Comparing OMI-based and EPA AQS in situ NO$_2$ trends: Towards understanding surface NOx emission changes

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Abstract. With the improved spatial resolution than earlier instruments and more than ten years of service, tropospheric NO$_2$ retrievals from the Ozone Monitoring Instrument (OMI) have led to many influential studies on the relationships between socioeconomic activities and NOx emissions. Previous studies have shown that the OMI NO$_2$ data show different relative trends compared to in situ measurements. However, the sources of the discrepancies need further investigations. This study focuses on how to appropriately compare relative trends derived from OMI and in situ measurements. We retrieve OMI tropospheric NO$_2$ vertical column densities (VCDs) and obtain the NO$_2$ seasonal trends over the United States, which are compared with coincident in situ surface NO$_2$ measurements from the Air Quality System (AQS) network. The Mann-Kendall method with the Sen's slope estimator is applied to derive the NO$_2$ seasonal and annual trends for four regions at coincident sites during 2005-2014. The OMI-based NO$_2$ seasonal relative decreasing trends are generally biased low compared to the in situ trends by up to 3.7% yr^{-1}, except for the underestimation in the Midwest and Northeast during Dec-Jan-Feb (DJF). We improve the OMI retrievals for trend analysis by removing the ocean trend, using the MODerate-resolution Imaging Spectroradiometer (MODIS) albedo data in air mass factor (AMF) calculation. We apply a lightning flash filter to exclude lightning affected data to make proper comparisons. These data processing procedures result in close agreement (within 0.3% yr^{-1}) between in situ and OMI-based NO$_2$ regional annual relative trends. The remaining discrepancies may result from inherent difference between trends of NO$_2$ tropospheric VCDs and surface concentrations, different spatial sampling of the measurements, chemical non-linearity, and tropospheric NO$_2$ profile changes. We recommend future studies to apply these procedures (ocean trend removal and MODIS albedo update) to ensure the quality of satellite-based NO$_2$ trend analysis and apply the lightning filter in studying surface NOx emission changes using satellite observations and in comparison with the trends derived from in situ NO$_2$ measurements. With these data processing procedures, we derive OMI-based NO$_2$ regional annual relative trends using all available data for the West (-2.0%±0.3 yr^{-1}), the Midwest (-1.8%±0.4 yr^{-1}), the Northeast (-3.1%±0.5 yr^{-1}), and the South (-0.9%±0.3 yr^{-1}). The OMI-based annual mean trend over the contiguous United States is -

1.5%±0.2 yr^{-1}. It is a factor of 2 lower than that of the AQS in situ data (-3.9%±0.4 yr^{-1}); the difference is mainly due to the fact that the locations of AQS sites are concentrated in urban and suburban regions.

1 Introduction

Nitrogen dioxide (NO$_2$) is an air pollutant. At high concentrations, it aggravates respiratory diseases and can lead to acid rain formation (e.g., Lamsal et al., 2015). It is also a key player to produce another pollutant, ozone (O$_3$), through photochemical reactions in the presence of Volatile Organic Compounds (VOCs) under sunlight. Tropospheric NO$_2$ is emitted both anthropogenically and naturally (e.g., Gu et al., 2016). Anthropogenic fossil fuel combustions and biomass burnings emit mostly nitrogen monoxide (NO) under high temperature, which is later oxidized by O$_3$ into NO$_2$. Major natural NO$_2$ sources include lightning and soils.

Surface NO$_2$ concentrations are regulated by the U.S. Environmental Protection Agency (EPA) through the National Ambient Air Quality Standards (NAAQS). NO$_2$ is measured routinely at the EPA Air Quality System (AQS) sites (Demerjian, 2000). Although the AQS network continually provides valuable hourly NO$_2$ measurements, AQS sites are mostly located in urban and suburban regions, leaving large regions of rural areas unmonitored. Satellite data provide a better spatial coverage than the in situ measurements.

Several satellites were launched to monitor tropospheric NO$_2$ vertical column densities (VCDs), such as the SCanning Imaging Absorption spectroMeter for Atmospheric CHartographY (SCIAMACHY), the Global Ozone Monitoring Experiment–2 (GOME-2), and the Ozone Monitoring Instrument (OMI). For trend analysis, the tropospheric NO$_2$ products from OMI surpass the others for a relatively high spatial resolution and over one decade of continuous operation (Boersma et al., 2004; Boersma et al., 2011). Thus, OMI NO$_2$ retrievals are widely applied in NO$_2$ and NOx emission trend studies (e.g., Lin et al., 2010, 2011; Castellanos et al., 2012; Russell et al., 2012; Gu et al., 2013; Lamsal et al., 2015; Lu et al., 2015; Tong et al., 2015; Cui et al., 2016; Duncan et al., 2016; de Foy et al., 2016a, 2016b; Krotkov et al., 2016; Liu et al., 2017). Tong et al. (2015) reported that the reduction rates calculated from OMI NO$_2$ VCDs and AQS surface NO$_2$ data at eight cities were -35% and -38% from 2005 to 2012, respectively. Lamsal et al. (2015) also found the divergence between the annual trends inferred from the two datasets, i.e. -4.8% yr^{-1} vs -3.7% yr^{-1} during 2005-2008, and -1.2% yr^{-1} vs -2.1% yr^{-1} during 2010-2013. There are several potential factors attributing to the discrepancies between trends from satellite and ground-based measurements: interferences by the oxidation products of NOx from the chemiluminescent instruments (Lamsal et al., 2008, 2014, 2015), the differences of sampling time between OMI (~13:30 local time) and AQS (hourly) measurements (Tong et al., 2015), a high sensitivity of NO$_2$ VCDs to high-altitude NO$_2$ in contrast to the high sensitivity of surface NO$_2$ concentrations to surface NOx emissions (Duncan et al., 2013; Lamsal et al., 2015), spatial representativeness of satellite pixels (Lamsal et al., 2015), and high uncertainties of satellite retrievals in clean regions (Lamsal et al., 2015).

To understand how various factors and the retrieval procedure affect the differences between the OMI derived trends and those derived from the surface AQS measurements, we utilize a regional 3-D chemistry transport model (CTM), a radiative

transfer model (RTM), and the Mann-Kendall method (Mann, 1945; Kendall, 1948) to calculate OMI-based NO_2 seasonal relative trends during Dec-Jan-Feb (DJF), Mar-Apr-May (MAM), Jun-Jul-Aug (JJA), and Sept-Oct-Nov (SON) (Section 2). We find that two procedures are essential to ensure the quality of trend analysis using OMI tropospheric NO_2 VCDs, including the ocean trend removal and the MODerate-resolution Imaging Spectroradiometer (MODIS) albedo update in calculating the

5   air mass factors (AMFs). The lightning filter (Section 3.1) is necessary for comparing OMI-based and in situ AQS NO_2 trends. With these procedures implemented, the differences between OMI-based and AQS in situ annual relative trends are within 0.3% yr^{-1} of coincident measurements for all the four regions. Finally, we estimate the OMI-based annual relative trends across the nation in Section 3.2. Conclusions are given in Section 4.

