

Answers to reviewer's comments

Reviewer 1.

(1) Page 6, Lines 16–17: What is the sampling time? 10-min?

A: The sampling time was indeed 10 minutes. The information was added to the text.

Text was modified on page 6, line 23: Static and dynamic collections were performed during 10 minutes.

(2) Page 7, Line 2: Does it lead to only “underestimation”? Doesn't it happen to overestimate?

A: In the particular case of α -pinene, an underestimation will be observed in comparison with the other studied monoterpenes that are less volatile. However, it is true that an overestimation will be most likely observed if other monoterpenes with higher volatility than α -pinene are measured.

Text was modified on page 7, line 10: An underestimation of the measured amounts of a compound with higher volatility or an overestimation relatively to the most volatile compounds are then expected when quantifying monoterpenes under field conditions where temperature changes can be significant.

(3) Page 7, Lines 15–24: What is the sampling time and the temperature in the calibration? And which type of SPME was calibrated? In Table S1, the calibration curves for only one type of SPME are listed. Later, field data obtained from all types of SPME were quantified. Therefore, calibration curves for all types of SPME should be shown.

A: The calibration of the instrument response during field measurements was performed by using liquid standard solutions at different concentrations. The SPME systems were not calibrated to allow their comparison.

Text was modified on page 7, line 30: The calibration of instrument response was performed for the field measurement of monoterpenes (α -pinene, Δ^3 -carene and limonene) and aldehydes (octanal, nonanal and decanal), to estimate the mass adsorbed on the coating materials of the different SPME-based systems (Table S1). Liquid standard solutions at different concentrations were used for this purpose.

Text was modified in page 8, line 16: However, when the calibration of analytes response was performed with standard solutions (but not the calibration of SPME collection), Δ^3 -carene and limonene levels increased relatively to α -pinene (Fig. 6A).

(4) Page 9, Line 4: Is it “the ratio between nonanal and decanal”? I think that the authors did not show the results of nonanal in the laboratory experiments.

A: The first mentioned ratio is the one obtained during field measurements between decanal and nonanal. Octanal was not measured with SPME fiber. However, during laboratory experiments, the tests were performed for octanal and decanal and the second mentioned ratio is the one obtained between decanal and octanal. Even though these ratios could probably still be compared because analytes belong to the same homologous series, the sentence was eliminated since additional tests would be required to confirm this possibility.

The sentence was removed.

(5) Page 9, Line 17–Page 11, Line 5: I felt that the discussions here were vague and subjective, and were not based on statics. What is the reason of the difference between SPME Arrow (PDMS/DVB) and SPME Arrow (PDMS/Carbon WR) in Figs. 6 and 7?

A: the whole section 3.4 was completely rewritten to emphasize that the purpose of the section was to evaluate preliminarily the effect of atmospheric parameters on SPME sampling. In Figs 6 and 7, the trends between SPME Arrow (PDMS/DVB) and SPME Arrow (PDMS/Carbon WR) were similar. The differences in the extracted masses are likely related with a combination of several factors, including the distinct extraction efficiencies, competitive adsorption and different effects of atmospheric parameters (e.g. temperature and relative humidity).

Text was modified on page 10, line 5: The effects of meteorological parameters (Table S3) on the measured atmospheric levels of monoterpenes (Fig. 9) and aldehydes (Fig. 10) were also tentatively evaluated in this study, since some of these parameters can influence not only the VOC emissions/atmospheric mixing ratios but also the adsorption on the SPME sorbent materials. The ratios between the amounts of monoterpenes and aldehydes sampled with PDMS/Carbon WR and PDMS/DVB were also compared with the referred parameters, in order to understand if meteorological conditions affect differently the sampling with these materials (Fig. S8).

Temperature has two opposing effects in field measurements performed at boreal forest sites. Increased temperature enhances VOC emissions from Scots pine (Tarvainen et al., 2005). In addition, it reduces the distribution constant of the analytes because adsorption is an exothermic process (de Fatima Alpendurada, 2000). During the sampling campaign, temperature remained almost constant, limiting the comparison between this parameter and the measured amounts of monoterpenes. The effect of temperature was then expected to be small when compared to other parameters, which was verified in our results. However, longer data sets encompassing periods of greater temperature variation are needed to perform this evaluation.

Relative humidity (RH) and precipitation also have two opposing effects, since monoterpene emission rates not only increase at high humidity levels and during and after precipitation (Llusà and Peñuelas, 1999; Schade et al., 1999) but also causes a small decrease in the SPME extraction capacity (section 3.1.3). Oppositely to temperature, relative humidity varied considerably during the sampling campaign, which allowed to study the effect of this parameter in our results. An increase in monoterpene amounts was observed during a precipitation event when relative humidity was high, showing that the decreasing effect of these parameters on the sorbent collection is less significant when compared to their increasing

effect on VOC emissions. A correlation was not found when considering the ratio between the amounts of monoterpenes sampled with PDMS/Carbon WR and PDMS/DVB, which is consistent with the small humidity dependence observed for the sampling with these materials under laboratory controlled conditions.

Due to the constancy of temperature during the sampling campaign, ozone and PAR were also expected to affect significantly the measured amounts of monoterpenes. Indeed, some anti-correlation was found between the measured monoterpenes and these parameters. This result is likely to reflect the increased photooxidation during periods of the day when PAR is high, since the effects of temperature and/or light on monoterpene emissions have been described previously (Aalto et al., 2014). Nonetheless, on-fiber oxidation might occur during SPME collection. For that reason, the effect of oxidants must be assessed further by performing complementary laboratory experiments under controlled conditions. Ozone and PAR did not affect distinctively the adsorption on the two different materials used in this study. PNC also seemed to increase with the amounts of monoterpenes present in the ambient air. This result is expected, since monoterpene oxidation in the atmosphere and consequent formation of low volatile compounds have been recognized to contribute to aerosol particle formation (Laaksonen et al., 2008). However, other factors can also contribute to the increase in PNC, such as long-range transport.

With reference to aldehydes, the studied parameters seemed not to influence negatively the SPME collection. A similar trend between aldehyde amounts and temperature was observed during the most of the sampling period, showing that temperature did not affect the SPME sampling and suggesting the existence of a temperature-dependence on aldehyde emissions. However, as referred previously, temperature remained almost constant during the sampling campaign and additional studies under controlled conditions are required to confirm this evidence. Relative humidity and precipitation also coincided with a burst in aldehyde atmospheric amounts, excluding a negative effect of this parameter on the SPME sampling, but seemed to anti-correlate with these parameters when aldehyde amounts were low. This observation can be a consequence of the solubility of these compounds in water at low concentrations, but additional studies are also required to confirm this hypothesis. A correlation was not found between PAR and aldehydes. However, contrary to monoterpenes, some correlation with ozone was observed, which was expected since increased emissions have been reported in another study when vegetation was exposed to ozone (Wildt et al., 2003). On-fiber oxidation studies are also still required for aldehydes. The effect of aldehyde amounts on PNC was not very clear, which might be a consequence of the lower atmospheric reactivity of these compounds. No correlation was observed between atmospheric parameters and the ratio between the amounts collected on PDMS/Carbon WR and PDMS/DVB, which agrees with the non-dependences on temperature and relative humidity verified in laboratory studies.

Even though the effects of atmospheric parameters on the SPME sampling were preliminarily evaluated with our method under atmospheric relevant conditions, longer data sets and quantitative data are needed to estimate accurately the correlation of these parameters with BVOC mixing ratios.

(6) Page 10, Lines 6–7: What is the reason of the anti-correlation between the measured monoterpenes and PAR?

A: At summer time, the main factor determining monoterpene emissions at the boreal forest is temperature due to the temperature-dependent residence in specific storage structures located internal or external to

the leaf. However, a PAR effect on monoterpene emissions was also found to be important (e.g. at spring-time when temperature is still low). In our study, the anti-correlation between PAR and monoterpenes is most likely not related with lower emissions from vegetation. Higher PAR is usually observed in the beginning of afternoon, which usually coincides with the period of the day with higher atmospheric reactivity. This reactivity is most likely the main reason why a correlation between PAR and monoterpenes is not observed. A discussion was added to the text.

Text was modified in page 10, line 27: Due to the constancy of temperature during the sampling campaign, ozone and PAR were also expected to affect significantly the measured amounts of monoterpenes. Indeed, some anti-correlation was found between the measured monoterpenes and these parameters. This result is likely to reflect the increased photooxidation during periods of the day when PAR is high, since the effects of temperature and/or light on monoterpene emissions have been described previously (Aalto et al., 2014).

(7) Page 10, Lines 13–14: This sentence is inconsistent with the previous two sentences. Explain the reason more carefully.

A: The sentence was revised.

Text was modified in page 10, line 33: PNC also seemed to increase with the amounts of monoterpenes present in the ambient air. This result is expected, since monoterpene oxidation in the atmosphere and consequent formation of low volatile compounds have been recognized to contribute to aerosol particle formation (Laaksonen et al., 2008). However, other factors can also contribute to the increase in PNC, such as long-range transport.

(8) Page 10, Lines 21–22: Is negative effect of RH and precipitation on the amount of aldehydes surely “evident”? Discuss more carefully.

A: A more careful discussion was added to the text.

Text was modified in page 11, line 8: Relative humidity and precipitation also coincided with a burst in aldehyde atmospheric amounts, excluding a negative effect of this parameter on the SPME sampling, but seemed to anti-correlate with these parameters when aldehyde amounts were low. This observation can be a consequence of the solubility of these compounds in water at low concentrations, but additional studies are also required to confirm this hypothesis.

(9) Page 10, Line 26: Insert “not”: . . .any correlation was “not” observed. . . Am I right?

A: The sentence was corrected.

Text was modified in page 11, line 15: The effect of aldehyde amounts on PNC was not very clear, which might be a consequence of the lower atmospheric reactivity of these compounds. No correlation was

observed between atmospheric parameters and the ratio between the amounts collected on PDMS/Carbon WR and PDMS/DVB, which agrees with the non-dependences on temperature and relative humidity verified in laboratory studies.

(10) Page 14, Figs. 1 and 2: In the results of SPME Arrow (PDMS/DVB), the values of peak area of α -pinene are larger than those of β -carene. According to kinetic data shown in Figs. S2 and S3, the values of peak area of α -pinene are smaller than those of β -carene. Are they consistent?

How were the error bars calculated? Define them.

A: The concentrations of analytes were not changed during all experiments and it is true that α -pinene was expected to be slightly higher during SPME Arrow (PDMS/DVB) kinetic studies and during comparison between extraction efficiencies. However, considering the error bars, the amounts of α -pinene and β -carene were quite similar when using a PDMS/DVB sorbent and results are still consistent with the SPME fiber coated with the same extraction material (Figs. S2, S4 and S7).

The error bars in Figs. 1 and 2 are the standard deviations from 3 repetitions. In Figs. S2-S3, three repetitions were performed at 10 minutes and the obtained RSD was used for all the calibration. The RSD will be likely higher at lower sampling times and lower for the longer sampling times. However, because most of the laboratory tests were performed at 10 minutes (except for temperature and RH studies that were performed at 20 minutes) and field measurements were performed at longer sampling times (45 minutes), the used RSD is representing a worst-case scenario in the kinetic studies.

Text was modified in page 5, line 24: For an estimation of standard deviations during kinetic studies, 3 repetitions were performed at 10 minutes, corresponding to the minimum sampling time used during all other experiments and consequently to the higher expected variation.

Text was modified in page 6, line 32: Three replicates were performed for each temperature and relative humidity studied.

(11) Figs. S2–S7: How were the error bars calculated? Define them.

A: The error bars in Figs. S2-S3 were calculated as described in the previous question. In Figs S4-S7, three repetitions were performed and the obtained standard deviations were used as error bars.

Text was modified on page 5, line 24: For an estimation of standard deviations during kinetic studies, 3 repetitions were performed at 10 minutes, corresponding to the minimum sampling time used during all other experiments and consequently to the higher expected variation.

Text was modified on page 6, line 15: Three repetitions were performed with each SPME system.

Text was modified on page 6, line 32: Three replicates were performed for each temperature and relative humidity studied.

(12) Fig. S11: The data of meteorological parameters seems to be different from those in Figs. 6 and 7. Is it just a careless mistake?

The title of the left axis should be “ratio”. It should not be “mass adsorbed (ng)”..

A: There was a mistake in the x-axis (date) that was corrected. The title of the y-axis was modified to Carbon WR/DVB (ng/ng).

