An instrument for the rapid quantification of PM-bound ROS: the Particle Into Nitroxide Quencher (PINQ)

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Abstract. Reactive oxygen species (ROS) present on or generated by particulate matter (PM) have been implicated in PM-induced health effects. Methodologies to quantify ROS concentrations vary widely both in detection and collection methods. However, there is currently an increasing emphasis on rapid collection and measurement due to observations of short half-life ROS. To address this problem, this manuscript details the design and characterization of a novel instrument for the measurement of PM-bound ROS named the Particle Into Nitroxide Quencher (PINQ). This instrument combines the 9,10-bis (phenylethynyl) anthracene-nitroxide (BPEAnit) ROS assay in conjunction with a purpose-built aerosol collection device, the insoluble aerosol collector (IAC). The IAC continuously collects PM regardless of size or chemistry directly into a liquid sample with a collection efficiency of > 0.97 and a cut-off size of < 20 nm. The sampling time resolution of the PINQ is one minute, with a limit of detection (LOD) of 0.08 nmol.m^{-3} in equivalent BPEAnit-Me concentration per volume of air. This high sample time resolution and sensitivity is achieved due a combination of the highly concentrated IAC liquid sample, minimized liquid sample volume and the rapid reaction and stability of the BPEAnit probe.

1. Literature Review

1.1 Introduction

The World Health Organization (WHO) found air pollution was responsible for more than 7 million premature deaths in 2012. A large contributor to this mortality is atmospheric particulate matter (PM), which has been linked to: cardiovascular disease (Chow et al., 2006; Donaldson et al., 2001; Nel, 2005); increases in the prevalence of chronic respiratory disorders (Nel, 2005; Penttinen et al., 2001); and adverse effects in both embryonic and adult neuron activity (Morgan et al., 2011). Furthermore, studies have shown a link between exposure to PM in diesel exhaust and an increased lung cancer risk (Silverman et al., 2012).

A consequence of these and other discoveries was that in 2012, diesel exhaust was classed as a type 1 carcinogen by the International Agency for Research on Cancer (IARC) (Benbrahim-Tallaa et al., 2012).

A key mechanism used to explain these adverse health effects is oxidative stress (Ayres et al., 2008; Li et al., 2002). Reactive oxygen species (ROS) are a group of free radicals which can be either: present on the surface of PM (PM-bound ROS)(Hung
5   and Wang, 2001; Venkatachari et al., 2007); or generated through chemical reactions between PM and cells (endogenous ROS)(Donaldson et al., 2002). When inhaled, PM-associated ROS interact with cells to create oxidative stress, interrupting cell function and potentially leading to cell inflammation and death. The ultrafine particle range have been shown to be particularly hazardous in this respect. Their ability to penetrate deeper into tissues than their larger counterparts allows them to collect inside mitochondria, causing major structural damage to cells (Li et al., 2002).

10   Aerosols generated through combustion processes are of particular concern for oxidative stress as they are a major source of ultrafine particles (Brines et al., 2015; Posner and Pandis, 2015) and their potential for high ROS content (Cho et al., 2005; Kao and S., 2002; Mudway et al., 2005; Ristovski et al., 2012; Zhao and Hopke, 2012). Combustion emissions associated with vehicle emissions, power generation and biomass burning significantly contribute to the aerosol burden in many heavily populated urban areas (Harrison and Yin, 2000; Ma et al., 2017; Simoneit et al., 2004); making PM-associated ROS a key
15   issue in the assessment and understanding of the health impacts of air pollution. There have been many in vivo investigations involving both the monitoring of human exposure, and the exposure of animals and lab-cultivated cells to ROS-containing aerosols (Donaldson et al., 2002; Morgan et al., 2011; Sa et al., 2014; Shi et al., 2006; Shima et al., 2006). Whilst these studies provide us with an understanding of the aforementioned health impacts of PM-associated ROS exposure, they lack the ability to quantify the oxidative potential of aerosols. In order to achieve quantification, a range of different in vitro methodologies
20   have been developed.

1.2 Chemical Assays

The measurement of oxidative potential is a complex issue in terms of both sample collection and chemical analysis. In order for measurement to take place, PM must in most cases be collected into a liquid and mixed with a chemical probe. The degree to which this probe reacts with the sample is then measured in order to ascertain a value related to oxidative potential. Many
25   different chemical probes have been used for this purpose, with a detailed review found elsewhere (Hedayat et al., 2015). The reason for this diversity of probes is twofold. First, each probe is only sensitive to a specific range of ROS. Second, it is not known which ROS contribute to oxidative potential, and if they contribute equally. As such, the use of the term oxidative potential is contentious in the field of in vitro ROS measurement. While these probes measure a certain contribution to oxidative potential, none are not true measures of total oxidative potential. This does not mean results are not valuable, as there
30   have been good correlations between probe responses and oxidative stress response. However, all results must be considered in the context of the probe and methodologies used, and discussed accordingly.

Some probes, including p-hydroxyphenylacetic acid (POHPAA) (Hasson and Paulson, 2003) and ascorbic acid (AA) (Fang et al., 2016; Stoeger et al., 2009), are sensitive to transition metals. The dithiotreitol (DTT) assay is commonly used (Eiguren-

Fernandez et al., 2017; Fang et al., 2014; Gao et al., 2017; Li et al., 2002, 2009; Puthussery et al., 2018; Sameenoi et al., 2012, 2013) due to its ability to simulate the reaction responsible for the generation of ROS when PM interact with cells; making its quantification a measure of endogenous ROS. It is sensitive quinones and transition metals, as well as some water soluble organic compounds (WSOC) and humic-like substances (HULIS). DTT typically requires sample reaction times of between

5   60 minutes (Puthussery et al., 2018)~~of up to~~and 90 minutes ~~with a sample~~for accurate quantification (Cho et al., 2005), which potentially limits the time resolution of real time applications. However, ~~new~~developments in microfluidic sensors have ~~more recently~~allowed for DTT measurements with reaction times as low as 18 minutes (Koehler et al., 2014; Sameenoi et al., 2012). 2,7-dichlorofluorescein diacetate (DCFH-DA) combined with a horse radish peroxide (HRP) catalyst is currently the most commonly used ROS probe in literature for the measurement of PM-bound ROS (Fuller et al., 2014; Huang et al., 2016b,

10   2016a; King and Weber, 2013; Venkatachari et al., 2007; Venkatachari and Hopke, 2008; Wang et al., 2011; Wragg et al., 2016; Zhou et al., 2017). This is due to its relatively simple fluorescence-based quantification, potential for semi-continuous monitoring and sensitivity to several different ROS species. These species include organic peroxides, hydroperoxides, aldehydes, alcohols and hypochlorite. However, DCFH is prone to several issues, including: auto-oxidation upon exposure to air and sunlight (Stevanovic et al., 2012a); a high background fluorescence and a lack of sensitivity (Pal et al., 2012); a

15   relatively complex chemistry setup for implementation, especially in the case of online measurements (Huang et al., 2016a; Wang et al., 2011; Wragg et al., 2016; Zhou et al., 2017); the catalytic activity of the HRP catalyst is dependent upon sample composition (Pal et al., 2012); and a minimum reaction time of 11 minutes for quantification with HRP (Zhou et al., 2017), and a 60 minutes without HRP (Pal et al., 2012). These attributes make this probe very difficult to work with in the field, and highlight the need for an alternative ROS probe for accurate oxidative potential measurements.