2 Methods

10   **2.1 EPA AQS surface NO_2 measurements**

The in situ surface NO_2 measurements from the U.S. EPA AQS network are used in this research. Sites with a continuous measurement gap of more than 50 days are removed and the observations of 140 remaining cites are used (Fig. 1). The AQS chemiluminescent analyzers are equipped with molybdenum converters to measure ambient NO_2 concentrations. These analyzers are known to have high biases, since the converters are not NO_2 specific and they measure some fractions of

15   peroxyacetyl nitrate, nitric acid and organic nitrates (Demerjian, 2000; Lamsal et al., 2008). In addition to chemiluminescent analyzers, several NO_2 specific photolytic instruments were deployed since 2013. By utilizing the data from both chemiluminescent and photolytic measurements at coincident sites during the overpassing time of OMI, we calculate the observed NO_2 concentration ratio between both measurements in Fig. S1 in the Supplement. The ratio peaks at 2.3 in June and decreases to 1.3 in November, indicating that the chemiluminescent analyzers overestimate by 27%-132% than photolytic

20   instruments. This finding is in agreement with Lamsal et al. (2008). We correct the chemiluminescent NO_2 data by the observed ratio assuming that the inter-annual change is small and the high bias of the chemiluminescent measurements is identical at all sites. This correction may contribute to the differences between in situ and OMI based absolute NO_2 trends but do not significantly affect the relative trends (since the correction is canceled out in computing relative trends). In this study, we only examine the relative trends and therefore the analysis results are not affected by the uncertainties in the in situ NO_2

25   measurement corrections.

2.2 REAM model

We use a 3-D Regional chEmical trAnsport Model (REAM) in the simulation of NO_2 profiles. REAM has widely been used in atmospheric NO_2 studies, including vertical transport (Choi et al., 2005; Zhao et al., 2009a; Zhang et al., 2016a), emission inversions (Zhao et al., 2009b; Yang et al., 2011; Gu et al., 2013, 2014, 2016), and regional and seasonal variations (Choi et

30   al., 2008a, 2008b). The model has a horizontal resolution of 36 km with 30 vertical layers in the troposphere, 5 vertical layers in the stratosphere, and a model top of 10 hpa. In this study, the domain of REAM is about 400 km larger on each side than

the contiguous United States (CONUS). Meteorology inputs driving transport process are simulated by the Weather Research and Forecasting model (WRF) assimilations constrained by National Centers for Environmental Prediction Climate Forecast System Reanalysis (NCEP CFSR, Saha et al., 2010) 6-hourly products. The KF-eta scheme is used for sub-grid convective transport in WRF (Kain and Fritsch,1993). We run the WRF model with the same resolution as in REAM but with a domain 10 grids larger on each side than that of REAM. REAM updates most of the meteorology inputs every 30 minutes while those related to convective transport and lightning parameterization are updated every 5 minutes. The chemistry mechanism expands that of a global CTM GEOS-Chem (V9-02) with aromatics chemistry (Bey et al., 2001; Liu et al., 2010, 2012a, 2012b; Zhang et al., 2017). For consistency, the GEOS-Chem (V9-02) simulation with $2° \times 2.5°$ resolution is used to generate initial and boundary conditions for chemical tracers.

Anthropogenic emissions of NOx and other chemical species are from the U.S. National Emission Inventory 2008 prepared using the Sparse Matrix Operator Kernel Emission (SMOKE) model. Biogenic emissions are simulated online using the Model of Emissions of Gases and Aerosols from Nature (MEGAN) algorithm (v2.1, Guenther et al., 2012). We parameterize lightning emitted NOx as a function of convective mass flux and Convective Available Potential Energy (CAPE) (Choi et al., 2005). NOx production per flash is set to 250 moles NO per flash, and the emissions are distributed vertically following the C-shaped profiles by Pickering et al. (1998). For recent model evaluations of REAM with observations, we refer readers to Zhang et al. (2016a, 2016b), Cheng et al. (2017), and Zhang et al. (2017).

2.3 OMI-based NO₂ VCDs

We retrieve the tropospheric NO_2 VCDs using the tropospheric slant column densities (SCDs, without destriping) from the Royal Dutch Meteorological Institute (KNMI) Dutch OMI NO_2 product (DOMINO v2, Boersma et al., 2011). OMI onboard the Aura satellite was launched in July 2004 and is still active. OMI overpasses the equator at about 13:30 Local Time (LT) and obtains global coverage with a 2600 km viewing swath spanning 60 rows. It has a ground level spatial resolution up to 13 km x 24 km (at nadir). The spatial extent of the OMI pixels will not affect our analysis as we focus on regional trend analysis. SCDs are retrieved by matching a modeled spectrum to an observed top-of-atmosphere reflectance with the Differential Optical Absorption Spectroscopy (DOAS) technique within a fitting window of 405-465nm. The stratospheric portion of SCDs are estimated and subsequently removed with a global CTM TM4 with stratospheric ozone assimilation (Dirksen et al., 2011). Deriving tropospheric VCDs from the remaining tropospheric SCDs requires the calculation of AMFs. Being an optically thin gas, tropospheric AMF for NO_2 can be calculated from AMF for each vertical layer (AMF_l) weighted by NO_2 partial VCDs at the corresponding layer (x_l) (Boersma et al., 2004), as shown in equation (1).

$$\text{tropospheric AMF} = \frac{tropospheric\ SCD}{tropospheric\ VCD} = \frac{\int AMF_l x_l dl}{\int x_l dl} \qquad (1)$$

As the vertical distribution of NO_2 is usually unknown, we typically substitute x_l by an a priori profile ($x_{l,apriori}$) from a CTM. AMF_l is the sensitivity of NO_2 SCD to VCD at a given altitude (Eskes and Boersma, 2003), and is computed using the

Double Adding KNMI (DAK) RTM (Boersma et al., 2011). As a result, the retrieved tropospheric NO_2 VCD computation depends on the a priori NO_2 vertical profile, the surface reflectance, the surface pressure, the temperature profile, and the viewing geometry (Boersma et al., 2011). Previous studies have addressed the sources of uncertainties in NO_2 retrievals, including surface reflectance resolutions (Russell et al., 2011; Lin et al., 2014), lightning NOx (Choi et al., 2005a; Martin et al., 2007; Bucsela et al., 2010), a priori CTM uncertainties (Russell et al., 2011; Heckel et al., 2011; Lin et al., 2012; Laughner et al., 2016), surface pressure and reflectance anisotropy in rugged terrain (Zhou et al., 2009), cloud and aerosol radiance (Lin et al., 2014, 2015), and boundary layer dynamics (Zhang et al., 2016a). In this study, we find that the first two factors are essential in NO_2 VCD trend analysis and we will discuss these in the following sections.

AMFs are derived using the pre-computed altitude-dependent AMF lookup table, which is generated by the DAK RTM. We use the NO_2 profiles from REAM, temperature and pressure from CSFR, viewing geometry and cloud information from DOMINO v2 product. We use the REAM results of 2010 to avoid the uncertainty introduced by yearly variation of NO_2 profiles. The yearly variations of meteorology and anthropogenic emission changes have little impact in polluted areas on trend analysis results using OMI data (Lamsal et al., 2015). We use the surface reflectance from DOMINO v2 product as default (Kleipool et al., 2008), and update it using a surface reflectance product with a higher temporal resolution (Section 2.3.2). The derived tropospheric NO_2 VCD relative trends with default surface reflectance are referred as "Standard".

2.3.1 Ocean trend removal

For trend and other analyses of OMI tropospheric VCDs, the data of anomalous pixels must be removed. The row anomaly initially occurred in June 2007 and subsequently in later years affected rows 26-40 (Schenkeveld et al., 2017). Additional anomalies can be found in some years in rows 41-55. For trend analysis from 2005-2014, we exclude rows 26-55, consistent with our understanding of the row anomaly (Schenkeveld et al., 2017). In addition, the data of coarse spatial resolution from rows 1-5 and rows 56-60 are also excluded, as suggested by Lamsal et al. (2015). The selection of rows 6-25 used in this research is stricter than the data flags in the DOMINO v2 product. Furthermore, we exclude OMI data with cloud fraction > 0.3 to minimize retrieval uncertainties due to clouds and aerosols (Boersma et al., 2011; Lin et al., 2014).

