

Interactive comment on “Improved Atmospheric Characterization through Fused Mobile Airborne Surface In Situ Surveys: Methane Emissions Quantification from a Producing Oil Field” by Ira Leifer et al.

Anonymous Referee #1

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Comments on Leifer et al: Improved atmospheric characterization:

General comments: This is an interesting and innovative paper, describing how both mobile surface measurement across strong local topography and parallel aircraft measurements have been used to derive methane emission fluxes. This is an important problem, and has wide applications. By using aircraft and vehicle to validate each other, both methodologies are improved. In the US, instrumented aircraft are available. In many less wealthy nations that is not the case and measurement will have to depend on vehicles supplemented by light drones. The topographic problem is also widely applicable, in the many locations where the boundary layer is poorly mixed. Thus although aspects of the experiment can be criticized and the introductory section is very inadequate, the paper is an interesting and useful contribution, that should be published after revision.

Specific details: The Introductory section is incoherent, and depends in part on outdated papers. It needs to be revised substantially.

Line 35 should probably cite up to date NOAA results, perhaps from Dlugokencky’s papers. For lifetime (also line 43-44), be careful not to mix up the various definitions of lifetime and again best to use a recent Dlugokencky paper or use IPCC.

[Yes, Ed’s work now cited.](#)

Line 38 – Nisbet et al 2015 is an error. No such paper. It’s 2014. Maybe better cite the major Nisbet et al 2016 Glob. Biogeo. Cycles paper, and also Schwietzke et al. (in Nature recently) on fossil fuel emission. Either here or in L50 cite Saunio, M. et al (2016) The global methane budget 2000–2012, Earth System Science Data, 8, 697-751, doi:10.5194/essd-8-697-2016.

[We have cited Saunio et al 2016 and Nisbet et al 2016. There was an error in the endnote database for Nisbet 2014 that has been corrected.](#)

Line 43 Claim of a 40% drop in lifetime - There is a major discussion on whether or not OH is changing and maybe that’s far beyond the scope of this paper. Also may be confusing lifetime definitions (perturbation/replacement) in general comment.

[We agree that lifetime, although worthy of mentioning in general, is well beyond the scope of the paper and thus the lifetime aspects are now greatly reduced.](#)

Line 49. See for example Rigby, M. and 18 others (2017) Role of atmospheric oxidation in recent methane growth. Proc. Natl. Acad. Sci. USA. 114, 5373-5377.

[Now cited in the OH loss paragraph with a summary sentence.](#)

Line 50. Maybe cite Saunio, M. et al (2016) again.

[Done](#)

Line 51 on – several more recent papers to cite on US fossil fuel emissions. Bruhwiler et al (2017) in JGR

very effectively questioned Turner et al. , so I think this could be rewritten as the Turner et al conclusions should be discounted.

Deleted.

Line 58-59 – The inventory discussion could be picked up around line 55 and should lead into Kirschke et al, and Saunio et al (2 papers, 2016, 2017) and maybe mention Nisbet and Weiss’s earlier top down/bottom up comment in Science, 2010.

Line 62 – many more recent papers than White 1976!

True, but they really do not make this point as specifically or as well. Also, much of the science of my father’s generation seems to be in the process of disappearing down the memory well – and so I would like to keep White highlighted. That said, the works of Peischl et al. (2015; 2016) is added. I note that both Peischl et al., (2015;2016), also cite White et al. (1976).

This paragraph L61-70 needs updating. Also it could mention some of the European work – for example in several papers by Bergamaschi et al.

Have added Saunio et al. (2017), who is working these days with Bergamaschi on inversion modeling.

Line 92-3 could be written by a PR firm selling dodgy goods. Most people live on the plains, not in the high Himalayas.

Point taken. Deleted. Although most people lived on the plains, wherever feasible, the wealthy have always lived on more defensible hills above the plains where the poor lived (and were flooded, marauded, etc.

Line 120 – reference for air flow to 1km only?

Zhong et al., (2004) again. Added.

Figure 1a looks like a photo of some poster on the floor: : terrible, key is unreadable. Why the angled projection???

Redone

Line 168 ‘stranded CH₄ clouds’ – we see these too, but don’t call them stranded. They are not washed up on the beach (strand) – they are making their own independent progress as yet unmixed. They are important as they come from single event sources, which can be keys to the emissions from a gas field – e.g., water processing.

Agreed, they are not stranded on a beach. Rephrased as “upwind CH₄ plumes”

Line 196 - precision and accuracy please, calibration etc etc. (though I accept the point in L252)

Added.

Lines 228-247 calculation and also 248-253. Maybe cite some of the Aliso canyon work here?

Done. Thanks!

L271 – pre/post-surveys. Good. When was this? How comparable in meteorology?

Added. “Primary changes were development of near surface winds, and a slight increase in the PBL”

L300-25 This section is very interesting. I’m surprised the PBL is located to a 20-m precision in the

complex topography. Seems very precise.

Its sharp – and something that we often see in California. The Marine layer PBL is often visualized by clouds and is equally sharp (almost every day on my morning commute), sometimes to a few meters!

L312 – tall pine trees? How tall - Not being Californian, are these 5m (tall in N.East Norway hills) or 150m redwoods?

They are tall in a California sense. Added that they are 30m+

L342 – really a town called Bodfish? are they inland codfish? or an obscure mermaid-body genus? Is this paper really about Bodfish bubbleology? The mind boggles.

It was a gold rush town named after a gentleman last name bodfish. The history of California is weird!

L365 – there is a lot of agriculture in the area – cows, wet fields, drains. This needs to be discussed. Are there any large cattle feed lots in and around Bakersfield? Landfills?

The one nearby dairy in the upwind direction is now pointed out on fig. 7A, and discussed. Potential plumes from the only nearby upwind dairy (Fig. 7a, white arrow) were directed by winds to pass to the west of the oil fields. The only Landfill is to the south (and not along the flow path. There are no wet fields in this part of the San Joaquin Valley – that is only in the Sacramento area. Water is a little too precious here, and it gets too hot!

Fig 8 – really interesting figure could have a bit more discussion.

L396 – how rapid is the rise? Any guesses?

I think the answer is pretty darn fast, methane has half the density of air, but mixing is not going to be similar to that for a 600K plume. In the Aliso Canyon leak (see <https://www.youtube.com/watch?v=exf8VPQDTY>) it looks like many meters per second. That said, it seems a bit overly speculative to put a number in print....



L402 – nice to get some d13C(CH4) isotopes to pin down the methane. See comment above about cattle and landfills in the area. Also Bakersfield is a big city – sewage.

Agreed, however, no isotope instruments were available.

L435 – intermittent activity – what the paper calls ‘stranded plumes’ – any idea how common this is?

For CO2, this would be more from cycling of the co-generation power plant being on and off, the language is now clarified to “the co-generation plant only being active some of the time, confirmed by data from the GOSAT-COMEX campaign.”

L440,442 – collected Lagrangian – English language disjunction. ?in Lagrangian mode??

Changed to “in a Lagrangian sense”

L441 – poor English again –a platform? Or platforms?

Changed to “measurement platforms”

L471 – faux precision. 1.6km

Changed. Thank you, that is something I usually catch in reviews I do!

L473-480 – good.

L503 – mountain peak – hitherto hidden in the discussion, now pops out. I hope you told the pilot: : : could be mentioned earlier as a problem.

The pilot flies more or less where we ask, but of course they can change their route as they see fit. Also, they are military pilots flying a converted military jets, and tend to do things their way a bit, For example, in one of the flight segments (not here), the pilot realized he needed. Now discussed in section 2.1

L532 – maybe cite some of the NOAA work in other gasfields – intermittent sources.

Apologies, but the intermittency for CO₂ was from cycling the power plan on and off. Actually, and this is the subject of another paper that we are writing, on the field level, CH₄ emissions are not intermittent, but stochastic – law of large numbers of leaks. In that paper, though, we do show that at the well level, they are intermittent. The text has been adjusted slightly.

L533 – CO₂ – maybe, or just a lot of trucks and all sorts of other sources?

Thanks, this needed extra clarification. Added “Additionally there are no upwind (non-oil field) roads, only the foothills of the Sierra Nevada Mountains.”

L540 and 551 – would be nice to see some isotopes on this.

Agreed, however, no isotope instruments were available.

L556 – repeat comment – no, we do NOT all live on the tops of Everest and Aconcagua. We live in the swamps of Calcutta and Shanghai. Schiphol runway, the center of Holland, is 11 ft under the ocean.

Agreed, this comment is very useful for Tehran and La Paz (both with major local gasfields) but it’s irritating to hammer it twice.

Having lived a year in Den Haag, I understand your point. Deleted.

Overall – interesting paper. Accept with revision.

Improved Atmospheric Characterization through Fused Mobile Airborne & Surface In Situ Surveys: Methane Emissions Quantification from a Producing Oil Field" Leifer et al.

Reviewer #2 Comments:

This manuscript describes a campaign conducted using a surface mobile platform along with an airborne platform, to estimate CH₄ and CO₂ emissions from California's Kern Oil Field using one day of measurements. In principle, this is a valid attempt to use two different platforms and merge the data sets to create a better picture of emissions, which would be worthwhile and could help improve current methodologies. But in practice, in my opinion, this manuscript does not accomplish that. For a submission to AMT, there is very little precision in the method description or the description of the measurements themselves. There is no quantification of how well the concentration measurements from the two platforms compare, nor is there any description of the interpolation method for the sparse measurements in space and time.

The description of the flux calculation is confusing and the description of the uncertainty analysis is too short, especially given the focus of this journal on measurement techniques and methods.

In the conclusion several statements are made that are, in my opinion, overly broad and have not been shown in the work, including the interpretation of the final emissions estimate in relation to a bottom-up inventory estimate and in relation to global trends.

In addition, it is not clear to me even that this method works in such complex terrain and conditions (the authors do spend a large portion of the paper describing the terrain, which they use to their advantage), which the authors do note. It is certainly not shown that this method might be better than any other method in terms of accuracy - this may not be required for publication, but it is claimed by the authors.

The manuscript would be improved by focusing on the methodology itself and justifying the various methods used to perform the estimate, and remove the focus on the estimated emissions result and the global methane budget.

- We agree with this assessment and plan to address flux and global implications in a future paper. Thus, we have deleted the discussion as recommended by the reviewer. We have clarified our argument that the approach is better and under what conditions, based on the philosophy that reducing extrapolation improves outcomes. We now devote the first paragraph of section 4.2 to this important issue. Additionally, we have re-arranged section 4.2 to remove some duplication. Also, we have toned down the title.

- WRT to the approach being better – our contention is simple – more information or more complete characterization of the PBL means less interpolation and extrapolation which cannot help but reduce uncertainty for any method. As to whether the improvement is significant or trivial, that of course depends on the situation. In any case, the text no longer makes the claim.

- With respect to leveraging terrain, I fully accept it would be better if I had an airplane and could fly and collect data. But most scientists on this planet very seldom have access to an airplane, particularly on a regular basis, and as a result, by leveraging terrain, scientists like myself, can collect useful vertical atmospheric profile data. In the San Joaquin Valley, agreement between the airplane data and car data was very good over a wide range of altitudes indicating that at a minimum, the method has significant promise. We have added a paragraph at the beginning of section 4.3 that now recommends further research and highlights that it may not be appropriate elsewhere.

- Based on the comments from the reviewer, we decided to add a paragraph to the methodology that explains why we are using an anomaly approach rather than an upwind/downwind mass balance approach. See the paragraph, which is presented below. Since this is key, a short summary now also appears in the study motivation section.

- “Unfortunately, the upwind data showed a lateral gradient, which coupled with uncertainty in precisely where the downwind air originated (given the topography, which features a gentle

incline towards the northeast, this gradient is unsurprising, in retrospect). Thus a very small shift in the winds between the upwind and downwind curtains results in a significant shift in C_B , with a very large effect on Q . As a result, the more traditional upwind/downwind mass balance approach was abandoned for an anomaly approach.”

Specific comments:

L 54-55 "most" should be "largest" L55 Does the EPA inventory discuss the global budget? What about agriculture? (The latest Global Methane Budget Saunio et al may be a better citation here, and they place agriculture and waste to be significantly larger contributor than fossil fuel, so this statement does not seem to be supported by the literature).

- As this is background material, we have added to the current text the Saunio reference to highlight that there is uncertainty in the CH4 budget.

L56-60 - Turner and Bruhwiler disagree as to whether a methane trend is detectable over the US; the Bruhwiler paper specifically refutes the results in Turner. So the phrasing here is not really correct, but could be phrased to simply emphasize that there is disagreement in the literature over whether a methane emissions trend in the US exists. More recent literature on the global methane is also available suggesting that the OH sink is the cause of the global increase (as mentioned in lines 41-46). References to a more recent Turner et al and Rigby et al, both in PNAS, 2017 should be made if this is to be discussed. [However, as noted above, perhaps devoting a large portion of the paper to the global methane budget is not really in the scope of this work].

- We agree and delete the sentence as ancillary to the main focus of the study.

L62 - Peischl, White and Karion all use essentially the same method (mass balance) - not sure what is meant by "direct assessment". Perhaps because of the aircraft-based winds used in the first two while Karion relied on model or ground observations?

- Yes, the difference is whether transport (i.e., transport) are measured in the study domain, or need to be modeled correctly. The sentence has been re-written to clarify what is meant by direct assessment.

L69 - This discussion should include a citation to Smith et al, 2015, ES&T where the ethane/methane ratio was not assumed a priori but determined from the airborne data, but still used to apportion emissions in the Barnett.

- Agreed. We missed Smith et al 2015 due to the long time frame between writing the paper and submission – the intro material was written in 2015! We have also added a citation to the more recent Peischl et al., 2016, too. In the same vein, we have added a citation to a recent paper by Schwietzke et al. 2016 to bring the introduction more up to date.

Fig 1: a) Panel in figure is not clear - much too small to read. (b) North should be indicated.

- Replaced by a higher resolution figure. North is now indicated on panel b.

L163-165 repetitive? awkward.

- Thanks, rewritten less repetitive

L167 CGE should be GCE?

- Thanks, yes. Fixed.

Stranded plumes: This should be more clearly described as plumes that are coming into the domain from upwind or outside the domain. At least this seems like what is being described here, that there is a criterion of a "clean" or at least relatively uniform upwind condition.

- Stranded plumes are plumes that due to wind shifts are no longer contiguous to their source (at least in the transit type of data we collect). Stranded plumes have been observed frequently in the SJV by AMOG.

- That said, whether they are stranded or connected to their source, is not relevant to whether they can disrupt the experiment, so the term stranded is dropped.

L170: What is the specific criterion for "too light or variable" on the wind speeds? In my opinion, this is very subjective in the description, including "flush nocturnal accumulations before the overpass" - so this is a restriction on wind speed in terms of transit time?

- Added detailed criteria. Winds speeds typically less than $\sim 2 \text{ m s}^{-1}$, and variability less than 30° . Flush the nocturnal (i.e., no CH_4 cloud at or nearby upwind of the site, which means that winds could not have been light as recently as several hours prior; however, winds are not measured several hours prior),

L179: These studies or most of them used compressed gases as standards either on- board or prior and after the campaigns - calibration is still required with CEAS systems.

- True, and we do calibrations daily. Text reflected to note that the calibration gas does not need to be onboard the platform – it can be at the lab / hotel/ base site, etc.

L193 At what height is the air sample drawn relative to the roof and the anemometer?

- Sample is drawn from between 2 m above the roof and 3 m above the roof, depending on speed. This has been clarified by rewriting the paragraph.

L184-205 Have the environmental variables been compared with local weather stations or other sensors for validation (i.e. of wind measurements, which as is noted in the text, can be difficult because of the need to account for vehicular motion)?

- Pressure: We have compared with the Bakersfield airport, and agreement is within our uncertainty on the altitude of the height of the sensor, and pressure changes over an hour (the airport reporting time).

- Winds are much more challenging they are always changing and always spatially varying. Our best efforts have been to compare short data sets collected at a range of speeds on an open road, early in the morning, to compare wind measurements for driving the road in both directions (with the wind at an angle). Worst performance is at around 45 mph for winds of less than 1 m/s where errors are on the order of 20% in speed and direction. At higher wind speeds and/or lower driving speeds, error decreased rapidly. The GPS correction to real speed error is much smaller as it corrects itself after a few readings, which we distribute across the data by spatial filtering that limits accelerations in the along travel direction to physical limits and to near zero in the direction transverse to the direction of travel. Discussion with Vaisala, indicate that there is no need to send their wind sensor in for annual calibration (barring it being hit by a large branch). We also optimized the wind sensor positioning to minimize uncertainty for winds within $\sim 30^\circ$ from the front of the vehicle. Since the relative wind always has a very strong along travel direction component, this criteria is almost always met for driving at all but the slowest speeds and/or the strongest cross winds. We have not spent effort at looking at accuracy for very high cross wind data (have measured to 17 m/s) because we do not analyze such collected data in our studies to date. Additionally, such strong winds in California tend to be very strongly modified by topography – making them particularly challenging to validate.

- This above discussion has been added to the supplementary material.

I was looking for this reference, which is cited for the instrumentation: "Leifer, I., Melton, C., Manish, G., and Leen, B.: Mobile monitoring of methane leakage, Gases and Instrumentation, July/August 2014, 20-24, 2014.", not clear what journal or no way to find this?

- This paper is in a trade journal (it was peer reviewed) and is attached.

Information should be given in a similar fashion as they are in 2.3 for AJAX on calibration methods. (see later comments on the supplement).

- Done. We have also added our linear cell pressure calibration to the supplemental material.

When merging the two data sets for a single analysis it becomes even more important to show that measurements of methane are on the same scale relative to the same standards, etc. and have been intercompared. The vertical profile indicates that they compare "well" at high altitudes, but no quantity is given.

L230-231 U_n and U_N are both representing perpendicular winds.

- Thanks. Typo corrected.

Section 2.4: This is not clear enough. There should be an equation here for Q as a function of U and C - initially it is reported that Q is simply the product but later that C is converted to a mass (density?) to derive an emission rate. What are the units of Q (flux of what? grams or moles?). If the emission rate units are in moles, then this is not required (as in Cambaliza et al).

- Equation added, units added, and it is noted that there is a conversion factor between ppm and moles.

The derivation of the background is also not very clear here - why it must be split into a right and left half? Might be more clear to describe x as the coordinate from the beginning of a transect to the end, and $x_{max}/2$ is the midpoint for each transect upwind? Is there no x -dependence of CBL and CBR? (L235 indicates they are only functions of z ?). An example would be nice here.