Text was modified on page 6, line 1 (supplement):

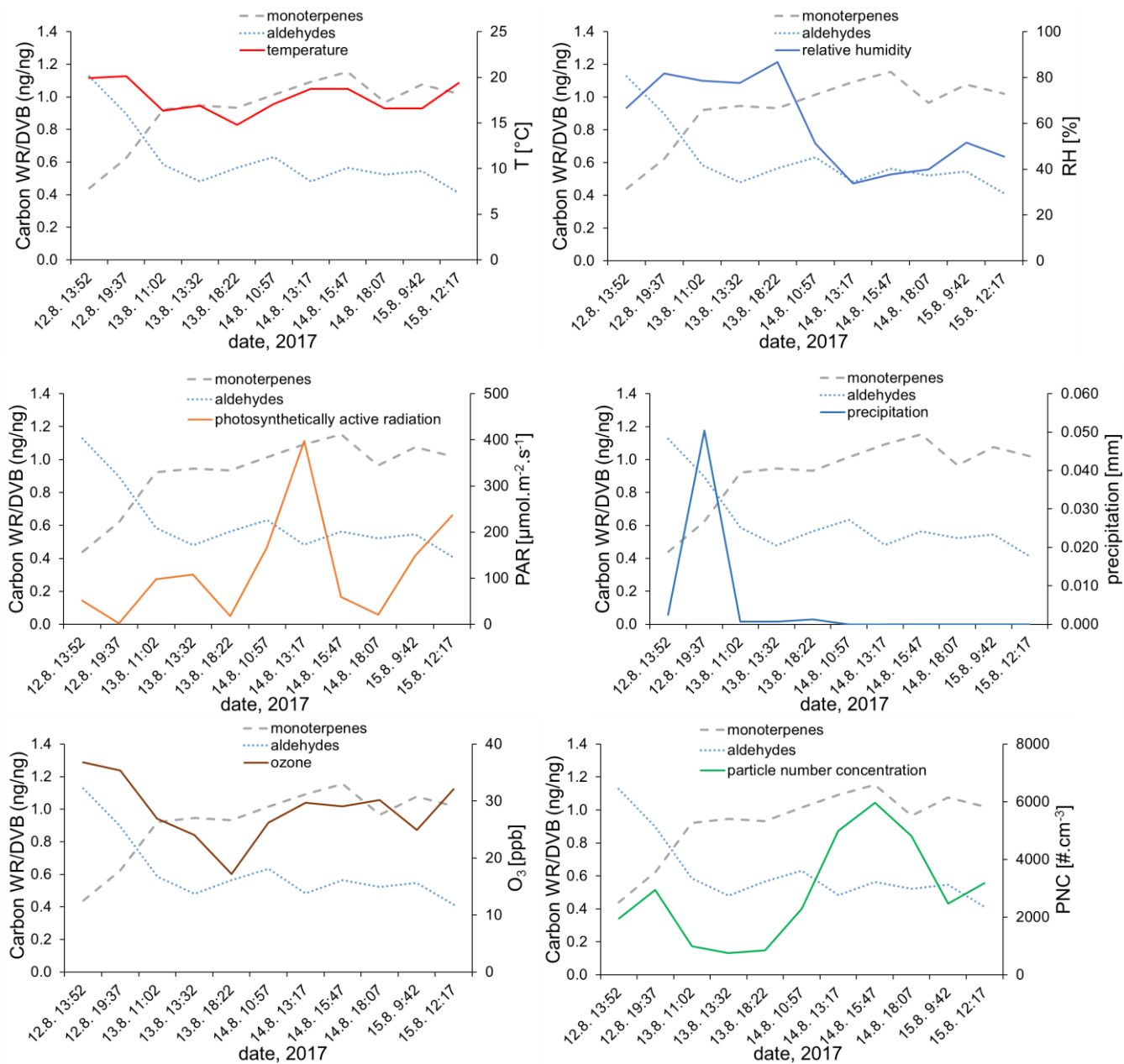


Figure S8: Effect of temperature (°C), relative humidity (%), photosynthetically active radiation ($\mu\text{mol}\cdot\text{m}^{-2}\cdot\text{s}^{-1}$), precipitation (mm), ozone (ppb) and particle number concentration ($\#\cdot\text{cm}^{-3}$) on the ratio between the amounts of monoterpenes (grey line) and aldehydes (blue line) sampled with PDMS/Carbon WR and PDMS/DVB and measured by GC-MS.

Reviewer 2.

General comments

(1) I would have liked to see a detailed characterization of the temperature effect as this C1 parameter is the main regulating force of monoterpene mixing ratios in a forested environment. The authors indicate considerable differences for α -pinene and D3-carene but I am still a bit confused on how one can deal with these effects under field conditions. I would assume that a reasonable approach would be to characterize the extraction efficiency along a wider range of temperatures and apply the respective correction under field conditions.

A: The proposed approach is one of the possibilities. However, the ideal solution would be to eliminate this effect. For that purpose, we would suggest the construction of a chamber that can be field portable and that is able to keep the temperature constant during both sampling and calibration. However, additional research is still needed to study both of these possibilities.

Text was modified on page 7, line 13: The effect of temperature on the amounts of analytes collected by SPME must consequently be assessed or avoided during quantitative field measurements.

(2) It would have been valuable if the authors could demonstrate the advantages of increased sampling capacity as a function of detection limits.

A: Indeed, it is important to develop a calibration method so that we can determine the detection limits for all used SPME systems. However, our results showed already the benefit of having increased capacity for both qualitative (e.g. we detected octanal with SPME Arrow but not with SPME fiber) and semi-quantitative measurements (e.g. errors related to the analysis of target compounds are usually higher when responses are closer to the baseline).

No modification was done to the manuscript.

(3) The amount of field data extremely low (15 points) and the assumptions on correlation (or not) with meteorological and environmental parameters should be more carefully approached and discussed. This is of particular importance especially when considering the title of the manuscript that at this stage could be misleading.

A: the whole section 3.4 has been rewritten and more carefully discussed.

Text was modified in page 10, line 5: The effects of meteorological parameters (Table S3) on the measured atmospheric levels of monoterpenes (Fig. 9) and aldehydes (Fig. 10) were also tentatively evaluated in this study, since some of these parameters can influence not only the VOC emissions/atmospheric mixing ratios but also the adsorption on the SPME sorbent materials. The ratios between the amounts of monoterpenes and aldehydes sampled with PDMS/Carbon WR and PDMS/DVB were also compared with the referred parameters, in order to understand if meteorological conditions affect differently the sampling with these materials (Fig. S8).

Temperature has two opposing effects in field measurements performed at boreal forest sites. Increased temperature enhances VOC emissions from Scots pine (Tarvainen et al., 2005). In addition, it reduces the distribution constant of the analytes because adsorption is an exothermic process (de Fatima Alpendurada, 2000). During the sampling campaign, temperature remained almost constant, limiting the comparison between this parameter and the measured amounts of monoterpenes. The effect of temperature was then expected to be small when compared to other parameters, which was verified in our results. However, longer data sets encompassing periods of greater temperature variation are needed to perform this evaluation.

Relative humidity (RH) and precipitation also have two opposing effects, since monoterpene emission rates not only increase at high humidity levels and during and after precipitation (Llusà and Peñuelas, 1999; Schade et al., 1999) but also causes a small decrease in the SPME extraction capacity (section 3.1.3). Oppositely to temperature, relative humidity varied considerably during the sampling campaign, which allowed to study the effect of this parameter in our results. An increase in monoterpene amounts was observed during a precipitation event when relative humidity was high, showing that the decreasing effect of these parameters on the sorbent collection is less significant when compared to their increasing effect on VOC emissions. A correlation was not found when considering the ratio between the amounts of monoterpenes sampled with PDMS/Carbon WR and PDMS/DVB, which is consistent with the small humidity dependence observed for the sampling with these materials under laboratory controlled conditions.

Due to the constancy of temperature during the sampling campaign, ozone and PAR were also expected to affect significantly the measured amounts of monoterpenes. Indeed, some anti-correlation was found between the measured monoterpenes and these parameters. This result is likely to reflect the increased photooxidation during periods of the day when PAR is high, since the effects of temperature and/or light on monoterpene emissions have been described previously (Aalto et al., 2014). Nonetheless, on-fiber oxidation might occur during SPME collection. For that reason, the effect of oxidants must be assessed further by performing complementary laboratory experiments under controlled conditions. Ozone and PAR did not affect distinctively the adsorption on the two different materials used in this study. PNC also seemed to increase with the amounts of monoterpenes present in the ambient air. This result is expected, since monoterpene oxidation in the atmosphere and consequent formation of low volatile compounds have been recognized to contribute to aerosol particle formation (Laaksonen et al., 2008). However, other factors can also contribute to the increase in PNC, such as long-range transport.

With reference to aldehydes, the studied parameters seemed not to influence negatively the SPME collection. A similar trend between aldehyde amounts and temperature was observed during the most of

the sampling period, showing that temperature did not affect the SPME sampling and suggesting the existence of a temperature-dependence on aldehyde emissions. However, as referred previously, temperature remained almost constant during the sampling campaign and additional studies under controlled conditions are required to confirm this evidence. Relative humidity and precipitation also coincided with a burst in aldehyde atmospheric amounts, excluding a negative effect of this parameter on the SPME sampling, but seemed to anti-correlate with these parameters when aldehyde amounts were low. This observation can be a consequence of the solubility of these compounds in water at low concentrations, but additional studies are also required to confirm this hypothesis. A correlation was not found between PAR and aldehydes. However, contrary to monoterpenes, some correlation with ozone was observed, which was expected since increased emissions have been reported in another study when vegetation was exposed to ozone (Wildt et al., 2003). On-fiber oxidation studies are also still required for aldehydes. The effect of aldehyde amounts on PNC was not very clear, which might be a consequence of the lower atmospheric reactivity of these compounds. No correlation was observed between atmospheric parameters and the ratio between the amounts collected on PDMS/Carbon WR and PDMS/DVB, which agrees with the non-dependences on temperature and relative humidity verified in laboratory studies. Even though the effects of atmospheric parameters on the SPME sampling were preliminarily evaluated with our method under atmospheric relevant conditions, longer data sets and quantitative data are needed to estimate accurately the correlation of these parameters with BVOC mixing ratios.

Specific comments

(4) P1L19. If there is a technical possibility to additionally evaluate the effects of ozone under relevant atmospheric conditions it would have been a valuable addition to this study.

A: A technical possibility to evaluate the effects of ozone under relevant atmospheric conditions would be to generate different oxidants in a chamber/flow tube and perform measurements, maintaining all other conditions constant. This approach would be very challenging, since adding oxidants to the chamber will also decrease the analytes concentration. There are also some possibilities that could possibly be used with our method to avoid the effect of ozone and other oxidants on the measured amounts of VOCs (e.g. the use of an ozone scrubber or a heated stainless steel tube connected to our sampling device). These studies must be performed further.

Text was modified on page 10, line 31: Nonetheless, on-fiber oxidation might occur during SPME collection. For that reason, the effect of oxidants must be assessed further by performing complementary laboratory experiments under controlled conditions.

Text was modified on page 11, line 14: On-fiber oxidation studies are also still required for aldehydes.

(5) P2L26-L27. I don't think that the conventional GC-MS techniques are "laborious, expensive and prone to contamination". GC-MS is the most widely used technique for quantification of ambient monoterpenes, with no interferences on temperature or wind speed as demonstrated for SPME. Therefore, I would suggest removing this sentence completely.

A: The sentence was removed and the main advantage of SPME compared to TD-GC-MS was referred.

Text was modified on page 2, line 24: Alternatively, monoterpenes have been successfully sampled on tubes packed with an adsorbent material (such as Tenax TA/Carbopack-B), and subsequently desorbed into a thermal desorption-gas chromatograph-mass spectrometer for off-line or on-line analysis (e.g. Haapanala et al., 2012; Hakola et al., 2012; Wong et al., 2013). The main limitation of this method is the requirement of sophisticated instrumentation that is less convenient for field measurements (e.g. thermal desorption unit and cryofocusing). Solid-phase microextraction (SPME) has also been used for the collection of monoterpenes (e.g. Yassaa et al., 2010; Zini et al., 2001). This technique combines sampling and pre-concentration of analytes in a single step and allows for direct thermal desorption into a heated gas chromatograph injection port (Koziel et al., 1999).

(6) P4L16. I was wondering if the sampling time could be reduced and if you could explain the reasons that you have chosen this approach. 45minutes is long time for sampling ambient mixing ratios that can dramatically change in such time frame.

A: The sampling time could likely be reduced, especially when using SPME Arrow. The main reasons for using a 45 minutes sampling time were to sample detectable amounts of target analytes with all SPME systems and to reduce the significance of errors associated with the time lag between sampling and injection (even though sample modification was not expected to be significant due to the retraction of extraction materials under the SPME needle and the fact that SPME syringes were closed with a cap).

Text was modified on page 4, line 19: Samples were collected in static mode for 45 minutes, to sample detectable amounts of target analytes with all SPME systems and to reduce the significance of errors associated with the time lag between sampling and injection.

(7) P4L27. Have you tried to develop a method under SIM mode? It is commonly known that the sensitivity is much reduced when using a mass scan.

A: A SIM mode increases the sensitivity of the method. However, at this stage of development, we would like to know what type of compounds can be identified and a scan mode was then chosen for that purpose. This approach allowed to study compounds that were not considered during the laboratory tests (limonene and nonanal).

Any modification was done in the manuscript.

(8) P5 and P6. I would suggest bringing supplementary figures S2, S3 and S4 into the main paper.

A: The figures were brought to the main paper.

Text was modified on page 17, line 1: Figs. S2, S3 and S4 modified to Figs. 1, 2 and 3.

(9) P7L13-14. If the extraction efficiency is reduced at higher temperatures (where we expect higher emissions) the final results will be heavily influenced. Please elaborate in detail.

A: The sentence was corrected.

Text was modified on page 7, line 13: The effect of temperature on the amounts of analytes collected by SPME must consequently be assessed or avoided during quantitative field measurements.

Text was modified on page 7, line 25: Even though the effect of relative humidity was negligible in our study, this parameter might greatly influence the SPME collection when using other sorbents or when sampling other analytes.

(10) P7L16. A plot (rather than a table) would have been more useful.

A: In order to compact the information obtained during calibration studies, results were presented as a table. All relevant parameters were carefully provided in the table and in the text.

No modification was done to the manuscript.

(11) Fig. 4. How many samples were used? Please add errorbars.

A: The results are the average amounts measured during all the campaign. This information was added to the caption. Repetitions were not performed during field measurements, since a satisfactory repeatability was obtained during laboratory measurements and 4 different systems were compared simultaneously in the field (which would require 12 systems for 3 repetitions and a long time to finalize the measurements of each experiment). For that reason, error bars were only added to the laboratory experiments. The detailed results during all sampling campaign are represented in Fig. S8 (Fig. S5 in the new version of the manuscript) and Table S2, which show the consistent improvement in extraction efficiency when using SPME Arrow comparatively to SPME fiber during all sampling period.

Text was modified on page 23, line 4: Figure 7: Comparison between the average mass of identified monoterpenes (α -pinene, Δ^3 -carene and limonene) and aldehydes (octanal, nonanal and decanal) collected with different PDMS/DVB SPME devices (fiber and Arrow) from ambient air and measured by GC-MS.

(12) Section 3.4: The effects of meteorological parameters should be addressed in a more comprehensive and detailed manner. Some conclusions (e.g. P9L29-31, P10L3-4, P10L7-15) are drawn very easily and without strong evidence driven by data. I would suggest to completely re-write this section, using softer language and presenting the available data in a different manner. Maybe xy plots (where x is the SPME arrow PDMS/DVB and y the PDMS/Carbon WR) colored by a 3rd dimension which would be the respective meteorological parameter, would depict better both the differences between the materials and the impact of the parameters. It is however understandable that eg. temperature did not vary significantly during the sampling period but I would have liked to understand why the two materials match on the 12.8 and have 100% difference on the 14.08.