Fig. 2a shows that there is an apparent increasing trend of the averaged tropospheric SCDs in the remote ocean region (Fig. 2b) with minimal marine traffic. This trend may reflect the inaccurately simulated stratospheric SCDs (A. Richter, personal communication, 2018) or the increase in the magnitude of the stripes (step-wise SCD variability from one row to another) in time, which originates from the use of a constant (2005-averaged) solar irradiance reference spectrum in the DOAS spectral fits throughout the mission and the weak increase of noise in the OMI radiance measurements (F. Boersma, personal communication, 2017; Zara et al., 2018). Fig. 2a shows that there is a positive annual trend of $1.75\pm0.45\times10^{13}$ molecules cm^{-2} yr^{-1}. The ocean trend is insensitive to the region selection in the remote North Pacific (varies within 10%). We only analyze OMI tropospheric column trends over the CONUS for grid cells with 2005-2014 averaged VCDs > 1×10^{15} molecules cm^{-2}, which tends to minimize the effect of the background noise. However, removing this background ocean (absolute) trend has a

non-negligible effect in reducing the OMI relative trend (Fig. 1). We treat this trend as a systematic bias. We calculate the contribution from the ocean (absolute) annual trend to SCDs for each year and subtract it from OMI tropospheric NO_2 SCDs uniformly in the following analysis. Since the origin of this trend is not yet clear, the ocean trend removal method may need updates in future studies. We refer to such derived (relative) trend data as "Ocean". An alternative method is to subtract

5   monthly SCD trends of the remote ocean region (Fig. S2 in the Supplement) from the OMI tropospheric SCD data. Although the end results (Fig. S3 in the Supplement) are essentially the same as the annual trend removal method, noises are added to the SCD data, making it more difficult to understand the effects of the MODIS albedo update and the lightning filter (next sections). We therefore choose to use the (absolute) annual trend removal method here.

2.3.2 MODIS albedo update

10   The albedo data used to calculate the AMF_l in "Standard" and "Ocean" versions of trend analysis are from the DOMINO v2 products, which are the climatology of averaged OMI measurements during 2005-2009 with a spatial resolution of $0.5° \times 0.5°$ (Kleipool et al., 2008) and is valid for 440 nm. We recalculate the AMF_l using the MODIS 16-day MCD43B3 albedo product with 1km spatial resolution, which combines data from both MODIS onboard Aqua and Terra satellites (Schaaf et al., 2002; Tang and Zhang, 2007). Aqua and Terra have an equatorial overpassing time of 13:30 LT and 10:30 LT, respectively. The

15   band 3 (459nm-479nm) is used to match the NO_2 fitting window (405nm-465nm). The albedo is spatially integrated to the geometry of OMI pixels and is temporally interpolated to match OMI overpassing dates. In order to maintain the consistency of the DOMINO retrieval algorithm (Boersma et al., 2011), we only use the MODIS data to improve the temporal variations of albedo data used in the retrieval. We scale the MODIS albedo data such that the mean albedo during 2005-2009 is the same as the OMI climatology at $0.5° \times 0.5°$. We recalculate OMI tropospheric VCDs using the MODIS albedo data as described. We

20   recalculate the relative OMI trend and remove the ocean (absolute) trend (Section 2.3.1). We refer to this version of OMI relative trend data as "MODIS".

2.3.3 Lightning event filter

Over North America, lightning is a major source of NOx in the free troposphere and its simulations in CTMs are uncertain (e.g., Zhao et al., 2009a; Luo et al., 2017). The large temporospatial variations of lightning NOx make it difficult to compute

25   satellite based NO_2 trends by changing the vertical distributions of NO_2 affecting the AMF calculation (e.g., Choi et al., 2008b; Lamsal et al., 2010) and the SCD values. Furthermore, accompanying lightning occurrences, the presence of cloud significantly affects the lifetime of NOx and the partitioning of NO_2 to NO in daytime and convective transport exports NOx from the surface and boundary layer to free troposphere, changing surface and column NO_2 concentrations. Given the difficulty to simulate lightning NOx accurately across different years and meteorological effects (vertical mixing) of lightning, we use a

30   lightning filter to remove potential effects of lightning NOx on the basis of the flash rate observations of cloud-to-ground (CG) lightning flash data detected by the National Lightning Detection Network[TM] (NLDN) (Cummins and Murphy, 2009; Rudlosky and Fuelberg, 2010). NLDN only reports the ground point of a CG lightning flash, while the CG lightning flash can extend

horizontally to tens of kilometers. A CG lightning flash can affect the NO_2 retrievals not only in the model grid cell where the CG lightning is located but also the nearby model grid cells. The atmospheric lifetime of NOx in the free troposphere can be up to 1 week. Therefore, we exclude the OMI NO_2 data within a radius of 90 km of the NLDN-reported CG lightning location (about two model grid cells around the grid cell where the CG lightning is located) for a period of 72 hours after the lightning occurrence. Since lightning usually occur along the track of a thunderstorm, the 90 km radius is more a constraint on lightning NOx effects across the track. The extended period of 72 hours is to ensure that we exclude data affected by lightning NOx. Figure 4 shows the distribution of the number of days of 2005-2014 with lightning detection. The Southwest monsoon and the South regions have more lightning days than the other areas. While there are fewer lightning flashes in the Northeast than the South (Fig. 3), large amounts of lightning NOx can be produced by high flash ratios of severe thunderstorms and they can be transported northward from the South to the Northeast (Choi et al., 2005). We therefore further filter OMI NO_2 data in the Northeast on the basis of CG lightning flash rates in the South. If the average CG flash rate in the South exceeds the 95th percentile value of the NLDN observations, which is 0.035 flash km^{-2} day^{-1} (Fig. S4 in the Supplement), we exclude in the analysis the Northeast OMI data in the following 72 hours. Excluding the OMI data based on CG lightning data implicitly removes the data affected by cloud-to-cloud lightning collocated with CG lightning. The lightning filter removes about 2%, 27%, 20%, and 19% of OMI data, which are coincident with AQS data, for the West, the Midwest, the Northeast and the South, respectively. We refer to this version of OMI relative trend data as "Lightning filter".

3 Results and discussion

We group the analysis results into different regions: (a) West, (b) Midwest, (c) Northeast, and (d) South (Fig. 1), following the regional divisions by the United States Census Bureau. To make a fair comparison between the in situ and OMI-based trends, we only use spatially and temporally coincident in situ and OMI NO_2 observations in Section 3.1. The AQS data are temporally interpolated based on the overpassing time of the available OMI pixels which cover the corresponding AQS sites. Similarly, only OMI pixels covering the corresponding available AQS sites are used. The data from each dataset are then aggregated and averaged on a regional basis into time series to calculate the regional trends.

We apply the Mann-Kendall method with the Sen's slope estimator to calculate the relative trend of NO_2 for each season, i.e. DJF, MAM, JJA, and SON, during 2005-2014. We compute the uncertainties of the trends with the 95th percentile confidence intervals using the Mann-Kendall method. Note that when we compare in situ and OMI-based trends, the lightning filter also removes in situ NO_2 data, which are coincident with the OMI NO_2 data affected by lightning. This leads to slightly different in situ NO_2 trends between Fig. 4 and Fig. 6 (Section 3.2.3). We first compute the trends using the "Standard" OMI VCD data. The ocean trend removal, MODIS albedo update, and lightning filter are then added in sequence to compute three different OMI-based NO_2 trends (in addition to "Standard") to compare to the AQS in situ results. A subtlety in the comparison is that the coincident data change when the lightning filter is applied. As a result, the AQS in situ results in this set of comparison differ from those in the other three sets.