20   Profluorescent nitroxides (PFNs) are a group of chemicals consisting of a nitroxide group bound to a fluorophore (Fairfull-Smith and Bottle, 2008). Initially these molecules are only weakly fluorescent, however upon reaction with free radicals they become highly fluorescent (Blinco et al., 2011). Quantification of these reactions can be achieved by comparing the fluorescence intensities before and after reactions. Several different PFNs have been developed at the Queensland University of Technology, one of which is BPEAnit. When placed into solution with dimethyl sulfoxide (DMSO), BPEAnit has been

25   shown to be sensitive to a broad range of PM-bound ROS (Stevanovic et al., 2012b), in particular those generated in combustion emissions (Miljevic et al., 2010; Stevanovic et al., 2012a). These include: carbon-centered radicals; nitrogen-centered radicals; sulphur-centered radicals; and peroxyl and hydroxyl radicals when in solution with Dimethyl Sulfoxide (DMSO) (Hedayat et al., 2015). Furthermore, its reaction is diffusion limited, allowing for comparatively quick measurement quantification. These characteristics make the BPEAnit a promising assay for the rapid online measurement of PM-bound

30   ROS.

1.3 Online Collection Techniques

Beyond chemical probes, there are several properties of ROS and combustion aerosols which complicate quantification. PM-bound ROS react readily with the atmosphere and other surroundings (Fuller et al., 2014). This is a significant issue with

standard filter capture techniques, as the concentration of PM-bound ROS can be skewed due to the decay of collected ROS during long periods of collection and storage prior to measurement (Fuller et al., 2014; Zhou et al., 2017). Therefore, methodologies involving either: long periods of collection; or delays between collection and measurement, risk severely underestimating the concentration of PM-bound ROS in an aerosol. Additionally, extraction processes to remove PM from

5   filters for analysis can introduce further positive and negative sample artefacts (Miljevic et al., 2014). To address these issues, methodologies have been developed to rapidly collect PM directly into liquid for more accurate quantification of PM-bound ROS.

Instrument	Chemical Assay	Collection Method	Insoluble Particles	Sample Flowrate	Time Resolution	Limit of Detection
ROS Sampling-Analysis System	DCFH	PILS	No	16.7 Lpm	10 min	n/a
OPROSI	DCFH	PC	No	5 Lpm	≤ 12 min	4 nmol.m^3
GAC-ROS	DCFH	GAC	No	16.7 Lpm	20 min	0.12 nmol.m^3
ROS Analyser	DCFH	PC	No	1.7 Lpm	8 min	2 nmol.m^3
o-MOCA	DTT	LSS	Yes	3 Lpm	3 hr	0.15 nmol.min^{-1}
Online DTT Monitoring System	DTT	PILS	No	16.7 Lpm	3 min	n/a
Online PM OP Instrument	DTT	PC	No	42 Lpm	1 hr	0.24 nmol.min^{-1}

Table 1 List of online ROS instruments and their key characteristics including: ROS Sampling-Analysis System (Venkatachari

10 *and Hopke, 2008; Wang et al., 2011); OPROSI (Fuller et al., 2014; Wragg et al., 2016); GAC-ROS (Huang et al., 2016a); ROS Analyser (Zhou et al., 2017); o-MOCA (Eiguren-Fernandez et al., 2017); ~~and~~ Online DTT Monitoring System (Koehler et al., 2014; Sameenoi et al., 2012) and Online PM OP Instrument (Puthussery et al., 2018). Limit of detection (LOD) for DCFH systems is given as equivalent H_2O_2 concentrations per cubic meter of air. O-MOCA LOD is given as DTT consumption per minute.*

1.3.1 Steam Collection Devices

Table 1~~Table 1~~ provides a list of the relevant online instruments developed for the quantification of different ROS species contributing to oxidative potential. Each instrument uses a different combination of either the DTT assay or DCFH assay with a method for direct capture of PM into liquid. The most predominant of these particle collection methods used are steam collection devices (SCD) (Khlystov, 1995; Kidwell and Ondov, 2010; Simon and Dasgupta, 1995), including: the Particle Into Liquid Sampler (PILS) (Orsini et al., 2003; Weber et al., 2001) and the Gas Aerosol Collector (GAC) (Dong et al., 2012). SCDs utilize condensational growth to achieve high efficiency collection of ultrafine particles. When in operation, a sample aerosol is continuously mixed with a precisely controlled flow of steam and cooled, creating a supersaturated aerosol. Particles inside this aerosol undergo condensational growth, increasing them to a size typically between 1 and 4 µm in diameter (Orsini et al., 2003). These grown particles are then collected into liquid with a very high efficiency through wetted impaction methods. The direct injection of steam allows for effective supersaturations to be achieved with high flowrates (Hering et al., 2014). For this reason, steam collection devices can achieve highly concentrated samples, as the ratio of aerosol sample flow to liquid collection flow can be maximized.

The PILS reports a collection efficiency of 97% particle number collection efficiency for particles between 0.03 – 10 µm (Orsini et al., 2003), while the GAC shows a total mass collection efficiency of > 99.5% (Huang et al., 2016a). The distinction between number collection efficiency and mass collection efficiency is important. Whilst the PILS system collects efficiency over all particle modes from fine to coarse, the GAC has a low collection efficiency for particles < 150 nm (Dong et al., 2012). This is due to high ultrafine losses inside a wetted annular denuder (WAD) (Simon and Dasgupta, 1993, 1995) used for the GACs gas phase measurements.

The collection method typically used to capture particles in SCDs involves the continuous deposition and washing off of particles on a solid surface. Hence, this method relies on particles being soluble in the collection liquid for its reported high collection efficiency. An alternative collection method was proposed in a publication focused on developing a system for the collection of virus aerosols. The impactor plate of a PILS was replaced with a wetted-wall PM cyclone which collects particles into a standing liquid vortex (Orsini et al., 2003). This prevents particles from depositing on a solid surface, allowing them to be captured into the sample liquid for analysis regardless of solubility. Another miniature wetted cyclone system designed to continuously wash the walls of the cone rather than directly collect particles into liquid has also been coupled to a PILS system (Peltier et al., 2007). Neither of these alternative collection methods have been adapted for ROS measurement.

1.3.2 Particle Collectors

The particle collector (PC) (Takeuchi et al., 2005) is another particle collection method shared by published PM-bound ROS devices. Names of similar devices include the mist chamber and aerosol collector. Variants of the PC are used in ~~both~~ the Online Particle-bound ROS Instrument (OPROSI)~~and~~, the ROS analyser, and the Online PM OP Instrument systems. The PC operates by collecting PM onto a hydrophilic filter which is continuously wetted with a fine mist of a capture solution. The

capture solution drips off the filter into a collection reservoir underneath, removing the water-soluble portion of the collected PM for analysis. The use of a filter allows for the collection of ultrafine PM with a very high efficiency (97.7 % (Takeuchi et al., 2005)) without the need of a condensational growth stage. A disadvantage of the PC system that the collection reservoir limits the time resolution of the system due to the continuous mixing of new and old sample.