- It is split into a right and a left half due to gradients across the field. Since CBL and CBR are average values, they do not have an x dependency

– Just looking now at the list of definitions (thank you, this clarifies things!), and it becomes clear that when the authors refer to concentration they are actually referring to a mole fraction, i.e. micromoles of CH_4 per mole of dry air (this should be defined), or ppm. Concentration is usually (if molar concentration) in units of moles per meter cubed (in SI units - the authors use mol per cm cubed), which could make it a "molar mass anomaly" for the authors ($N^?$). These should be re-defined correctly in the future draft for section 2.4 - call C a mole fraction and N a concentration.

- An equation has been added, and text now notes that C is in ppm and that there is a conversion factor to moles per volume.

- When there is a gradient as there almost always is in nature, it is unclear as to which part of the background is transported to the measurement plane, introducing uncertainty. We address this by derive background from the downwind dataplane.

This section is unclear with equation (1) not clear to me why the integral of a Gaussian distribution would be zero. Not clear how C' is related to Φ_P . A reader has to work way too hard to make sense of this method. Also, from Figure S6, it seems that Φ is a distribution for each transect, but in the equation C is a function of z . How is the vertical interpolation done, and how is C' defined?

- Equations have been added and the section has been clarified. This is a distribution for each left or right transect at each altitude (hence the z dependency). Interpolation, prior to integration, is linear, now stated

Equation 2 does make sense, although x1 and x2 should be defined.

- Equation 2 was rewritten to not use x1 and x2

Figure S5: What is this figure telling us? What are the colors in the tracks (yellow/green?). Could an elevation map suffice here?

- Data key added. The purpose of this figure is to show the typical surface obstacles to surface winds along the profile, elevation is not particularly relevant.

Figure S6: Is there a transect upwind at 2200 m (as in (c)?), but there is no dashed line in (a) corresponding to this one? The data should be shown as well as the interpolated curtain, to show if there is spatial structure in x and z of this background that is being smoothed?

- Since 2200 m is background, there is no need to separate the background concentration from a plume and it is used without smoothing or analysis in the linear interpolation.

Supplement S5: L168 should read section 2.4 in text. L167. not clear. So the peak of the distribution (is this the mode?, i.e. the value most commonly seen in this upwind transect?), is used as the value for the entire transect, and then the background was interpolated vertically - how?

- There was a mistake in figure S6 caption – the probability distributions are all for CH₄, half for the right side of the data field and half for the left side of the data field. This has been corrected. Additionally, the methodology of filling in the background data plane is now also described in the supplemental material.

Is x in Fig S6 going from west to east?

- The caption for Fig S6 has been improved to include the definition of x and z, as in the main text.

L253, I agree, but wouldn't call this "appropriateness" - more specific. Maybe appropriateness of the measurements for the assumptions made for the analysis? What assumptions are being made that need to be satisfied? I would call it representativity or just say that spatial and temporal variability are the dominant sources of uncertainty. L276: to 1800 masl (from what base elevation?)

- There is a good suggestion. Changed to representative and to spatial and temporal variability. Also changed to 1800 masl.

L277: at 2258 is this part of the profile? Isn't this higher then?

- It was a typo – corrected to 2058 m here and elsewhere. This was above the airplane profile. Data were collected in the open to compare with the direction and speed of winds near the top of the overlapping profiles where AMOG was surrounded by tall trees, and showed good agreement This clarification now is added.

Fig 4 (indicate masl rather than just meters for clarity)

- Done.

L314, this is nice to note, but should also be included in the supplementary section on measurements, as well as to what precision they agreed (within X ppb agreement on average, or something quantitative).

- Moved to section 2.2 where calibration also is mentioned for the GHG analyzer.

L325, don't remnant structures from the prior day make the mass balance or emissions estimate not correct, according to earlier text about flushing out prior days' emissions? (from reading on, we see this is the "upwind" profile, so this should be mentioned here or somewhere nearby).

- The upwind profile is to characterize the atmospheric structure, not to provide input to the mass balance, and air from the profile will pass to the east of the oil fields. That this is the upwind profile is now mentioned.

L329 (alpha - alpha' should be used here for clarity for the reader).

- This would require doing so for all greek letters in all captions or the entire text, which seems excessive. We will discuss with the copy editor.

Fig 4: For the upwind profile, alpha-alpha', the CH₄ is lower in the PBL than above. However, in Fig S6, the upwind "curtain" or plane, is showing higher CH₄ at lower altitudes. How are these two figures consistent?

- The upwind profile is at alpha-alpha', Supp. Fig. S6 shows the background curtain (not upwind) curtain, correctly labeled in the caption and section heading, and is at delta-delta' and is derived from are data outside the plume. The methodology is described in Section 2.4, now corrected in the supplemental to refer to the right section.

In Figure S6, which of the transects are AJAX, and which AMOG?

- AMOG is on the surface, now labeled.

Fig 5 why only is the north wind shown? These are very low wind speeds indeed, esp. for doing an emissions estimate.

- As noted above, the profile is not for directly estimating emissions, but to characterize the atmospheric profile structure, primarily the location of the PBL. The figure was very busy already, so only the north component (ascent and descent) are shown. The purpose of showing the winds is to note they are not (in this case, useful for identifying the PBL, and that they show a change between the earlier and later profiles. The east component conveys similar (lack of useful) information.

- Text now notes "Winds were not useful for deriving the location of the PBL."

L345 yes at 4 m/s - is that the wind speed? It's not shown. Was that the wind speed for 5 hours?

- Yes – text clarified.

L347 how is growing from 100 to 1675m a stable PBL? Also, is stable referring to the atmospheric stability class (i.e turbulence) or the fact that the PBL depth is not changing much in time?

- For clarity, text changed to "~100 m growth to ~1675 m,"

L368 Westerlies?

- Changed.

Figure 8: what is the time difference between when these transects were measured, as well as when the transects for the background (shown in Figure S6) sampled? Was the background plane subtracted point by point, i.e. in x, z space so that a higher background was subtracted on the east side, (L386)? Still don't understand where the distributions Phi come in to this picture.

Fig. 8 how was the vertical interpolation done, and the extrapolation above the highest flight transect at ~1100 masl? It seems like a different method was used for U_n and CH₄, noting where they drop off in the vertical. Figure 4 indicates that AMOG was driving the surface transect much earlier than the AJAX transects (or perhaps I misinterpreted this), so how can we combine them when we know the PBL is growing?

L400: Extrapolating these emissions to an annual average is a stretch and not at all defensible. This is one of the reasons that recent similar studies that are performed over a short time frame report their emission rates in moles per second or kg per hour or such. The section on the uncertainty estimate is short and not thorough - the distributions that go into the Monte Carlo would need to be explained better.

- Agreed. Results are now reported in Mol s⁻¹ (with the Gg yr⁻¹ reported in parentheses for comparison to inventory). We have also expanded the Monte Carlo approach section

L431: Could you look at a slope of the CH₄ to CO₂ tracer plot in the plume to show this consistency with the reported ratio?

- Its really the anomaly, and we are comparing the emissions, which accounts for this in a way that a scatterplot would not.

L496 I would say that these complexities also challenge this method because of the variability that you are not measuring - and the model you are using assuming some constancy in wind.

Overall, this method does not fully account or try to discount the possibility of un-steadiness in the winds between upwind and downwind transects that could lead to accumulation of emissions during slower wind speed periods (night time but also could just be earlier in the day). Perhaps this is dealt with in the uncertainty calculation but that is not clear in the text as written here.

L499-501, Please indicate some quantification of the differences here. This is a methods paper - how well did the concentrations (mole fractions) agree (above the PBL), in ppb? Were any calibration tanks measured on both systems?

- Calibration tanks for the two platforms were different, but from the same vendor. Agreement was 99.7% for CH₄ and 99.9% for CO₂. Comparison of the median winds showed is 38% for U_{east} is 27%. A figure of the altitude variations is shown as Supp. Fig. S7.

L502-505, yes it is true that we could not expect the winds to agree, but what does this indicate for the interpolation of wind in the vertical from the different platforms? Is that variability captured at all?

- As noted in the text, this disagreement above the PBL is due to AJAX data being influenced by being collected in the lee of a mountain peak. This paragraph is only for winds above the PBL, The next paragraph is for winds in the PBL, where there was good agreement.

L523+ What about plumes of CH₄ that are following these complex winds and topography? the simple interpolation and treatment of the surface data is troublesome under these conditions.

A mass balance equation is a conservation of mass and the equations (although not written out here) assume some steady uniform wind condition. This is clearly violated here. Perhaps the uncertainty calculation deals with this problem but it is not clear.

- The upwind profile and similarly the downwind profile are not for the purpose of characterizing the upwind or background concentrations, but solely for the purpose of identifying the PBL. We do compare extensively a range of parameters between AJAX and AMOG to show that surface profiles using topography can be used to characterize atmospheric structure. This is a second very important purpose of the manuscript, because not everyone has an airplane or access to an airplane, particularly in the developing world. We fully agree it is nice to have an airplane and I wish I had one of my own. Additionally, the upwind and downwind profiles are “inconveniently” located with respect to the study area (which is approximately 50 km to the south of the upwind profile. In the case of the flux calculation, stranded layers are not evident in the airborne profile data. This could be because they are more likely to occur in the mountains used for the surface profile, which is speculative, and thus is not mentioned in the manuscript. Given the typical air flow patterns, such features will typically not return to the SJV center when formed on the east mountains, but would for formation on the west mountains.

L541: Are these factors not accounted for in the inventory? What about temporal variability? Also, what about the uncertainties on both numbers, assuming they are 1- sigma (which should also be noted incidentally)? Seems to me the emissions estimate actually overlaps with the inventory quite well given the uncertainties that are reported.

- Text adjusted to note that the derived flux lies within the inventory uncertainty, but is higher, and then continues to note this is consistent with the metastudy of Brandt et al., that inventories likely are too low.

L552. In my opinion, this should be toned down - this one measurement supports the conclusion is that the global loss rate of CH₄ to OH (or soils) is underestimated? What percentage of the global methane budget does 25 or 32 Gg/year actually represent?

- Agreed, rewritten – see response above.

Conclusion

L559. This statement implies that the uncertainty has been reduced from other methods, which is not the case, and has not been shown.

- Removed

L562 But this method relies on the aircraft measurements as well as the surface, so could not be applied in the absence of those resources!

- Rewritten.

L564 - The flux quantification is "direct", meaning measured winds and concentrations were used, but that is the flux through a point in space and time ($Q(x,z)$) - the rest is a simple model: you must integrate that flux based on an interpolation (in space and time), and must subtract a background that has its own model and interpolation, and the attribute that flux to a surface emission which requires some Eulerian conditions - steady flow through a control volume. All are a "model" - just a simple one.

However this point can be made differently - that one should measure before adding one more assumption to the model, which is that of a vertically well-mixed plume. Other studies have moved away from this assumption of vertical well-mixedness as well: Cambaliza et al., Heimbürger et al. (Elementa 2017), Lavoie et al. (ES&T 2015 and 2017), Conley et al (both 2016 as well as 2017: <http://www.atmos-meas-tech-discuss.net/amt-2017-55/>), and numerous others, especially when sampling in the near field. I agree that this is a valid point to make using these observations.

- Good points, and an additional sentence has been added to highlight.

Supplement:

L26 cfm should be given in metric

- Yes, done.

What are some estimated uncertainties on the FGGA CH₄ measurements based on the calibration standard - how often is it sampled, is there noise/drift, etc? A sentence or two on this is warranted beyond just the statement that a calibration was performed. Was there a water correction, or were the dry values reported by the FGGA used?

- Added

The additional accuracy of the 450C sentence should go where it is first discussed, before the sentence about the FGGA. Earlier it says it achieves 1ppb accuracy, but now it says that it can achieve 50ppt if calibrated with hourly zero gas measurements - which number applies here? Where do the authors get the accuracies reported for the other analyzers (ozone, etc)? Manufacturer?

- This has been clarified. The accuracies are from the manufacturer and include 24 hour drift.

If the main paper is not about these auxiliary gases, this information should not really be mentioned and could be removed.

- We prefer to describe the system completely, and to include (with better explanation) the improvement in accuracy of the 450i by hourly zero measurement, as this could be of interest to other researchers

Interestingly, no accuracy or uncertainty is reported for CO₂ or CH₄, the main gases of interest in this work (for the AMOG measurements).

- Added

S2.2: Is there a reference for the MMS wind system? There is no information given here, and this is a key measurement for flux studies. Uncertainty on winds should be reported for both platforms.

- The MMS is a NASA developed system that has not been published. We provide a link to the homepage, and report its accuracies. Additionally, information on AMOG winds is now included, and explained as it depends on velocity of AMOG and the velocity of the winds.

1 **Atmospheric Characterization through Fused Mobile Airborne**
2 **& Surface *In Situ* Surveys: Methane Emissions Quantification**
3 **from a Producing Oil Field**

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5 Laura Iraci³, Josette Marrero³, Ju-Mee Ryoo³, Tomoaki Tanaka³, Emma Yates³

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9
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11
12 **Abstract.** Methane (CH₄) inventory uncertainties are large, requiring robust emission derivation
13 approaches. We report on a fused airborne/surface data collection approach to derive emissions from an
14 active oil field near Bakersfield, central California. The approach characterizes the atmosphere from the
15 surface to above the ~~Planetary Boundary Layer~~ (PBL) and combines downwind trace gas concentration
16 anomaly (plume) above background with normal winds to derive flux. This approach does not require a
17 well-mixed PBL, allows explicit, data based, uncertainty evaluation, and was applied to complex
18 topography and wind flows.

19
20 *In situ* airborne (collected by AJAX – the Alpha Jet Atmospheric eXperiment) and mobile surface
21 (collected by AMOG – the AutoMOBILE trace Gas – Surveyor) data were collected on 19 August 2015 to
22 assess source strength. Data included an AMOG and AJAX intercomparison transect profiling from the
23 San Joaquin Valley (SJV) floor into the Sierra Nevada Mountains (0.1-2.2 km altitude), validating a novel
24 surface approach for atmospheric profiling by leveraging topography. The profile intercomparison found
25 good agreement in multiple parameters for the overlapping altitude range from 500 to 1500 m, for the
26 upper 5% of surface winds, which accounts for wind-impeding structures, i.e., terrain, trees, buildings,
27 etc. Annualized emissions from the active oil fields were 31.3±16 Gg methane and 2.4±1.2 Tg carbon
28 dioxide. Data showed the PBL was not well-mixed at distances of 10-20 km downwind, highlighting the
29 importance of the experimental design.
30

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35 1. Introduction

36 1.1. Methane Trends and Uncertainty

37 On decadal timescales, methane (CH₄), affects the atmospheric radiative balance more strongly than
38 carbon dioxide (CO₂), (IPCC, 2007, Fig. 2.21). Since pre-industrial times, CH₄ emissions have risen by a
39 factor of 2.5 (Dlugokencky et al., 2011; Khalil and Rasmussen, 1995), while estimates of its lifetime has
40 decreased and now is estimated at ~8.5 years (Sonnemann and Grygalashvyly, 2014). Atmospheric CH₄
41 growth almost ceased between 1999 and 2006, but has resumed since 2007 (Nisbet et al., 2014;
42 Schwietzke et al., 2016). Several processes are proposed to underlie this trend (Ghosh et al., 2015; John et
43 al., 2012) with recent isotopic shifts suggesting wetlands are the dominant driver (Nisbet et al., 2016);
44 however, high uncertainty in emission inventories (IPCC, 2013) complicates interpretation of the
45 underlying mechanism(s).

46
47 The dominant CH₄ loss arises from reaction with hydroxyl (OH), whose concentration has been
48 increasing in recent decades (John et al., 2012), causing a decrease in the estimated CH₄ lifetime of 0.5%
49 yr⁻¹ (Karlsdóttir and Isaksen, 2000). Overall, the estimate of the CH₄ lifetime has decreased by ~40% from
50 an estimated 12 years in 2007 (IPCC, 2007). Rigby et al. (2017) suggest a decline in OH is likely (66%)
51 to have contributed to increasing CH₄ since 2007. The recent discovery of a new significant CH₄ loss
52 mechanism, terrestrial uptake (Fernandez-Cortes et al., 2015), illustrates the need to understand loss
53 mechanisms better (Allen, 2016).

54
55 Large CH₄ budget uncertainties remain for many sources (IPCC, 2013), with greater uncertainty in future
56 trends from global warming feedback (Rigby et al., 2008) and increasing anthropogenic activities
57 (Kirschke et al., 2013; Saunio et al., 2016; Wunch et al., 2009). Emphasizing these uncertainties are
58 recent studies that suggest underestimation by a factor of 1.5 in the important anthropogenic CH₄ source,
59 Fossil Fuel Industrial (FFI) emissions (Brandt et al., 2014). Tellingly, this discrepancy only was noted
60 recently (Miller et al., 2013), in part because the US CH₄ monitoring network is too sparse to constrain
61 emissions at “regional to national scales” (Dlugokencky et al., 2013), with isotopic data indicating even
62 larger underestimation by a factor of 1.6-2.1 (Schwietzke et al., 2016). FFI emissions are the largest
63 (Brandt et al., 2014; EPA, 2017) or second largest after agriculture (Saunio et al., 2016) anthropogenic
64 contributor to the global CH₄ budget. These uncertainties strongly argue for the need for new, robust
65 methodologies for flux derivation.

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76 **1.2. Methane Flux Estimation**

77 Various approaches have been developed to derive surface emissions from CH₄ concentration
78 measurements including direct flux assessment – i.e., measurement of winds and concentrations through a
79 plane, and/or by the comparison of upwind and downwind mass budgets (Peischl et al., 2016; Peischl et
80 al., 2015; White et al., 1976), data-driven mass balance, e.g., Karion et al. (2013), tracer-tracer ratio
81 (LaFranchi et al., 2013), and assimilation inverse models, e.g. Jeong et al. (2013); Jeong et al. (2012);
82 (Saunio et al., 2017). Challenges for the latter approach include the needs for accurate meteorological
83 transport models and good *a priori* emission distributions (Miller et al., 2013; Peischl et al., 2016; Smith
84 et al., 2015). Miller et al. (2013) concluded that bottom-up inventories (EPA, 2013; European
85 Commission, 2010) significantly underestimate husbandry and FFI emissions. To apportion CH₄ to FFI
86 versus biological sources, the tracer-tracer approach has been applied using ethane, whose emission ratio
87 to CH₄ requires tight constraint (Peischl et al., 2013; Simpson et al., 2012; Wennberg et al., 2012). In
88 practice, this emission ratio is an *a priori* assumption in the assessment.