3.1 In situ and "Standard" OMI-based trends

Fig. 4 shows that both AQS in situ and "Standard" OMI-based seasonal relative trends are negative for all seasons across the regions. OMI-based trends generally underestimate the decreasing trends by up to 3.7% yr^{-1} (the absolute difference between relative trends) except for the large overestimation in the Midwest and the Northeast regions during DJF. The overestimates in these two regions are 3.0% yr^{-1} and 1.1% yr^{-1}, respectively. On average, the differences between OMI-based and in situ seasonal relative trends are 1.6% yr^{-1}, -0.3% yr^{-1}, 1.0% yr^{-1}, and 1.4% yr^{-1} for the West, the Midwest, the Northeast, and the South regions, respectively. Note that the relative trends are calculated using coincident measurements for the comparisons. The focus of this work is to examine if the differences between AQS in situ and OMI-based trends can be reduced.

3.1.1 Improvement due to ocean trend correction

After removing the ocean trend as discussed in Section 2.3.1, the OMI-based NO$_2$ decreasing trends are more pronounced as shown in Fig. 4 ("Ocean", blue line) by 0.1-0.9% yr^{-1}. The regional relative trends have different sensitivities to the ocean trend removal due to different tropospheric VCDs levels. In general, the discrepancies between OMI-based and in situ trends are reduced except for the Midwest and the Northeast regions during DJF, which are already biased low. The averaged differences between OMI-based and in situ seasonal relative trends for the West, the Midwest, the Northeast, and the South regions are 1.2% yr^{-1}, -1.1% yr^{-1}, 0.4% yr^{-1}, and 1.0% yr^{-1}. Only in the Midwest region, removing the ocean trend enlarges the difference due to the large winter bias.

3.1.2 Improvement due to MODIS albedo update

The adoption of the up-to-date MODIS albedo (Section 2.3.2) greatly reduces the difference of relative trends in the Midwest during DJF from -3.6% yr^{-1} ("Ocean") to 1.3% yr^{-1} ("MODIS"), the improvement of DJF trend difference is more moderate from -1.7% to 0.5% (Fig. 4). There are no significant changes of the comparisons in other regions or other seasons. Fig. 5 shows the albedo seasonal relative trends for the 4 regions coincident with AQS in situ NO$_2$ data. The OMI DOMINO v2 incorporates a climatology albedo dataset (Kleipool et al., 2008) with snow/ice albedo adjustment, in which the albedo value is reset to be 0.6 if snow/ice is reported in the NASA Near-real-time Ice and Snow Extent (NISE) dataset (Boersma et al., 2011). The climatology albedo data have no trends. Thus, the trends of albedo from the DOMINO v2 product mainly originate from the yearly variation of NISE detected snow/ice and to a lesser extent the OMI sampling variation. The noticeable seasonal trend of the OMI DOMINO v2 albedo dataset is the 3.9% yr^{-1} increase in DJF of the Midwest and a smaller DJF increase (1.0%) of the Northeast. In contrast, the MODIS albedo dataset exhibits a smaller positive DJF trend (0.8% yr^{-1}), 3.1% yr^{-1} less than the trend from DOMINO v2, in the Midwest, and a small negative DJF trend (-0.8%) in the Northeast. These differences suggest that using a fixed snow/ice albedo and climatology albedo data are inadequate. The comparison to the AQS data shows that the MODIS albedo update leads to better agreement between satellite and in suit trends in winter in these regions (Fig. 4).

3.1.3 Comparison after lightning event filter

As discussed in Section 2.3.3, lightning NOx affects the retrievals of satellite tropospheric NO_2 VCDs and NO_2 vertical mixing. Fig. 6 shows that the lightning filter significantly reduces the difference between the OMI-based relative trend and that of the AQS data by 0.5-1.4% yr^{-1} in the Northeast and 0.9-1.3% yr^{-1} in the South. As a result, the seasonal trend differences are within 0.9% yr^{-1} in these two regions except during SON. The Northeast is affected by the lightning filter due to lightning in this region and transport of lightning NOx from the South (Section 2.3.3). The lightning filter has little effect on the West and the Midwest. While lightning NOx can be significant during the monsoon season in some regions of the West (Fig. 3), the average tropospheric NO_2 VCDs are usually $< 1x10^{15}$ molecules cm^{-2} and lightning affected regions are therefore excluded in trend analysis.

The effect of lightning filter (Fig. 6) cannot be shown in Fig. 4 because the coincident OMI and AQS data points are fewer after applying the lightning filter. We examine the improvements of ocean trend removal, MODIS albedo update, and data screening with the lightning filter by comparing the differences of different OMI-based seasonal relative trends from the AQS in situ trends in Fig. 7. The previously discussed improvements such as OMI albedo update for the Midwest and the Northeast during DJF are shown. By subtracting the AQS trends, we can now find clear improvements of the lightning filter for the South and the Northeast. There remains seasonal variation of OMI-based trend biases relative to in situ data but the discrepancies of the annual trends after the three discussed procedures are relatively small at 0.3% yr^{-1}, -0.3% yr^{-1}, -0.1% yr^{-1}, and 0.0% yr^{-1}, in the West, the Midwest, the Northeast, and the South regions (Fig. 1), respectively. The remaining seasonal difference of the trends reflects in part the nonlinear photochemistry (Gu et al., 2013), the effects of NOx emission changes on NO_2 retrievals (Lamsal et al., 2015), different spatial coverages of the two measurements, and the inherent difference between trends of NO_2 tropospheric VCDs and surface concentrations.

3.2 OMI-based NO_2 trends

Table 1 summarizes the regional annual trends of coincident AQS in situ and OMI data. The "Standard" OMI data (following the DOMINO v2 algorithm) tend to show less NO_2 reduction than AQS data. After applying the three data processing procedures discussed in the previous section to the OMI data, the agreement with the AQS trends is within the uncertainties of the trends. While lightning NOx is part of OMI NO_2 observations, we treat the influence of lightning on the OMI tropospheric VCD trend as a bias for comparison purposes in this study. Table 1 shows the effects of data sampling when both AQS and OMI data are analyzed and when the lightning filter is applied.

Without the lightning filter, AQS decreasing trends are stronger than the decreasing trends of OMI data (Fig. 7). The lightning trend in the NLDN data is unclear due in part to the changing instrument sensitivity (Koshak et al., 2015). If lightning NOx is not accounted for in OMI retrieval, tropospheric NO_2 VCDs are overestimated. On the other hand, lightning accompanies low pressure systems which mix the atmosphere vertically and tend to reduce surface NO_2 concentrations when anthropogenic emissions are high such as urban and suburban regions. Therefore, lightning has opposite effects on surface and

satellite trends. The low-pressure dilution effect on surface NO_2 concentrations depends on anthropogenic emissions (since the end point of dilution is the background NO_2 value). Therefore, the weaker decreasing surface trends likely reflects a reduction of low-pressure dilution effect. Similarly, as anthropogenic emissions decrease, the positive bias of tropospheric VCDs due to lightning NOx becomes larger, likely resulting in weaker decreasing trends. We consider the lightning effects on surface NO_2 trends to be mostly meteorological driven not by lightning NOx directly (e.g., Ott et al., 2010; Lu et al., 2017) and hence the filtered OMI NO_2 data are likely closer to emission related concentration changes.