1.3.3 Liquid Spot Sampler

The o-MOCA system (Eiguren-Fernandez et al., 2017) utilizes a Liquid Spot Sampler (LSS) (Hering et al., 2014) to collect particles for analysis. Similar to the SCDs, the LSS system utilizes a condensational growth process to collect ultrafine particles. However, the method of implementation is significantly different. Rather than mixing the sample aerosol with steam, the aerosol instead undergoes a three-stage water condensational growth process (Hering et al., 2014). The sample aerosol is drawn through a wet-walled tube where it passes through three independently-controlled temperature regions. The induced temperature differentials combined with diffusion of water vapour from the wetted walls of the tube lead to condensational growth of < 10 nm particles (Hering et al., 2014). The grown particles are impacted into a small liquid volume over the period of 3 hours before being pumped off for analysis.

The reliance on diffusion to introduce water vapour to the sample limits the operational flowrate, leading the LSS to use three separate 1 L.min^{-1} growth tubes for a total sample flow of 3 L.min^{-1}. The direct impaction into liquid removes the reliance of the PM to be soluble in the collection liquid, allowing for solubility-independent collection.

1.3.4 Measurement Technique Comparisons

Ultrafine PM concentrations are heavily dependent upon source proximity and atmospheric conditions (Sabaliauskas et al., 2013) which can lead to significant variations over short time periods and distances. This, coupled with the short half-life of some PM-bound ROS species (Fuller et al., 2014; Zhou et al., 2017), indicates that the concentration of PM-bound ROS is dynamic and prone to significant changes over short distances and times. Therefore, in order to accurately measure and understand the health impacts of PM-bound ROS, monitors must have time resolutions sufficient to accurately quantify these variations.

The time resolution of the discussed instruments varies from widely from 3 minutes to 3 hours. It is clear that the SCD and PC based instruments offer much higher potential time resolutions than that of the LLS-based OPROSI. Within the SCDs, the PILS-based systems have, at a minimum, half the time resolution of the GAC-ROS due to its dual gas and particle phase measurements. This dual measurement also causes the GAC-ROS system to have a low collection efficiency in the ultrafine PM range, indicating that a SCD without a gas phase measurement stage is a superior choice for PM-bound ROS measurements. The PILS-based Online DTT Monitoring System has the highest time resolution of all the discussed instruments. This is in part due to the total sample liquid volume being much lower than the PC-based systems as it does not require a reservoir. For this reason SCDs currently offer the highest potential for high sampling resolution instruments for PM-bound ROS measurements.

Aside from the OPROSI, a common disadvantage shared by all current systems is their inability to reliably collect PM insoluble in the collection solution for analysis. This is problematic as a significant portion of primary combustion emissions are hydrophobic (Popovicheva et al., 2008), making them insoluble in the aqueous-based solutions used for the DCFH and DTT probes. This can result in an underestimation in measurements which is not possible to correct for in post-analysis of the data using known losses. This is because the water insoluble fraction represents a distinct group of PM with both a separate physiological impact (Delfino et al., 2010), and a different ROS contributions (Verma et al., 2012). Future systems should endeavour to adapt or create new methodologies to allow for the collection of insoluble particles to improve the understanding and toxicity of aerosol oxidative capacity.

1.4 Manuscript Focus

This paper discusses the design and testing of a novel instrument called the Particle Into Nitroxide Quencher (PINQ). The PINQ has been developed to address the need for an accurate and repeatable method of measuring the PM-bound ROS present in aerosols. Informed from the current literature on the field, the instruments collection mechanism is based on a steam collection device, with PM collected directly into a miniature liquid vortex containing the BPEAnit ROS assay. The fluorescence increase of the BPEAnit probe is measured using a purpose-build flow through fluorimeter with a low internal volume and minimal flow-path length to ensure rapid quantification of PM-bound ROS.

2. Instrument Description

2.1 PINQ Layout
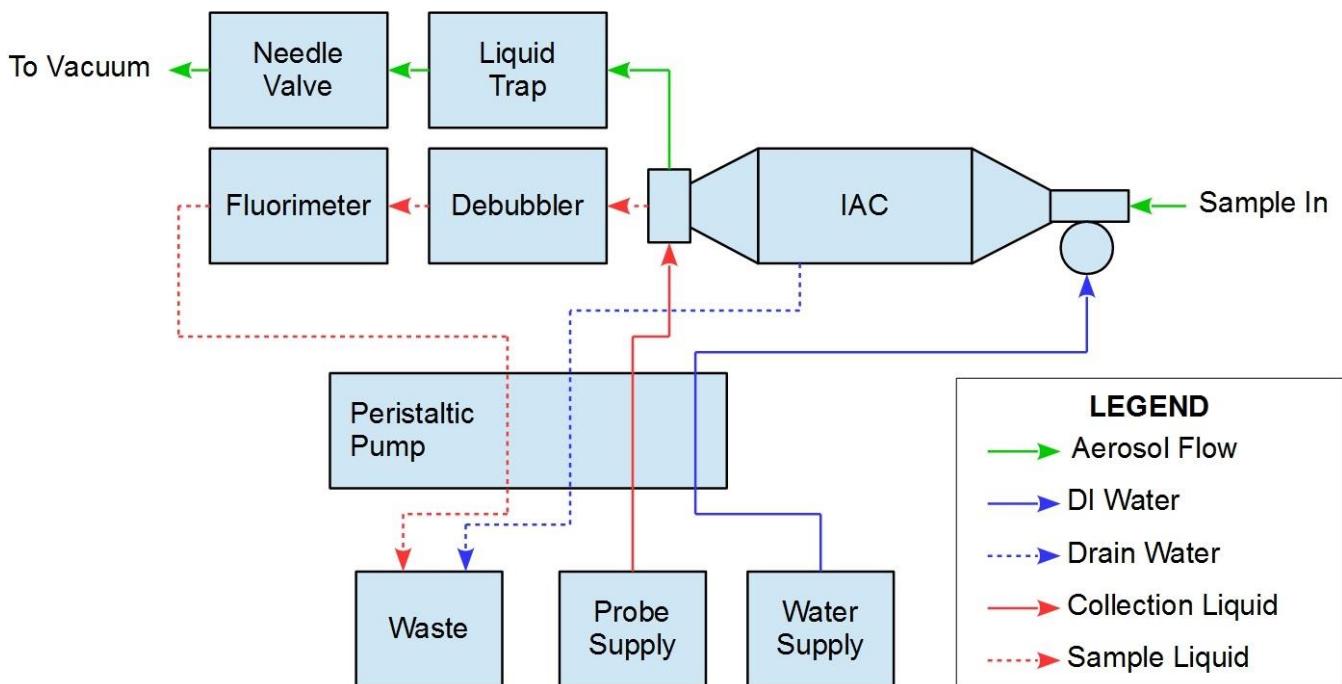
Figure 1 Diagram of the PINQ system showing all major components and the aerosol and liquid flow paths linking them.

5 The flow diagram of the PINQ can be seen in *Figure 1*~~Figure 1~~, including all key components of the system and their corresponding connections through liquid and aerosol flows. The aerosol is collected in a novel instrument called the Insoluble Aerosol Collector (IAC), in which PM is collected continuously into a liquid sample independent of size or chemical composition. The sample solution is then debubbled and passed through a purpose-build flow-through fluorimeter to quantify PM-bound ROS.

10 The standard sample flow rate for the instrument is 16.7 L.min^{-1} (1 m^3.hr^{-1}), which is set prior to sampling using a needle valve connected to vacuum. This has been found stable over 24 hour sampling periods. The corresponding steam generator water supply for this flowrate is 1.5 mL.min^{-1}. All liquid flows are controlled using a peristaltic pump and hence are given in volumetric rather than mass units for completeness. The influence of temperature variation on liquid flow during normal laboratory conditions (20 - 30 °C) results in a negligible impact on instrument performance. In more extreme temperature 15 environments it may be necessary to regulate liquid reservoir temperatures and correct for variations.