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90 Direct assessment approaches have advantages over inversion approaches. Direct approaches allow
91 explicit uncertainty evaluation and do not require an *a priori* emission spatial distribution, which may be
92 unknown. Direct approaches also do not require the ability to model atmospheric transport accurately
93 across the study region. In areas of complex topography or highly variable winds, this transport can
94 challenge assimilation approaches, which also are challenged in areas with poorly characterized (or
95 unknown) or highly variable sources, particularly if the measurement network is sparse. For direct
96 assessment approaches, data collection should be rapid if winds and/or emissions are variable, and at
97 adequate data density to characterize fine-scale structure.

98 **1.3. Study Motivation**

99 Herein we report on a novel application of fused airborne and surface *in situ* data to directly estimate CH₄
100 emissions using an anomaly approach rather than a more typical mass balance approach due to a lateral
101 gradient in the upwind data. A direct approach does not require accurate winds over the study domain,
102 only in the measurement plane. The approach was applied to 1164 km of airborne data collected on 19
103 August 2015, by NASA's Alpha Jet Atmospheric eXperiment (AJAX), while AMOG (AutoMOBILE
104 greenhouse Gas) Surveyor collected 1074 km of contemporaneous mobile surface data. Both platforms
105 measure carbon dioxide (CO₂), CH₄, water vapor (H₂O), and ozone (O₃), as well as winds, pressure,
106 relative humidity (RH), and temperature (T). The surface and airborne datasets were collected in a

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112 downwind curtain or plane oriented approximately orthogonal to the winds, to characterize the full
113 planetary boundary layer (PBL) from surface to above the PBL.

114
115 Additionally, the survey route was designed to include an ascent to ~2.2 km above sea level to include
116 surface PBL characterization. Data fusion between measurement platforms was validated by a vertical
117 profile intercomparison for 0.5 to 1.5 km altitude by AMOG Surveyor, leveraging topographic relief.

118 **1.4 The South San Joaquin Valley, California**

119 Most of California oil production lies in the San Joaquin Valley (SJV), as does most of California
120 agriculture, including many intensive dairies (Gentner et al., 2014), and the major north-south
121 transportation artery. For this study, data were collected for the Kern River oil fields (Kern Front oil field,
122 Kern River oil field and the Poso Creek oil field, referred to herein as the Kern Fields), located adjacent to
123 northwest Bakersfield (**Fig. 1A**). These adjacent oil fields create a strong CH₄ source that largely is
124 isolated from confounding plumes from other SJV sources. This area includes complex wind flow
125 patterns across and around the “toe” of Sierra Nevada Mountain foothills, which extend into the Kern
126 Front and Kern River oil fields. Here, topographic steering ensures predictable prevailing northwesterly
127 winds blow across the Kern Fields.

128
129 Strong orographic forcing also arises from tall bluffs (~100 m) on the Kern River Valley’s south bank,
130 which also separates the Kern River oil field from the urban city of Bakersfield (pop. 364,000 in 2013).
131 The fine-scale wind structure that results from orographic forcing on transport dictated an anomaly
132 approach for flux derivation, as did the presence of strong CH₄ structures (plumes) in the valley’s lowest
133 air. In the anomaly approach, transects must extend beyond a reasonably well-defined plume.

134
135 Topography (i.e., mountain ranges) plays a locally dominant role in overall southern California air flows
136 where upper level winds locally force the lower level flows that transport pollutants (Bao et al., 2008).
137 The SJV is delimited on the east by the Sierra Nevada Mountains and on the west by the Transverse
138 Coastal Mountain Range (**Fig. 1A**). Transport between the SJV and adjacent air basins is poor due to
139 California’s mountain ranges. The SJV features weak surface winds (Bao et al., 2008) with the worst air
140 quality in the United States occurring in the cities of Bakersfield and Delano (American Lung
141 Association, 2016) in the SJV.

142
143 Pacific Ocean air primarily enters the SJV through the San Francisco Bay area and the Carquinez Strait,
144 where it splits north into the Sacramento Valley and south into the SJV (Zhong et al., 2004). This flow

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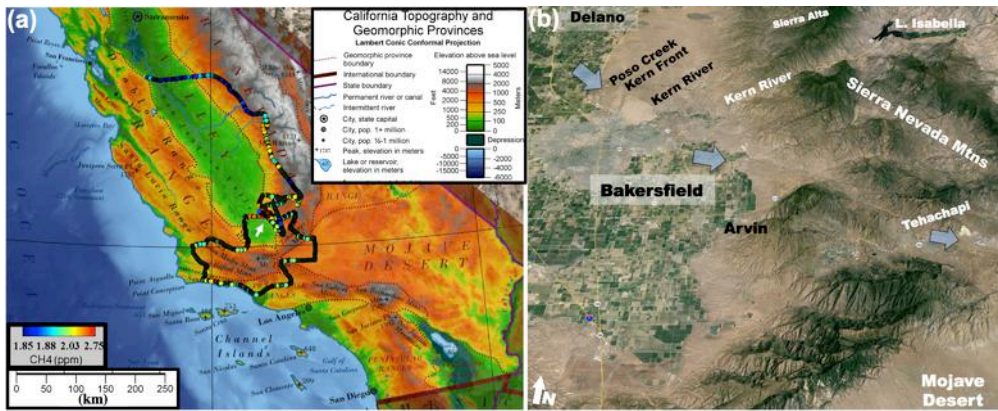
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156 | extends up to ~1 km altitude (Zhong et al., 2004). These winds are near orthogonal to the 600-km long
157 | central valley of California - i.e., cross-slope. South of Bakersfield, winds shift to from the west due to
158 | mountains that guide SJV air out into the Mojave Desert, where it affects air quality for up to hundreds of
159 | kilometers distance (VanCuren, 2015). Although the Tehachapi Pass is the main exit pathway of SJV air,
160 | other passes also transport air into the Mojave Desert. These flows are augmented by high inland
161 | temperatures relative to the Pacific Ocean, which creates a horizontal pressure gradient that drives local
162 | upslope flows during the day and returning downslope nocturnal flows (Zhong et al., 2004). The pressure
163 | gradient is maximal around sunset, although winds peak ~4 hours later, shortly before midnight. This
164 | pressure gradient is controlled by the semi-permanent Pacific high, situated offshore central California,
165 | which diverts storms far to the north during summer. This pressure feature drives prevailing west-
166 | southwesterly winds at the regional scale in the California south coast air basins (Boucouvala and
167 | Bornstein, 2003).

168 | **2. Methodology**

169 | **2.1. Experimental design**

170 | Data were collected as part of the *GOSAT-COMEX Experiment* (Greenhouse gases Observing SATellite -
171 | CO₂ and Methane Experiment - GCE) Campaign. GCE was developed to characterize emissions on
172 | spatial scales from decameter (*in situ* surface, imaging spectroscopy) to kilometer (*in situ* airborne) to
173 | deca-kilometer (satellite) in an area of complex topography. GCE design combined *in situ* mobile surface
174 | and airborne data with GOSAT satellite data. *In situ* data serve to assess the satellite pixel / plume
175 | overlap. Key GCE requirements are relatively steady, strong, isolated emissions and predictable and
176 | steady winds. Prevailing study area winds are from the west-northwest, veering to westerly winds to the
177 | southeast of Bakersfield (**Fig. 1**). Prevailing wind directions are highly reliable due to topographic
178 | control.



179

180 **Figure 1. (a)** Full surface and airborne data for 19 Aug. 2015 mapped over California topography. White
 181 arrow shows Bakersfield. Data key on panel. **(b)** Study area map showing direction of daytime prevailing
 182 winds and nearby mountain topography (Google Earth, 2016). See Supp. Fig. S1 for a high-altitude (20-
 183 km) photo of the entire study area and surrounding terrain.

184 GCE developed from the COMEX Campaign (Krautwurst et al., 2016), which combined *in situ* airborne
 185 and surface observations with both imaging and non-imaging spectroscopy to explore synergies for GHG
 186 emission estimation (Thompson et al., 2015). COMEX focused on southern California CH₄ sources
 187 including husbandry, landfills, natural geology, and petroleum hydrocarbon refining and production.

188

189 GCE combines airborne and surface data collected at dramatically different speeds. AJAX collects data at
 190 ~500 km hr⁻¹, capturing a snapshot of atmospheric winds and plume structure. Surface GCE data are
 191 collected quasi-Lagrangian, starting northwest (upwind) and proceeding southeast and then east
 192 (downwind). This enables useful data collection even when a CH₄ plume drifts into the study area after
 193 the upwind survey – data collection proceeds downwind faster than advection. The surface route was
 194 designed carefully to traverse all targeted GOSAT pixels using rarely used (low traffic) surface roads and
 195 requires ~100 minutes.

196

197 Airborne and surface surveys are timed so that the downwind data plane (Kriings et al., 2011) is surveyed
 198 concurrent with the satellite overpass. Data planes extend from the surface (AMOG) to above the PBL
 199 (AJAX), reducing uncertainty by providing a more complete atmospheric characterization including
 200 below where airplanes are permitted to fly (~500 m in an urban area). AJAX and AMOG profile data are
 201 fused by an interpolation approach that imposes the observed vertical structure and the flux through the
 202 data curtain is calculated (Sect. 2.5).

6

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210 GCE first incorporates an AMOG Surveyor upwind transit from Delano (100 m) on the SJV floor to
 211 Sierra Alta (1800 m) and higher to confirm that upwind CH₄ plumes do not threaten to impact the study
 212 area during the experiment, otherwise the survey is aborted. A key mission abort criterion is wind
 213 compliance. Specifically, winds must not be too light (typically less than ~2 m s⁻¹) or variable (>30°),
 214 must flush nocturnal accumulations before the GOSAT overpass (i.e., no CH₄ cloud at or nearby upwind
 215 of the site, which means that winds could not have been light as recently as several hours prior; however,
 216 winds are not measured several hours prior), and must be prevailing. The upwind transit provides vertical
 217 profile information including PBL height and vertical structure. AJAX repeats this upwind transect to
 218 compare wind profiles with AMOG; however, discrepancies in the transects arise from the road following
 219 terrain, and the airplane needing to avoid peaks along the ridge.

220 **2.2. AutoMOBILE trace Gas (AMOG) Surveyor**

221 Mobile atmospheric surface measurements have been conducted for many years using a customized van
 222 (Lamb et al., 1995) or a recreational vehicle (Farrell et al., 2013; Leifer et al., 2013). Recently, the
 223 development of cavity enhanced absorption spectroscopy (CEAS) analyzers has opened the way for rapid
 224 and highly accurate trace gas measurements (Leen et al., 2013) without the need for onboard compressed
 225 gases as in gas chromatography (Farrell et al., 2013), although periodic calibration with gas standards is
 226 important, albeit typically not onboard the platform. This allows for smaller vehicle survey platforms at
 227 lower logistical overhead (Leifer et al., 2014; McKain et al., 2015; Pétron et al., 2012; Yacovitch et al.,
 228 2015). A competing sensor technology that has been used in mobile survey data collection is open path
 229 spectroscopy (Sun et al., 2014). Mobile survey platforms can incorporate older technology such as
 230 fluorescence to, for example, measure ozone, O₃.

231
 232 Mobile surface data were collected by the AMOG Surveyor (Leifer et al., 2014) (see Supp. Sect. S2.1 for
 233 additional details), a modified commuter car. AMOG Surveyor provides mobile high-speed, high-spatial
 234 resolution observations of meteorology (winds, temperature, pressure), trace gases (greenhouse and
 235 others), and remote sensing parameters. AMOG uses a range of trace gas analyzers and careful design
 236 with respect to wind flow around the vehicle to characterize strong spatial heterogeneity at up to highway
 237 speeds.

238
 239 Two-dimensional winds are measured by a sonic anemometer (VMT700, Vaisala) mounted 1.4 m above
 240 the roof, which is at 1.6 m height, above vehicle flow streamlines for slow to highway speeds. Estimated

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252 accuracy is approximately 10° and 0.3 m s^{-1} for wind speeds above 1.5 m s^{-1} (see supplement for further
253 details).

254
255 A high-flow vacuum pump (GVB30, Edwards Vacuum) draws air down a sample lines from 5 and 3 m
256 above ground for GHG and ozone (O_3) analyzers. The 5-m sample line height references low speed /
257 stopped ($< \text{a few m s}^{-1}$) AMOG sample collection. At high speed ($> 10 \text{ m s}^{-1}$) the sample tube flexes
258 backwards to 3 m height to avoid destructively hitting obstacles at high speed. This protects the sample
259 line from hitting bridges, tree branches, etc. Greenhouse gases, CO_2 , CH_4 , and H_2O , are measured at up to
260 10 Hz by an Integrated Cavity Offaxis Spectrometer-Cavity Enhanced Absorption Spectroscopy analyzer,
261 with a 1 s accuracy of 1 ppb for CH_4 (ICOS-CEAS, 911-0010, Los Gatos Research, Inc.). Calibration is
262 with a Scott-Marin CH_4 and CO_2 atmospheric standard. A fluorescence analyzer measured O_3 at 0.25 Hz
263 (49C, ThermoFischer Scientific, MA). This difference does not arise from calibration differences: the
264 AMOG Surveyor O_3 analyzer was cross calibrated with the AJAX calibration source to 1 ppb accuracy.
265 AMOG Surveyor's full trace gas suite (carbonyl sulfide, carbon monoxide, nitric oxide, nitrogen dioxide,
266 hydrogen sulfide, sulfur dioxide, total sulfur, ammonia) was not deployed on 19 Aug. 2015.



267
268 **Figure 2.** Study platforms. (a) AutoMOBILE trace Gas (AMOG) Surveyor, Kern River oil field in
269 background. Photo courtesy Ira Leifer. (b) The Alpha Jet Atmospheric eXperiment (AJAX) aircraft, photo
270 courtesy Akihiko Kuze, JAXA. See Supplemental Material Section 1 for further details.

271
272 The greenhouse gas analyzer is calibrated using a Scotty's whole-air standard before and/or after each
273 data collection with the calibration factor closest to the day of flight being applied to each raw CO_2 and
274 CH_4 measurement. Calibration factors have been shown to agree within less than 1 ppb. The calibration
275 factor includes a linear correction for cell pressure, which can drop at higher altitudes. This pressure
276 calibration has been shown to be linear from 140 mtorr down to 28 mtorr.
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300 Relevant recent AMOG Surveyor improvements since Leifer et al. (2014) include a high speed
 301 thermocouple (50416-T, Cooper-Atkins) and a high accuracy (0.2 hPa) pressure sensor (61320V RM
 302 Young Co.). Both are mounted in a roof passive radiation shield (7710, Davis Instruments) to largely
 303 eliminate dynamic pressure effects from the airflow. Position information is critical to accurate wind
 304 measurements and is provided by redundant (two) Global Navigation Satellite Systems (19X HVS,
 305 Garmin) that use the GLONASS, GPS, Galileo, and QZSS satellites at 10 Hz (WGS84). AMOG
 306 analyzers and sensor data are logged asynchronously on a single computer. Custom software integrates
 307 the data streams and provides real-time visualization of multiple parameters in the Google Earth
 308 environment.

309 2.3. Alpha Jet Atmospheric eXperiment (AJAX)

310 AJAX (Fig. 2b) collected airborne *in situ* measurements of CO₂, CH₄, H₂O by cavity ring down
 311 spectroscopy (G2301-m, Picarro Inc.), O₃, (Model 205, 2B Technologies Inc.), and meteorological
 312 parameters including 3D winds by the Meteorological Measurement System
 313 (<https://earthscience.arc.nasa.gov/mms>), a NASA developed system with accuracy of ±1 m s⁻¹. The
 314 greenhouse gas analyzer is calibrated using NOAA whole-air standards; calibrations are performed before
 315 and/or after each flight with the calibration factor closest to the day of flight being applied to each raw
 316 CO₂ and CH₄ measurement. Further corrections include applying water vapor corrections provided by
 317 Chen et al. (2010) to calculate CO₂ and CH₄ dry mixing ratios. Data also are filtered for quality control for
 318 deviations in instrument cavity pressure, to improve inflight precision.

319
 320 Overall CH₄ measurement uncertainty is typically <2.2 ppb, including contributions from accuracy of the
 321 standard, precision (1-σ over 6 min), calibration repeatability, inflight variance due to cavity pressure
 322 fluctuations, and uncertainty due to water corrections and pressure dependence (based on environmental
 323 chamber studies). See Hamill et al. (2015); Tanaka et al. (2016), and Yates et al. (2013) for further
 324 aircraft and instrumentation details, and Supp. Sect. S2.2.

325 2.4. Background estimation and data fusion

326 The flux ($Q(x, z)$) in moles s⁻¹ m² with respect to lateral transect distance (x) and altitude (z) through the
 327 x, z plane is the product of the normal winds ($U_N(x, z)$) in m s⁻¹ and the plume concentration anomaly
 328 ($C'(x, z)$) or mole fraction in ppm (Leifer et al., 2016).

329 $Q(x, z) = k(z) U_N(x, z) C'(x, z) = k(z) U_N(x, z) (C(x, z) - C_B(x, z))$ (1)

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338 $k(z)$ converts from ppm to moles. Interpolation of C' and U_y is linear within the PBL and is assumed
339 uniform above the PBL. To calculate $Q(x, z)$ requires C' relative to background ($C_B(x, z)$). Initially surface
340 data that was collected for an upwind surface transect was used to derive C_B , using the assumption of
341 vertical uniformity for "background."

342
343 Unfortunately, the upwind data showed a lateral gradient, which coupled with uncertainty in precisely
344 where the downwind air originated (given the topography, which features a gentle incline towards the
345 northeast, this gradient is unsurprising, in retrospect). Thus a very small shift in the winds between the
346 upwind and downwind curtains results in a significant shift in C_B , with a very large effect on Q . As a
347 result, the more traditional upwind/downwind mass balance approach was abandoned for an anomaly
348 approach.