A: All section 3.4 was completely rewritten to emphasize that the main purpose of the section was to evaluate preliminarily the effect of atmospheric parameters on the SPME sampling. In Figs 6 and 7, the trends between SPME Arrow (PDMS/DVB) and SPME Arrow (PDMS/Carbon WR) were similar. The differences in the extracted masses are likely related with a combination of several factors, including the distinct extraction efficiencies, competitive adsorption and different effects of atmospheric parameters (e.g. temperature and relative humidity).

Corrected as described above.

(13) P10L11. While this may be true, no real evidence is provided that PNC was increased due to a particle formation event of not due to some transport. Please provide some evidence or revise the text accordingly. P10L25-26. As above.

A: The sentences were revised.

Text was modified on page 10, line 33: PNC also seemed to increase with the amounts of monoterpenes present in the ambient air. This result is expected, since monoterpene oxidation in the atmosphere and consequent formation of low volatile compounds have been recognized to contribute to aerosol particle formation (Laaksonen et al., 2008). However, other factors can also contribute to the increase in PNC, such as long-range transport.

Text was modified on page 11, line 15: The effect of aldehyde amounts on PNC was not very clear, which might be a consequence of the lower atmospheric reactivity of these compounds.

(14) P10L31. The authors correctly recognize but only briefly discuss the limitations of their dataset.

A: The limitations of our data set were clarified.

Text was modified on page 11, line 19: Even though the effects of atmospheric parameters on the SPME sampling were preliminarily evaluated with our method under atmospheric relevant conditions, longer data sets and quantitative data are needed to estimate accurately the correlation of these parameters with BVOC mixing ratios.

Text was modified on page 13, line 26: Longer data sets are also required to study in more detail the effects of atmospheric parameters on the SPME sampling under atmospheric relevant conditions.

(15) P11L20. It is a bit strange to see a technical paper entitled “Field measurements. . .” to use as final statement the fact that “more studies are needed to develop a proper calibration method for field measurements”. It denotes that the current manuscript did not sufficiently fulfill its purpose. Please revise.

A: The sentence was revised.

Text was modified on page 13, line 25: More studies under controlled conditions are needed to understand the influence of co-adsorbed species and oxidants on the SPME sampling and to develop a proper calibration method for quantitative field measurements.

Field measurements of biogenic volatile organic compounds in the atmosphere using solid-phase microextraction Arrow

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Abstract. Biogenic volatile organic compounds (BVOCs) emitted by terrestrial vegetation participate in a diversity of natural processes. These compounds impact both on short-range processes, such as on plant protection and communication, and on ~~high~~long-range processes, by e.g. participation on aerosol particle formation and growth. The biodiversity of plant species around the Earth, the vast assortment of emitted BVOCs, and their trace atmospheric concentrations contribute to the high remaining uncertainties about the effects of these compounds on atmospheric chemistry and physics, and call for the development of novel collection devices that can offer portability with improved selectivity and capacity. In this study, a novel solid-phase microextraction (SPME) Arrow sampling system was used for the static and dynamic collection of BVOCs from the boreal forest, and samples were subsequently analysed on-site by gas chromatography-mass spectrometry (GC-MS). This system offers higher sampling capacity and improved robustness ~~than-when compared to the~~traditional equilibrium-based SPME techniques, such as SPME fibers. Field measurements were performed in summer 2017 at the Station for Measuring Ecosystem-Atmosphere Relations (SMEAR II) in Hyytiälä, Finland. Complementary laboratory tests were also performed to compare the SPME-based techniques under controlled experimental conditions and to evaluate the effect of temperature and relative humidity on their extraction performance. The most abundant monoterpenes and aldehydes were successfully collected. A significant improvement on sampling capacity was observed with the new SPME Arrow system ~~when compared~~ ~~to~~over SPME fibers, with collected amounts being approximately 2 times higher for monoterpenes and 7-8 times higher for aldehydes. BVOC species exhibited different affinities for the type of sorbent materials used (PDMS/Carbon WR vs. PDMS/DVB). Higher extraction efficiencies were obtained with dynamic collection prior to equilibrium regime, but this benefit during the field measurements was small probably due to the natural agitation provided by the wind. An increase in temperature and relative humidity caused a decrease in the amounts of analytes extracted under controlled experimental conditions, even though the effect was more significant for PDMS/Carbon WR than for PDMS/DVB. Overall, results demonstrated the benefits and challenges of using SPME Arrow for the sampling of BVOCs in the atmosphere.

1 Introduction

Vegetation covering Earth landmasses release a diversity of biogenic volatile organic compounds (BVOCs), which comprise a large variety of molecules that differ in size, physicochemical properties and metabolic origin (Laothawornkitkul et al., 2009; Peñuelas and Llusà, 2001; Peñuelas and Staudt, 2010). Terrestrial biosphere function as one of the key regulators of atmospheric chemistry, and is fundamental for sustainability of air quality and climate (Arneeth et al., 2010; Bryan and Steiner, 2013). BVOCs participate in many natural processes, including plant metabolism, growth, reproduction, protection, and communication ~~between plants,~~ within plant communities, and between plants and insects (Laothawornkitkul et al., 2009; Peñuelas and Staudt, 2010). ~~BVOC~~These emissions vary considerably in time, space and between species, and are strongly influenced by temperature and light (Kesselmeier and Staudt, 1999; Schollert et al., 2014; Tarvainen et al., 2005). Once in atmosphere, BVOCs participate in atmospheric reactions, which leads to the formation of numerous secondary products (Atkinson and Arey, 2003). The lifetime of BVOCs vary to a large extent, depending on the compound and oxidants involved (Atkinson and Arey, 2003). The low volatile secondary products formed in these photo-oxidation reactions can subsequently ~~result in~~lead to the formation of secondary organic aerosols (SOA) (Jimenez et al., 2009). Aerosols are recognized to affect climate, both directly by reflecting or absorbing solar radiation and indirectly by acting as a cloud condensation nuclei (Kulmala et al., 2004). BVOCs are believed to be the largest source of SOA on a global scale (Henze and Seinfeld, 2006).

Monoterpenes are a class of naturally occurring compounds with great importance to atmospheric physics and chemistry. These compounds participate in photochemical reactions that affect ozone and carbon monoxide concentrations, and contribute to secondary organic aerosol formation and growth through their oxidation products (Kavouras et al., 1999; Lerdau et al., 1997). They also have an important biological role, e.g. as allelopathic and defense compounds against pathogens and herbivores (Kesselmeier and Staudt, 1999). Measurement of monoterpenes is usually performed by on-line proton-transfer-reaction-mass spectrometry (PTR-MS) (e.g. Aalto et al., 2014; Rantala et al., 2015). This technique offers fast detection of VOCs, ~~high~~good sensitivity, good time resolution and low detection limits (Graus et al., 2010). However, PTR-MS cannot differentiate compounds with the same molecular mass and it usually requires the use of long sampling lines that ~~can~~ cause sample alteration. Alternatively, monoterpenes ~~can~~have been successfully sampled on tubes packed with an adsorbent material (such as Tenax TA/Carbopack-B), and subsequently desorbed into a thermal desorption-gas chromatograph-mass spectrometer for off-line or on-line analysis (e.g. Haapanala et al., 2012; Hakola et al., 2012; Wong et al., 2013). ~~However, this technique is laborious, expensive, prone to contamination and~~The main limitation of this method is the requirement of ~~sophisticated instrumentation that is less convenient for field measurements~~requires additional resources/installations (e.g. thermal desorption unit and cryofocusing). ~~Alternatively,~~solid-phase microextraction (SPME) ~~can~~has also been used for the ~~collection of~~monoterpene sampling (e.g. Yassaa et al., 2010; Zini et al., 2001). This technique combines sampling and pre-concentration of analytes in a single step and allows for direct thermal desorption into a heated gas chromatograph injection port (Kozziel et al., 1999).

Carbonyl compounds also play an important role in the atmosphere due to their involvement in photochemical reactions and contribution to aerosol particle formation and growth (Jang and Kamens, 2001; Kesselmeier and Staudt, 1999). Aldehydes have been sampled in forest air with ~~C₁₈~~-cartridges coated with a 2,4-dinitrophenyl hydrazine (DNPH) derivatization reagent and analyzed by liquid chromatography-mass spectrometry (LC-MS) (e.g. Hellén et al., 2004). The main drawbacks of this method are the laborious sample preparation and long sampling times.

The trace amounts of BVOCs found in ambient air and their wide variety demands for the development and application of portable, more-selective and robust sampling and pre-concentration techniques. In our previous research, solid-phase microextraction (SPME) fibers and needle trap microextraction (NTME) syringes, combined with portable GC-MS, have been successfully used for the sampling and analysis of BVOCs in the boreal forest (Barreira et al., 2015; Barreira et al., 2016).

These methods have several advantages, such as full portability, low infrastructure/resource demands, high pre-concentration, no sample preparation and fast on-site analysis. However, the low mixing ratios of ~~some of these compounds~~ ~~some of these compounds~~ in forest atmosphere claim for additional improvements in pre-concentration. In this study, SPME Arrow was tested for the collection of BVOCs from boreal forest ambient air. This novel SPME-based system consists of a steel rod coated with a larger amount of sorbent material than the traditional SPME fibers, offering increased capacity but maintaining the compatibility for direct thermal desorption and analysis in a conventional GC-MS due to its shape and dimensions (Helin et al., 2015). The coated rod can be withdrawn in a steel ~~tub~~ ~~needle~~, which makes the device more robust. Samples were collected simultaneously by SPME fibers for comparison purposes. The effect of meteorological parameters at the sampling place on the ~~SPME sampling measured amounts of BVOCs~~ was tentatively evaluated. The inherent characteristics of SPME-based sampling techniques and coating materials used in this work (polydimethylsiloxane/divinylbenzene (PDMS/DVB) and PDMS/Carbon WR) were studied in the laboratory before the field campaign. ~~Also~~, static and dynamic SPME Arrow collection modes were also compared.

2 Material and Methods

2.1 Chemicals and materials

α -Pinene (98%), Δ^3 -Carene ($\geq 98.5\%$), ~~l~~-limonene ($\geq 99\%$), octanal (99%), nonanal (98%), decanal ($\geq 98\%$) from Sigma-Aldrich (St. Louis, USA) were used as standards. Stock solutions were prepared in dichloromethane (99.99%, Fisher Scientific, Loughborough, UK); and subsequently diluted with the same solvent to obtain needed concentrations. ~~These solutions were used~~ for the calibration of instrument response. For laboratory studies, diffusion vials were prepared by adding small amounts of standards to headspace vials (20 mL) and inserting a piece of deactivated fused silica retention gap (1.5 m \times 0.53 mm (i.d.), Agilent Technologies, Palo Alto, USA) through the septa, to allow a constant diffusion of the compounds from the vials. These vials were inserted into a home-made permeation oven, and the calibration gas flow was diluted with nitrogen, and subsequently transferred to an additional chamber from where samples were collected. The diffusion rates were measured by weighting the vials in five different days and determining the amount of analyte losses per time. The obtained diffusion rates

were 0.252 mg/h for α -pinene, 0.129 mg/h for Δ^3 -carene and 0.070 mg/h for octanal, which correspond to concentrations of 149 ppbv for α -pinene, 76 ppbv for Δ^3 -carene and 44 ppbv for octanal. Decanal diffusion rate was not possible to determine, probably due to slow evaporation from the diffusion vial. SPME fibers coated with PDMS/DVB (65 μ m, Supelco, Bellefonte, PA, USA) and PDMS/Carbon WR (95 μ m, CTC Analytics AG, Zwingen, Switzerland), and SPME Arrows coated with the same PDMS/Carbon WR (120 μ m) and PDMS/DVB (120 μ m) type₂ of sorbents (CTC Analytics AG, Zwingen, Switzerland) were used for analyte collection. The SPME fibers had a sorbent length of 10 mm, while SPME Arrows sorbent length was 20 mm. The diameter of Arrow needle was 1.1 mm. All SPME fibers and SPME Arrows were pre-conditioned according to the manufacturer's instructions.

2.2 Measurement site

BVOC sampling was performed at the SMEAR II station (Station For Measuring Ecosystem-Atmosphere Relations, 61°50.845' N, 24°17.686' E, 179 m above sea level) in Hyytiälä, Southern Finland (Hari and Kulmala, 2005). The station is situated in an approximately 55 years old and relatively homogeneous Scots pine stand, of about 21 m canopy height and 1170 stems ha⁻¹ of average tree density. The forest around the station is dominated by conifers (mainly Scots pine and Norway spruce). Tampere is the largest neighboring region, with approximately half-million inhabitants, and is located 60 km southwest from the SMEAR II station. The sampling site was situated about 1 m from a 127 m high mast for atmospheric and flux measurements mounted 2 m above the average forest floor.

2.3 Sampling and analysis

Ambient air samples were collected and analysed on-site from 11th to 15th of August, 2017. For comparison purposes, one PDMS/DVB SPME fiber, two PDMS/DVB SPME Arrows and one PDMS/Carbon WR SPME Arrow were used. Samples were collected in static mode for 45 minutes, to sample detectable amounts of target analytes with all SPME systems and to reduce the significance of errors associated with the time lag between sampling and injection. Additionally, a home-made dynamic sampling system for SPME Arrow was used for comparison with static SPME collection (Fig. S1). This device was adapted and modified from the sampling system developed in previous research for SPME fibers (Barreira et al., 2015). Samples were measured using a conventional GC-MS, consisting of an Agilent 6890 N gas chromatograph equipped with an Agilent 5973 mass selective detector (Agilent Technologies, Palo Alto, USA). The analytical column was a HP-5MS (30 m \times 0.25 mm \times 0.25 μ m, Agilent Technologies, Palo Alto, CA, USA). The initial oven temperature was 70°C (1 min), and it was increased to 250°C (1 min) at 20°C/min. The total run time was 11 minutes. Helium (99.996%, AGA, Espoo, Finland) was used as carrier gas in a constant flow mode (1.5 mL/min). SPME Arrow and fibers were desorbed in splitless mode (2 min) with a 2.0 mm internal diameter (i.d.) split/splitless inlet liner. Desorption temperature was 270°C for all the SPME ~~devices~~ devices. A standard inlet septum was used for SPME Arrow, while a 23-gauge Merlin Microseal and a Merlin nut (Merlin Instrument Company, Half Moon Bay, USA) were used in the injection port for conventional SPME fibers. The temperature of GC-MS transfer line was 250°C and the ion source and quadrupole temperatures were kept at 230°C and 150°C,

respectively. Electron ionization (70 eV) was used. The scan mass range was from 30–400 amu. The mass spectra and retention times of each analyte were obtained with standard solutions and used for identification of studied compounds in the collected samples. For semi-quantitation, extracted ion chromatograms with base ions were used (m/z 93 for α -pinene, Δ^3 -carene and limonene; m/z 43 for octanal and decanal; and m/z 57 for nonanal).