2.2 The Insoluble Aerosol Collector (IAC)

The PINQs aerosol collection stage, the IAC, is categorized as a steam collection device. The sample aerosol is mixed with water vapour to generate a supersaturated aerosol in which the PM undergo condensational growth. These grown particles are

collected with a high efficiency into a continuously flowing sample solution inside a solvent resistant vortex collector. The system can be divided into four main sections: inlet; growth chamber; vortex collector; and steam generator.
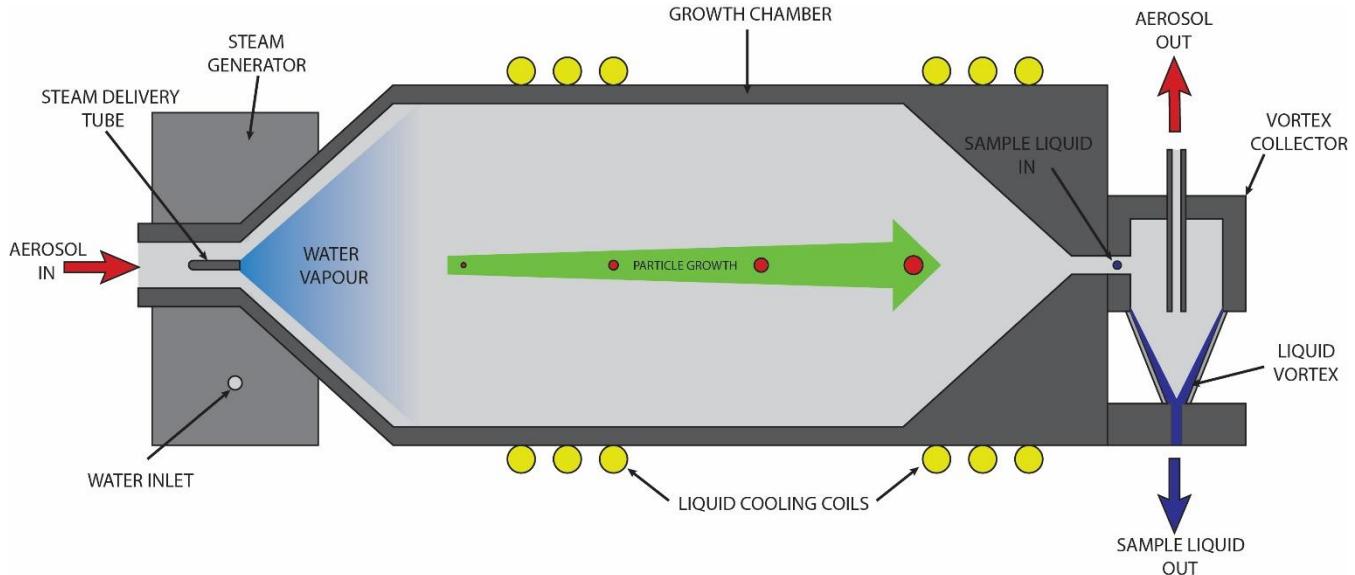
Figure 2 Diagram of the PINQs Insoluble Aerosol Collector (IAC) showing key components including: the aerosol inlet, steam generator, growth chamber and vortex collector.

2.2.1 Inlet

The IAC inlet is a 10 mm inner diameter tube containing a 1/16 in outer diameter aluminium steam delivery tube concentric to the inlet, oriented in the direction of flow. A precisely controlled mass flowrate of water vapour is injected into the aerosol flow through this delivery tube. The aerosol and water vapour are then turbulently mixed through a 45 degree expanding cone, creating a homogenous supersaturated mixture.

2.2.2 Growth Chamber

The supersaturated aerosol is passed through a cylindrical growth chamber (50 mm inner diameter, 140 mm length) to cool the mixture and allow time for particle condensational growth to occur. The chamber is actively liquid cooled to ambient temperature using a liquid cooling system and copper tubing compressed around the chamber to prevent any excess heat build-up in the system from the steam supply or aerosol sample. A drain at the base of the chamber removes excess condensation on the chamber walls. Grown particles exiting the growth chamber are accelerated through a 30 degree contracting cone into the vortex collector.

2.2.3 Solvent-Resistant Vortex Collector

The vortex collector is a specially designed aerosol capture device which collects particles into liquid regardless of solubility. The concept and dimensions are based on a wetted-wall cyclone described previously in literature (Orsini et al., 2008), with name vortex collector introduced here to differentiate it from the design of a standard miniature PM cyclone (Gussman et al.,
5   2002; Kenny et al., 2000). Both share similar geometries which are selected in order to create a stable air cyclone inside the device. However, whilst the PM cyclone deposits particles onto the solid surface of the cone, the vortex collector deposits particles directly into a collection liquid. This is achieved through the continuous injection and removal of liquid at the top and base of the cone, respectively. The inertial force of the air cyclone acts on the liquid as it passes through the cone, causing the liquid to spread on the cone wall and form a standing liquid vortex. Grown particles entering into the collector from the growth
10   chamber impact into this liquid vortex and are captured directly into the liquid stream.

The condensational growth stage prior to the collector removes the significance of the cut-off size and shape which characterize PM cyclones. Instead, the main requirement is that particles are grown above the size at which maximum collection efficiency is achieved. As such, the dimensions specified in the literature regarding PM cyclones can be altered to achieve the desired standing liquid vortex without risking performance. In particular, the aerosol inlet and exit holes can be narrowed to accelerate
15   the aerosol flow into the collector in order to better form a liquid vortex (Orsini et al., 2008). The location of liquid injection is also important, with the most stable vortex achieved when the liquid is injected into the aerosol stream before entering into the collector itself. Without these modifications the resulting collector does not form a standing liquid vortex for direct collection of PM into liquid (Peltier et al., 2007). Instead it could be more aptly described as a washed-wall cyclone, with a lower reported collection efficiency for insoluble particles.

20   Beyond these design points other parameters governing the performance of the vortex collector, including: liquid and aerosol flowrates; sample liquid; temperature; materials; and surface finish; making prediction of vortex formation difficult. To this end, the vortex collector was constructed using a combination of aluminium and stainless steel with a transparent fluorinated ethylene propylene (FEP) cone. This ensures the instrument is solvent resistant for use with DMSO in the PINQ system, whilst also allowing for visual confirmation of vortex formation.

25   This device was chosen over simpler impaction methods for two reasons. First, the vortex collector collects particles directly into liquid allowing for solubility-independent particle collection. Second, the liquid vortex ensures that the collected PM is homogenously mixed with the sample liquid exiting the cone base. When integrated into the PINQ, this factor coupled with the diffusion limited reaction between BPEAnit and collected ROS leads to the sample liquid exiting the base being fully reacted and ready for quantification.

30   ### 2.2.4 Steam Generator

A stable and precisely controlled steam flowrate is essential for effective condensational growth of ultrafine particles. Initially a design similar to that described in literature was used (Orsini et al., 2003), in which a regulated water flow was input into a

heated stainless steel tube. A similar design was attempted with both one and two stage heating, but was fraught with stability issues. This was attributed to the system having a low thermal mass, causing an unstable equilibrium point in the system. To remedy this issue, a new generator was proposed with a larger, insulated thermal mass.