349
350 In the anomaly approach, $C_B(x, z)$ was derived from evaluating $C_B(x < x_{max}/2, z)$ and $C_B(x > x_{max}/2, z)$,
351 denoted $C_{BL}(z)$ and $C_{BR}(z)$, respectively, where x_{max} is the lateral extent of the data curtain. Then, $C_B(x, z)$ is
352 derived from a first order linear polynomial fit of $C_{BL}(z)$ and $C_{BR}(z)$.

353
354 Both $C_{BL}(z)$ and $C_{BR}(z)$ are derived from the amplitude of a Gaussian fit to the left and right probability
355 density functions ($\Phi_L(C(x < x_{max}/2, z))$ and ($\Phi_R(C(x > x_{max}/2, z))$), respectively, for each flight transect level.
356 Specifically, for Φ_L and Φ_R , Gaussian functions are fit to model the plume distribution (Φ_P) and the
357 background distribution (Φ_B). In these data, Φ_B is well-fit by a single Gaussian, while Φ_P is best described
358 by multiple Gaussian functions. Then, $C_{BL}(z)$ and $C_{BR}(z)$ are defined such that,

359 $\int \Phi_{BL}(C_{BL}(z)) = 0$ and $\int \Phi_{BR}(C_{BR}(z)) = 0$.

360 where Φ_{BL} and Φ_{BR} are the background Φ_B for the left and right halves of the data plane, respectively.
361 Concentration is not a conserved value, thus C' is converted into mass (N') by the ideal gas law (k in Eq
362 1) for spatial integration to derive the total emissions (E), which is the integration of the flux through the
363 plane, Q ,

364
$$E = \int_{x_1}^{x_2} \int_0^{z=PBL} Q(x, z) dz dx$$

365 Interpolation, prior to integration, is linear.

366 2.5. Uncertainty evaluation for emission calculation

367 A flux estimate requires two types of assumptions with respect to the flux calculation: representativeness
368 and representativeness. Specifically, background concentration profiles may be incorrect, while winds,
10

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378 which are measured accurately, could be un-representative, as could concentrations due to temporal
379 variability over the period needed to make the measurements. Monte Carlo simulations based on observed
380 data variability were run to assess uncertainty. Instrumental uncertainty is far less than spatial and
381 temporal variability and hence spatial and temporal variability is the dominant source of uncertainty
382 (Leifer et al., 2016).

383

384 Monte Carlo simulations were based on 1 standard deviation in the observed $U_N(z)$ around the mean for
385 each flight transect altitude level on the right and left sides, i.e., $U_{N,L}(z)$ and $U_{N,R}(z)$. Gaussian
386 distributions with half-widths of 1σ based on the values of $U_{N,L}(x,z)$ and $U_{N,R}(x,z)$ was formed for each
387 transect altitude. The distribution was randomly sampled to populate $U_N(x,z)$, and then interpolated as
388 described above. Other variables were Monte Carlo simulated in the same manner - i.e., a Gaussian
389 distribution was calculated for the left and right portions of the data based on 1 standard deviation in the
390 observations of the variable around its mean. Variables then were randomly sampled and interpolated.
391 Specifically, Monte Carlo simulations also addressed C_B , and C . Because instrumentation error is so much
392 less than spatial and temporal variability, Monte Carlo simulation of C_B represents uncertainty in the
393 source of the background (upwind) air, which could have some veering from the east or west coupled
394 with convergence in the horizontal plane. One million Monte Carlo simulations were run for a flux
395 uncertainty calculation.

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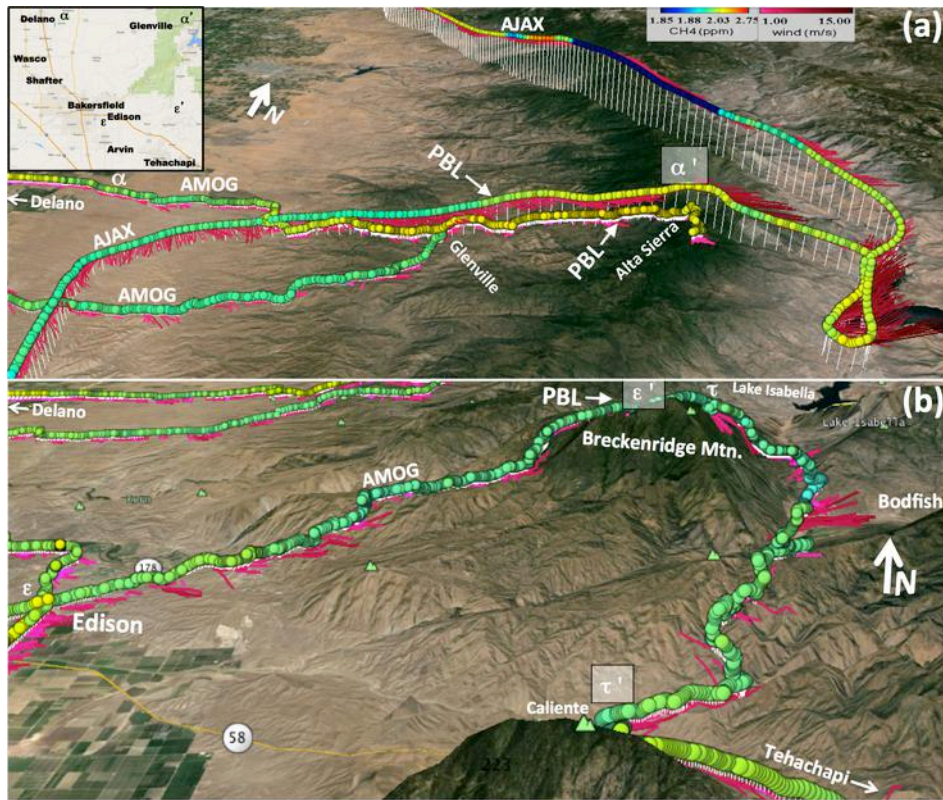
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425
 426 **Figure 3.** (a) Pre-survey, upwind AMOG surface and AJAX airborne methane (CH₄) and winds for
 427 vertical profile on the Delano – Alta Sierra transect (α-α'). Inset shows area map. (b) Post survey,
 428 downwind AMOG surface profile ascent Edison-Breckenridge (ε-ε') and descent Breckenridge-Bodfish-
 429 Caliente (τ-τ'). Upwind profile visible top left. Planetary boundary layer (PBL) identified.

430 3. Results

431 3.1. Profile data

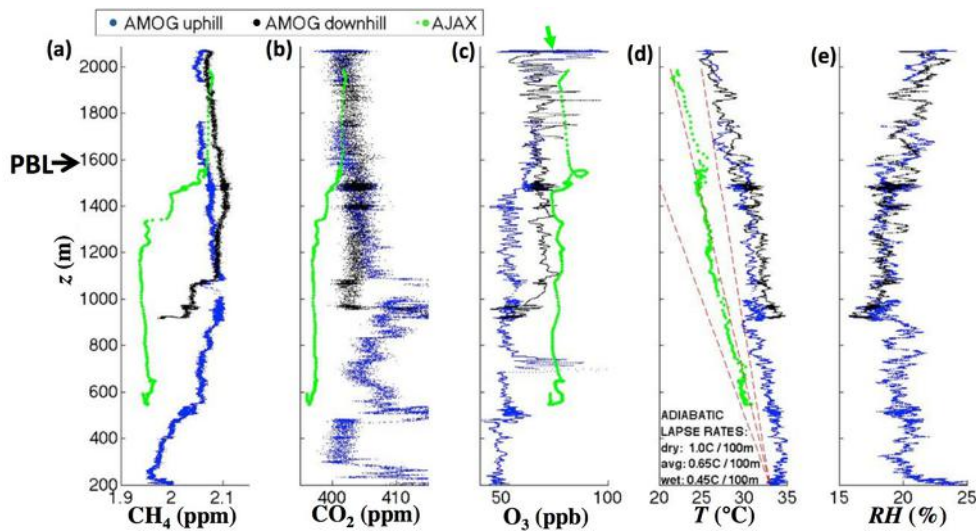
432 Four vertical profiles (surface and airborne) were collected to understand PBL evolution during the
 433 survey (2 hrs.) and across the survey domain spanning the experiment. Primary changes were
 434 development of near surface winds, and a slight increase in the PBL. AMOG and AJAX collected pre-
 435 survey intercomparison vertical profiles ~30 km north of the Kern Fields between the small town of
 436 Delano on the SJV floor (100 m) up to a meadow above Shirley Meadows (2058 m) on a ridge of the
 12

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438 Greenhorn Mountains in the Sierra Nevada Mountain Range (Fig. 3). This profile spans a wide range of
 439 topography, from grasslands on rolling hills, to tall pine trees near Alta Sierra, see Supp. Fig. S5 for
 440 surface images along the profile. AMOG also conducted a post-survey, downwind vertical atmospheric
 441 profile to 1800 m_{asl}. Approximately 15 minutes of data were collected in an open (200–300 m) field
 442 above Shirley Meadows (2058 m) that was fairly exposed with only thin stands of pine trees on terrain
 443 falling steeply off to both sides. The wind direction and speeds for Shirley Meadow was consistent with
 444 winds at Alta Sierra, several hundred meters below, where AMOG was surrounded by tall trees. Shirley
 445 Meadows was slightly above the top of the AJAX profile.

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446
 447 The AMOG vertical ascent was collected before the AJAX profile to enable concurrent AMOG/AJAX
 448 data collection for the Kern Fields. The AMOG ascent/descent was from 18:48 to 21:09 (20:08 UTZ at
 449 crest), while AJAX flew a descent pattern from 20:58 to 21:04 UTC. The AMOG descent was shortened
 450 to ~1000 m altitude (Glenville, CA) to allow AMOG to reach the Kern Fields nearly concurrent with
 451 AJAX and GOSAT.



452
 453 **Figure 4.** Surface altitude (z) above mean sea level profiles for west-east Delano-Alta Sierra transect (Fig.
 454 3A, α - α') for AMOG and AJAX (a) methane (CH_4), (b) carbon dioxide (CO_2), (c) ozone (O_3), (d)
 455 temperature (T), and (e) relative humidity (RH). Also shown on (d) are the dry, average, and wet adiabatic
 456 lapse rates. Data key on panel, planetary boundary layer (PBL), labeled. Green arrow shows extrapolation
 457 of AJAX trend to Shirley Meadows altitude (2058 m).

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460 Overlapping AMOG and AJAX profile data were collected between 500 and 2000 m. There was very
461 good agreement between the two platforms for CO₂ and CH₄ for altitudes between 1.55 and 2 km (Fig. 4a
462 and 4b), 99.9% and 99.7%. AMOG and AJAX CH₄ concentrations decreased notably from the well-
463 mixed PBL to the near surface layer, from ~2.07 ppm (500-750 m) to ~1.93 ppm (250-300 m). AJAX also
464 showed a decrease in CO₂ from 403 ppm to below 400 ppm. The CO₂ decrease was consistent with a
465 shift to agricultural air where CO₂ vegetative uptake reduces CO₂ concentrations. The PBL grew from
466 600 to 900 m between AMOG's ascent and descent and then to 1500 m by the time of AJAX's descent
467 based on the CH₄, CO₂, and O₃ data.

468
469 The PBL was identified at ~1580-1600 m based on both surface and airborne relative humidity (*RH*) and
470 temperature (*T*) vertical profiles. Winds were not useful for deriving the location of the PBL. Diurnal
471 heating is apparent between the two AMOG Surveyor *T* profiles, but does not change the lapse rate.
472 Because AJAX flies above the surface where AMOG collects data, AJAX temperatures are lower. In the
473 lower atmosphere, the lapse rate was 6.9°C km⁻¹ for AJAX between 500-900 m, while the AMOG lapse
474 rate from 200-900 m was a similar 5.6°C km⁻¹. Between 950 and the top of the PBL, AMOG lapse rates
475 were much shallower, 2.5 °C km⁻¹, with a jump in temperature at 900 m. Above the PBL, the AMOG
476 lapse rate was 3.5°C km⁻¹, close to the wet adiabatic lapse rate (Fig. 4d).

477
478 Above the PBL, O₃ concentrations between AMOG and AJAX were ~20 ppb different although the
479 AMOG and AJAX profile slope (dO_3/dz) were the same. If the trend in AJAX O₃(*z*) from 1600 to 1850 m
480 is extended to $z = 2058$ m (Fig. 3C, green arrow), there is agreement with AMOG Shirley Meadows (open
481 field) O₃ concentrations. This similar slope but different absolute value could indicate O₃ loss as it
482 diffused down through the pine canopy to the surface (and AMOG). Tall pine trees (30+ m) dominate
483 above ~1700, except for Shirley Meadows where, as noted, there was good agreement. For $900 < z <$
484 1400 m, AJAX - AMOG agreement was better for the descent, which was closer in time to AJAX than
485 the ascent. This shift likely was associated with formation of the daytime PBL.

486
487 In this upwind profile, AJAX observed elevated O₃ that was well mixed down to 500 m, while earlier
488 AMOG showed well-mixed O₃ down to only 1100 m. There also was a small (~10 ppb) O₃ enhancement
489 at the top of the PBL in both the airborne and surface profiles. The highest O₃ concentrations were
490 observed by AMOG in Shirley Meadows, where visibility was low due to smoke aerosols from the Rough
491 Fire (NASA, 2015). Air above the PBL was more humid than elsewhere in the profile, except for the
492 lowest 50 m above the valley floor, which was enriched in CH₄, CO₂, and *RH*, possibly from nocturnal
493 accumulation and agriculture including irrigation *RH* inputs. There were thin, atmospheric layers that

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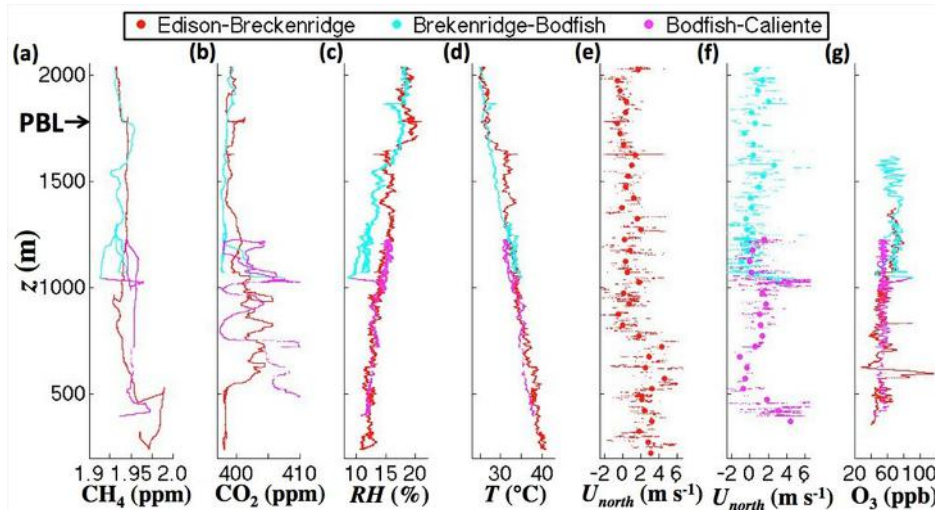
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499 | suggest remnant structures from the prior day. For example, at ~550 m the air changed character, with a
500 | jump in CO₂ by ~10 ppm, and of O₃ by ~ 10 ppb, and a decrease in the CH₄ altitude gradient (dCH₄/dz).

501

502 | Air was more polluted at greater altitude above the PBL in the upwind (Delano – Alta Sierra) profile for
503 | O₃ for both platforms with air 10-20 ppb greater than in the PBL. Additionally, AJAX CH₄ and CO₂ were
504 | significantly higher above the PBL. The AMOG CH₄ and CO₂ data are less clear, presumably because
505 | AMOG data were prior to the disappearance of the nocturnal, stably stratified PBL. This was consistent
506 | with visual observations of haze by AMOG from Shirley Meadows as well as by the AJAX pilot. Also,
507 | air above the PBL was more humid.



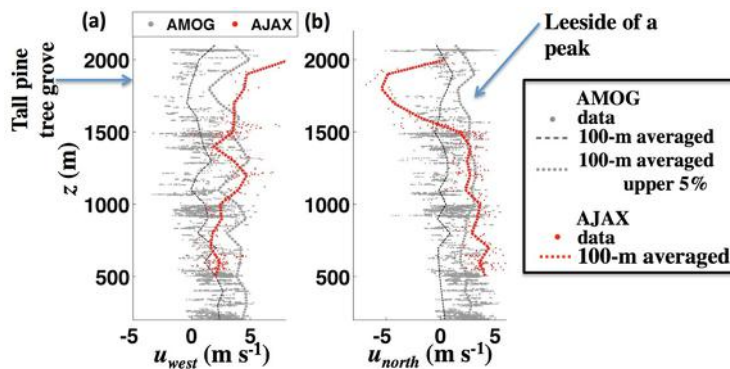
508

509 | **Figure 5.** Surface altitude (z) above sea level profiles for Edison-Breckenridge ascent (red) and descent
510 | (blue) to Bodfish and then Caliente profile (magenta) (Fig. 3b) for AMOG Surveyor (a) methane (CH₄),
511 | (b) carbon dioxide (CO₂), (c) relative humidity (RH), (d) temperature (T), north wind (U_{north}), for (e)
512 | ascent and (f) descent, dots shown 50-m altitude binned averaged, and (g) ozone (O₃). Planetary
513 | Boundary Layer (PBL) labeled.

514

515 | A downwind ascent profile in the SJV was collected from Edison, CA to the high flanks of Breckenridge
516 | Mountain, followed by a descent behind the Breckenridge Mountain to Caliente, CA through the tiny
517 | town of Bodfish (Fig. 3b). This descent was separated from the SJV by a ridge and includes dryer, clean
518 | air is that is representative of air from around Lake Isabella, a fairly isolated mountain valley. The
519 | downwind profile was collected quasi-Lagrangian in that the time separating the two profiles (about four

520 | hours) is comparable to the transport time (75 km at [a mean wind speed of 4 m s⁻¹](#), implies 5 hours for
 521 | transport). Thus, the downwind profile was for close to the same air. Over these hours, there was some
 522 | additional PBL development, ~100 m [growth](#) to ~1675 m, with highly uniform CH₄ between 1000 m and
 523 | the top of the PBL (**Fig. 5a**). Thus, the PBL remained fairly stable over the course of the study. Air in
 524 | both the upper PBL and above was cleaner with lower humidity and CH₄ concentrations. Unfortunately,
 525 | the O₃ analyzer overheated during the ascent and resumed collecting data on the descent at ~1500 m.



