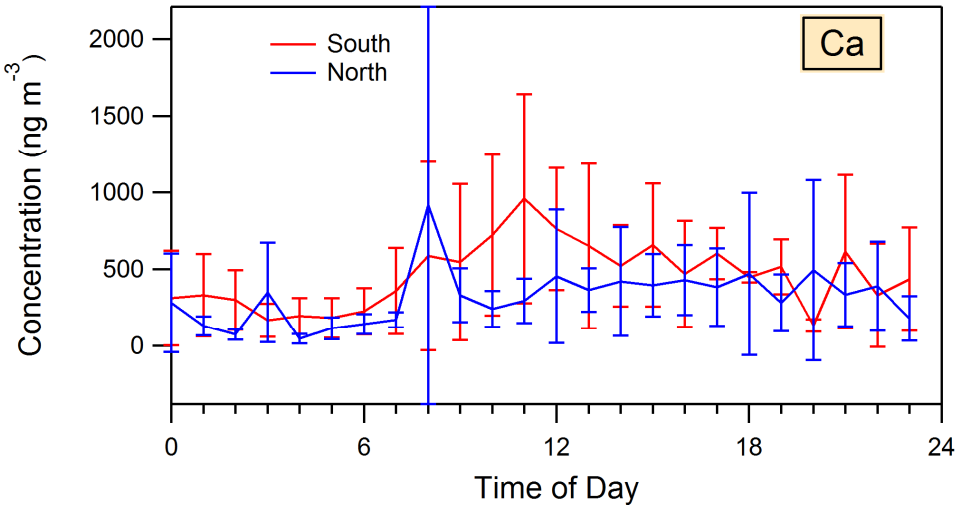
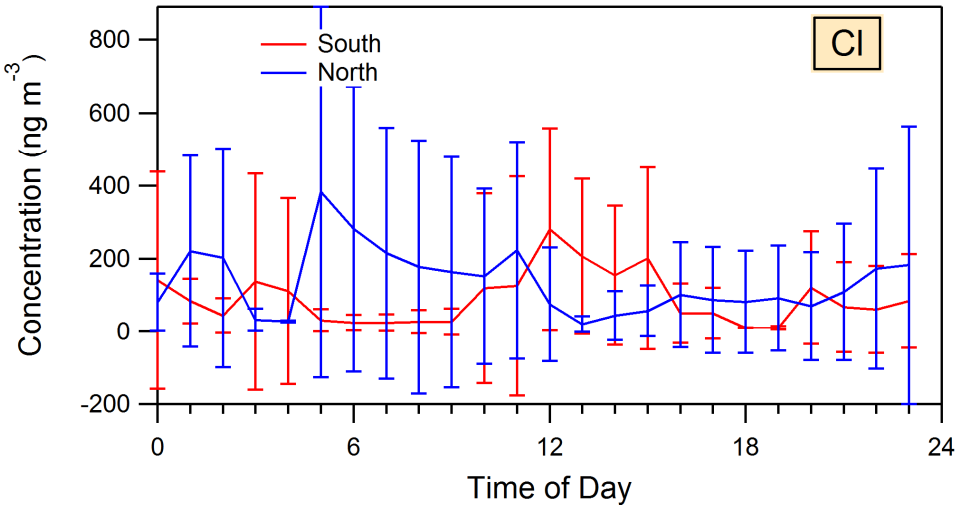
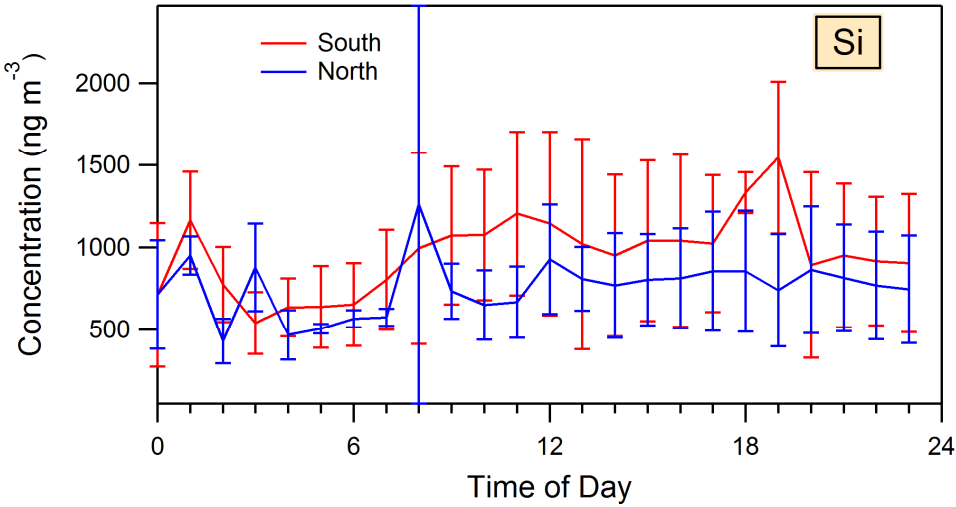