The new generator design consisted of a cylindrical aluminium block sheathed in thermally insulating polytetrafluoroethylene (PTFE). The aluminium block contains a sealed internal chamber with narrow inlet and exit holes bored perpendicularly at the base and top, respectively. The block is maintained at a constant temperature by the use of two heater cartridges and a thermocouple driven with a Novus N1020 PID controller. A flow of water regulated by a peristaltic pump is continuously input into the base of the chamber through the inlet hole, where it vaporizes and expands, before being ejected through the exit hole. The aluminium steam delivery tube from the inlet section is inserted directly into the exit hole to minimize cooling and reduce system size. This design greatly improved stability and allows for long periods of unattended sampling.

2.3 Flow-Through Fluorimeter

A compact flow-through fluorimeter was designed to provide fast and accurate measurement of sample fluorescence for ROS quantification. The sample liquid fluorescence is continuously measured in a small flow-through quartz cell contained within a stainless steel housing. The excitation source is a 450 nm diode laser, and the fluorescence response measured perpendicularly with an Oceanoptics USB2000+ spectrometer. The laser and spectrometer are directly mounted to the stainless steel housing to remove the need for optical fibre connections, reducing setup complexity and minimizing size. The setup is controlled using a dedicated LabVIEW-based application.

3. Methodology

This study is divided into two investigations. The first is the measurement of the IAC collection efficiency. The second is the characterisation of the PINQ system fluorescence response. Flow diagrams of experimental setups, further methodology details, and calculations can be found in the supplementary material.

3.1 IAC Characterization

3.1.1 Particle Mass Collection Efficiency

The collection efficiency of the IAC for both fine ($PM_{2.5}$) and ultrafine ($PM_{0.1}$) particles was investigated through the comparison of mass concentration of ammonium sulphate collected with the IAC and those collected onto membrane filters. Sample aerosol was generated using a solution of ammonium sulphate in DI water in a Mesa Labs 6 Jet Collision Nebulizer. Sample was dried and diluted to provide the required sample flowrate before being sampled by the IAC, filter (Whatman Nuclepore polycarbonate membrane, 25 mm diameter, 0.2 µm pore size) and scanning mobility particle sizer (SMPS) (TSI 3071 Classifier, TSI 3772 Condensation Particle Counter, Aerosol Instrument Manager (AIM) software). IAC samples were collected into sealed quartz sample vials. Sample filters were collect and placed in quartz vials where they were sonicated for

5 minutes in 1 mL of solution. Samples were refrigerated until ion chromatography (IC) analysis was performed by the Central Analytical Research Facility (CARF) based at the Queensland University of Technology (QUT). Measurements were performed using a Dionex Integrion High Precision Ion Chromatography setup with a Dionex AS-AP autosampler. An AS18 Column (150 mm x 2 mm) with an isocratic 0.30 mL.min^{-1} flow was used. 0.1, 1, 5, 10, 20 and 100 ppm standards were used as calibration points.

3.1.2 Size Dependent Particle Number Collection Efficiency

The size dependent particle number collection efficiency was investigated using Di-Ethyl-Hexyl-Sebacat (DEHS) particles. This was chosen as the droplet activation diameter is not dependent on the composition of the particles if they are hydrophobic (Andreae and Rosenfeld, 2008). Particles were generated using a Mesa Labs Single Jet Collision Nebulizer containing a solution of DEHS in ethanol. Monodisperse particle sizes were size-selected from the resulting polydisperse aerosol using a TSI 3071 Electrostatic Classifier for sizes between 30 and 300 nm. The aerosol was then diluted to provide sufficient sample aerosol flowrate. A 3-way valve was used to rapidly switch between pre and post instrument sampling, with particle concentrations for each size found by averaging a 60 s sample from a TSI 3025 butanol-based CPC after system had stabilized. A nafion dryer was attached to the CPC inlet to prevent the high humidity of the sample influencing measurements.

3.2 PINQ Characterization

The PINQ was assembled for characterization as per the layout described in *Figure 1*~~Figure 1~~. A 1 µM solution of BPEAnit in DMSO was prepared as the sample solution for the PINQ. A 1.5 mL.min^{-1} steam generator water feed rate and a 1 mL.min^{-1} sample flowrate was used. The PINQ inlet was connected to a three way valve, with one port connected to a continuous supply of high purity nitrogen gas and the other connected to a combustion sample chamber. A cigarette was left to smoulder in the sample chamber for 5 min before being extinguished. The PINQ then sampled continuously, with the supply switched between nitrogen gas and cigarette smoke every minute for a total of eight samples.

The fluorescent response of the BPEAnit probe is expressed as an equivalent nanomolar increase in the concentration of the methyl adduct of the probe (BPEAnit-Me). The conversion factor for this is found through the slope of a calibration curve generated from fluorescence measurements of BPEAnit-Me standards of various concentrations. For PINQ measurements this equivalent response is then normalized against the ratio of liquid supply flowrate to aerosol sample flowrate; resulting in a measure of PM-bound ROS in equivalent concentration of BPEAnit-Me per volume of air in nmol.m^{-3} using *Equation 1*~~Equation 1~~. The limit of detection was calculated as three times the standard deviation of a set of 20 s blanks collected while sampling nitrogen.

Equation 1

$$C_{ROS} = FR * CF * \frac{q_{ls}}{q_A}$$

Where: C_{ROS} is the concentration of ROS in nmol.m^{-3}; FR is the fluorescence response of the spectrometer in unitless counts; CF is the calibration factor calculated from the calibration curve calculated as nmol.L^{-1}.counts^{-1}; q_{ls} is the liquid sample flowrate in L.min^{-1}; and q_A is the aerosol sample flowrate in m^3.min^{-1}.

4. Results and Discussion

4.1 IAC Characterization

4.1.1 Particle Mass Collection Efficiency

The IAC investigation was initially focused on mass collection efficiency in order to directly measure the fraction of aerosol collected into the sample liquid. As particle mass is proportional to the third order of particle size it was necessary to consider aerosols in terms of mass distribution when evaluating results. To this end, the averaged SMPS particle size distribution of the ammonium sulphate aerosol sample was fitted with a log-normal distribution and a corresponding mass distribution calculated. These distributions along with those of the ultrafine samples discussed below can be found in the supplementary material. The ultrafine particles (PM$_{0.1}$) correspond to 3 ± 5 % of the total mass of the sample, whilst the entire mass distribution was in the fine particle size range (PM$_{2.5}$), hence the collection efficiency calculated using this aerosol is referred to fine particle collection efficiency. The averaged collection efficiency and standard error for the collected efficiency for fine particles is calculated as:

$$CE_{fine} = 1.00 \pm 0.04$$

As the majority of the mass in the previous tests was contained in the particles larger than 100 nm (~97 %) this result does not necessarily indicate a high collection efficiency of ultrafine particles. Therefore, these experiments were repeated with the larger ammonium sulphate particles removed from the sample using an impactor as detailed in the supplementary Sect. S2.1. The averaged SMPS particle size distribution over the sample collection period was fitted with a log-normal distribution multiplied by a logistic function to emulate the size cut-off generated by the impactor, and a corresponding mass distribution was calculated. The ultrafine particles (< 100nm) correspond to 80 ± 10 % of the total mass of the sample, which is considerably higher than the previous size distribution.