The AQS in situ NO_2 annual relative trends (coincident with OMI data with lightning filter) are most significant in the Northeast (-5.2±0.6% yr^{-1}) and the West (-4.2±0.5% yr^{-1}), followed by the South (-3.0±0.5% yr^{-1}) and the Midwest (-2.8±0.6% yr^{-1}) regions. The nationwide annual trend is -4.1±0.4% yr^{-1}, which is consistent with the previous studies (Lamsal et al., 2015; Lu et al., 2015; Tong et al., 2015; de Foy et al., 2016b; Duncan et al., 2016; Krotkov et al., 2016). The significant NO_2 reductions result from updated technologies and strict regulations (Krotkov, et al., 2016). The OMI-based NO_2 trends with the discussed procedures (coincident with AQS data) show similar reduction rates in the West (-3.8±0.4% yr^{-1}), the Midwest (-3.1±0.5% yr^{-1}), the Northeast (-5.3±0.7% yr^{-1}) and the South (-3.0±0.5% yr^{-1}) regions. The nationwide annual trend is -3.9±0.3% yr^{-1}.

One advantage of satellite observations over a surface monitoring network is spatial coverage. The processed OMI data ("Lightning filter") coincident with the AQS data show a national annual trend of -3.9±0.3% yr^{-1} similar to the AQS in situ trend of -4.1±0.4% yr^{-1}. Using all data available (Fig. 8, Table 1), the OMI data ("Lightning filter") show a much lower trend of -1.5±0.2% yr^{-1}, about half of the AQS trend (-3.9±0.4% yr^{-1}). Fig. 9 shows that the AQS sites, which are mostly urban and suburban sites, tend to be located in regions with high tropospheric NO_2 VCDs. The OMI decreasing trend is a function of tropospheric NO_2 VCDs, increasing from 0% yr^{-1} to -6% yr^{-1} (Fig. 9). The national annual trend is close to the value of clean regions which contribute much more than polluted regions. The larger decrease near the anthropogenic source regions reflect in part the nonlinear photochemistry (Gu et al., 2013) and in part to a stronger influence of NOx sources such as soils in rural regions.

4. Conclusions

Using data from the DOMINO v2 algorithm, we find that the computed OMI-based seasonal NO_2 (relative) trends underestimate the decreasing trends of the EPA AQS data by up to 3.7% yr^{-1}. While lightning NOx is part of OMI NO_2 observations, we treat the influence of lightning on the OMI tropospheric VCD trend as a bias for comparison purposes in this study. Furthermore, lightning NOx effects need to be removed when using satellite observations to understand the effects of changing anthropogenic emissions.

In this study, we show that removing the background ocean trend, adopting MODIS albedo data (with better temporospatial resolutions and characterization of snow/ice), and excluding lightning influences can bring OMI tropospheric NO_2 VCD trends in close agreement (within 0.3% yr^{-1}) with those of the AQS data. Among the corrections, the background ocean trend removal

is not as significant as the latter two. Since the origin of this trend is not yet clear, the ocean trend removal method may need updates in future studies. The remaining differences may result from the inherent differences between trends of NO_2 tropospheric VCDs and surface concentrations, different spatial sampling of the measurements, chemical-nonlinearity, and tropospheric NO_2 profile changes. The largest effects of MODIS albedo update are in winter in Midwest and Northeast and

5 those of lightning filter are in the South and the Northeast. After applying these data processing procedures, the derived OMI-based annual regional NO_2 trends change by a factor of > 2 for the South, the Midwest, and the West and seasonal changes can be even larger. We derive OMI-based NO_2 regional annual relative trends using all available data for the West (-2.0%±0.3 yr^{-1}), the Midwest (-1.8%±0.4 yr^{-1}), the Northeast (-3.1%±0.5 yr^{-1}), and the South (-0.9%±0.3 yr^{-1}).

The national annual trend of the processed OMI data is -1.5±0.2% yr^{-1}, about half of the AQS trend (-3.9±0.4% yr^{-1}). It

10 reflects that the AQS sites are mostly located in the urban and suburban regions, where OMI data show much larger decreasing trends (up to -6% yr^{-1}) than rural regions (down to 0% yr^{-1}). The reasons for the dependence of OMI derived trends on tropospheric NO_2 VCDs and the seasonal/regional trend differences are still not completely understood. Further studies are necessary to improve our understanding of these trends. The observation-based lightning filter implemented in this study is preliminary. Incorporating chemical transport modeling may improve this filter. Moreover, the results presented here represent

15 an alternative and indirect way to assess the importance of lightning NOx for National Climate Assessment (NCA) analyses described in Koshak et al. (2015), and Koshak (2017). Inversion studies (e.g., Zhao and Wang, 2009; Gu et al., 2013, 2014, 2016) will be needed to quantify the emission and AMF changes corresponding to the OMI tropospheric NO_2 VCD trends.

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25 **Data access**

The datasets used in this research have been obtained online as follows:

- DOMINO v2 NO_2 retrievals: http://www.temis.nl/airpollution/no2.html
- EPA AQS NO_2 data: US Environmental Protection Agency. Air Quality System Data Mart [internet database] available at http://www.epa.gov/ttn/airs/aqsdatamart.
30 - NLDN lightning data: https://lightning.nsstc.nasa.gov/data/data_nldn.html
- MODIS MCD43B3 data: https://lpdaac.usgs.gov/dataset_discovery/modis/modis_products_table/mcd43b3