526 |
 527 | **Figure 6.** Altitude (z) profiles for (a) west (upslope) and (b) north (cross slope) wind components from
 528 | AMOG and AJAX for overlapping altitudes of the Delano–Alta Sierra transit (Fig. 3, α – α'), 100-m
 529 | altitude rolling-averaged data for AJAX, AMOG, and AMOG upper 5% of winds. Data key on figure.

530 |
 531 | Direct comparison between AMOG and AJAX winds is inappropriate because AMOG winds are affected
 532 | strongly by obstacles including hills, trees, and buildings. However, in many instances, terrain is open, or
 533 | gently rolling hills, and there tend to be regions of stronger winds that we propose are representative of
 534 | free atmosphere winds. AMOG data were altitude binned and the strongest winds in each bin were
 535 | compared with AJAX (Fig. 6). Agreement is generally good (within 15-20%) between the upper 5% of
 536 | AMOG cross-slope (west) winds in each altitude-averaged band (Fig. 6a). For the upslope wind (north)
 537 | agreement is better (within 5-10%) for a larger range of altitudes (Fig. 6b). This allows fusions of the
 538 | upper 5% of AMOG winds with AJAX winds. [Over the full altitude range, the median differences were](#)
 539 | [38% and 27% for the north and east wind components. The altitude variation in the agreement is shown in](#)
 540 | [Supp. Fig. S7.](#)

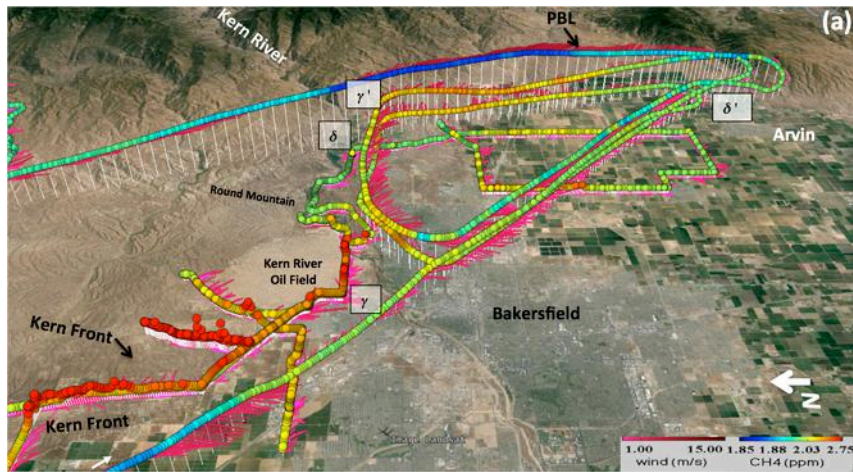
541 **3.2. Kern Fields and Bakersfield Greenhouse Gas Emissions**

542 **3.2.1 Methane**

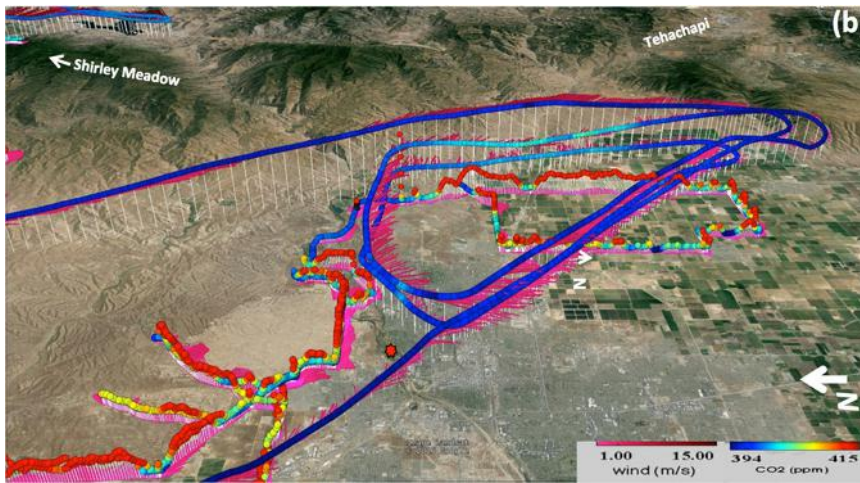
543 On 19 Aug. 2015, winds over the Kern Fields were prevailing (northwesterly) and fairly strong ($\sim 3 \text{ m s}^{-1}$)
544 on the ground and somewhat stronger aloft (Fig. 7). Potential plumes from the only nearby upwind dairy
545 (Fig. 7a, white arrow) were directed by winds to pass to the west of the oil fields, agricultural fields in this
546 part of the SJZ are dry. As a result, surface topography like the Kern River Bluffs imposed only small
547 wind modification at the surface and at altitude. Southeast of Bakersfield, winds veered to westerlies
548 towards passes in the Sierra Nevada Mountains that connect to the Mojave Desert. The downwind survey
549 included two plume transits on agricultural roads with negligible to no traffic. These transits clearly
550 showed the plume's eastward drift, passing to the north of the small town of Arvin, CA.

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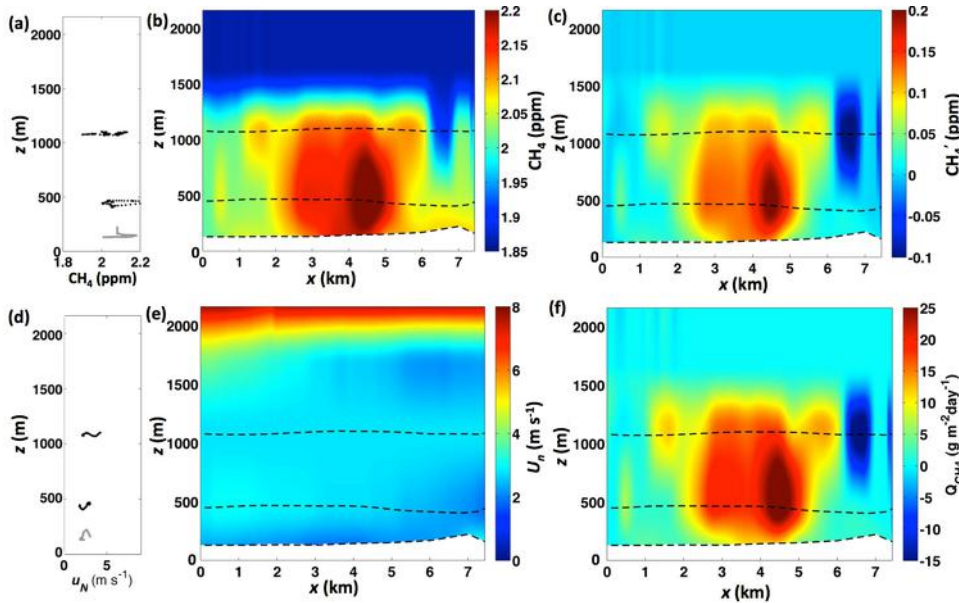


552



553

554 **Figure 7.** Combined AJAX and AMOG winds and *in situ* (a) methane (CH₄) and (b) carbon dioxide
 555 (CO₂) for the Kern Fields on 19 Aug. 2015 for prevailing wind conditions. White arrow to the west of
 556 Kern Front oil field shows location of nearby dairy. Greek letters identify two downwind curtains. Red
 557 star on (b) locates origin for transect $\gamma-\gamma'$. Data keys on figure.



558

559 **Figure 8.** (a) Methane (CH₄) altitude (z) profiles for 19 Aug. 2015 for AJAX (black) and AMOG (gray)
 560 data. (b) Interpolated, fused AJAX and AMOG CH₄ data, with respect to lateral east distance (x) relative
 561 to 119.0023°W, 35.3842°N for data plane γ - γ' (Fig. 7). Dashed lines show data locations. (c) CH₄
 562 anomaly (CH₄') relative to the background data plane (Supp. Fig. S6A). (d) Vertical normal wind profile
 563 (U_n) from AJAX (black) and AMOG (gray) data during ascent/descent, (e) interpolated, fused U_n , and (f)
 564 CH₄ flux (Q_{CH_4}) for the Kern Fields. Data key on panels.

565

566 The background CH₄ plane $C_B(x,z)$ was extracted from the CH₄ data outside the plume – $C_{BL}(z)$ and
 567 $C_{BR}(z)$, see Eqn. (2) – immediately downwind of the Kern Fields (transect γ - γ'). C_B showed a slight
 568 increase towards the east of ~20 ppb (Supp. Fig. S6a). The normal wind (U_n) was fairly uniform across
 569 the data plane, including downwind of the canyon (Fig. 8e). Thus, the CH₄ flux ($Q_{CH_4}(x, z)$ shows similar
 570 spatial patterns to CH₄'(x, z). Emissions from the Kern Fields' were dominated by a large, focused CH₄
 571 plume (or group of plumes) in the core of a much broader, dispersed, and poorly defined plume. This
 572 structure is evident in both surface AMOG data and in the lowest AJAX altitude for plane γ - γ' with both
 573 showing the strongest peak at $x = 4.5$ km (Fig. 8b, dashed lines). Total estimated emissions (E) were
 574 63.5±50% Mol s⁻¹ (equivalent to 32 Gg yr⁻¹). Uncertainty is from the Monte Carlo simulations, described
 575 in section 2.5.

576 Within the plume, concentrations are elevated at 1200 m altitude relative to 500 m and the surface,
577 indicating buoyant rise. Additional evidence for buoyant rise is provided by two small plumes at $x \sim 1.7$
578 and 5.7 km were centered at the top of the PBL but were not also observed in surface and mid altitude
579 data. The upper AJAX flight line was several hundred meters below the top of the PBL (at ~ 1580 m, Fig.
580 4), which constrains the main plume and was centered vertically in the PBL. Concentrations above the
581 PBL were determined from AJAX descent and ascent data (Fig. 4), in agreement with AMOG data above
582 the PBL. These observations show that the plume was not well mixed across the PBL. Another important
583 feature is the upper altitude clean air intrusion at $x \sim 6.5$ km lies downwind of Round Mountain Canyon to
584 the east of the Kern River oil field (Fig. 8b, Fig. 7a for location). This intrusion does not penetrate down
585 to 500 m and represents a downslope airflow of cleaner upper level air.

587
588 For comparison, a recent bottom-up estimate of CH_4 emissions based on production data for the Kern
589 Fields estimated 10-40 Gg $\text{CH}_4 \text{ yr}^{-1}$ (68% Confidence Level), by combining oil and gas production data
590 with US-EPA emissions factors for associated wells (Jeong et al., 2014). Other CH_4 sources are unlikely
591 to confuse this interpretation as petroleum system emissions are ~ 20 times larger than estimated nearby
592 livestock and landfill CH_4 emissions of ~ 2.3 and 1.4 Gg yr^{-1} , respectively (Calgem, 2014).

593 3.2.2. Carbon Dioxide

594 Background CO_2 for data curtain $\gamma - \gamma'$ (Supp. Fig. S6b) was highly uniform. Given the strong crosswinds
595 and care taken to avoid trailing other vehicles on the low-trafficked China Loop Road, these data passed
596 quality review— CO_2 exhaust contamination manifests as a dramatic increase in the standard deviation as
597 AMOG intersects a turbulent vehicle exhaust plume. There was a shallow CO_2 layer constrained to the
598 lower 100 to 200 m with ~ 10 ppm enhancement (Fig. 9a), also observed in the CO_2 vertical profile (Fig.
599 4b), a layer that was characterized by elevated relative humidity. Further evidence that these broad spatial
600 CO_2 emissions are real is from the spatial similarity to CO_2 enhancements in the lowest AJAX flight data
601 (Fig. 9c). For example the surface CO_2 plume was strongest at $x \sim 4.5$ km in AMOG and AJAX data. The
602 broad spatial extent of these emissions, similar to the broad CH_4 emissions suggests a relationship to
603 field-scale (engineering or geological) processes. Overall CO_2 emissions were $1730 \pm 50\% \text{ Mol s}^{-1}$
604 (equivalent to $2.4 \pm 1.2 \text{ Tg yr}^{-1}$).

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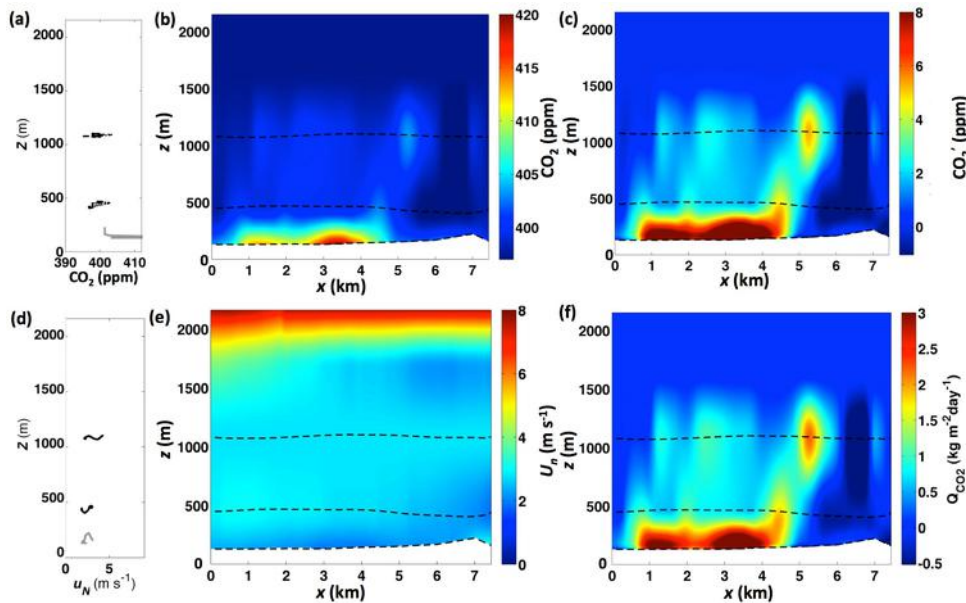
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626 **Figure 9.** (a) Vertical carbon dioxide (CO_2) altitude (z) profile data for 19 Aug. 2015 for AJAX (black)
 627 and AMOG (gray) data. (b) Interpolated, fused AJAX and AMOG CO_2 data curtain with respect to lateral
 628 east distance, x , relative to 119.0023°W , 35.3842°N for curtain $\gamma-\gamma'$ (Fig. 7b). Dashed lines show data
 629 locations. (c) CO_2 anomaly (CO_2'). (e) Vertical normal wind profile (U_n). (e) Interpolated, fused U_n , and
 630 (f) CO_2 flux (Q_{CO_2}) for the Kern River and Kern Front oil fields for 19 Aug. 2015. Data key on panels.

631 There was a strong CO_2 anomaly in a focused plume at $x = 5$ km and $z = 1$ km. This plume likely relates
 632 to the two cogeneration power plants located in the Kern River oil field. Further support for this
 633 interpretation is its co-location with a similarly focused CH_4 plume at the same location. This power
 634 plant-related feature is a persistent feature that has been observed in other surveys (Leifer – unpublished
 635 data). The upper clean air intrusion in the CH_4 data curtain also is apparent in the CO_2 data (Fig. 9b), in
 636 front of Round Mountain Canyon (Fig. 7).

637

638 Based on a reservoir $\text{CO}_2:\text{CH}_4$ gas ratio of 92.2%:1.7% (Lillis et al., 2008) and 32 Gg yr^{-1} CH_4 emissions,
 639 the Kern Fields' CO_2 emissions were predicted to be 1.8 Tg yr^{-1} , which is fairly consistent with the
 640 directly derived emissions of 2.4 Tg yr^{-1} . Both these values are somewhat lower than the inventory for the
 641 cogeneration plants in Kern River oil field, 3.1 Tg yr^{-1} (CARB, 2016). The disagreement with inventory
 642 likely arises from the co-generation plant only being active some of the time, confirmed by data from the
 643 GOSAT-COMEX campaign.

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649 4. Discussion

650 4.1. Experimental design and real-time visualization

651 Ideally, GCE airborne and surface data are collected first upwind and then downwind. However, AJAX
652 airborne data are not collected in a Lagrangian sense as would be necessary for slower, less maneuverable
653 airborne platform thanks to its extreme speed and maneuverability. This allows collection of near
654 snapshot (~30 minutes) data. Slower, AMOG surface data were collected in a quasi-Lagrangian sense,
655 reducing the likelihood of confounding interference in the study area from non-FFI SJV inputs due to
656 wind shifts after the pre-survey (for non-nominal winds the collection is aborted). Given the AJAX-
657 AMOG speed difference, concurrent surface and airborne data could not be collected both upwind and
658 downwind, and thus, concurrency was prioritized for downwind. For flight efficiency and to provide
659 downwind concurrency with AMOG, AJAX flew a triangle that allowed AJAX to complete transects at
660 three altitudes in close to AMOG's upwind-downwind survey time.

661

662 After the Kern Fields survey, AJAX returned to base, while AMOG collected additional surface data,
663 exploring the fate of emissions from the Kern Fields. The word, "exploring" is significant, as real-time
664 visualization of winds, CH₄, and O₃ guided the downwind surveying. Data were collected to test the
665 hypothesis that there was a relationship between wind strength and the specific outflow path from the SJV
666 to Mojave Desert - specifically, that more northerly passes, which require greater wind veering from
667 prevailing are preferred at lower winds speeds. The AMOG survey first confirmed that outflow was not
668 up the Kern River Valley, and then collected a downwind vertical profile into the Sierra Nevada
669 Mountains to search for outflow through a pass near Breckenridge Mountain. After confirming its
670 absence, AMOG then investigated in the Tehachapi Pass, where the outflow was identified. Thus, on 19
671 Aug. 2015, when winds were strong, the outflow was by the most direct pathway - the Tehachapi Pass.

672 4.2. Experimental design and uncertainty reduction

673 The experimental design reduced uncertainty by characterizing the PBL through surface and airborne data
674 fusion so that a well-mixed PBL is not required. Note, for a well-mixed PBL, surface-airborne data fusion
675 does not reduce uncertainty. The benefit arises for a not well-mixed PBL where a significant fraction of
676 the plume mass lies below the lowest altitude the airplane can fly. In such case, surface data inclusion

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678 adds information to the PBL characterization. For example, flights are often face airspace restrictions for
679 a number of conditions including cities, approach pathways, military airspace, and/or safety.