The generated particle mass concentrations with the impactor in front of the atomiser (ultrafine particle experiment) were over 60 times lower than the PM$_{2.5}$ experiments. In order to ensure the collected samples contained measurable concentrations of sulphate ions, the flow through the vortex collector was reduced from 1.00 ml.min^{-1} to 0.15 ml.min^{-1}. This increased the residence time of the sample liquid in the vortex collector for the same aerosol flowrate, resulting in a more concentrated sample.

The averaged collection efficiency and corresponding standard error are calculated as:

$$CE_{ultrafine} = 1.05 \pm 0.06$$

The error of the ultrafine collection efficiency is larger than that of the fine collection due to the significantly lower aerosol sample concentration. However, the results are within error of each other, indicating that the IAC collects both fine and

ultrafine particles with a very high efficiency. This is an expected result for steam collection devices, which will typically collect all particle sizes with equally high efficiency due to condensational growth of particles in the growth chamber well into the ultrafine size range.

4.1.2 Steam Dilution Factor

5   The sample liquid entering into the IAC was doped with a known quantity of sodium chloride. The steam dilution factor (SDF) was determined from the change in the chloride ion concentrations before and after the liquid feed passed through the vortex collector. Relevant measurements and individual calculated steam dilution factors for each sample can be found in the supplementary material. The average steam dilution factor for a standard liquid supply of 1 mL.min^{-1} (SDF_{SF}) was calculated as:

$$SDF_{SF} = 0.882 \pm 0.004$$

This corresponds to a 0.134 ± 0.007 mL.min^{-1} contribution of condensed water into the sample flow.

The averaged steam dilution factor for a lowered liquid supply flowrate of 0.15 mL.min^{-1} (SDF_{LF}) was calculated as:

$$SDF_{LF} = 0.527 \pm 0.007$$

SDF_{LF} is lower than that that of the SDF_{SF}. This reduction indicates that the ratio of condensed water to sample flowrate has

15   increased in the liquid sample exiting the PINQ. Despite this, the effective contribution of condensed water to the sample flow for SDF_{LF} was calculated as 0.136 ± 0.007 mL.min^{-1}; which is within error of the condensation water contribution calculated for SDF_{SF}. This shows that the condensation rate in the sample is independent of: liquid sample flowrate; PM mass; and particle number concentration (PNC). This indicates the principle source of condensation in the sample is from inside the cyclone itself, which is directly proportional to the water flowrate input into the steam generator. This ensures that the steam dilution

20   factor can be accurately corrected for in all measurements.

Lowering the liquid supply flowrate allows for higher concentrations of aerosol in the sample stream. In theory this could be further applied to improving the sensitivity of the PINQ. However, the corresponding increasing influence of steam dilution presents a limitation in the potential application of this. It is essential to keep the contribution of condensed water below 30% of the total sample (SDF > 0.7) when using a DMSO solution of the BPEAnit probe as larger percentages of water in DMSO

25   will cause a nonlinear change in the fluorescent signal.

4.1.3 Size Dependent Hydrophobic Particle Collection Efficiency

Ammonium sulphate was selected as the test aerosol for the mass collection efficiency investigation due to its shape being close to spherical (Biskos et al., 2006) and detectability with ion chromatography. However, due to its high hygroscopicity it will undergo condensational growth at a much lower supersaturation than those required for hydrophobic aerosols

30   (Popovicheva et al., 2008). To ensure that the collection efficiency is independent on the chemical composition of particles, measurements of the number concentration collection efficiency with hydrophobic particles were conducted.

Size dependent hydrophobic particle number collection efficiency was measured through the size preclassification of DEHS aerosol, and corresponding measurements of PNC entering and exiting the PINQ. Collection efficiency was corrected for inlet and chamber losses which were estimated to be 1 minus the lower limit of the ultrafine mass collection efficiency. The calculated particle number collection efficiencies at each tested size are shown in Figure 3.
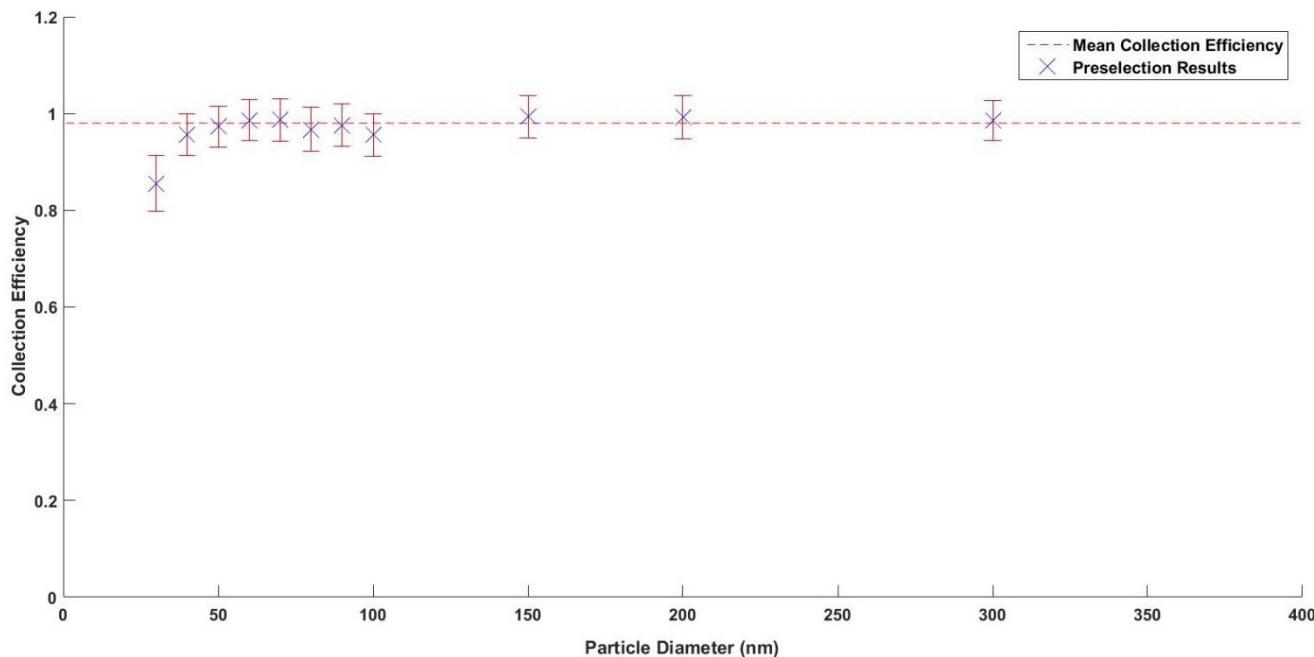
Figure 3 Calculated collection efficiencies for DEHS preselection measurements with standard error bars and mean collection efficiency

The mean collection efficiency was calculated as:

$$CE_{PNC} => 0.97$$

10   As it was not possible to generate sufficient DEHS test particles smaller than 30nm, the cut-off size corresponding to 50% of the maximum collection efficiency can only be extrapolated from the above graph to be < 20 nm. This is an improvement over the earlier vortex-based wetted-wall cyclone design which reported a collection efficiency of > 0.88 with a cut-off size of 30 nm (Orsini et al., 2008). It is also higher than the 0.69 collection efficiency reported by the washed miniature cyclone developed for organic carbon measurements (Peltier et al., 2007).