5 The same method was employed for the laboratory tests, although the initial oven temperature was 50°C (1 min) and the final temperature 250°C (1 min) at 20°C/min. For the laboratory determination of analytes extraction time profiles and to compare the extraction efficiencies of the different SPME-based sampling techniques, an Agilent 5975 C mass selective detector (Agilent Technologies, Palo Alto, USA) was also used, while all other laboratory tests were performed with the same GC-MS used for the field measurements. Samples were collected during 10 minutes, except for studying the effects of temperature and
10 relative humidity on the collection of BVOCs in which a 20 minutes sampling time was chosen.

3 Results

In this work, ~~the a novel~~ SPME Arrow system was optimized and tested in the laboratory to study its applicability for the field measurement of BVOCs in forest atmosphere. The characterization of SPME-based techniques, including kinetics of extraction, comparison of techniques and adsorbents extraction performances, and influence of temperature and relative
15 humidity on the extracted amounts are described in the first sections. The measurements performed in the field and comparison with atmospheric temperature, relative humidity, ozone (measured at 4.2 m height), precipitation, photosynthetically active radiation (PAR) and particle number concentration (PNC) (available at <http://avaa.tdata.fi/web/smart> and provided by Junninen et al., 2009) are then further discussed.

3.1 Characterization of SPME-based sampling techniques

20 3.1.1 Extraction time profile

The extraction time profiles for α -pinene, Δ^3 -carene, octanal and decanal were obtained in this study, to evaluate the occurrence of equilibrium and/or competitive adsorption during an experimental sampling time of one hour at laboratory ambient temperature. These compounds have been reported as [some of](#) the most abundant monoterpenes and aldehydes at the sampling site (e.g. Barreira et al., 2016). [For an estimation of \$\alpha\$ -Pinene standard deviation during kinetic studies, 3 repetitions were performed at 10 minutes, corresponding to the minimum sampling time used during all other experiments and consequently to the higher expected variation.](#) α -Pinene reached equilibrium after 10 minutes of extraction when a SPME Arrow coated with PDMS/DVB was used, and after 20 minutes with ~~the a~~ SPME fiber coated with the same material (Fig. [1S2](#)). For PDMS/Carbon WR, α -pinene equilibrium was reached already after 5 minutes of sampling for both SPME-based techniques. Δ^3 -Carene did not reach equilibrium when using PDMS/DVB for one hour, while with PDMS/Carbon WR equilibrium was
25 reached after 20 minutes for SPME Arrow and 40 minutes for SPME fiber. These results show that kinetics of extraction are
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faster with PDMS/Carbon WR comparatively to PDMS/DVB. The extraction time profiles for the aliphatic aldehydes studied showed that these compounds did not reach equilibrium after 60 minutes of sampling.

In the obtained extraction time profiles, analyte amounts did not decrease with time when using a SPME fiber and SPME Arrow coated with PDMS/DVB and a SPME fiber coated with PDMS/Carbon WR. This evidence suggests that interanalyte displacement due to competitive adsorption was not observed during the period of sampling, which is particularly interesting for PDMS/DVB since the uniformity of its micropores has been reported to potentially result in the displacement of analytes with less affinity by the ones with highest (Pawliszyn, 2011). However, ~~when using SPME Arrow coated with PDMS/Carbon WR,~~ some displacement was observed for Δ^3 -carene ~~when using SPME Arrow coated with PDMS/Carbon WR.~~

The extraction time profile for a dynamic sampling with ~~a~~ SPME Arrow coated with PDMS/Carbon WR was also obtained (Fig. ~~2S3~~). As expected, equilibrium was reached much faster than in static mode. This fact is clearly observed for octanal, that reached equilibrium in 40 minutes while in static mode it was not achieved during the experimental time. However, displacement effects were as well observed for Δ^3 -carene.

3.1.2 Extraction efficiencies of different SPME-based techniques

A comparison between SPME Arrow and SPME fiber extraction efficiencies was performed under the laboratory conditions described in the previous section. Three repetitions were performed with each SPME system. As shown in Fig. ~~3S4~~, extraction efficiency of SPME Arrow was approximately 2 times higher than for SPME fiber for PDMS/DVB and 3 times higher for PDMS/Carbon WR after 10 minutes of sampling, even though ~~this improvement~~ was slightly different depending on the analytes. Compound specific extraction efficiencies were distinct for PDMS/DVB and PDMS/Carbon WR. PDMS/DVB had the best affinity towards monoterpenes, while there was no statistically significant difference between materials for aldehydes.

Both materials adsorbed more Δ^3 -carene than α -pinene. At first glance, this selective adsorption seems to be greater for PDMS/Carbon WR than for PDMS/DVB. However, the equilibrium was not reached for all the studied compounds and a longer sampling time, such as the one used in the field measurements (45 min), will then impact on the relative amounts of analytes collected with both materials (Fig. ~~S31~~). Extraction efficiencies of sampling modes were also compared. Static and dynamic collections were performed during 10 minutes. As observed in Fig. ~~S25~~, α -pinene amounts were similar with both static and dynamic sampling. This result is expected since when equilibrium is reached the analyte collection is not anymore influenced by the sampling mode used. However, kinetics of extraction are much faster with dynamic sampling, which results in higher amount of analyte extracted in shorter time. On the other hand, the collected amounts of Δ^3 -carene and studied aldehydes were higher with dynamic mode under pre-equilibrium conditions.

3.1.3 Effect of temperature and relative humidity on the extraction

The effects of temperature and relative humidity on the SPME Arrow extraction efficiencies for the studied analytes were ~~studievaluated~~ under controlled conditions. The total sampling time was 20 minutes, which was considered ~~to be~~ enough to note any effect of ~~these~~ parameters on the extracted amounts. Three replicates were performed for each temperature and

[relative humidity studied](#). An increase in temperature is recognized to affect SPME collection by decreasing the distribution constant. The change in distribution constant with temperature is also dependent on the molar change in enthalpy of the analyte when it moves from the gas phase to the fiber sorbent (Pawliszyn, 2011). According to our results, the temperature effect changes significantly depending on the coating material and the analyte. As observed in Fig. 44, the effect of temperature was more significant for α -pinene extraction, especially when using a PDMS/Carbon WR material. On the other hand, temperature had a smaller effect on the extraction of Δ^3 -carene, also more pronounced when using a PDMS/Carbon WR adsorbent. Extraction efficiencies of octanal and decanal were not influenced by temperature. Similar results were obtained for SPME fibers coated with the same sorbents (Fig. S36). These results are expected due to the differences in molar change in enthalpy of the analytes when they move from the air to the sorbent, which causes distinct changes in the partition coefficient at different temperatures. [An underestimation of the measured amounts of a compound with higher volatility or an overestimation relatively to the most volatile compounds are then expected when semi-quantifying monoterpenes under field conditions where temperature changes can be significant. An underestimation of \$\alpha\$ -pinene measured amounts is then expected when semi-quantifying monoterpenes under experimental conditions where temperature changes are quite significant. The effect of temperature on the amounts of analytes collected by SPME must consequently be assessed or avoided during quantitative field measurements.](#)

The effect of relative humidity on the extraction efficiency of SPME fibers has been observed previously (e.g. Namieśnik et al., 2003). However, in our study, this effect was expected to be small due to the high hydrophobicity of SPME coating materials used. Indeed, relative humidity had a small influence on the extraction performances of both SPME Arrows (Fig. 52) and SPME fibers (Fig. S47). The small effect of relative humidity when using hydrophobic materials has been also observed previously, where the extraction of benzene, toluene, p-xylene and ethylbenzene with a PDMS/DVB coated SPME fiber at different humidity showed a maximum reduction of the mass adsorbed by approximately 21% after 1-h sampling (Koziel et al., 2000). A small effect of relative humidity on the SPME extraction when using hydrophobic coatings was also shown in another study, where two carbon-based SPME coatings were used for the sampling of 1,1,1-trichloroethane and carbon tetrachloride (Chai and Pawliszyn, 1995). In the referred publication, relative humidity reduced the amounts of extracted analytes at ambient temperature by less than 10% at up to 75% RH. [Even though the effect of relative humidity was negligible in our study, this parameter might greatly influence the SPME collection when using other sorbents or when sampling other analytes. Longer sampling times will probably enhance the effects of both temperature and relative humidity on the extracted amounts of analytes, but additional studies are still required to verify this hypothesis.](#)

3.2 Calibration of instrument response

The calibration of instrument response was performed for [the field measurement of monoterpenes](#) (α -pinene, Δ^3 -carene and limonene) and aldehydes (octanal, nonanal and decanal), to estimate the mass adsorbed on the coating materials of the different SPME-based devices systems (Table S1). [Liquid standard solutions at different concentrations were used for this purpose.](#) A linear 4-point calibration curve (0.1 ng to 10 ng) was obtained for monoterpenes, while a linear 5-point calibration curve (0.1

ng to 50 ng) was obtained for aldehydes. Three repetitions were done for each concentration level. The intermediate reproducibility (Rw), expressed as relative standard deviation (RSD), was from 0.2 to 19.5% for monoterpenes and from 2.0 to 18.7% for aldehydes, with higher RSD for lower concentration levels. A good linearity and sufficient correlation coefficients were observed for the mass ranges used. The limits of detection (LOD), which are also given in Table S1, varied from 17.7 pg to 28 pg for monoterpenes, while the LOD values obtained for aldehydes were from 61.1 pg to 155.2 pg.

3.3 Atmospheric levels of organic volatile compounds identified in air samples

In this study, three monoterpenes were identified during field measurements, namely α -pinene, Δ^3 -carene and limonene, ~~were identified and measured~~ (Fig. S58 and Table S2). The extracted amounts of BVOCs were in the order of few ng, which are in line with our preliminary measurements done in the previous year using an SPME Arrow coated with PMDS/Carbon WR and the same method as in this work (Fig. S69). The atmospheric levels of these BVOCs have been intensively determined at the SMEAR II boreal forest station, and are known to be dominated by α -pinene and Δ^3 -carene (Rinne et al., 2000; Yassaa et al., 2012). Similar results have been as well found in our previous research by using a SPME fiber (PDMS/DVB) combined with portable GC-MS (Barreira et al., 2015). Limonene has been also found in previous research, but at relative amounts that were considerably smaller when compared to the most abundant monoterpenes (e.g. Barreira et al., 2015; Rinne et al., 2000; Yassaa et al., 2012). The dominance of α -pinene and Δ^3 -carene was observed in this study, when considering the peak areas obtained by conventional GC-MS (Fig. S74). However, when the calibration of analytes response was performed with standard solutions ~~was performed~~ (but not the calibration of SPME collection), Δ^3 -carene and limonene levels increased relatively to α -pinene (Fig. 63A). This fact is particularly significant for limonene, which even overcomes the levels of α -pinene when SPME Arrow ~~RR~~ROW is used as a sampling device. Furthermore, PDMS/Carbon WR enhanced ~~these~~ differences when compared to PDMS/DVB. These results show that the materials used in this study are particularly selective for these compounds, especially for limonene. The preferential adsorption of both materials for Δ^3 -carene comparatively to α -pinene was also demonstrated in the laboratory studies described earlier (section 3.1.2). Furthermore, Δ^3 -carene amounts are enhanced by the temperature and relative humidity effects on α -pinene extraction described in section 3.1.3, this effect being more marked for PDMS/Carbon WR.

Aliphatic aldehydes, particularly octanal, nonanal and decanal, were identified in this study (Fig. S58 and Table S2). These aldehydes have been as well reported in another research performed in the boreal forest (Hellén et al., 2004). The most abundant aldehydes measured during the sampling campaign were nonanal and decanal, while octanal amounts were relatively small and only measurable when a SPME Arrow systemdevice was used (Fig. 63B). These aldehydes were more adsorbed with PDMS/DVB than with PDMS/Carbon WR, which was not observed in the laboratory experiments. The reason for this result might be related with the different times of collection and kinetics of adsorption, since equilibrium has not been reached in those experiments. During field measurements, the presence of wind speed or air bulk movement significantly affects the mass transfer process from bulk air to the sorbent (Pawliszyn, 2011), which can reveal differences in the adsorption of these compounds.