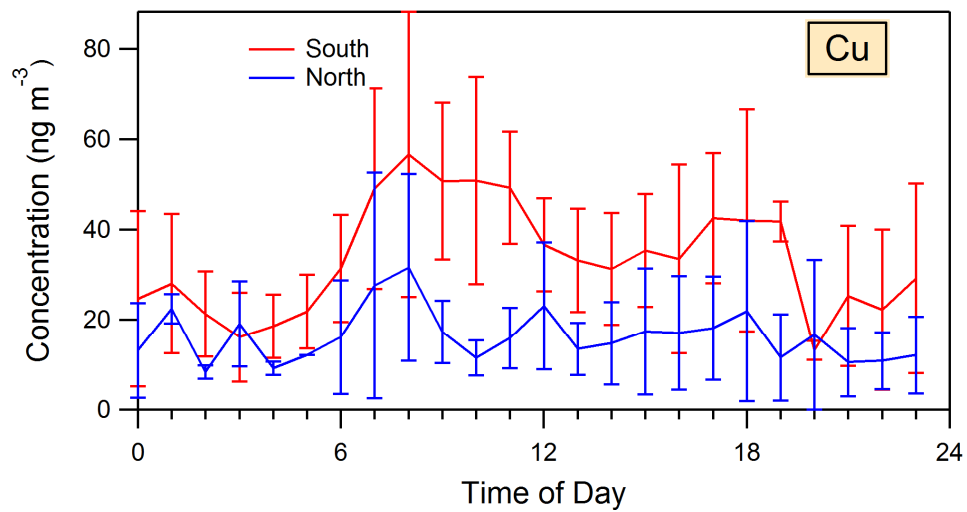
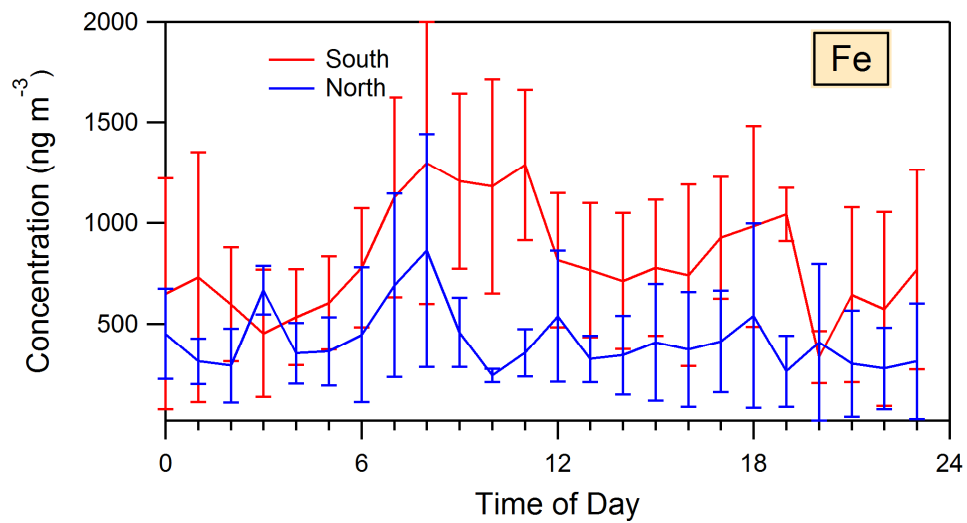
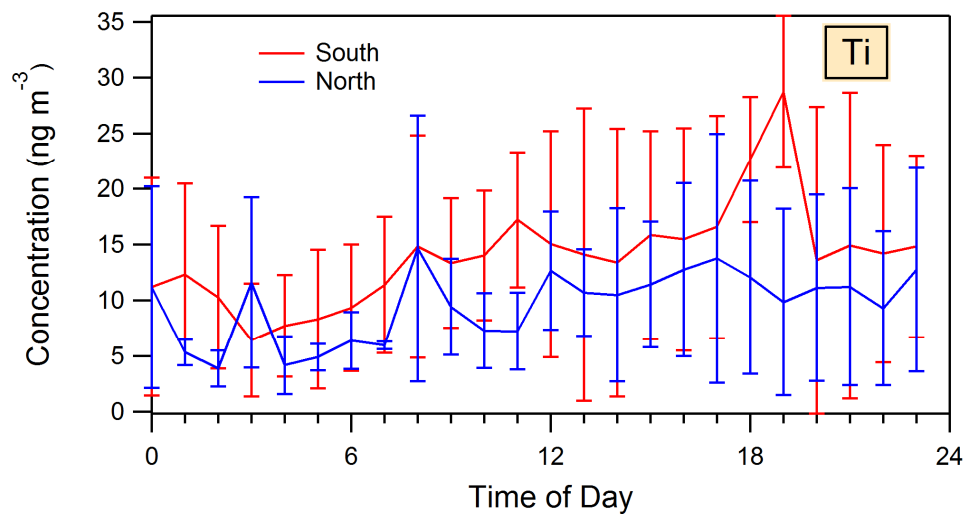