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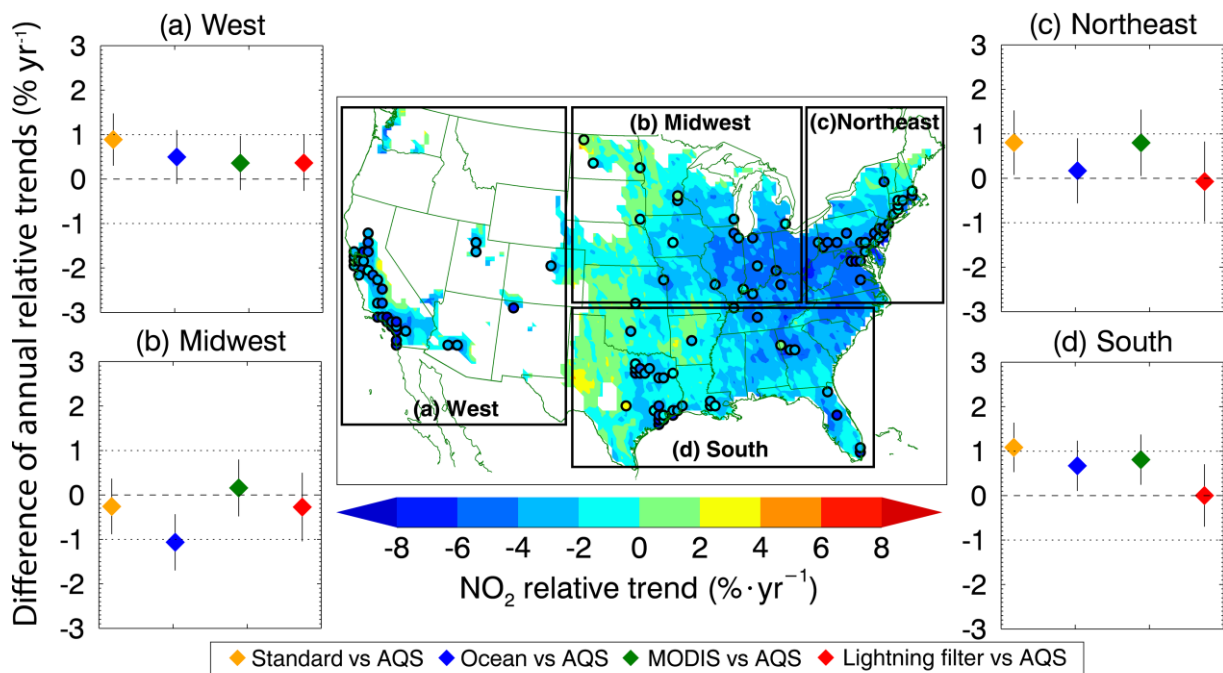
Figure 1. The solid black borders in the center map define the four regions used in this study. The colored background shows the OMI-based NO_2 annual relative trends of the "Lightning filter" data. Grid cells with 2005-2014 mean NO_2 VCD values < 1×10^{15} molecules cm^{-2} are excluded in this study and are shown in white. The black bordered circles represent the locations of AQS sites. Panel (a) through (d) show the regional difference (OMI-based relative trends minus AQS relative trends) of annual relative trends between coincident OMI-based and AQS in situ data. The colored diamonds are for "Standard" (orange), "Ocean" (blue), "MODIS" (green), and "Lightning filter" (red) OMI data, respectively. The different OMI VCD data are described in Section 2.4.
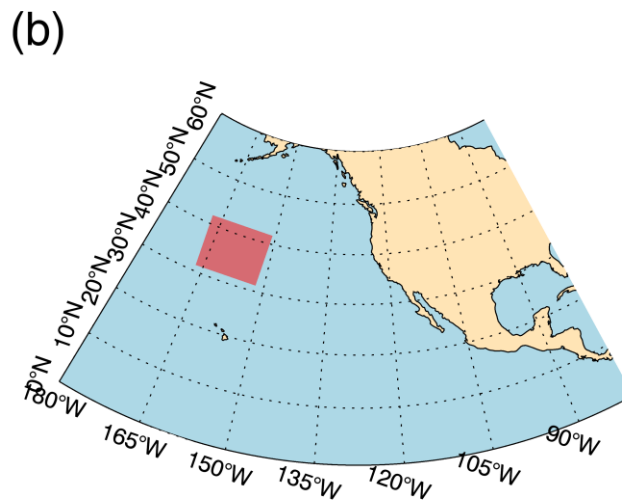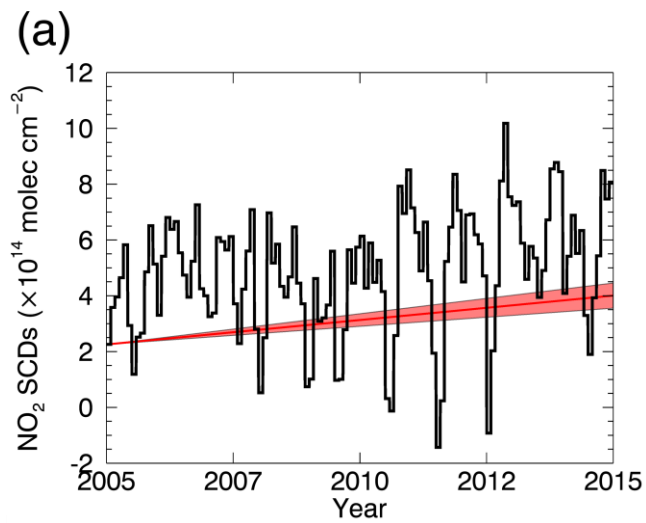
Figure 2. The black line in panel (a) shows the monthly averaged OMI tropospheric NO_2 VCD values in the North Pacific region (red box in panel (b)) from 2005 to 2014. The red line in panel (a) represents the ocean trend used in this research, with the 95th percentile confidence intervals shaded in red.
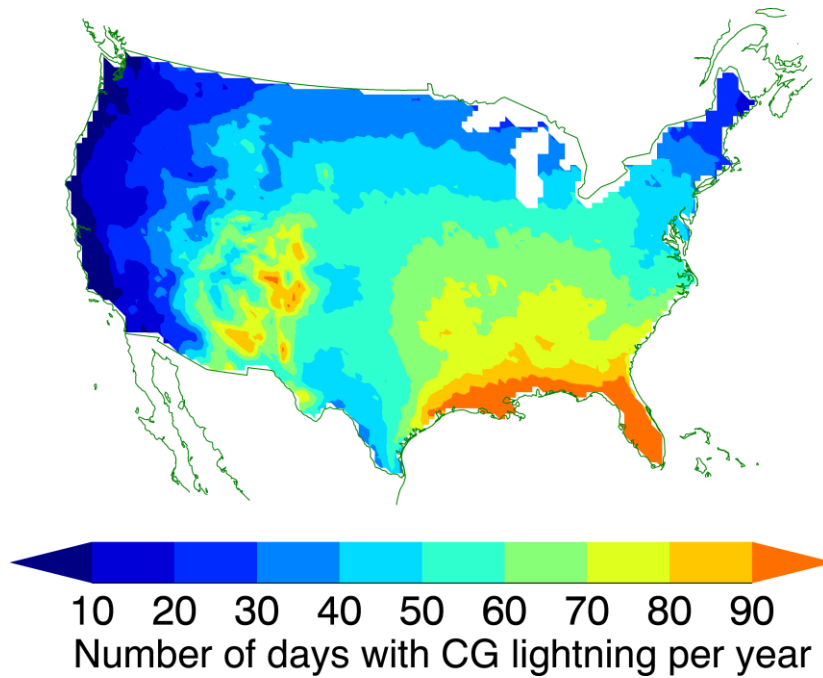
Figure 3. Number of days with NLDN detected cloud-to-ground (CG) lightning per model grid cell per year during 2005-2014. The lightning occurrences are calculated using the REAM grid resolution.
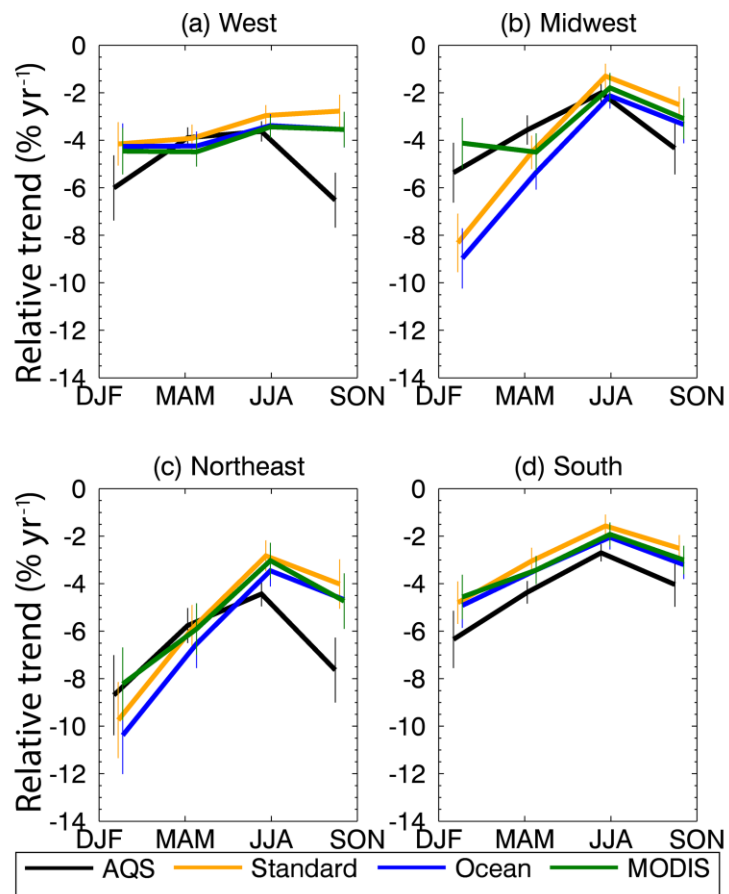
Figure 4. Seasonal relative trends of NO₂ calculated from the AQS in situ measurements ("AQS", black line) and those derived from different OMI VCD data ("Standard", orange line; "Ocean", blue line; "MODIS", green line). The error bars represent 95ᵗʰ percentile confidence intervals.
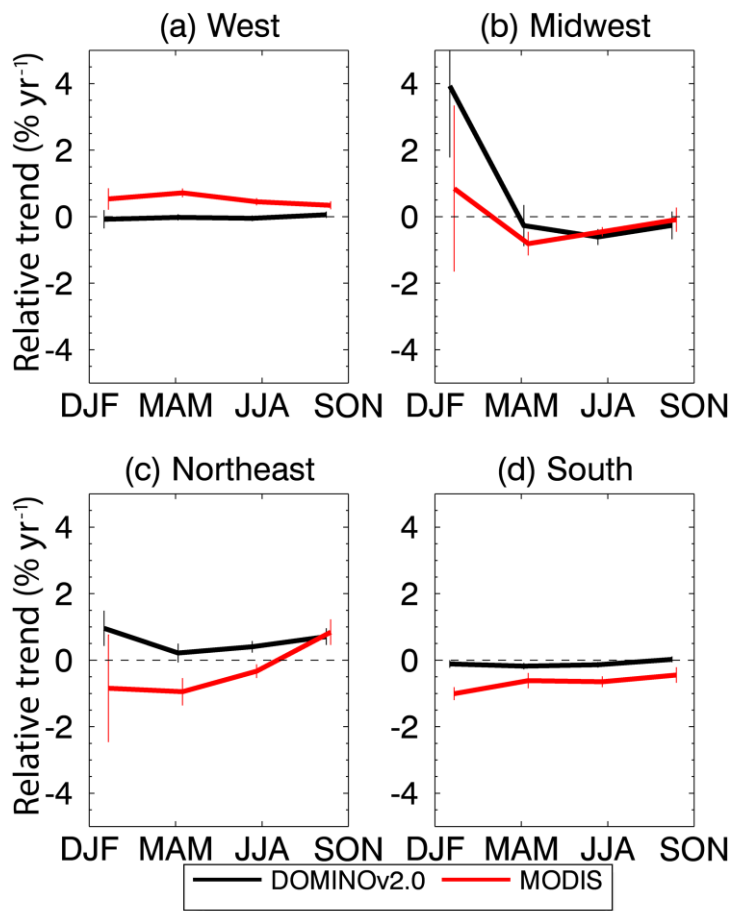