A comparison between SPME fiber and SPME Arrow was also performed, to evaluate quantitatively the enrichment provided by the SPME Arrows. Only PDMS/DVB was used for comparison in this study, since it was the less sensitive material to changes in temperature and relative humidity during the laboratory experiments. The monoterpenes were highly enriched when using SPME Arrow instead of the conventional SPME fiber (Fig. 74A). The amounts of monoterpenes measured when using SPME Arrow were approximately 2 times more than with the SPME fiber of the same material, which resembles the results obtained in the laboratory experiments. The enhancement was also slightly different depending on the analytes. However, compound specific extraction efficiencies were observed ~~between these devices~~. The ratio between Δ^3 -carene and α -pinene was approximately 1.4 when using ~~a~~ SPME fiber, but raised to 1.8 with SPME Arrow. This fact was found as well in the laboratory experiments, where a ratio of 0.9 was obtained for these compounds with the SPME fiber while a ratio of 1.3 was observed for SPME Arrow. The difference is less significant between limonene and α -pinene, with a ratio of 1.0 for the SPME fiber and 1.1 for the SPME Arrow; and between Δ^3 -carene and limonene, with ratios of 1.5 and 1.6 respectively. The SPME Arrow system also improved the collection of ~~the~~ aliphatic aldehydes ~~studied~~ when compared to SPME fibers (Fig. 74B). Interestingly, the improvement effect of SPME Arrow was much higher than the one verified in the laboratory studies, with amounts increasing 7 to 8 times comparatively to the SPME fiber. However, the sampling time for the laboratory experiments was shorter and equilibrium has not been reached for aldehydes. Furthermore, the wind influences the mass transfer process from air to the sorbent during the field experiments. Both of these facts impact on the amounts adsorbed in ~~the SPME devices~~ and can cause the observed differences in the enrichment with SPME Arrow. ~~The compound specific extraction was also evaluated in this study for nonanal and decanal, since octanal has not been detected with SPME fiber. Accordingly, the ratio between decanal and nonanal was 2.8 with SPME fiber and 2.4 with SPME Arrow, which indicates a small increase in nonanal extraction when using SPME Arrow. In the laboratory experiments, the ratio decreased from 0.18 to 0.13, which corroborates the field results.~~

Static and dynamic collection ~~modes~~ were compared for ~~the~~ SPME Arrow coated with PDMS/DVB. The extraction amounts were slightly higher when dynamic sampling was used. This result was observed for ~~the~~ all monoterpenes (Fig. 85A) and aldehydes (Fig. 85B) identified and measured during the sampling campaign. However, the differences were relatively small. ~~This result suggests~~ the proximity to or the attainment of the equilibrium state, where an increase in the time of extraction does not result in higher amounts of analyte extracted on the SPME materials. ~~A likely reason for this evidence is related with the fact that during field measurements the sampling wind improves the mass transfer from air to the sorbent effect of wind speed cannot also be neglected when using these devices, since it will increase the flow during the sampling increases in a similar way to the sampling devices used for dynamic extraction. However, in fact, VOC mass loading on the sorbent increases with an increase in wind velocity from 0 to 5 cm/s (Pawliszyn, 2011) and for that reason dynamic sampling is recommended to eliminate this effect. The main drawback when collection is performed in dynamic mode is the fact that~~ compounds with lower affinity for the coating are more susceptible to displacement effects (e.g. Tuduri et al., 2002). A decrease on extraction time can partially prevent ~~these~~ effects, even though might also cause a decrease in sensitivity that can compromise the possibility to measure some of the BVOCs present at trace levels in the atmosphere. ~~The effect of~~

~~wind speed cannot also be neglected when using these devices, since it will increase the flow during the sampling in a similar way to the sampling devices used for dynamic extraction. Besides, VOC mass loading on the sorbent increases with an increase in wind velocity from 0 to 5 cm/s (Pawliszyn, 2014).~~

3.4 Effect of meteorological parameters on the atmospheric levels of VOCs

5 The effects of meteorological parameters (Table S3) on the measured atmospheric levels of monoterpenes (Fig. 9) and aldehydes (Fig. 10) were also tentatively evaluated in this study, since some of these parameters can influence not only the VOC emissions/atmospheric mixing ratios but also the adsorption on the SPME sorbent materials. The ratios between the amounts of monoterpenes and aldehydes sampled with PDMS/Carbon WR and PDMS/DVB ~~were~~ also compared with the referred parameters, in order to understand if meteorological conditions affect differently the sampling with these materials (Fig. S:8).

10 Temperature has two opposing effects in field measurements performed at boreal forest sites. ~~Increasing the temperature enhances VOC emissions from Scots pine (Tarvainen et al., 2005). In addition, but will also it reduces the distribution constant of the analytes because adsorption is an exothermic process (de Fatima Alpendurada, 2000). During the sampling campaign, temperature remained almost constant, and this fact limited the comparison between this parameter and the measured~~ amounts of monoterpenes. The effect of temperature was then expected to be small when compared to other parameters, which was verified in our results. ~~Longer data sets encompassing periods of greater temperature variation are needed to perform this evaluation.~~

15 Relative humidity (RH) and precipitation also have two opposing effects, since monoterpene emission rates not only increase at high humidity levels and during and after precipitation (Llusà and Peñuelas, 1999; Schade et al., 1999), but also causes a small decrease in the SPME extraction capacity (section 3.1.3). Oppositely to temperature, relative humidity varied considerably during the sampling campaign, which allowed to study the effect of this parameter in our results. An increase in monoterpenes amounts was observed during a precipitation event when relative humidity was high, showing that the decreasing effect of these parameters on the sorbent collection is less significant when compared to their increasing effect on VOC emissions. A correlation was not found when considering the ratio between the amounts of monoterpenes sampled with PDMS/Carbon WR and PDMS/DVB, which is consistent with the small humidity dependence observed for the sampling with these materials under laboratory controlled conditions.

20 Due to the constancy of temperature during the sampling campaign, ozone and PAR were also expected to affect significantly the measured amounts of monoterpenes. Indeed, some anti-correlation was found between the measured monoterpenes and these parameters. This result is likely to reflect the increased photooxidation during periods of the day when PAR is high, since ~~the effects of temperature and/or light dependent on monoterpene emissions have~~ been described previously (Aalto et al., 2014). Nonetheless, on-fiber oxidation might occur during SPME collection. For that reason, the effect of oxidants must be assessed further by performing complementary laboratory experiments under controlled conditions. Ozone and PAR did not affect distinctively the adsorption on the two different materials used in this study. PNC also seemed to increase with the

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amounts of monoterpenes present in the ambient air. This result is expected, since monoterpene oxidation in the atmosphere and consequent formation of low volatile compounds have been recognized to contribute to aerosol particle formation (Laaksonen et al., 2008). However, other factors can also contribute to the increase in PNC, such as long-range transport. With reference concern to aldehydes, the studied parameters seemed not to influence negatively the SPME collection. A similar trend between aldehyde amounts and temperature was observed during the most of the sampling period, showing that temperature did not affect the SPME sampling and suggesting the existence of a temperature-dependence on aldehyde emissions. However, as referred previously, temperature remained almost constant during the sampling campaign and additional studies under controlled conditions are required to confirm this evidence. Relative humidity and precipitation also coincided with a burst in aldehyde atmospheric amounts, excluding a negative effect of this parameter on the SPME sampling, but seemed to anti-correlate with these parameters when aldehyde atmospheric amounts were low. This observation can be a consequence of the solubility of these compounds in water at low concentrations, but additional studies are also required to confirm this hypothesis. A correlation was not found between PAR and aldehydes. However, contrary to monoterpenes, some correlation with ozone was observed, which was expected since increased emissions have been reported in another study when vegetation was exposed to ozone (Wildt et al., 2003). On-fiber oxidation studies are also still required for aldehydes. The effect of aldehyde amounts on PNC was not very clear, which might be a consequence of the lower atmospheric reactivity of these compounds. No Any correlation was not observed between atmospheric parameters and the ratio between the amounts collected on PDMS/Carbon WR and PDMS/DVB, which agrees with the non-dependences on temperature and relative humidity verified in laboratory studies.

Even though the effects of atmospheric parameters on the SPME sampling were preliminarily evaluated with our method under atmospheric relevant conditions, longer data sets and quantitative data are needed to estimate accurately the correlation of these parameters with BVOC mixing ratios.

The effect of meteorological parameters (Table S3) on the measured atmospheric levels of monoterpenes was evaluated in this study, in order to understand their influence on the amounts adsorbed on the coating materials of the SPME-based sampling devices used (Fig. 96). The ratio between the amounts of monoterpenes and aldehydes sampled with PDMS/Carbon WR and PDMS/DVB was also compared with the referred parameters, in order to understand if meteorological conditions affect differently these two different types of extraction materials (Fig. S.811). Extraction temperature has two opposing effects in field measurements performed at boreal forest sites, since increasing the temperature enhances VOC emissions from Scots pine, but because adsorption is an exothermic process, increasing temperature will reduce the distribution constant of the analytes (de Fatima Alpendurada, 2000; Tarvainen et al., 2005). However, the temperature remained almost constant during the campaign, with a variation from 15 to 20 °C. Due to this reason, the effect of temperature on the measured amounts is expected to be small when compared to other parameters influencing the analyte emissions and atmospheric concentrations. This small effect was found in our study, where any significant correlation was found between temperature and the adsorbed amounts of monoterpenes. However, some anti correlation was found for the ratio between the amounts sampled with

PDMS/Carbon WR and PDMS/DVB, which is explained by the fact that PDMS/Carbon WR is more affected by changes in temperature than PDMS/DVB (section 3.1.3).

Relative humidity (RH) and precipitation also have two opposition effects, since monoterpene emission rates increase at high humidity levels and during and after precipitation (Llusà and Peñuelas, 1999; Schade et al., 1999), but also causes a small decrease in the SPME extraction capacity (section 3.1.3). As can be verified in Fig. 26, some correlation was observed when relative humidity was high. This correlation is more noticeable during and after rain episodes. However, the ratio between the amounts of monoterpenes sampled with PDMS/Carbon WR and PDMS/DVB increased at low humidity and decreases when humidity is high. This result also reflects the more pronounced effect of relative humidity on PDMS/Carbon WR material.

Due to the referred constancy of temperature during the sampling campaign, ozone and PAR were expected to be the most significant meteorological parameters affecting the measured monoterpene amounts. Indeed, an anti correlation was found between the measured monoterpenes and these parameters. However, both ozone and PAR did not affect distinctively the adsorption on the two different materials used in this study, which demonstrates that they impact is mostly on the monoterpene mixing ratios in the atmosphere. Nonetheless, some oxidation might occur after collection on the SPME devices, even though the total collection time was inferior to the lifetime of the studied compounds when exposed to ozone (Atkinson and Arey, 2003). PNC increased with the amount of monoterpenes present in the ambient air. This result is expected, since monoterpene oxidation in the atmosphere and consequent formation of low volatile compounds have been recognized to contribute to aerosol particle formation and growth (Laaksonen et al., 2008). Additionally, low amounts of monoterpenes were found in days when particle number concentration was extremely high. This result is also indicative of the oxidation of monoterpenes in the atmosphere and partition into aerosol particle phase, under favorable ambient conditions.

The effect of meteorological parameters on the measured amounts of aliphatic aldehydes was also studied (Fig. 107). The ratio between the amounts sampled with PDMS/Carbon WR and PDMS/DVB was as well used for comparison (Fig. S811). The atmospheric levels of these compounds seemed to increase with ambient temperature. Some correlation of carbonyl compounds with temperature has been described by Hellén et al. (2004). However, as referred previously, temperature remained almost constant during all sampling campaign. Furthermore, studies about the temperature effect on carbonyl emissions from forest are still scarce. On the other hand, some negative effect of relative humidity and precipitation on the amounts of aldehydes was evident during the campaign period, even though the effect is not visible when concentrations are high. This result might be explained by the fact that these compounds dissolve in water at low concentrations. However, also more studies under controlled conditions are required to demonstrate this hypothesis. A correlation was also found between aldehydes and ozone, which is expected since ozone is known to increase aldehyde emissions from vegetation (Wildt et al., 2003). On the other hand, any correlation was observed between PAR or PNC and the measured amounts of aldehydes. This result is explained by the lower reactivity of aldehydes in the atmosphere when compared to terpenoid compounds (Hellén et al., 2004). The studied aldehydes might yet contribute to aerosol formation and growth, but any inference about these contribution is very challenging without additional studies performed under controlled conditions. The ratio between the amounts of aldehydes sampled with PDMS/Carbon WR and PDMS/DVB was not affected by any of meteorological parameters (Fig. S811).

~~Even though the measured amounts of studied BVOCs followed the meteorological parameters registered at the sampling site, the amounts of the analytes extracted did not vary markedly during the sampling campaign to really obtain a proof for the correlation. This result is logical since temperature, which also did not vary significantly, is the main driving force of monoterpene emissions during summer and also influences aldehyde emissions (Tarvainen et al., 2005; Wildt et al., 2003). An exception was verified in the second day of campaign, where a peak on both monoterpene and aldehyde concentrations was observed. This peak coincided with a period of high temperature, high humidity, precipitation and the lowest PAR registered during the sampling periods. Since a similar trend in concentrations was observed for monoterpenes and aldehydes, this finding might indicate that studied aldehydes respond predominantly to the same atmospheric conditions contributing to monoterpene emissions. However, additional studies are still required to confirm this possibility.~~

10 4 Conclusion

A novel SPME Arrow system was tested in this study for the collection of BVOCs at the boreal forest (SMEAR II, Hyttiälä, Finland). Conventional SPME fibers were used for comparison. Samples were successfully collected with both SPME-based sampling systems, and were analyzed by conventional GC-MS. Neither additional sampling line nor sample pre-treatment was needed, reducing analysis time, sample contamination and potential losses. The most abundant monoterpenes and aldehydes were measured. PDMS/Carbon WR had higher affinity towards Δ^3 -carene and limonene than PDMS/DVB, while PDMS/DVB enhanced the extraction of α -pinene. Nonanal and decanal were the most abundant aliphatic aldehydes. The extraction efficiency of ~~the~~ SPME Arrow[®] was about two times higher than the one of SPME fiber, with an exception for aldehydes during the field campaign where a 7 to 8-fold enhancement was observed. Dynamic sampling demonstrated higher extraction efficiencies than static mode prior to equilibrium, but the improvement during field measurements was small due to the effect of wind speed on the extraction and/or to the fact that extraction was near equilibrium. Meteorological parameters influenced the amounts of studied BVOCs in the atmosphere, but did not seem to influence significantly the SPME sampling. However, ~~l~~laboratory tests showed that temperature and relative humidity decrease the extracted amounts of BVOCs, especially the ones with higher volatility and when PDMS/Carbon WR is used. Overall, results demonstrated the potential of SPME Arrow for the in-situ measurement of BVOCs in the atmosphere and the challenges that need to be solved for using these devices for quantitative purposes. More studies under controlled conditions are needed to understand the influence of co-adsorbed species and oxidants on the SPME sampling and to develop a proper calibration method for quantitative field measurements. ~~Longer data sets are also required to study in more detail the effects of atmospheric parameters on the SPME sampling under atmospheric relevant conditions.~~