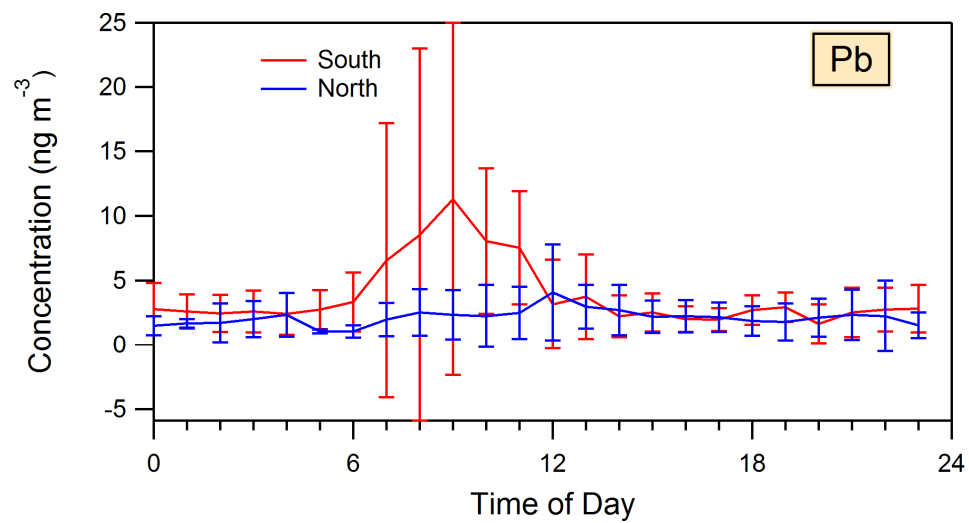
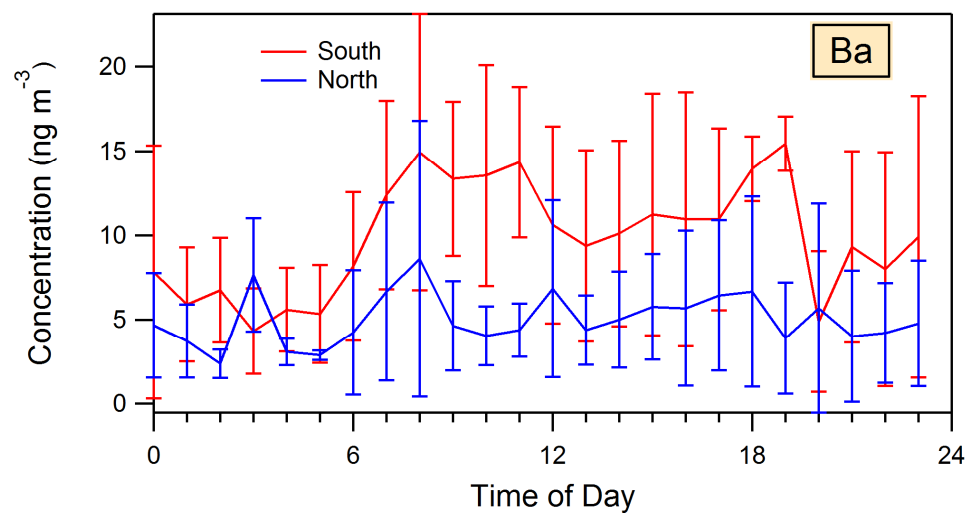
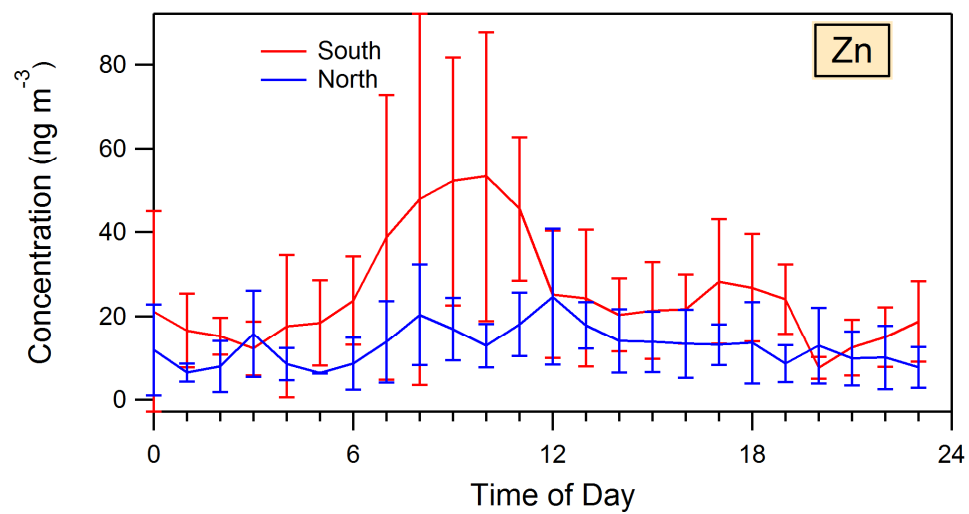


Figure S3S4: Diurnal variations of the Group A elements Si, S, Cl, K, Ti, Mn, Fe, Cu, Zn, and Pb. See Fig. 76.