680 Aerial survey altitudes were designed to span from near the top of the PBL to as low as permissible and
681 an intermediate level (0.5, 1, 1.2 km). Thus, surface data added information on the lowest third of the 1.6-
682 km thick PBL. This lower portion of the PBL is more important on days when the PBL is shallower.

684 Observations showed that the well-mixed PBL assumption was poor as far as 10-20 km downwind. One
685 solution is to collect data even further downwind, where the PBL should be better mixed (White et al.,
686 1976); however, secondary (potentially uncharacterized) sources downwind of the study area and upwind
687 of the downwind data plane add confounding anomalies. Also, wind flow complexity can lead to transport
688 orthogonal to the overall downwind direction, leading to flux leakage out of the plume. The likelihood of
689 plume loss increases over greater distances. And finally, as the PBL evolves with time, it imposes an
690 evolving structure on the wind and concentration vertical profiles, which also challenge the well-mixed
691 PBL assumption – particularly if transport to the downwind plane requires hours.

693 The *in situ* analyzers record concentration and winds with very high accuracy; however, only at a single
694 location and time. Thus, *in situ* uncertainty arises mostly from inadequate characterization of temporal
695 variability and spatial heterogeneity in winds and emissions over the survey time period. The best strategy
696 is to minimize study time; however, there is a necessary tradeoff between spatial resolution and study
697 time. AJAX collects data quickly, allowing survey completion within far less than typical atmospheric
698 change timescales. Similarly, the surface survey route was designed to minimize collection time,
699 primarily on rural/agricultural roads carefully selected to avoid traffic congestion and traffic lights. The
700 surface survey requires ~90 minutes to complete and is conducted quasi-Lagrangian.

702 GCE treats uncertainty explicitly, allowing improvements in the data collection strategy to reduce
703 uncertainty. For example, the east-west downwind transect was lengthened from earlier data collects to
704 characterize background concentrations better. GCE also does not require an *a priori* emission
705 distribution and thus incorporates explicitly emissions from super-emitters, normal emitters, and
706 distributed sources, improving robustness of the findings. In contrast, inversion models require a
707 reasonable spatial *a priori* emission distribution and the ability to model transport across the study
708 domain. However, complex wind flows from fine-scale topographic structures, as observed for the Kern
709 Fields, challenge transport modeling.

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743 **4.3. Profile intercomparison**

744 This study leveraged terrain to provide profile information with a surface mobile platform, which was
745 compared with airborne data. In this study, the two were combined to provide more complete coverage of
746 the atmosphere than a single platform could, at a fraction of the cost (not to mention logistical
747 complexity) of having two airborne platforms. Whereas the approach worked well in the San Joaquin
748 Valley, further research is needed to confirm its utility in other settings.

749
750 Above the PBL, there was excellent agreement between surface and airborne concentration profile data,
751 while concentration profiles within the PBL show significant differences between the two profiles, likely
752 related to air mass shifts and diurnal heating during the time between the profiles (Fig. 4). Winds above
753 the PBL were in poor agreement, with the north component in the opposite direction (Fig. 6). Underlying
754 this discrepancy was a mountain peak, which clearly caused large-scale alterations in the wind flow field.

755
756 Within the PBL, agreement between unfiltered surface AMOG winds and AJAX winds was poor,
757 unsurprising because surface winds are strongly affected by obstacles. However, by filtering AMOG
758 winds (collected 3-m above the surface) for the strongest 5%, agreement was within 15-20% for the
759 along-slope – i.e., north – winds, and better for upslope winds (west). Specific exceptions were when
760 AMOG was in a dense grove of pines, and when AJAX flew behind into the lee of a mountain peak.
761 Surface winds are modulated by a wide range of surface factors including trees, steep hills and hillocks,
762 blocking by a steep slope, rolling hills, and structures (Supp. Fig S5). However, a combination of gusts
763 (among thin wooded terrain on steep slopes) and the limited spatial extent of most obstacles underlies the
764 agreement between the filtered AMOG and AJAX wind profiles. Agreement is better for the upper
765 portions of the PBL (within 10-20%) where Sierra Nevada Mountain slopes are steeper. In contrast, the
766 slope lower in the PBL is gentle, and surface boundary layer effects are more pronounced, biasing wind
767 speeds slower.

768
769 The wind orientation to the slope affects the comparison because topography imposes wind field structure
770 at large and small scales. Where winds advect air upslope, transport incorporates a non-negligible vertical
771 component that is missed by the 2D sonic anemometer used in the study reported here. The current
772 AMOG configuration measures 3D winds, as does AJAX.

773
774 Some of the discrepancy between AMOG and AJAX wind profiles could have arisen from temporal
775 changes between the two profiles; however, this is unlikely for two reasons. First, the top of the PBL was

776 identified four times over the course of the study and remained stable within 100 m across the domain.
777 And second, surface wind observations remained relatively constant after the mid-morning shift to
778 daytime conditions (breakup of nocturnal stratification). However, the poor agreement between AJAX
779 and AMOG vertical concentration profiles within the PBL suggests significant air mass shifts –
780 highlighting the need for better concurrence.

781 4.4. GHG FFI emissions

782 Emissions for the Kern Fields were estimated at 32 ± 16 Gg CH_4 yr^{-1} with CH_4 emissions ~20% above
783 EPA inventories, and 2.4 ± 1.2 Tg CO_2 yr^{-1} . The broad CO_2 plume suggests emissions from the geologic
784 reservoir – likely along the same pathways associated with CH_4 leakage – in addition to the focused and
785 not continuous emissions from the co-generation power plants. On China Loop Road (where the CO_2
786 surface plume was transected), strong crosswinds and light traffic would have prevented significant
787 vehicular CO_2 contamination. Additionally there are no upwind (non-oil field) roads, only the foothills of
788 the Sierra Nevada Mountains.

790 For comparison, a recent bottom-up estimate of CH_4 emissions from the Kern Fields estimated 25 ± 15 Gg
791 CH_4 yr^{-1} by combining oil and gas production data with emissions factors for associated wells used by
792 US-EPA (Jeong et al., 2014), i.e. 19 Aug. 2015 CH_4 emissions were a third above inventories. The
793 derived flux lies within the inventory uncertainty, but is higher, consistent with a recent metastudy of
794 field studies of FFI production emissions, which showed significant underestimation in the EPA budget
795 (Brandt et al., 2014; Miller et al., 2013). A number of factors likely play a role including the age of the
796 Kern River oil field (over a century), production factors (steam injection), shallowness of the reservoir
797 (<300 m), location in a tectonically active area, which creates alternate migration pathways from the
798 reservoir (Leifer et al., 2013), and the recent expansion of the number of wells in the Kern Front oil field
799 (from GoogleEarth timeline imagery). Many of these factors are common to other production fields in
800 California, the US, and globally. Given the importance of FFI to the overall budget, even small
801 underestimation could be highly significant. Thus, this uncertainty highlights the need for improved
802 measurement tools to reduce the significant uncertainty in the CH_4 budget and for satellite measurement
803 validation, particularly for complex terrain and in the source's near field.

804 5. Conclusion

805 This study showed how to combine airborne and surface *in situ* data to improve emissions derivation, and
806 demonstrated the novel use of topography to characterize vertical atmospheric structure with a surface
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818 mobile platform. Given that mountains cover a significant fraction of the earth's land surface, further
 819 research should be undertaken to confirm that this approach applies in other settings. Data showed the
 820 PBL was not well-mixed, even 10-20 km downwind, highlighting the importance of the direct flux
 821 quantification approach. Direct quantification does not require accurate modeling of winds across
 822 complex terrain, but does require interpolation and data modeling to identify the background.
 823

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824 **Table of Nomenclature**

825	Units	Description
826	AJAX (-)	Alpha Jet Atmospheric eXperiment
827	AMOG (-)	AutoMOBILE trace Gas
828	Bbl (-)	Barrel (of oil) 1 bbl = 6.38 m ³
829	COMEX (-)	CO ₂ and MEthane eXperiment
830	EOR (-)	Enhanced oil recovery (techniques)
831	EPA (-)	Environmental Protection Agency
832	GCE (-)	GOSAT COMEX Experiment
833	GHG (-)	Greenhouse Gases
834	GOSAT (-)	Greenhouse gases Observing SATellite
835	GHG (-)	Greenhouse gas
836	PBL (-)	Planetary Boundary Layer
837	SJV (-)	San Joaquin Valley
838	Tg	Terragram (10 ¹² g)
839	UTZ (-)	Universal time
840	$C'(x,z)$ (ppm)	concentration anomaly (above C_B)
841	$C(x,z)$ (ppm)	concentration
842	$C_B(x,z)$ (ppm)	background concentration – outside plume
843	$C_{BL}(z)$ (ppm)	background concentration profile – left side of profile
844	$C_{BR}(z)$ (ppm)	background concentration profile – right side of profile
845	E (mol s ⁻¹)	Emission source strength
846	$k(z)$ (mol ppm ⁻¹)	Conversion factor from the ideal gas law
847	N' (mol cm ⁻³)	molar mass anomaly
848	$Q(x,z)$ (mol m ⁻² s ⁻¹)	Flux through the data plane
849	R^2 (-)	Correlation coefficient
850	RH (%)	Relative humidity
851	T (°C)	Temperature

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858	$U_n(x,z)$	(m s ⁻¹)	Winds normal to the data plane, a function of (x, z)
859	U_{north}	(m s ⁻¹)	North wind component
860	U_{west}	(m s ⁻¹)	West wind component
861	x	(m)	lateral distance – approximately cross-wind
862	x_L	(m)	left half of the transect ($x < x_{max}/2$)
863	x_{max}	(m)	length of a transect
864	x_R	(m)	right half of the transect ($x > x_{max}/2$)
865	y	(m)	lateral distance – approximately co-wind
866	z	(m)	altitude
867	$\Phi_L(C)$	(-)	concentration probability distribution for left side of transect
868	$\Phi_R(C)$	(-)	concentration probability distribution for right side of transect
869	$\Phi_P(C)$	(-)	concentration probability distribution for the plume
870	$\Phi_B(C)$	(-)	concentration probability distribution for the background
871	α, α'	(-)	designation for Delano – Alta Sierra surface transect
872	$\varepsilon, \varepsilon'$	(-)	designation for Edison– Breckenridge Mtn. surface transect
873	τ, τ'	(-)	designation for Breckenridge – Caliente surface transect
874	β, β', β_1'	(-)	designation for Wasco – Granite surface transect
875	γ, γ'	(-)	designation for Oildale – Oil City surface and airborne transects
876	δ, δ'	(-)	designation for Ming Park – Arvin surface and airborne transects

877

878 **Data Availability.** Data will be provided as per the data policy.

879

880 **Author Contribution.** I. Leifer prepared the manuscript with input from all co-authors. C. Melton
881 prepared figures and conducted data analysis. M. Fischer helped prepare the emissions budgets. J. Frash
882 helped with AMOG data collection. L. Iraci, J. Marrero, J-M. Ryoo, T. Tanaka, and E. Yates are part of
883 the AJAX team and worked to collect and analyze AJAX data.

884 There are no competing interests

885

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These results agree with a recent metastudy of field studies of FFI production emissions, which showed significant underestimation in the EPA budget (Brandt et al., 2014; Miller et al., 2013). Given the importance or dominance of FFI emissions in anthropogenic greenhouse gas budgets, an increase of 25-50% of the FFI contribution requires either reduction in another budget

category, and/or an increase in the loss rate. However, a recent husbandry emissions study also suggested significant underestimation (Gentner et al., 2014). Thus, the present study supports the hypothesis that CH₄ loss rates are underestimated. For example a recent study identified a new loss mechanism in near-surface soils (Fernandez-Cortes et al., 2015). In any case, this study

1 Supplemental Material

2 S1. Study Area



3
4 **Figure S1.** Photo of Bakersfield and the South San Joaquin Valley from the NASA (Earth
5 Research ER-2 airplane at 20-km altitude. Blue-white arrows show approximate direction of
6 prevailing winds, oil fields near Bakersfield labeled. Photo courtesy Stuart Broce, Pilot, NASA
7 Armstrong Flight Research Center.

8

9 S2. Platforms

10 S2.1. Surface – AMOG Surveyor

11 Mobile surface *in situ* measurements using Cavity RingDown Spectroscopy (CRDS) (Pétron et
12 al., 2012; Farrell et al., 2013) and open path spectroscopy (Sun et al., 2014) are becoming more
13 common. Surface data were collected for the *GOSAT COMEX Experiment* by the AMOG
14 (AutoMOBILE trace Gas) Surveyor (Leifer et al., 2014). AMOG Surveyor is a commuter car
15 (Versa SP, Nissan, Japan) that is modified for mobile high-speed, high-spatial resolution
16 observations of meteorology (winds, temperature, and pressure), gases (greenhouse and other
17 trace), and remote sensing parameters (Fig. S2).



18

19

20 **Figure S2.** (a) AMOG Surveyor in the Transverse Coastal Range (1300 m) – San Joaquin Valley
 21 in background. (b) Cockpit view of gauges, security video, rear video, real-time data display. V_A ,
 22 V_{FB} , V_{RB} , V_I – voltages for alternator, front battery, rear battery, inverter. T_b , T_o , T_w – temperatures
 23 for inverter, engine oil, and radiator water. P_T , P_o , P_w , P_s , P_C , P_R – pressures for tires, oil, water
 24 suspension, compressor, and regulated air for chemical scrubbers. (c) AMOG Surveyor in Sierra
 25 Nevada Mountains, roof package labeled.

26

27 Analyzers: AMOG Surveyor draws air down two ½” PFA Teflon sample lines from 5 and 3 m
 28 above ground into a configurable range of gas analyzers by a high flow (850 lpm, 30 cfm)
 29 vacuum pump (Edwards, GVSP30). The higher sample line connects to several analyzers
 30 including a Fast-flow, enhanced performance Greenhouse Gas Analyzer (FGGA, [enhanced](#)
 31 [model, Los Gatos Research, CA](#)), which uses Integrated Cavity Off-Axis Spectrometer-Cavity
 32 Enhanced Absorption Spectroscopy (ICOAS-CEAS) and measures carbon dioxide, CO₂,
 33 methane, CH₄, and water vapor, H₂O, at up to 10 Hz (Model 911-0010, Los Gatos Research, Inc.,
 34 Mountain View, CA). AMOG also measures carbonyl sulfide (COS) and carbon monoxide (CO)
 35 with an ICOAS-CRDS analyzer (Model 907-0028, Los Gatos Research, Inc., Mountain View,

36 CA). An additional sample line collects feeds an ICOAS-CRDS that measure ammonia (NH₃)
37 and hydrogen sulfide (H₂S). For all CEAS analyzers, dry values are used. Also, three
38 chemiluminescence trace gas analyzers measure nitric oxide (NO) and nitrogen oxides (NO_x) at
39 0.1 Hz at 25 ppt accuracy (42TL, ThermoFischer Scientific, Waltham, MA), and ozone (O₃) at
40 0.25 Hz at 1 ppb accuracy (42C, ThermoFischer Scientific, Waltham, MA), and sulfur dioxide
41 (SO₂) at 0.1 Hz at 1 ppb accuracy (450C, ThermoFischer Scientific, Waltham, MA). This

42 accuracy is from the manufacturer and is based on 24 hour drift. Better accuracy is achieved by
43 hourly zero gas measurements using chemically sparged air (Type CI, Cameron Great Lakes,
44 OH), which in the laboratory improved accuracy to 50 ppt. Given that SO₂ and H₂S atmospheric
45 concentrations are typically less than 1 ppb in California, this was an important improvement.

46
47 The FGGA is calibrated with an air calibration standard for greenhouse gases (CH₄: 1.981 ppmv;
48 CO₂: 404 ppmv; balance ultrapure air) and are stable to 1 ppb for CH₄ over 24 hours, and 0.12
49 ppm for CO₂ over 24 hours. Accuracy is <0.03%. Calibrations are performed before and after
50 each field collect. The 49i was cross-calibrated with the AJAX O₃ analyzer to 1 ppb, and during
51 a repeat cross calibration several months later had maintained its calibration to between 1 and 2
52 ppb.

53
54
55 Meteorology: A sonic anemometer (VMT700, Vaisala, Finland) is mounted 1.4 m above the roof
56 and measures two-dimensional winds. Estimated accuracy is approximately 10° and 0.3 m s⁻¹ for
57 wind speeds above 1.5 m s⁻¹; however, accuracy improves with vehicle velocity and wind speed
58 as vehicle flow stream line interferences are reduced. Accuracy was determined empirically by
59 driving several kilometers back and forth on a rural road in an open area in the early morning and
60 comparing measured winds in the two directions. Note, these accuracies are greater than the
61 manufacturer maximum error. At lower wind speeds, accuracy appears to be closer to 0.2 m s⁻¹,
62 and 15-20°; however, is extremely challenging to determine. Still, filtering, nocturnal wind data
63 generally agrees well (~10°) with expectations from topographic forcing at wind speeds of ~0.2 -
64 0.5 m s⁻¹ on large spatial scales (tens of kilometers) for highway speed (140 km hr⁻¹) data. In
65 general, winds are more accurate than stated if the winds are from within 30° of forward
66 direction, as stated if they are from the side, unless strong (>~4 m s⁻¹), in which case they are

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Moved up [1]: The 450C can achieve 50 ppt accuracy by hourly zero gas measurements using chemically sparged air (Type CI, Cameron Great Lakes, PA).

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79 equally accurate and very poor if from within ~15° of the behind direction. As a result, tail winds
80 are not evaluated.