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5 References

- Aalto, J., Kolari, P., Hari, P., Kerminen, V.-M., Schiestl-Aalto, P., Aaltonen, H., Levula, J., Siivola, E., Kulmala, M., and Bäck, J.: New foliage growth is a significant, unaccounted source for volatiles in boreal evergreen forests, *Biogeosciences*, 11, 1331-1344, doi: 10.5194/bg-11-1331-2014, 2014.
- Arneth, A., Harrison, S. P., Zaehle, S., Tsigaridis, K., Menon, S., Bartlein, P. J., Feichter, J., Korhola, A., Kulmala, M., and O'donnell, D.: Terrestrial biogeochemical feedbacks in the climate system, *Nat. Geosci.*, 3, 525-532, doi: <http://dx.doi.org/10.1038/ngeo905>, 2010.
- Atkinson, R. and Arey, J.: Gas-phase tropospheric chemistry of biogenic volatile organic compounds: a review, *Atmos. Environ.*, 37, 197-219, doi: [http://dx.doi.org/10.1016/S1352-2310\(03\)00391-1](http://dx.doi.org/10.1016/S1352-2310(03)00391-1), 2003.
- Barreira, L. M. F., Parshintsev, J., Kärkkäinen, N., Hartonen, K., Jussila, M., Kajos, M., Kulmala, M., and Riekkola, M.-L.: Field measurements of biogenic volatile organic compounds in the atmosphere by dynamic solid-phase microextraction and portable gas chromatography-mass spectrometry, *Atmos. Environ.*, 115, 214-222, doi: <http://dx.doi.org/10.1016/j.atmosenv.2015.05.064>, 2015.
- Barreira, L. M. F., Xue, Y., Duporté, G., Parshintsev, J., Hartonen, K., Jussila, M., Kulmala, M., and Riekkola, M.-L.: Potential of needle trap microextraction–portable gas chromatography–mass spectrometry for measurement of atmospheric volatile compounds, *Atmos. Meas. Tech.*, 9, 3661-3671, doi: 10.5194/amt-9-3661-2016, 2016.
- Bryan, A. M. and Steiner, A. L.: Canopy controls on the forest-atmosphere exchange of biogenic ozone and aerosol precursors, *Mich. J. Sustain.*, 1, 31-49, doi: <http://dx.doi.org/10.3998/mjs.12333712.0001.005>, 2013.
- Chai, M. and Pawliszyn, J.: Analysis of environmental air samples by solid-phase microextraction and gas chromatography/ion trap mass spectrometry, *Environ. Sci. Technol.*, 29, 693-701, doi: 10.1021/es00003a017, 1995.
- de Fatima Alpendurada, M.: Solid-phase microextraction: a promising technique for sample preparation in environmental analysis, *J. Chromatogr. A*, 889, 3-14, doi: [https://doi.org/10.1016/S0021-9673\(00\)00453-2](https://doi.org/10.1016/S0021-9673(00)00453-2), 2000.
- Graus, M., Müller, M., and Hansel, A.: High resolution PTR-TOF: quantification and formula confirmation of VOC in real time, *J. Am. Soc. Mass Spectrom.*, 21, 1037-1044, doi: <https://doi.org/10.1016/j.jasms.2010.02.006>, 2010.
- Haapanala, S., Hakola, H., Hellén, H., Vestenius, M., Levula, J., and Rinne, J.: Is forest management a significant source of monoterpenes into the boreal atmosphere?, *Biogeosciences*, 9, 1291-1300, doi: 10.5194/bg-9-1291-2012, 2012.
- Hakola, H., Hellén, H., Hemmilä, M., Rinne, J., and Kulmala, M.: In situ measurements of volatile organic compounds in a boreal forest, *Atmos. Chem. Phys.*, 12, 11665-11678, doi: <https://doi.org/10.5194/acp-12-11665-2012>, 2012.
- Hari, P. and Kulmala, M.: Station for measuring ecosystem-atmosphere relations, *Boreal Environ. Res.*, 10, 315-322, 2005.
- Helin, A., Rönkkö, T., Parshintsev, J., Hartonen, K., Schilling, B., Läubli, T., and Riekkola, M.-L.: Solid phase microextraction Arrow for the sampling of volatile amines in wastewater and atmosphere, *J. Chromatogr. A*, 1426, 56-63, doi: <https://doi.org/10.1016/j.chroma.2015.11.061>, 2015.
- Hellén, H., Hakola, H., Reissell, A., and Ruuskanen, T. M.: Carbonyl compounds in boreal coniferous forest air in Hyytiälä, Southern Finland, *Atmos. Chem. Phys.*, 4, 1771-1780, doi: 10.5194/acp-4-1771-2004, 2004.
- Henze, D. K. and Seinfeld, J. H.: Global secondary organic aerosol from isoprene oxidation, *Geophys. Res. Lett.*, 33, 1-4, doi: 10.1029/2006GL025976, 2006.
- Jang, M. and Kamens, R. M.: Atmospheric secondary aerosol formation by heterogeneous reactions of aldehydes in the presence of a sulfuric acid aerosol catalyst, *Environ. Sci. Technol.*, 35, 4758-4766, doi: 10.1021/es010790s, 2001.
- Jimenez, J., Canagaratna, M., Donahue, N., Prevot, A., Zhang, Q., Kroll, J. H., DeCarlo, P. F., Allan, J. D., Coe, H., Ng, N., Aiken, A., Docherty, K., Ulbrich, I., Grieshop, A., Robinson, A., Duplissy, J., Smith, J., Wilson, K., Lanz, V., Hueglin, C., Sun, Y., Tian, J., Laaksonen, A., Raatikainen, T., Rautiainen, J., Vaattovaara, P., Ehn, M., Kulmala, M., Tomlinson, J., Collins,

- D., Cubison, M., Dunlea, E., Huffman, J., Onasch, T., Alfarra, M., Williams, P., Bower, K., Kondo, Y., Schneider, J., Drewnick, F., Borrmann, S., Weimer, S., Demerjian, K., Salcedo, D., Cottrell, L., Griffin, R., Takami, A., Miyoshi, T., Hatakeyama, S., Shimono, A., Sun, J., Zhang, Y., Dzepina, K., Kimmel, J., Sueper, D., Jayne, J., Herndon, S., Trimborn, A., Williams, L., Wood, E., Middlebrook, A., Kolb, C., Baltensperger, U., and Worsnop, D.: Evolution of organic aerosols in the atmosphere, *Science*, 326, 1525-1529, doi: 10.1126/science.1180353, 2009.
- 5 Junninen, H., Lauri, A., Keronen, P., Aalto, P., Hiltunen, V., Hari, P., and Kulmala, M.: Smart-SMEAR: on-line data exploration and visualization tool for SMEAR stations, *Boreal Environ. Res.*, 14, 447-457, 2009.
- Kavouras, I. G., Mihalopoulos, N., and Stephanou, E.: Formation and gas/particle partitioning of monoterpenes photo-oxidation products over forests, *Geophys. Res. Lett.*, 26, 55-58, doi: 10.1029/1998GL900251, 1999.
- 10 Kesselmeier, J. and Staudt, M.: Biogenic volatile organic compounds (VOC): an overview on emission, physiology and ecology, *J. Atmos. Chem.*, 33, 23-88, doi: 10.1023/A:1006127516791, 1999.
- Koziel, J., Jia, M., Khaled, A., Noah, J., and Pawliszyn, J.: Field air analysis with SPME device, *Anal. Chim. Acta*, 400, 153-162, doi: [https://doi.org/10.1016/S0003-2670\(99\)00614-5](https://doi.org/10.1016/S0003-2670(99)00614-5), 1999.
- 15 Koziel, J., Jia, M., and Pawliszyn, J.: Air sampling with porous solid-phase microextraction fibers, *Anal. Chem.*, 72, 5178-5186, doi: 10.1021/ac000518l, 2000.
- Kulmala, M., Suni, T., Lehtinen, K., Maso, M. D., Boy, M., Reissell, A., Rannik, Ü., Aalto, P., Keronen, P., and Hakola, H.: A new feedback mechanism linking forests, aerosols, and climate, *Atmos. Chem. Phys.*, 4, 557-562, doi: 10.5194/acp-4-557-2004, 2004.
- Laaksonen, A., Kulmala, M., O'Dowd, C., Joutsensaari, J., Vaattovaara, P., Mikkonen, S., Lehtinen, K., Sogacheva, L., Maso, M. D., and Aalto, P.: The role of VOC oxidation products in continental new particle formation, *Atmos. Chem. Phys.*, 8, 2657-2665, doi: 10.5194/acp-8-2657-2008, 2008.
- Laothawornkitkul, J., Taylor, J. E., Paul, N. D., and Hewitt, C. N.: Biogenic volatile organic compounds in the Earth system, *New Phytol.*, 183, 27-51, doi: 10.1111/j.1469-8137.2009.02859.x, 2009.
- 25 Lerdau, M., Litvak, M., Palmer, P., and Monson, R.: Controls over monoterpene emissions from boreal forest conifers, *Tree Physiol.*, 17, 563-569, doi: <https://doi.org/10.1093/treephys/17.8-9.563>, 1997.
- Llusà, J. and Peñuelas, J.: Pinus halepensis and Quercus ilex terpene emission as affected by temperature and humidity, *Biol. Plantarum*, 42, 317-320, doi: <https://doi.org/10.1023/A:1002185324152>, 1999.
- Namieśnik, J., Jastrzebska, A., and Zygmunt, B.: Determination of volatile aliphatic amines in air by solid-phase microextraction coupled with gas chromatography with flame ionization detection, *J. Chromatogr. A*, 1016, 1-9, doi: [http://dx.doi.org/10.1016/S0021-9673\(03\)01296-2](http://dx.doi.org/10.1016/S0021-9673(03)01296-2), 2003.
- 30 Pawliszyn, J.: Handbook of solid phase microextraction, Elsevier, 2011.
- Peñuelas, J. and Llusà, J.: The complexity of factors driving volatile organic compound emissions by plants, *Biol. Plantarum*, 44, 481-487, doi: <https://doi.org/10.1023/A:1013797129428>, 2001.
- 35 Peñuelas, J. and Staudt, M.: BVOCs and global change, *Trends Plant Sci.*, 15, 133-144, doi: <https://doi.org/10.1016/j.tplants.2009.12.005>, 2010.
- Rantala, P., Aalto, J., Taipale, R., Ruuskanen, T. M., and Rinne, J.: Annual cycle of volatile organic compound exchange between a boreal pine forest and the atmosphere, *Biogeosciences*, 12, 5753-5770, doi: 10.5194/bg-12-5753-2015, 2015.
- Rinne, J., Hakola, H., Laurila, T., and Rannik, Ü.: Canopy scale monoterpene emissions of Pinus sylvestris dominated forests, *Atmos. Environ.*, 34, 1099-1107, doi: [http://dx.doi.org/10.1016/S1352-2310\(99\)00335-0](http://dx.doi.org/10.1016/S1352-2310(99)00335-0), 2000.
- 40 Schade, G. W., Goldstein, A. H., and Lamanna, M. S.: Are monoterpene emissions influenced by humidity?, *Geophys. Res. Lett.*, 26, 2187-2190, doi: 10.1029/1999GL900444, 1999.
- Schollert, M., Burchard, S., Faubert, P., Michelsen, A., and Rinnan, R.: Biogenic volatile organic compound emissions in four vegetation types in high arctic Greenland, *Polar Biol.*, 37, 237-249, doi: <https://doi.org/10.1007/s00300-013-1427-0>, 2014.
- 45 Tarvainen, V., Hakola, H., Hellén, H., Bäck, J., Hari, P., and Kulmala, M.: Temperature and light dependence of the VOC emissions of Scots pine, *Atmos. Chem. Phys.*, 5, 989-998, doi: 10.5194/acp-5-989-2005, 2005.
- Tuduri, L., Desauziers, V., and Fanlo, J. L.: Dynamic versus static sampling for the quantitative analysis of volatile organic compounds in air with polydimethylsiloxane-Carboxen solid-phase microextraction fibers, *J. Chromatogr. A*, 963, 49-56, doi: [https://doi.org/10.1016/S0021-9673\(02\)00222-4](https://doi.org/10.1016/S0021-9673(02)00222-4), 2002.
- Wildt, J., Kobel, K., Schuh-Thomas, G., and Heiden, A. C.: Emissions of oxygenated volatile organic compounds from plants Part II: emissions of saturated aldehydes, *J. Atmos. Chem.*, 45, 173-196, doi: 10.1023/A:1024030821349, 2003.

- Wong, G. K., Ng, S. J., and Webster, R. D.: Quantitative analysis of atmospheric volatile organic pollutants by thermal desorption gas chromatography mass spectrometry, *Anal. Methods*, 5, 219-230, doi: 10.1039/C2AY25982F, 2013.
- Yassaa, N., Custer, T., Song, W., Pech, F., Kesselmeier, J., and Williams, J.: Quantitative and enantioselective analysis of monoterpenes from plant chambers and in ambient air using SPME, *Atmos. Meas. Tech.*, 3, 1615-1627, doi: 10.5194/amt-3-1615-2010, 2010.
- 5 Yassaa, N., Song, W., Lelieveld, J., Vanhatalo, A., Bäck, J., and Williams, J.: Diel cycles of isoprenoids in the emissions of Norway spruce, four Scots pine chemotypes, and in Boreal forest ambient air during HUMPPA-COPEC-2010, *Atmos. Chem. Phys.*, 12, 7215-7229, doi: 10.5194/acp-12-7215-2012, 2012.
- 10 Zini, C. A., Augusto, F., Christensen, E., Smith, B. P., Caramão, E. B., and Pawliszyn, J.: Monitoring biogenic volatile compounds emitted by *Eucalyptus citriodora* using SPME, *Anal. Chem.*, 73, 4729-4735, doi: 10.1021/ac0103219, 2001.

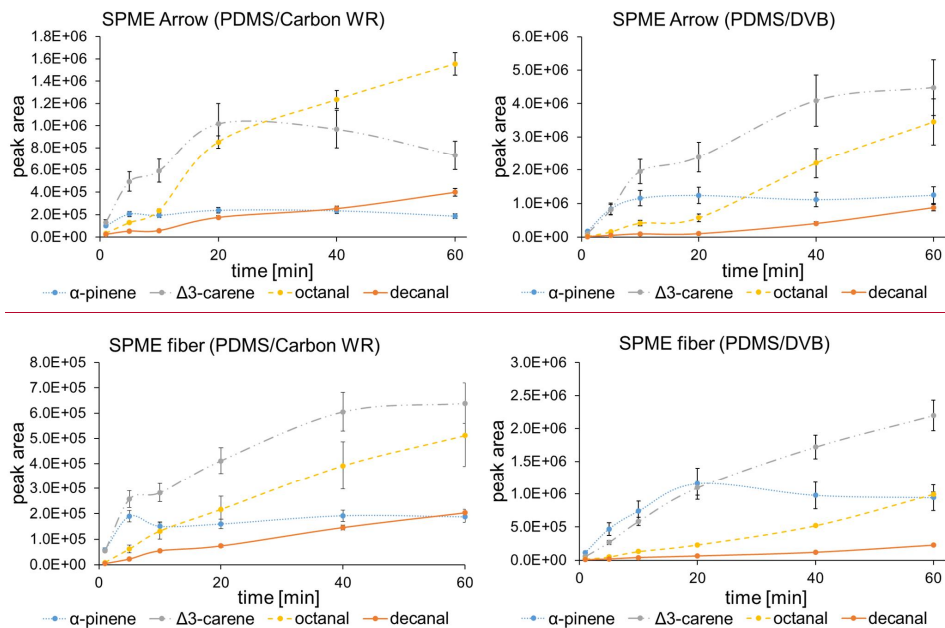


Figure 1: Extraction time profiles obtained for the studied analytes using SPME fiber and SPME Arrow coated with PDMS/Carbon WR and PDMS/DVB.