Figure 5. Seasonal relative albedo trends of OMI (black line) and MODIS (red line) surface reflectance products, coincident with AQS in situ data used in Figure 5. The error bars represent 95ᵗʰ percentile confidence intervals.
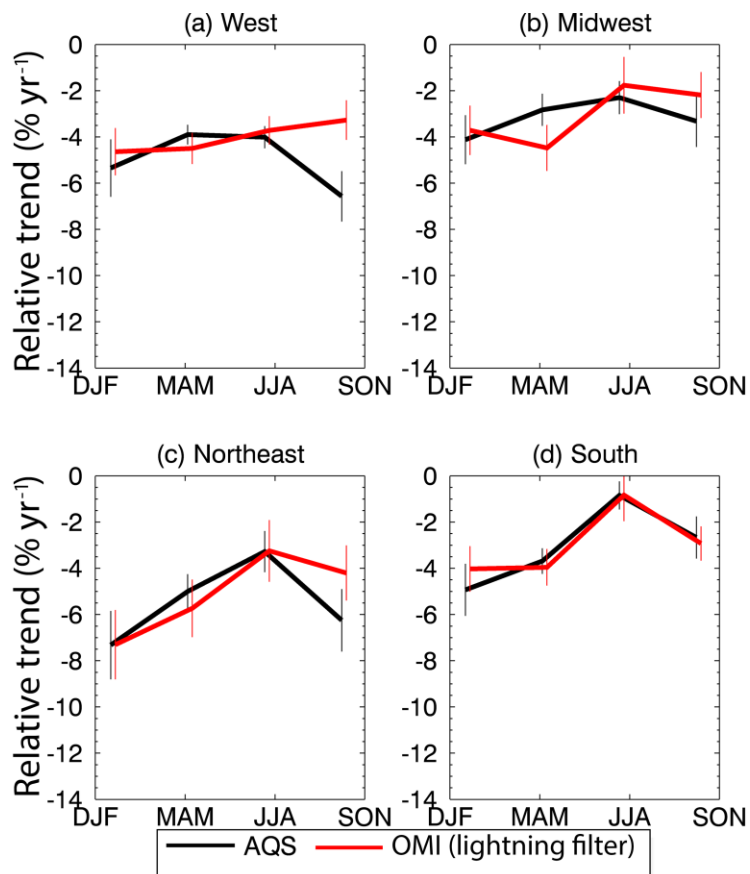
Figure 6. Seasonal relative trends of NO₂ calculated from the AQS in situ measurements ("AQS", black line) and those derived from OMI data after applying the lightning filter ("OMI (lightning filter)", red line). The error bars represent 95ᵗʰ percentile confidence intervals. The coincident data points are less than those used in Figure 5 and therefore the AQS trends are not the same.
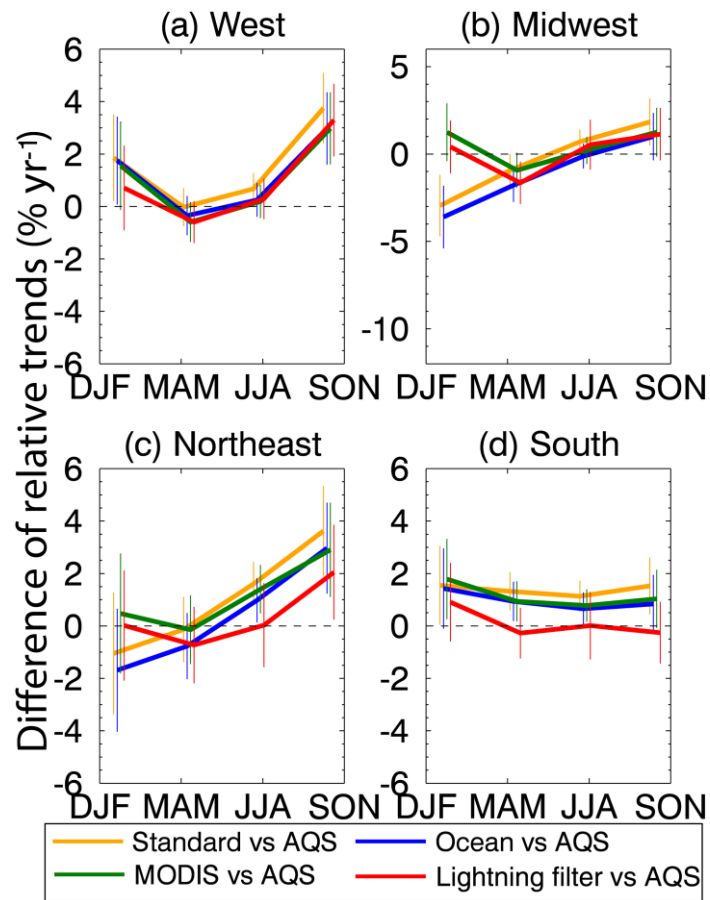
Figure 7. Seasonal differences of OMI-based relative trends from those computed from AQS in situ data. The error bars represent 95th percentile confidence intervals. The relative trends are shown in Figs. 4 and 6. The figure legends are the same as in Figs. 4 and 6 but with the AQS trends subtracted from the OMI-based trends.
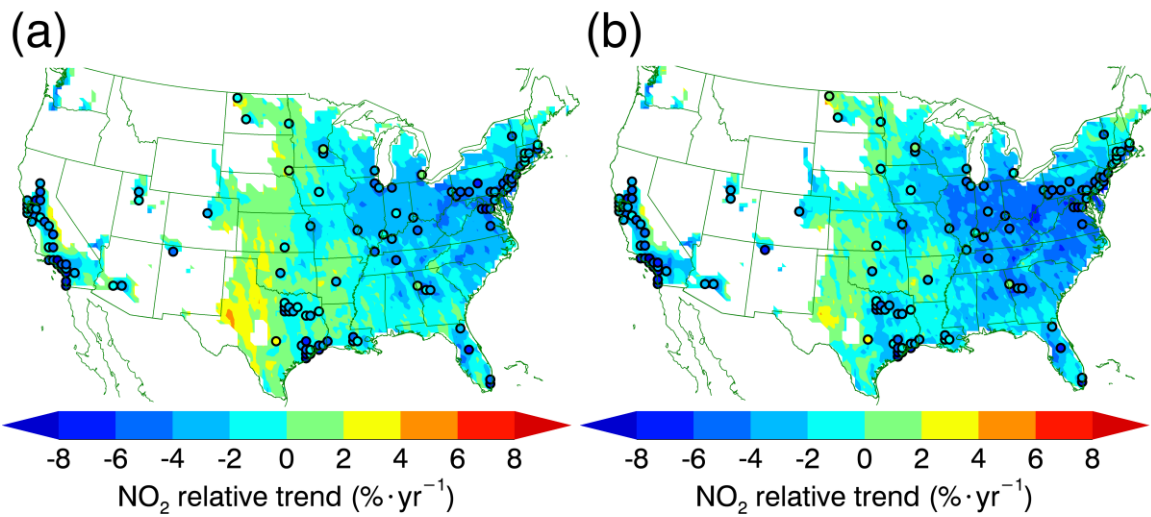
Figure 8: Annual relative trends of OMI-based NO₂ for "Standard" (a) and for "Lightning filter" (b) as the colored background. Black bordered circles indicate corresponding AQS NO₂ trends. Grid cells with 2005-2014 mean NO₂ VCDs < 1x10¹⁵ molecules cm⁻² are excluded in the analysis and are shown in white.
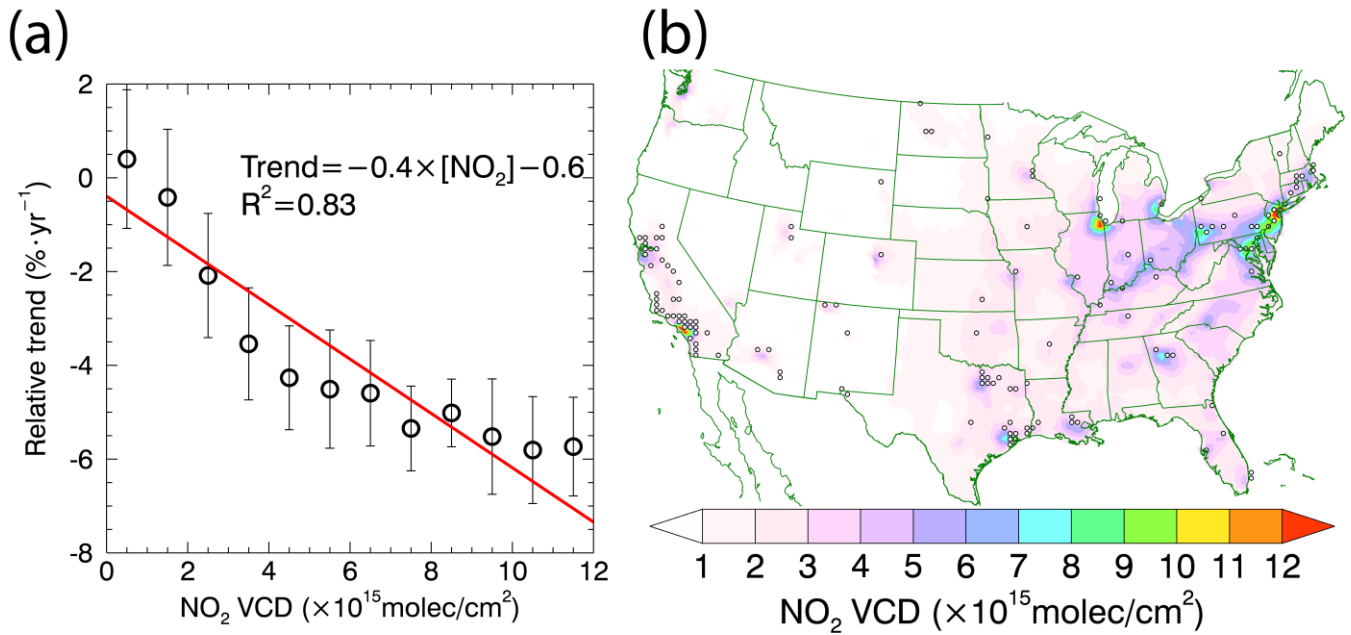
Figure 9. (a) The "Lightning filter" OMI-based NO₂ relative trend as a function 2005-2014 averaged OMI tropospheric NO₂ VCD binned every $1 \times 10^{15} molec/cm^2$. The error bars represent 95th percentile confidence intervals. The red line shows a least-squares regression. (b) The distribution of 2005-2014 averaged OMI tropospheric NO₂ VCD. Black bordered circles represent AQS sites. The OMI tropospheric NO₂ data ("Lightning filter") are used.