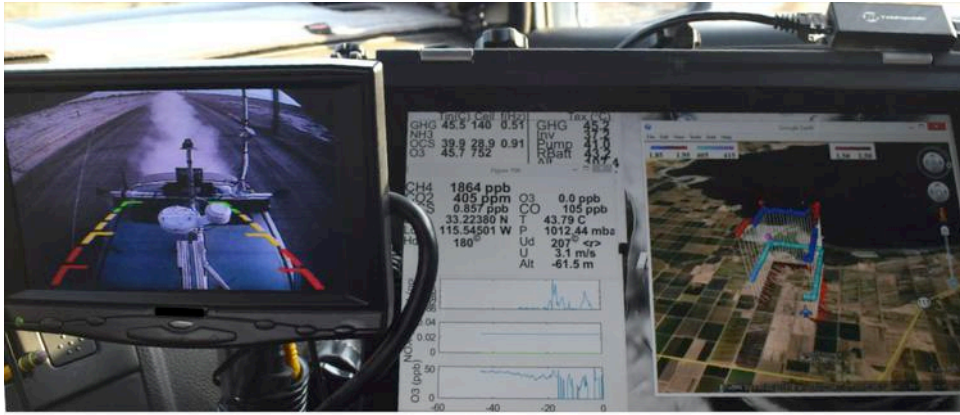
81
82 AMOG system improvements beyond (2014) include a high speed thermocouple (50416-T,
83 Cooper-Atkins, CT) and a high accuracy (0.2 hPa) pressure sensor (61320V RM Young Co.,
84 MI), connected by a stainless steel line into a roof passive radiation shield (7710, Davis
85 Instruments, CA) to reduce dynamic pressure effects. The radiation shield also includes a Type T
86 thermocouple (Omega, CT) digitized at 0.03°C resolution (CB-7018, Measurement Computing,
87 MA). A solar insolation sensor is digitized at 16 bit and 1 Hz (CB-7017, Measurement
88 Computing, MA). Two (redundant) Global Navigation Satellite Systems (19X HVS, Garmin,
89 KS) that use the GLONASS, GPS, Galileo, and QZSS satellites provide position information at
90 10 Hz.

91
92 Vehicle Power: To support the science package (~1.8 kW), with clean DC and AC power,
93 AMOG has a 3.3 kW alternator (Nations Alternator, Cape Girardeau, MO), with a 2.7 kW
94 inverter (2810M, Outback Power, Arlington, OR), and a dual voltage conversion 2.4kW
95 uninterruptible power supply (Tripp Lite SU3000RTXL3U) backed by three, deep cycle gel
96 batteries for a total of 250 Amp-hours (Lifeline Batteries, WI; 6FM100H, Vision, MO; PVX-
97 1040T, Sun Xtender, CA) with active isolation (Dual Rectifier Isolator, Stolper International,
98 Inc., San Diego, CA). The 100 A-hr batteries and inverter are mounted in the cabin floor center
99 to improve stability. The DC system includes a 1-farad capacitor to stabilize against surges.

100
101 AMOG Surveyor weighs ~1 ton above stock, with significant safety implications, which were
102 addressed by enhancements to handling, suspension, and braking. Specifically, front drilled and
103 slotted ceramic brakes (F2473, Black Hart). Suspension modifications include rear airbag
104 suspension (NV-NINV-RBK, X2 Industries, AZ), adjustable rear truck shocks (for a Ford F-
105 150), performance coil-over front struts (TSC123, Tanabe, Japan), strut tower bar, sway bar, and
106 ladder brace.

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108
 109 **Figure S3.** AMOG Surveyor cockpit view showing real-time display (right) and rear camera
 110 view in the Salton Sea, CA. Methane (CH_4), carbon dioxide (CO_2), and wind speed (U) and
 111 direction (U_d) are shown in the Google Earth visualization window. Rolling display (lower left)
 112 shows CH_4 , nitrogen oxides (NO_x) and ozone (O_3). Diagnostics window (upper left) shows cell
 113 pressures and temperatures and key temperatures.

114

115 Data Handling and Integration: A touchscreen tablet (SpectreX360, HP) logs data
 116 asynchronously from instruments and sensors through several serial Ethernet servers (5450
 117 NPort, Moxa, Brea, CA) and industrial switches (EDS205, Moxa, Brea, CA). Logged data are
 118 mirrored to a SSD LAN drive in AMOG. Acquisition time is identified to ~30 milliseconds from
 119 the position of the data in the serial server buffer queues.

120

121 Custom software integrates the data streams and creates real time visualizations of multiple
 122 parameters in the Google Earth environment to enable adaptive surveying (Thompson et al.,
 123 2015). In adaptive surveying, the survey route is modified based on real time environmental
 124 conditions (winds, new/unexpected sources, etc.). GoogleEarth visualizations are displayed on
 125 one to several computers in AMOG Surveyor (**Fig. S3**) and remotely through cloud mirroring.
 126 Viewing algorithms automatically follow the vehicle, rotated to display wind vectors, and adjust
 127 the view altitude based on vehicle velocity. Algorithms minimize track overlap confusion
 128 through selective use of transparency, i.e., when AMOG Surveyor returns on the same course, or
 129 loops around. Rolling history displays of gas concentrations are useful for identifying recently

130 transected plumes. Other windows display AMOG Surveyor and analyzer diagnostics, and real
131 time analyzer gas and meteorology values.

132

133 **S2.2. Airborne - AJAX**

134 Airborne *in situ* data were collected by AJAX (Alpha Jet Atmospheric eXperiment), operated
135 from NASA Ames Research Center (ARC) at Moffett Field, CA. The alpha jet aircraft, which
136 has been modified for science missions, measures carbon dioxide and methane (Picarro Inc.,
137 model G2301-m), ozone ([Model 205](#), 2B Technologies Inc.), formaldehyde (Compact
138 Formaldehyde Fluorescence Experiment, COFFEE), and meteorological parameters including
139 3D winds [by the Meteorological Measurement System \(MMS\), a NASA developed system](#)
140 [\(https://earthscience.arc.nasa.gov/mms/\)](https://earthscience.arc.nasa.gov/mms/), from two externally-mounted wing pods (**Fig. S4**). [MMS](#)
141 [accuracies are \$\pm 1 \text{ m s}^{-1}\$ horizontal, \$0.3 \text{ m s}^{-1}\$ vertical](#). The greenhouse instrument was calibrated
142 using whole-air (National Oceanic and Atmospheric Administration) standards before and after
143 aircraft deployment. The ozone sensor is frequently calibrated to a NIST- traceable standard.
144 Further details on the aircraft and instrumentation are reported by Hamill et al. (2015); Tanaka et
145 al. (2016) and Yates et al. (2013).

146



147
148 **Figure S4.** AJAX photo. Courtesy Warren Gore, NASA Ames Research Center.

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156 The Alpha Jet is owned by H211, LLC, a collaborative partner with NASA. It is a tactical strike
157 fighter developed by Dassault-Breguet and Dornier through a German-French NATO
158 collaboration. Dassault concurrently developed a trainer version of the Alpha Jet that is still in
159 service with the French Air Force. Carrying a crew of two, it has a length of 12.2 m, a wingspan
160 of 9.2 m, and a height of 4.2 m. Its empty weight is 3540 kg and a maximum takeoff weight of
161 8000 kg. It has a ceiling of 15,545 m, speed of 280 – 930 km/hr, and a range of approximately
162 1930 km with full fuel.

163

164 The Alpha Jet stationed at NASA Ames – Moffett Field is operated in accordance with an FAA
165 Experimental Certificate of Airworthiness. It has a 2 to 2.5 hr flight duration, permitting up to
166 two missions per day with appropriate crew changes. Three highly experienced H211 pilots are
167 FAA Type Certificated to fly the Alpha Jet, and science test flights began in September 2010.
168 Following a complete avionics update and installation of the NASA-specified payload
169 management and control system in early 2009, the Alpha has proven extremely robust and
170 reliable. Its fleet safety record as a twin-engine, all weather jet is excellent, and its modern
171 Snecma engines produce a noise signature equivalent to current generation Stage III noise
172 compliant turbofan aircraft.

173

174 H211 has provided significant upgrades to the aircraft to support scientific studies. Extensive
175 wiring and cabling provisions have been installed to both wing pod locations, as well as the
176 centerline pod, to allow for distribution of 120 and 26 volt AC and 28 volt DC to each wing pod,
177 as well as additional 120 volt AC and 28 volt DC service to the centerline pod. Redundant
178 heavy-duty Ethernet cables have been provided from the wing pods to the centerline pod and
179 backseat control console. An operator interface panel has been installed in the rear cockpit to
180 allow power on/off/failure interface to each scientific instrument. Additionally, the pilot has a
181 payload master power switch that can remove all electrical power from the NASA payloads in
182 the event an abnormal electrical condition is encountered.

183

184 Multiple redundant Garmin G600/G530/G430/G696 systems record and display position,
185 attitude, heading, altitude, true airspeed, groundspeed, true air temperature, wind speed, wind
186 direction, and a wide variety of additional data through dual digital air data computers. This

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188 information is recorded for science use. A digital autopilot system allows highly accurate
189 heading and track control via GPS steering, plus precise altitude control during air sampling
190 missions. AJAX flights can also be followed in real-time using the NASA Airborne Science
191 Mission Tool Suite.

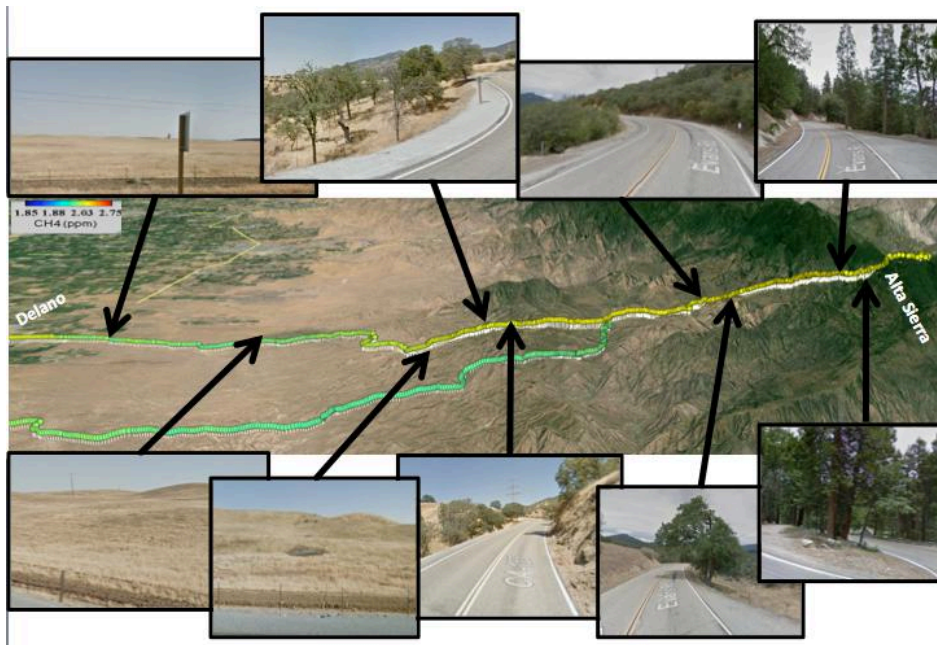
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193 Two wing-mounted pods have been modified by NASA-ARC to carry instrumentation, with
194 three down-looking window ports available on each pod. Each wing pod has an approximate
195 available volume of 0.1 cubic meter, with a maximum payload weight of 136 kg. The centerline
196 pod has two payload areas of approximately 86.4 x 25.4 x 30.5 cm and 68.6 x 16.5 x 25.4 cm,
197 carrying combined payloads up to 136 kg total.

198

199 **S4. Upwind Profile**

200 An upwind pre-survey east-west transect was conducted by AMOG from Delano (~70 m) on the
201 floor of the San Joaquin Valley to Alta Sierra (~1750 m) on the ridge of the Greenhorn
202 Mountains in the Sierra Nevada Mountain Range (Fig. S5). This survey passed through a range
203 of surface topography and vegetation and canopy types. Example Google Maps “street images”
204 show variation from flat grasslands to rolling grass covered hills, to scattered low oak trees, to at
205 the highest altitudes, dense, tall pine forests. The road shifts from an initial gradual rise while
206 following a primarily straight and gently curved pathway, to steeper climbs cut into steep slopes
207 with sharp curves, and even hairpin curves.



208
 209 **Figure S5** – Sierra Nevada Mountain Range vertical profile, and Google maps street images
 210 showing changing terrain. [Data key on figure.](#)

211

212 **S5. Derivation of the background data curtain**

213 The background data plane was calculated from the probability distributions on the left and on
 214 the right side of the transects at each altitude. Then, the CH₄ for the peak of each distribution is
 215 assigned to the left and right side of each altitude transect and linearly interpolated. Finally, the
 216 distribution is vertically interpolated to fill in the background data plane (Fig. S6).

217

218 The background data plane (Fig. S6) for transect $\gamma-\gamma'$ (Fig. 7) showed a trend of increasing CH₄
 219 towards the west, rising more than ~25 ppb, at both the surface and at 480 m altitude. In contrast,
 220 background CO₂ across the data curtain was quite uniform.

221

222 Anomaly concentration was relative to the background concentration curtain (Fig. S6a & 6b) and
 223 was derived by estimating the background concentration at each transect altitude from fitting a

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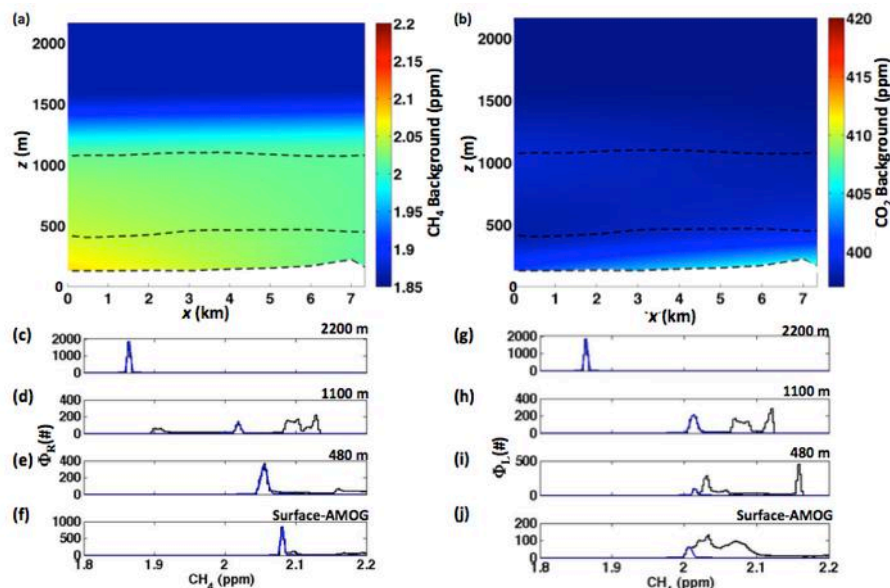
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224 Gaussian to the background occurrence concentration distribution and using the distribution peak
225 as the background concentration (Fig. S6c-S6f). The methodology is described in Sect. 2.4.



226
227 **Figure S6** – Background (a) methane (CH_4) and (b) carbon dioxide (CO_2) data curtain with
228 respect to lateral east distance (x) relative to 119.0023°W , 35.3842°N for data plane $\gamma\text{-}\gamma'$ and
229 altitude (z). Dashed line shows data altitudes. (c-f) CH_4 left side probability distribution and (g-j)
230 CH_4 right side probability distributions (Φ).

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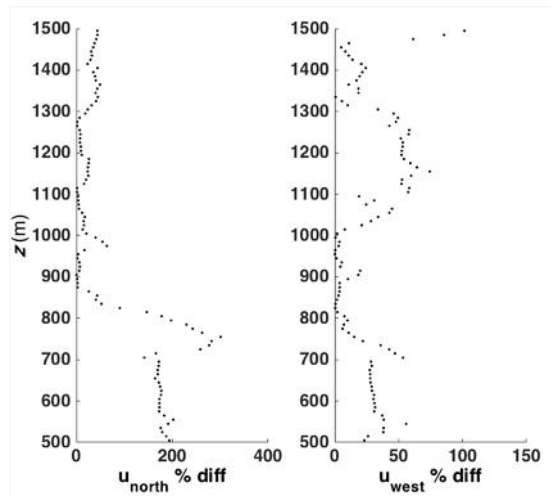
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Figure S7 – Comparison between AMOG (a) u_{north} , (b) and u_{west} .

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Mobile Monitoring of Methane Leakage

BY IRA LEIFER, PH.D., CHRISTOPHER MELTON, MANISH GUPTA, PH.D. AND BRIAN LEEN, PH.D.

Abstract

It is critical to monitor methane leakage from oil exploration and distribution pipelines for safety, climate-related issues, and profitability. Conventional monitoring technology, which involves collecting discrete air samples, provides neither spatial nor temporal resolution. With the recent advent of portable, laser-based analyzers, it is now possible to make real-time, high-fidelity mobile measurements of greenhouse gases and pollutants. In this paper, we describe the development of an AutoMOBILE greenhouse Gas Survey Platform (AMOG Surveyor) that includes trace gas analyzers, a global positioning system, a sonic anemometer, and a real-time data visualization package. Sample deployment data are presented to illustrate the ability of the AMOG Surveyor in identifying and locating natural gas leaks, and rapidly assessing their severity.

Introduction

With the dramatic increase in worldwide natural gas production and consumption, there has been strong, renewed interest in monitoring methane leakage for safety and climate-related issues as well as product loss and profitability. From years 2002 to 2012¹, there were approximately 800 significant natural gas pipeline incidents in the United States, including more than 250 explosions that killed over 100 people, injured over 450 others, and caused more than \$800M in damages. The number of incidents is expected to increase as the distribution infrastructure continues to age while being stressed further by expanding domestic production from unconventional gas sources. In addition to safety issues, methane is a potent greenhouse gas with a global warming potential that is 21 times larger than

carbon dioxide on a century timescale, and 100 times that of carbon dioxide on a decadal time-scale. In 2011, 69 billion cubic feet of methane is estimated to have leaked to the atmosphere in the U.S. alone from the natural gas distribution process² at a cost of approximately \$1B, with other estimates suggesting leaked amounts and costs that are 10 times larger. Moreover, a recent study³ suggests that the EPA may be underestimating methane emissions during the gas drilling phase by factors of 100 to 1000. This represents a significant impact on the bottom line, and may suggest that natural gas generated power would cause a greater climate impact than even that generated from coal burning, absent aggressive application of technology to prevent such leakages.

Traditionally, researchers have studied methane from leakage and other sources by either discrete, atmospheric flask samples, which are analyzed in the laboratory, or at a fixed measurement station with concentrations determined by gas chromatography coupled with a flame ionization detector (GC-FID). By their nature, such measurements are very limited in both spatial and/or temporal resolution.

Recently, with the advent of laser-based gas analyzers, continuous and rapid atmospheric methane measurements became possible, allowing researchers to measure temporal trends in methane concentrations. However, these analyzers are typically deployed in laboratories or environmental sheds, and, thus, still provide little to no spatial resolution. By correlating the methane readings with wind speed and direction (e.g., the eddy flux technique), a stationary analyzer can sometimes provide spatial resolution over a few square kilometers; however, more extensive spatial mapping is required to identify

and characterize methane leakage, and fairly strict locale, emission, and wind conditions must be met. Thus, for all of these measurements, deconvolving the effects of temporal and spatial variability in source strength from temporal and spatial variability in transport is very challenging. Similarly, temporal variability, like shifting winds, pulsed emissions, or multiple shifting emission sites, invalidates underlying assumptions of eddy flux measurements.