15   An important note in these comparisons is that the two variants found in literature both used direct measurements of insoluble particles collected in liquid after exiting the instrument. The collection efficiency calculated here is the efficiency of initial collection into the liquid, and would need to be corrected for losses in the liquid flow path between collector and detector for a full comparison. This was not explored as liquid line losses are not an issue in the application of the IAC in the PINQ. Reaction with the BPEAnit probe is effectively completed instantaneously upon collection, meaning particles lost in the liquid

20   path have already fully reacted with the BPEAnit probe.

4.2 PINQ Characterization – Response Time and LOD

The response time of the PINQ was separated into two parameters: a time delay between a source change and the corresponding start of change in fluorescence; and the mixing time taken for the fluorescence response to become completely independent of the previous source. An example of the fluorescence response of the BPEAnit probe when the instrument was switched from the ROS source (cigarette smoke) to nitrogen along with relevant key times is shown in Figure 4. In the 8 samples taken the signal strength of the cigarette smoke was approximately 60 times the magnitude of the background noise.
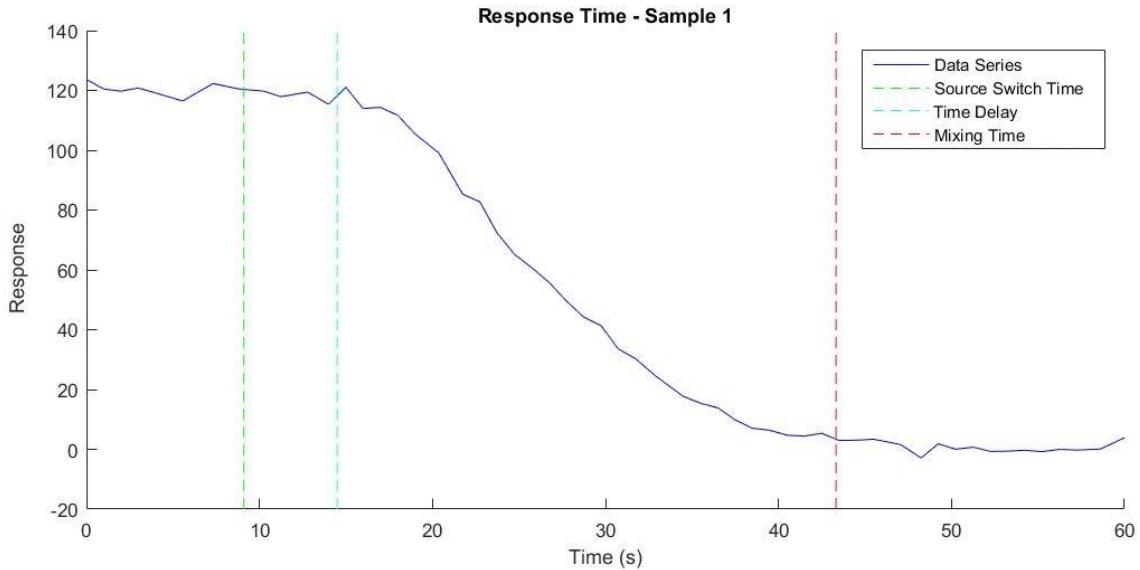
Figure 4 the fluorescence data series over time for Response Time - Sample 1, with: the time the source was switched from cigarette smoke to nitrogen (Source Switch Time); the time at which the fluorimeter began to measure a decrease in signal (Response Time Delay); and the time at which the signal reduced to background (Mixing Time Delay)

The averaged values of time delay and mixing time of the PINQ for the eight samples taken and corresponding standard errors were calculated as:

$$time\ delay = 6.3 \pm 0.6\ s$$
$$mixing\ time = 32 \pm 1\ s$$

The time delay of 6.3 s represents a small time correction factor which does not strongly influence instrument performance. In contrast, the mixing time of 32 s is a key value in determining the limits of instrument time resolution. This value indicates the minimum averaging time which provides independent data points, and is directly proportional to both: the liquid sample flowrate; and the internal liquid volume of the liquid sample flow path inside the instrument. An increased liquid sample flowrate above the tested 1 mL.min^{-1} will result in a faster time resolution. However, this will do so at the cost of a lower ratio of collection liquid flowrate to aerosol sample flowrate, diluting the aerosol and reducing instrument sensitivity. Internal liquid volume was minimized through the use of small bore 1/16" OD tubing for all flowpaths and a custom-made low internal

volume fluorescence cell. An unavoidable limiting factor in reducing the mixing time is the internal volume of the cyclone, which cannot be minimized further without impacting instrument performance.

The results of this experiment indicate that the system requires a total of ~ 40 seconds to reach equilibrium response when connected to an aerosol source. When performing measurements on the PM-bound ROS it is necessary to account for any contributions from the gas phase by alternating sampling between the total aerosol source, and the source after filtering to remove the particle-bound ROS. Therefore, the time resolution of the PINQ is determined by the sum of the total response time, and the time required to average the signal for an accurate measurement. An averaging period of 20 seconds was selected, leading the PINQ to have a time resolution of 1 minute when performing alternating gas and total phase measurements (See Sect. 4.3). A total of 234 blanks of 20 seconds averaging time sampled with nitrogen were collected and normalized as equivalent nmol of BPEAnit-Me per cubic meter of air as detailed in Sect. 0. The limit of detection of the PINQ was determined as three times the standard deviation of these blanks.

$$LOD = 0.08 \; \frac{nmol}{m^3}$$

The PINQ has a lower LOD than other online systems with responses normalized per volume of air sampled for two primary reasons. First, the IAC has a very small ratio of liquid sample flow to aerosol sample flow. The ensures a more highly concentrated sample than those of particle collector-based systems, which reported LODs between 4 nmol.m^{-3} (Wragg et al., 2016) and 2 nmol.m^{-3} (Zhou et al., 2017). The GAC-ROS system has similar aerosol and liquid flows, resulting in similarly concentrated sample and hence a closer LOD of 0.12 nmol.m^{-3} (Huang et al., 2016a). The second factor contributing to the lower PINQ LOD is the different ROS probe used. The BPEAnit probe does not autooxidise in the same manner that the DCFH probe used by the other system discussed here, resulting in more stable blanks and hence a lower LOD.

4.3 PINQ Response –Particle and Gas Phase Contributions

As has been shown in Sect 4.1. the IAC stage of the PINQ collects PM with a very high collection efficiency. However, experiments with the PINQ system also show sensitivity to gas phase products. In order to accurately quantify the PM-bound ROS the PINQ response must be corrected for the gas phase contributions during measurements. This is achieved by alternately sampling with and without a HEPA filter, resulting in both particle and gas phase measurements. This method was chosen over the use of a gas denuder for two reasons. First, gas denuders can cause significant losses of ultrafine particles (Stevanovic et al., 2015). Second, the gas phase can contain significant concentrations of ROS (Stevanovic et al., 2017) which cannot be investigated if scrubbed with a gas denuder.

The exact mechanism and efficiency with which the IAC collects the gas phase contribution is unknown. Likely it is a combination of: the collection of water-soluble gas species during the condensational growth stage; and DMSO soluble species during interactions with the collection liquid inside the vortex collector. The quantification of these effects is a complex issue as it will likely be dependent upon each gas species solubility in both water and DMSO. As such the gas phase response is

considered semi-quantitative, with its characterization beyond the scope of this manuscript. Hence the PINQ is presented here as an instrument for the measurement of PM-bound ROS only.