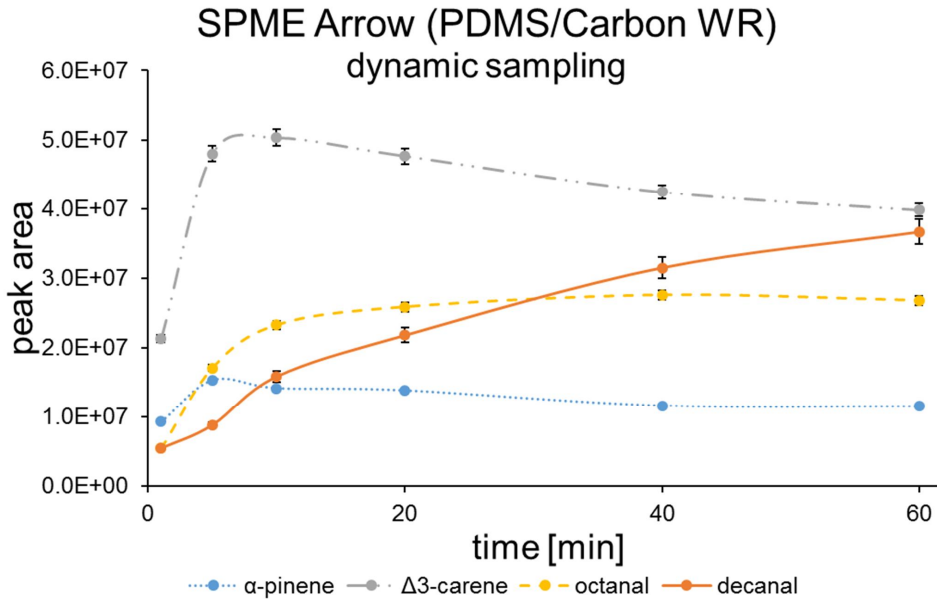


Figure 2: Extraction time profiles obtained for the studied analytes using dynamic sampling by SPME Arrow coated with PDMS/Carbon WR.

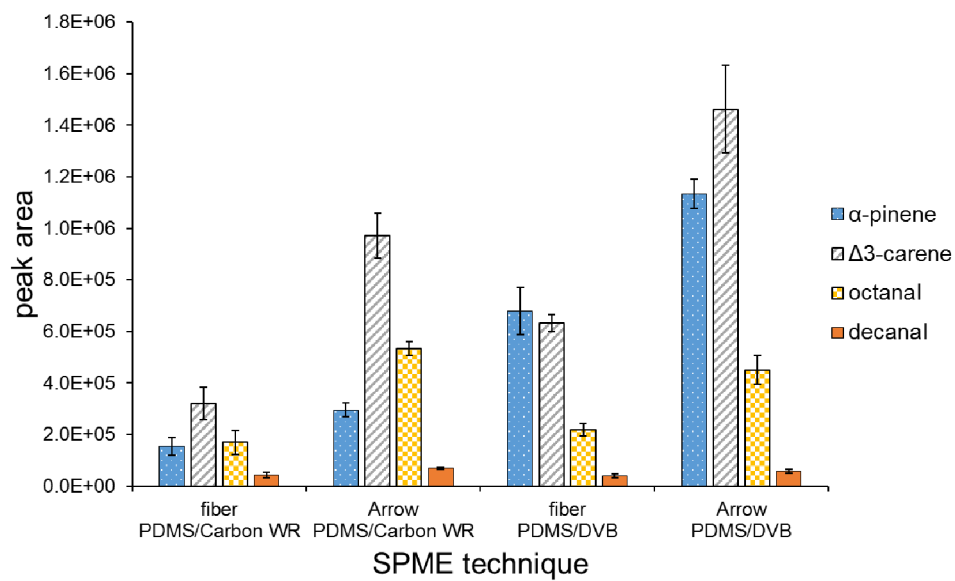


Figure 3: Comparison of the extraction efficiencies obtained with SPME Arrow and SPME fiber coated with PDMS/Carbon WR and PDMS/DVB.

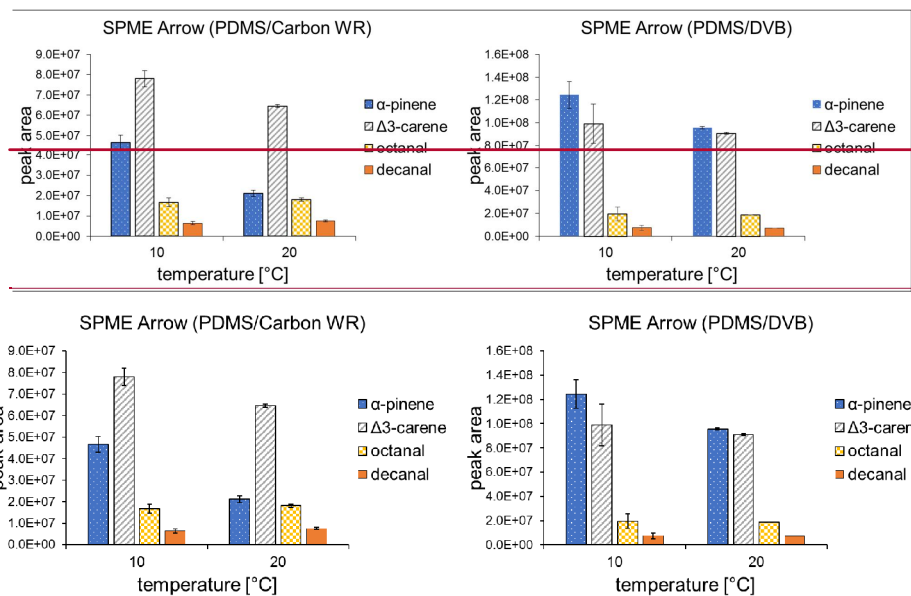


Figure 14: Effect of temperature ($^{\circ}\text{C}$) on the extraction efficiencies obtained with SPME Arrow ~~RRW~~ using PDMS/Carbon WR and PDMS/DVB sorbents.

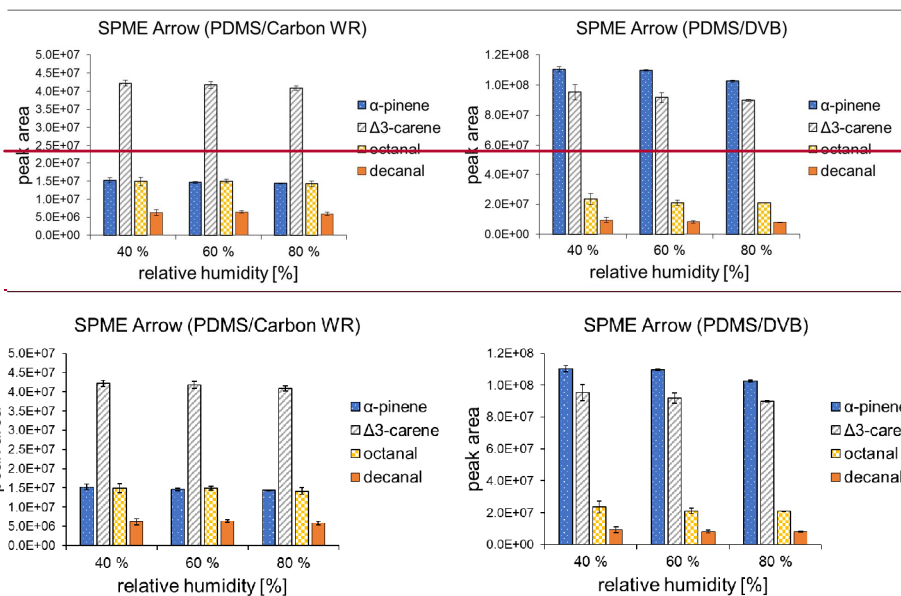


Figure 52: Effect of relative humidity (%) on the extraction efficiencies obtained with SPME Arrow[®] using with PDMS/Carbon WR and PDMS/DVB sorbents.

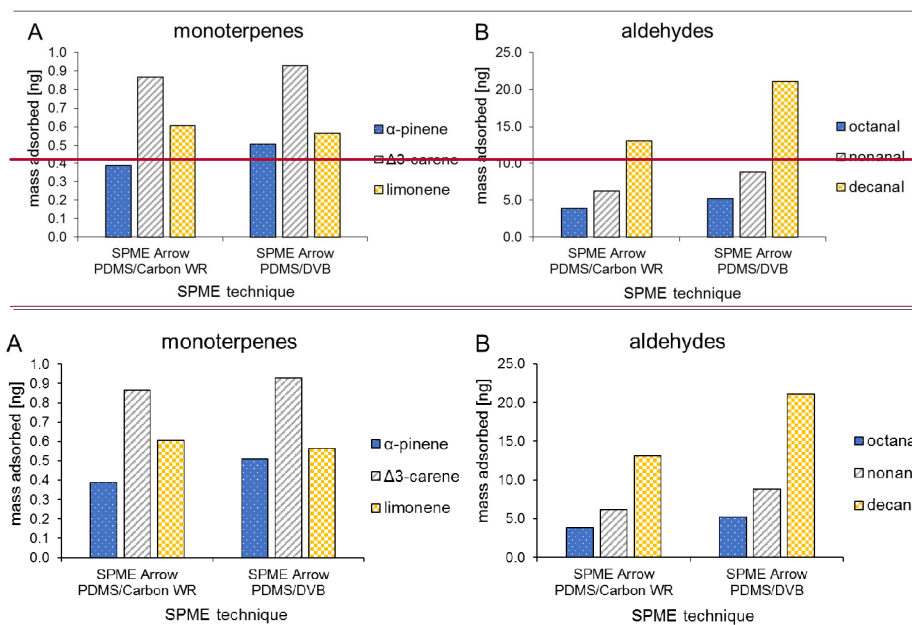
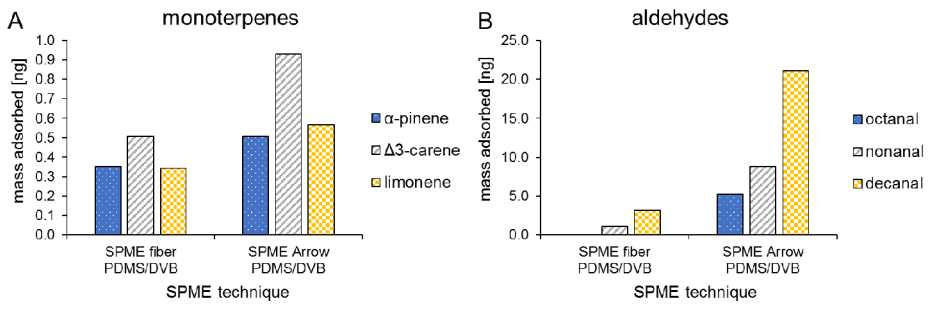
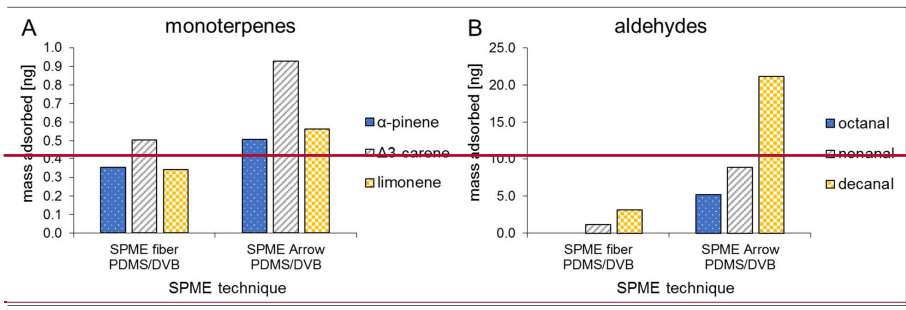


Figure 63: Comparison between the average mass of identified monoterpenes (α -pinene, Δ^3 -carene and limonene) and aldehydes (octanal, nonanal and decanal) collected with different SPME Arrow sorbents (PDMS/DVB and PDMS/Carbon WR) from ambient air and measured by GC-MS.



5 **Figure 74:** Comparison between the average mass of identified monoterpenes (α -pinene, Δ^3 -carene and limonene) and aldehydes (octanal, nonanal and decanal) collected with different PDMS/DVB SPME devices-devices (fiber and Arrow) from ambient air and measured by GC-MS.