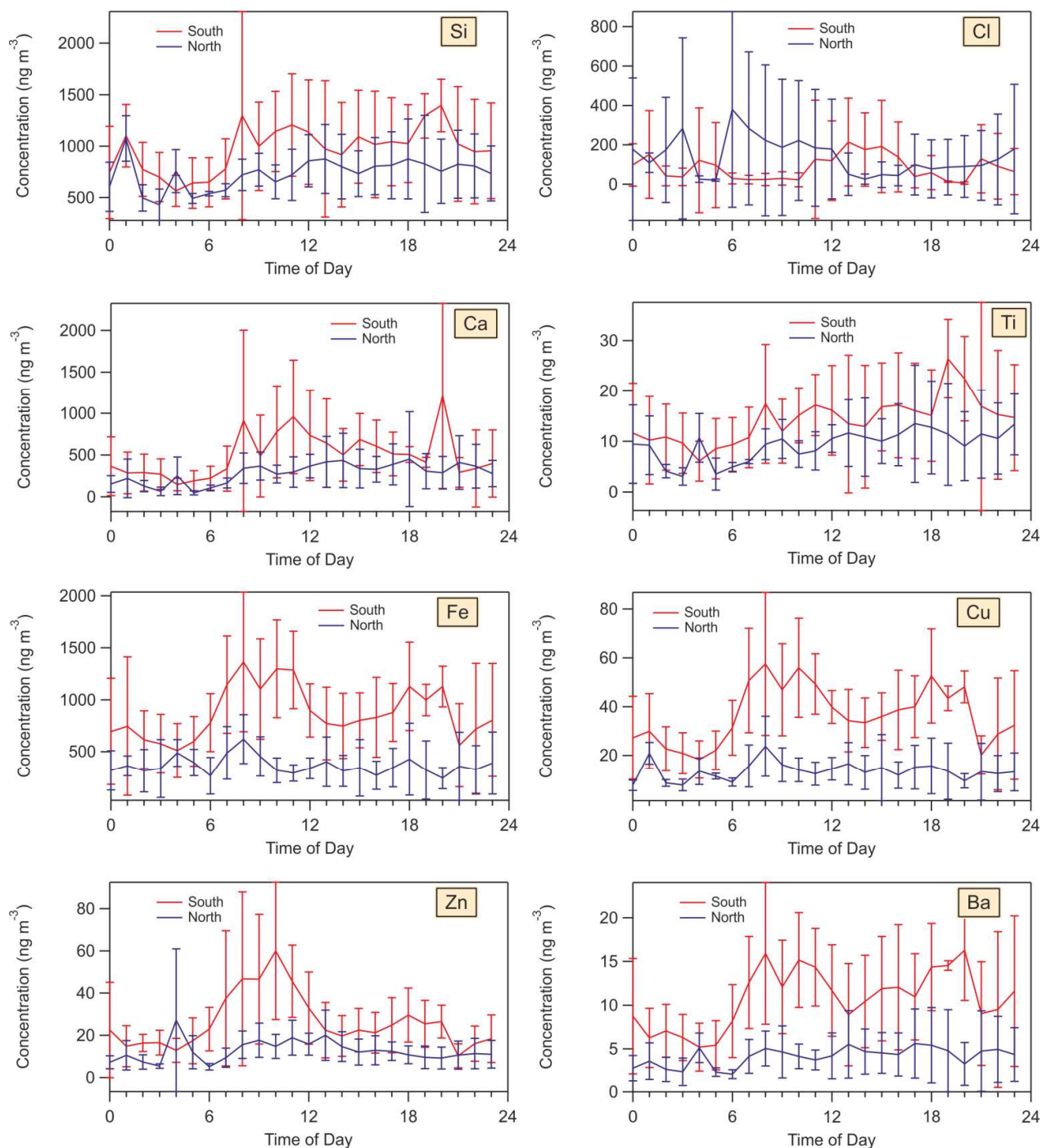


Figure S4S5: Diurnal variations of the Group A elements Si, Cl, ~~K~~, Ca, Ti, ~~Mn~~, Fe, Cu, Zn, and Ba, ~~and Pb~~. South means a wind from the freeway towards the station. See Fig. ~~408~~.

Reference

- Tanner, T. M., Young, J. A., and Cooper, J. A.: Multielement analysis of St. Louis aerosols by nondestructive techniques, *Chemosphere*, 3, 211-220, 1974.