Table 1. Annual relative trends calculated with coincident data and all available data. The 95th percentile confidence intervals from Mann-Kendall method are also listed.

Region	Annual relative trends of coincident data (% yr^{-1})				Annual relative trends using all data (% yr^{-1})			
	Standard		Lightning filter[a]		Standard		Lightning filter	
	AQS	OMI	AQS	OMI	AQS	OMI[b]	AQS	OMI[b]
West	-4.1±0.5	-3.2±0.4	-4.2±0.5	-3.8±0.4	-4.1±0.5	-0.9±0.4	-4.2±0.5	-2.0±0.3
Midwest	-3.4±0.5	-3.6±0.4	-2.8±0.6	-3.1±0.5	-2.5±0.5	-0.9±0.4	-2.2±0.5	-1.8±0.4
Northeast	-5.8±0.5	-5.0±0.5	-5.2±0.6	-5.3±0.7	-4.7±0.5	-3.0±0.4	-4.1±0.5	-3.1±0.5
South	-3.8±0.4	-2.7±0.3	-3.0±0.5	-3.0±0.5	-3.5±0.4	-0.2±0.4	-3.0±0.5	-0.9±0.3
Nationwide	-4.3±0.4	-3.5±0.3	-4.1±0.4	-3.9±0.3	-4.0±0.4	-0.7±0.3	-3.9±0.4	-1.5±0.2

[a] These data include the three data processing procedures of this study, namely, ocean trend correction, MODIS albedo update, and lightning filter screening.

[b] The spatial coverage is shown in Figure 1.