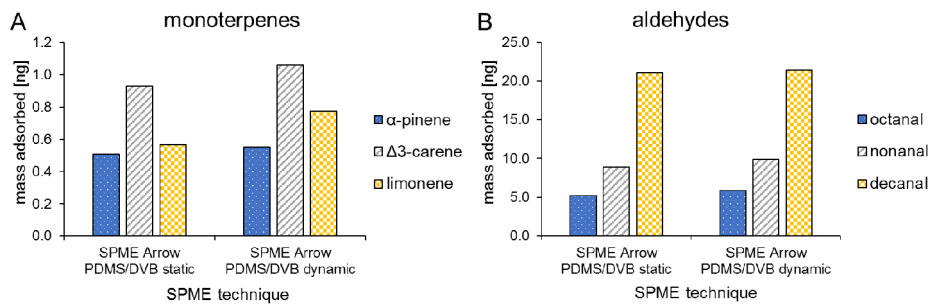
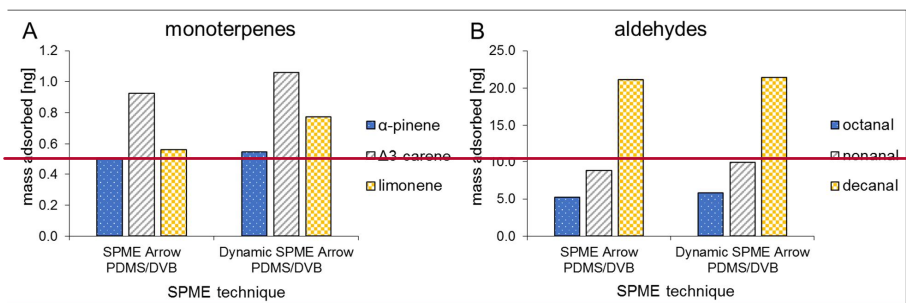
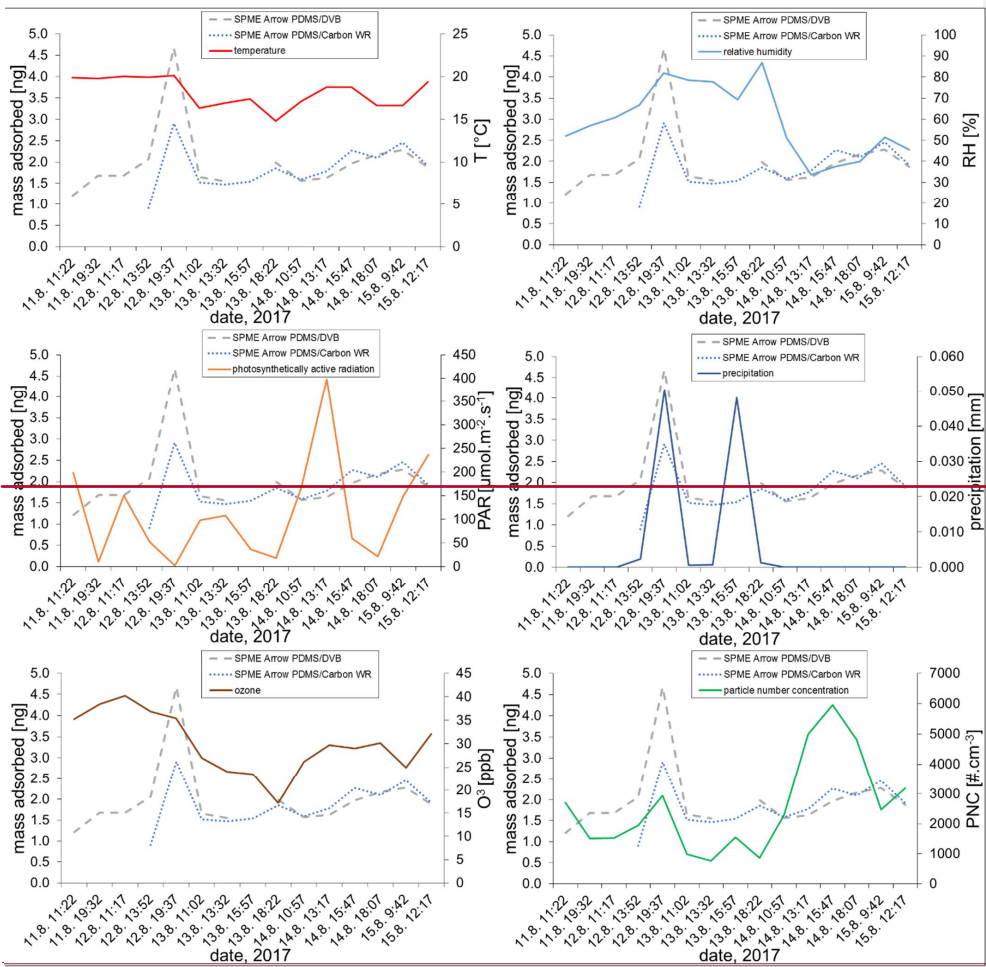
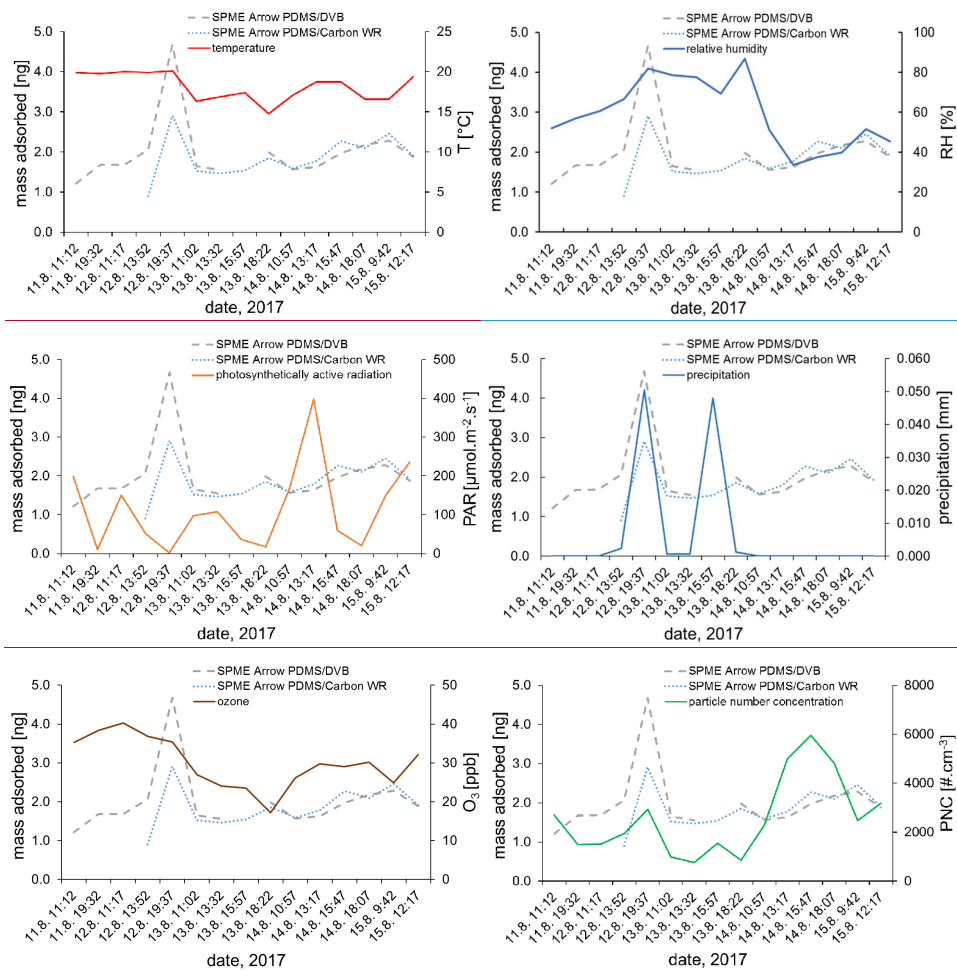
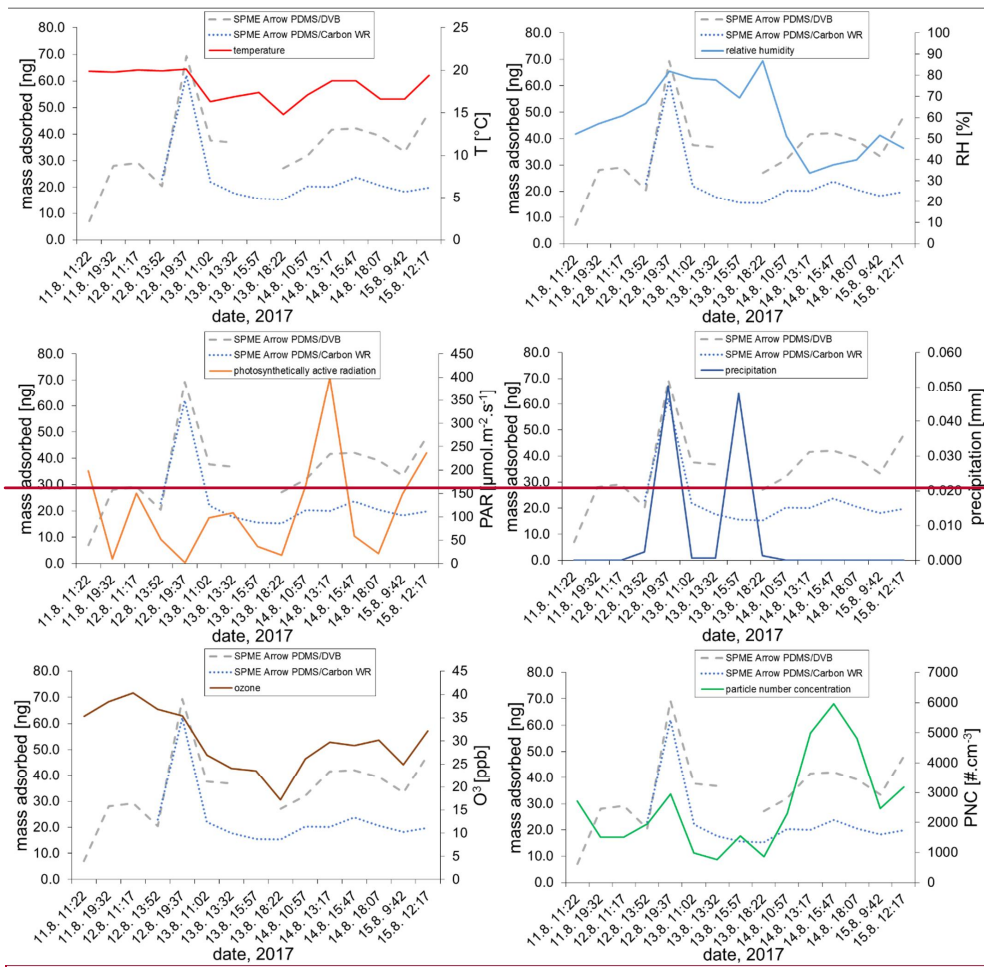


Figure 85: Comparison between the average mass of identified monoterpenes (α -pinene, Δ^3 -carene and limonene) and aldehydes (octanal, nonanal and decanal) collected with different **PDMS/DVB-SPME** sampling modes (static and dynamic) by SPME Arrow (PDMS/DVB) from ambient air and measured by GC-MS.





5 **Figure 96:** Effect of temperature ($^{\circ}\text{C}$), relative humidity (%), photosynthetically active radiation ($\mu\text{mol.m}^{-2}.\text{s}^{-1}$), precipitation (mm), ozone (ppb) and particle number concentration ($\#\text{.cm}^{-3}$) on the mass of monoterpenes adsorbed on the SPME Arrows used in this study (PDMS/DVB and PDMS/Carbon WR) and measured by GC-MS.



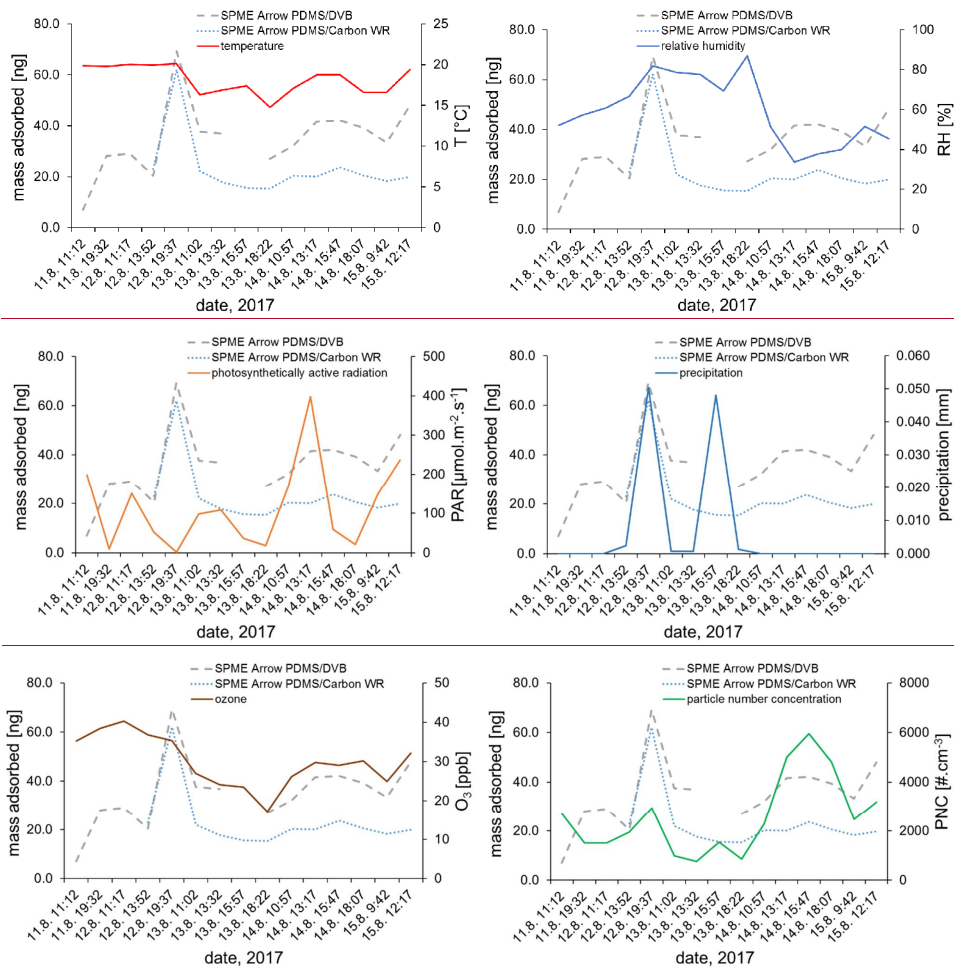


Figure 107: Effect of temperature ($^{\circ}\text{C}$), relative humidity (%), photosynthetically active radiation ($\mu\text{mol}\cdot\text{m}^{-2}\cdot\text{s}^{-1}$), precipitation (mm), ozone (ppb) and particle number concentration ($\#\cdot\text{cm}^{-3}$) on the mass of aldehydes adsorbed on the SPME Arrows used in this study (PDMS/DVB and PDMS/Carbon WR) and measured by GC-MS.