Development of a Mobile Monitor

To get at the underlying emissions and sources, separating the temporal and spatial variability from transport is critical, and addressed by rapid mobile measurements – installing a methane analyzer in a mobile platform (e.g., car, boat, aircraft, etc...). As noted above, GC-FID is poorly suited for this application due to its slow analysis speed (e.g., minutes), need for consumable gases, vibration sensitivity, and maintenance requirements. The former is especially limiting, since, in a vehicle travelling $60 \pm$ miles/hour, GC-FID provides spatial resolution of about a mile, useless for methane leak localization and mitigation, although useful for large-scale applications, like satellite validation⁴. Fast mobile spectral measurements, can provide the snapshot “image” of plumes of methane or other gases needed to detangle transport processes from the underlying leakage emissions. The mobile monitor described below combines an Off-Axis Integrated Cavity Output Spectroscopy (ICOS)⁵, laser-based Fast Greenhouse Gas Analyzer (FGGA) with a global positioning system (GPS), sonic anemometer (to measure wind direction and speed), system diagnostic monitoring, and custom realtime data visualization software installed in a small commuter vehicle for real-time, spatially-resolved methane measurements. Secondary trace gases are measured by additional ICOS instruments, such as ammonia and NO_2 , to discriminate

between methane sources, such as dairies and combustion i.e., diesel trucks and other heavy machinery; key given that road measurements have a bias towards combustion emissions. The system is termed an AutoMOBILE greenhouse Gas Survey or AMOG Surveyor.

Off-Axis ICOS Fast Greenhouse Gas Analyzer

The Fast Greenhouse Gas Analyzer (FGGA) manufactured by Los Gatos Research is suited for the necessary rapid detection at high accuracy required by mobile platforms. The analyzer, shown in Figure 1, uses Off-Axis Integrated Cavity Output Spectroscopy (Off-Axis ICOS) to measure methane (CH_4), carbon dioxide (CO_2), and water vapor (H_2O) at 5 to 10 Hz. Such high speed is critical to obtain sufficient spatial resolution while moving at typical highway velocities in order to collect repeat data on quicker than typical atmospheric change times (hour-scale). The technique has been described in detail previously⁶, and only a brief overview will be provided below.



Figure 1. LGR Fast Greenhouse Gas Analyzer that utilizes Off-Axis ICOS to accurately quantify methane, carbon dioxide, and water vapor aboard a mobile platform at rates up to 10 Hz

Two diode lasers operating near 1600 nm and 1650 nm for CO_2 and $\text{CH}_4/\text{H}_2\text{O}$ detection respectively are coupled into a high-finesse optical cavity consisting of two highly-reflective mirrors ($R \approx 99.99\%$). Light transmitting through the cavity is focused onto an amplified detector. A data control, acquisition, and analysis system tunes the lasers over a small spectral range ($1 - 3 \text{ cm}^{-1}$) at 300 Hz, digitizes the detector signal, averages 30–300 spectra (for 1–10 Hz response), and analyzes the cavity-

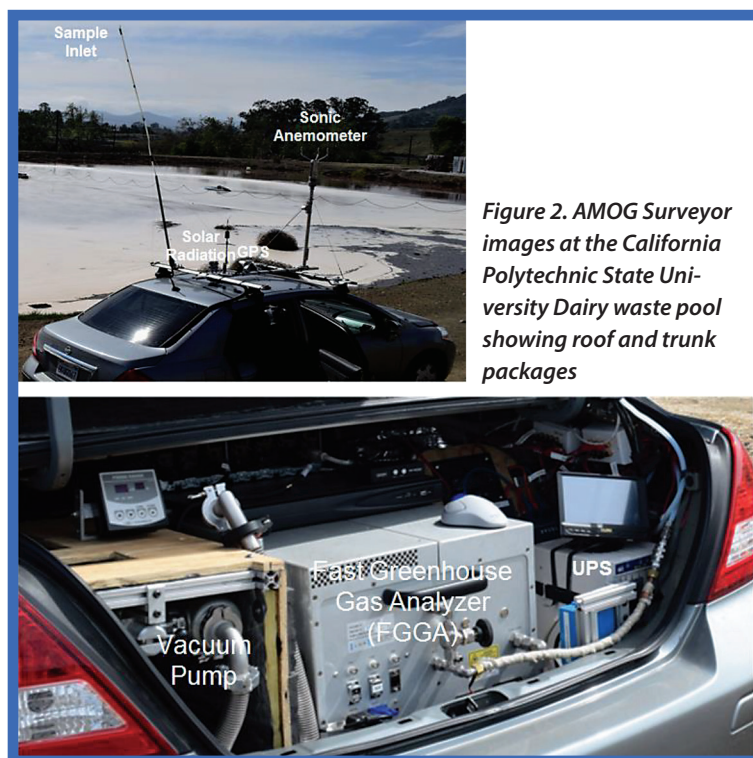


Figure 2. AMOG Surveyor images at the California Polytechnic State University Dairy waste pool showing roof and trunk packages

enhanced absorption spectra to determine the trace gas concentrations—in this case, CH₄, CO₂, and H₂O. A pump pulls sample through the cavity, while the cell pressure is actively regulated to attain high precision. By varying the pump flow, the analyzer can provide effective data rates ranging from 10–0.1 Hz, sufficient to meet both mobile and stationary (low power) monitoring applications.

In Off-Axis ICOS, the laser trajectory into the cavity is not critical and the resulting analyzer is not affected by small changes in optical alignment due to vibration, shock, and thermal stresses. The mirror coatings are made from metal oxides and do not degrade with time or chemical contact. Finally, although the gas sample is filtered, the mirrors can be removed, cleaned, and replaced in the field by a minimally trained operator if necessary.

AMOG Surveyor

The AMOG Surveyor is a Nissan Versa commuter car that has been modified for scientific trace gas surveys (see Figure 2). The Surveyor, which was developed based on tens of thousands of kilometers of data collection experience, has been designed to facilitate effective adaptive surveys for real-time, trace gas plume characterization. The Surveyor includes subsystems for power management, sample gas handling, gas analysis, thermal management, ancillary measurements, data communication, and real-time software.

Power Management

In order to obtain high spatial resolution, real-time data at high travel speed (e.g., 60 mph), the AMOG Surveyor utilizes a high-flow scroll pump (8 CFM or 30 CFM) to pull sample through the analyzers. This pump requires 600–900 W during operation and 3–4 kW during startup. In order to accommodate this pump and the other operating analyzers aboard the AMOG Surveyor, a 2.8 kW inverter is installed in the wheel well connected to two 100 Amp hr deep cycle – solar batteries that are capable of sourcing the startup surge power. All DC wiring utilizes 0/4A gauge cable, and the vehicle alternator is upgraded to provide 110 amps at idle and 220 amps (2.5 kW) at 1500 rpm. Except for the pump and 12V DC lighting and fan systems, all

other power is routed through an uninterruptible power supply, with a regulated 12VDC 120W power source for peripherals and other instruments to isolate from electrical noise sources

Gas Handling

As noted above, sample is pulled through the system using a scroll pump coupled to the analyzers using KF25 fittings to maximize gas conductance. Gas flow to the analyzers is controlled by adjustable throttle valves and bypass valves to optimize the sample measurement pressure, maintain a 5-10 Hz measurement rate and a short flow time from inlet to instrument. Sample is routed from a flexible mast atop the Surveyor to the analyzers using ½" stainless steel and Teflon® tubing. The latter is heated to 60°C to prevent condensation and allow for rapid sampling of gases that readily absorb onto surfaces (e.g., NH₃). The sampling height ranges from 3 to 5 m above ground level when the vehicle is moving fast or is stationary, respectively. In order to prevent the gas handling and analytical systems from fouling, a course filter is mounted at the sampling tube inlet, while a finer, 1 µm filter is mounted before splitting to the various analyzers.

Gas Analysis

As described in detail above, the ambient air is quantified using an Off-Axis ICOS Greenhouse Gas Analyzer that can provide measurements of CH₄, CO₂, and H₂O at up to 10 Hz with 1σ, 5s precisions of 1 ppb for CH₄, 0.15 ppm for CO₂, and 100 ppm for H₂O — the latter being useful for mapping air mass shifts. Additional cavity-enhanced analyzers also are simultaneously utilized in the AMOG Surveyor, including NH₃ and NO₂ analyzers (Los Gatos Research). The Surveyor has space, power, and gas handling support for a fourth and potentially fifth instrument as well.

System Thermal and Noise Regulation

In order to address the significant heat load generated by the scroll pump, an insulated compartment separates the pump from the other AMOG subsystems. This pump, inverter/battery compartment, and FGGA compartments are ventilated by 250, 90, and

250 CFM of forced air, respectively. Thermocouples continuously monitor the temperatures of the pump, inverter, analyzers, and vehicle compartments, and these readings are logged to confirm proper operation of cooling systems. For hot weather sampling missions, an auxiliary trunk air conditioner is incorporated, while windows are tinted.

Thermal insulation is coupled with noise insulation and reduction (e.g., pump mounts) that underlay vehicle choice. Specifically, an open vehicle (e.g., SUV or van) exposes the driver to significant noise, and as a result, the system may not be used as often. Current instrument cabin noise levels are comparable to road tire noise at 25 mph, highly tolerable, and thus facilitate continuous data collection.

Peripheral Measurements

Methane concentrations coupled with GPS coordinates alone are insufficient to identify the location and magnitude of methane sources. It is critical to correlate this spatially-resolved data with wind direction and speed. Thus, the AMOG Surveyor is equipped with a sonic anemometer that measures wind direction and speed with an accuracy of better than 2° and 0.1 m/s (or 1% of combined wind and vehicle velocity). To avoid vehicle flow streamline contamination of wind data, the anemometer is mounted 1.3 m above the vehicle roof (~3 m above ground) on a roof rack on a 1" diameter, stainless steel tube, which is three-way braced to reduce vibrations. The anemometer is surrounded by a wire cage, which does not affect wind measurements and provides protection against small branches and leaves. The AMOG Surveyor also includes a fiber-coupled solar spectrometer to support airborne remote sensing data over the UV-NIR range of 180-1080 nm.

Data Communication

In order to provide real-time mapping of the measured results, the AMOG Surveyor must be in constantly connected to the internet to download satellite imagery and maps; however, Google Earth allows up to 2 gigabytes of cached data. Thus, the Surveyor includes an amplified cellular modem to assure connectivity. Internal data communications are managed via high speed Ethernet connections and RS-232 serial servers. A separate data acquisition

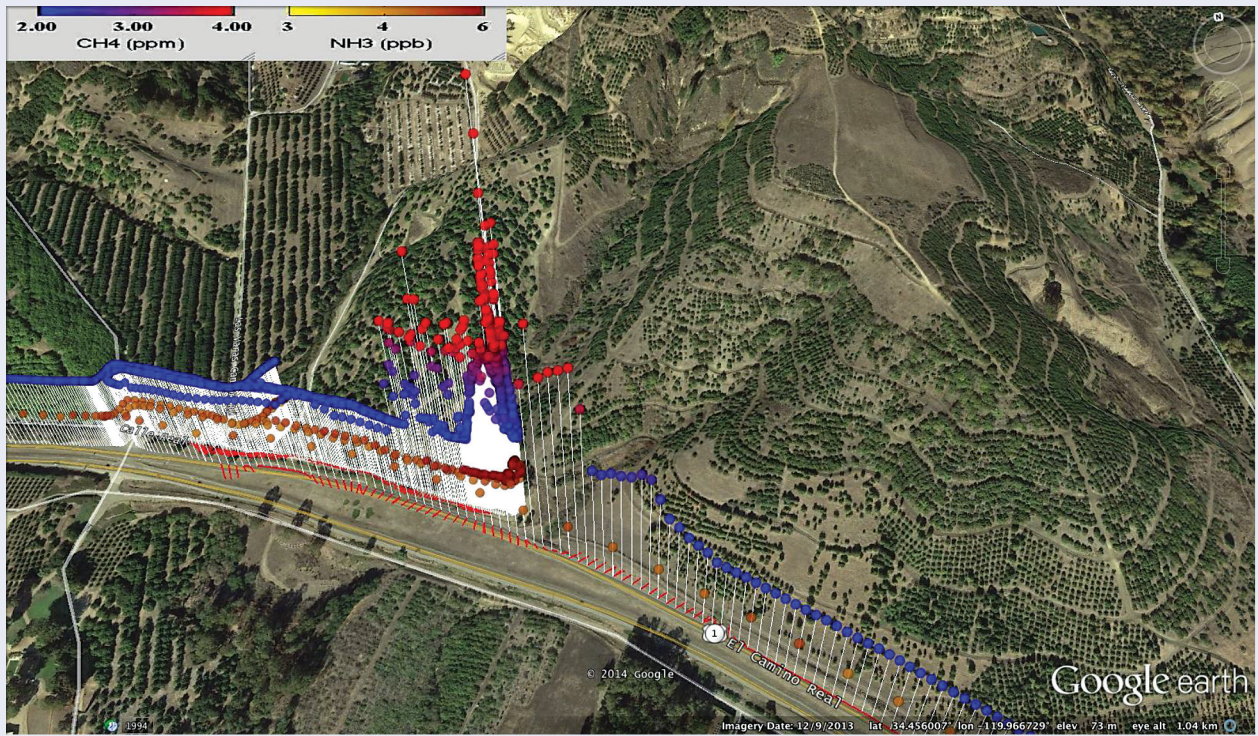


Figure 3. AMOG surveyor 2 Hz data during the discovery of a pipeline leak in an orchard while driving at highway speed on US 101; methane, CH₄, and ammonia, NH₃, are shown as color bars that are clamped to limits indicated in the legend. The altitude of the data point is also proportional to CH₄. Wind bar color and length scale with speed.

system monitors the battery voltages, current flows, and GPS output at 1 Hz. Finally, the Surveyor is equipped with four, high-definition video cameras that record continuously and provide fore, aft, and side views from the vehicle. These video data can be correlated to specific measurements to help interpret methane sources, as well as enhancing vehicle security.

Visualization and Data Logging Software

Realtime data visualization is key to enabling successful adaptive surveying. Visualization of the spatial relationships between gas concentrations, winds, and other key parameters must be easily comprehensible to enable rapid and effective survey decisions. AMOG software pushes key data to Google Earth at 0.25 Hz to enable real-time visualization of multiple measurement components centered on the vehicle with automatic camera view settings. Older data is represented more transparently to facilitate clarity when doubling back. Near publication quality data

visualizations are archived continuously, simplifying data reporting.

Flexibility and robustness in data logging are accomplished by custom software that communicates with all instruments aboard the Surveyor, using an asynchronous logging system based on a NMEA tagging protocol. This system provides robust data collection, even in cases where some signals are periodically lost (e.g., GPS) or mis-formatted. Serial buffers are frequently “flushed” (1 Hz or faster) to ensure that data recording times are well-defined: better than 0.1 s or equivalent to the best GPS accuracy of ~3 m while collecting data at fast highway speeds (30 m/s). Post-processing routines interpolate data to the FGGA data acquisition rate of 5 or 10 Hz, as well as identify and interpolate obvious outliers, and use more advanced digital filters to improve spatial and temporal resolution and suppress noise. In the case of lost temporary communication with instruments, internally logged data is merged with the tag stream in post processing to generate a seamless data product.

Sample Data

An example of a pipeline leak that was detected during a highway speed (25 m/s:~56mph) survey is shown in Figure 3, which was collected at 04:35 on the 10th of April 2014 UTC. While driving north on US 101, typical background methane levels suddenly rose to a peak of 8,900 ppb with the realtime wind vectors indicating a source to the north. Winds were light, ~0.5 m/s (~1.1 mph) and shifted from offshore to onshore shortly before the plume as the highway crossed into a valley area, highlighting a recirculation pattern where the methane plume was re-transported back onshore.

Based on Google Earth imagery, a potential turnoff was located ~1/2 a mile further, allowing AMOG to slow safely and exit the highway to investigate further. Driving along an access road, methane levels of up to 21,000 ppb were re-encountered in an orchard close to a sign indicating a buried pipeline. Ammonia levels rose slowly rather than sharply—likely due to fertilizer. CO₂ levels (not shown) suggested some CO₂ in the natural gas leak, while NO₂

data confirmed the leak was unrelated to combustion. The post-processed data shown in Figure 3 are for 2 Hz. In contrast, realtime data is 0.25 Hz, which lessens visualization clutter, particularly for visualizations with multiple gases.

Future Work

Mobile monitoring shows great promise for identifying and containing methane leakage and emissions for both climate change mitigation and minimization of product loss. Future efforts will include making more compact and lower power mobile analysis systems for applications where speed is less critical, while measuring a wider array of components. One very promising concept involves simultaneous methane, ethane, and ammonia measurements to distinguish or even detangle methane emissions from pipeline leakage, landfills, dairies, and other sources that each contains specific ethane-to-methane and other gas ratios. Similarly, adding measurements of other greenhouse gases (e.g., carbon dioxide and nitrous oxide) and pollutants (e.g., carbon monoxide and nitrogen dioxide) can help elucidate climate variables and distinct pollution sources, while discriminating against road vehicular emission biases. Finally, multiple mobile monitors may be synchronized to provide wide-scale, real-time mapping of gas sources and leaks yielding far more accurate source strength and location data. **G&I**

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7. 1σ , 5s represents one measurement standard deviation with 5 seconds of data averaging.



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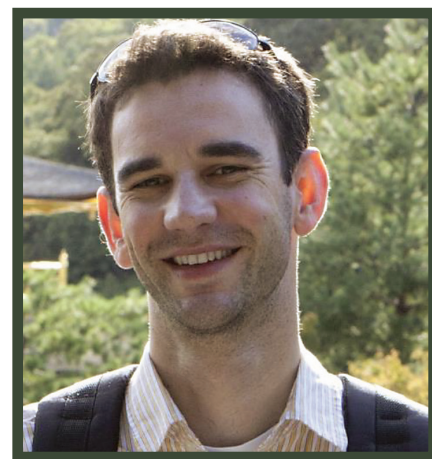


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