5. Summary

This manuscript's first focus was on the design and characterization of the IAC as a high efficiency aerosol collector for use in the PINQ system. It is defined as a steam collection device, in which the sample aerosol is continuously mixed with a stream of water vapour to generate a supersaturated mixture, growing the particles into large liquid droplets to ensure high efficiency capture independent of initial particle size. The grown droplets are collected continuously into a liquid vortex inside a specially designed vortex collector. This was designed over a simpler impaction system as it allows the capture of insoluble in the collection liquid. The vortex collector is also solvent resistant and allows for visual confirmation of the liquid vortex.

The IAC mass collection efficiency was determined to be within error of 1.00 for both fine and ultrafine ammonium sulphate particles. This result indicates that provided particles seed liquid droplets, they are collected with a very high collection efficiency. However, hydrophilic particles like ammonium sulphate will undergo condensational growth more readily than hydrophobic particles. As the IAC must collect particle independent of chemical composition for use in the PINQ system, the number collection efficiency of highly hydrophobic DEHS test particles was also investigated. Using these particles the number collection efficiency was found to be > 0.97 with a cut-off size of < 20 nm. This result shows that the IAC is capable of collecting particles with a high efficiency independent of particle size and composition.

The PINQ was developed to measure PM-bound ROS using the BPEAnit chemical probe in conjunction with the IAC. A 1 µM solution of BPEAnit in DMSO is used as the sample collection liquid, with particles collected directly into the probe solution inside the IAC vortex collector. The rapid mixing of liquid inside the vortex coupled with the diffusion limited reaction between the probe and any ROS collected ensures the liquid exiting the IAC is fully reacted. The sample liquid is then debubbled and input into a specially designed flow-through through fluorimeter. Finally, the fluorescence response measured is converted into PM-bound ROS through calibrations performed on known concentrations of BPEAnit-ME, and are expressed in nmol.m^{-3}.

The PINQ is also sensitive to gas phase ROS. Therefore, in order to quantify PM-bound ROS it is necessary to account for gas phase contributions by periodically filtering the aerosol sample. Experiments on response time with a standard sample flowrate of 1 mL.min^{-1} indicate that after switching sources the PINQ signal takes ~ 40 seconds to reach equilibrium response. Therefore, with a 20 second sample averaging time the sample time resolution of the instrument is one minute when alternating between filtered and unfiltered samples. With this time resolution the LOD of the instrument was determined to be 0.08 nmol.m^{-3}. Both the time resolution and LOD of the instrument are considerably lower than other instruments currently found in literature, indicating the PINQ is a viable candidate for the quantification of PM-bound ROS.

Data availability

Collection efficiency of PINQ data is available as part of an online Supplement. For a more detailed description on the design and construction of the instrument or interest in instrument evaluation, please contact the corresponding author.

5 **Author contribution**

R. A. B Improved and finalised the design of PINQ; Made experimental set-up, performed data analysis and wrote the manuscript

S.S Made initial design of the device; contributed to the experimental design, assisted with data interpretation; reviewed the manuscript

10 S.B Reviewed the manuscript; contributed to the experimental design and data interpretation.

Z. D. R Came up with the original idea; contributed to the experimental design, assisted with data interpretation; reviewed the manuscript

Competing interests

The authors declare that they have no conflict of interest.

15 **Acknowledgements**.

The authors wish to thank the following groups/institutions for their contribution to this work: the aerosol physics group at Tampere University of Technology (TUT) for their assistance in the characterisation of the initial version of the device; Central Analytical Research Facility (CARF) for their assistance with ion chromatography analysis, and the Design and Manufacturing 20 Centre (DMC) team from Queensland University of Technology for their help with the design and building of this device. This research was funded by: ARC Discovery grant DP120100126 Fundamental study into the role of the organic fraction on the toxicity of combustion generated airborne particles and EU FP7 Project ID: 308524, CITI-SENSE Development of sensor-based Citizens' Observatory Community for improving quality of life in cities.

6. References

Andreae, M. O. and Rosenfeld, D.: Aerosol-cloud-precipitation interactions. Part 1. The nature and sources of cloud-active aerosols, Earth-Science Rev., 89(1–2), 13–41, doi:10.1016/j.earscirev.2008.03.001, 2008.

Ayres, J. G., Borm, P., Cassee, F. R., Castranova, V., Donaldson, K., Ghio, A., Harrison, R. M., Hider, R., Kelly, F., Kooter, I. M., Marano, F., Maynard, R. L., Mudway, I., Nel, A., Sioutas, C., Smith, S., Baeza-Squiban, A., Cho, A., Duggan, S. and Froines, J.: Evaluating the toxicity of airborne particulate matter and nanoparticles by measuring oxidative stress potential--a workshop report and consensus statement., Inhal. Toxicol., 20(1), 75–99, doi:10.1080/08958370701665517, 2008.

Benbrahim-Tallaa, L., Baan, R., Grosse, Y., Secretan-Lauby, B., El Ghissassi, F., Bouvard, V., Guha, N., Loomis, D. and Straif, K.: Carcinogenicity of diesel-engine and gasoline-engine exhausts and some Nitroarènes, Pollut. Atmos., 13(SPEC. ISS.), 43–44, doi:10.1016/S1470-2045(12)70280-2, 2012.

Biskos, G., Paulsen, D., Russell, L. M., Buseck, P. R. and Martin, S. T.: Prompt deliquescence and efflorescence of aerosol nanoparticles, Atmos. Chem. Phys., 6(12), 4633–4642, doi:10.5194/acp-6-4633-2006, 2006.

Blinco, J. P., Fairfull-Smith, K. E., Morrow, B. J. and Bottle, S. E.: Profluorescent Nitroxides as Sensitive Probes of Oxidative Change and Free Radical Reactions, Aust. J. Chem., 64(4), 373–389, doi:10.1071/CH10442, 2011.

Brines, M., Dall'Osto, M., Beddows, D. C. S., Harrison, R. M., Gómez-Moreno, F., Núñez, L., Artíñano, B., Costabile, F., Gobbi, G. P., Salimi, F., Morawska, L., Sioutas, C. and Querol, X.: Traffic and nucleation events as main sources of ultrafine particles in high-insolation developed world cities, Atmos. Chem. Phys., 15(10), 5929–5945, doi:10.5194/acp-15-5929-2015, 2015.

Cho, A. K., Sioutas, C., Miguel, A. H., Kumagai, Y., Schmitz, D. A., Singh, M., Eiguren-Fernandez, A. and Froines, J. R.: Redox activity of airborne particulate matter at different sites in the Los Angeles Basin, Environ. Res., 99(1), 40–47, 2005.

Chow, J. C., Watson, J. G., Mauderly, J. L., Costa, D. L., Wyzga, R. E., Vedal, S., Hidy, G. M., Altshuler, S. L., Marrack, D., Heuss, J. M., Wolff, G. T., Pope, C. A., Dockery, D. W., Pope, rd C. A. and Dockery, D. W.: Health effects of fine particulate air pollution: lines that connect, J. Air Waste Manag. Assoc., 56(6), 709, doi:10.1080/10473289.2006.10464545, 2006.

Delfino, R. J., Staimer, N., Tjoa, T., Arhami, M., Polidori, A., Gillen, D. L., George, S. C., Shafer, M. M., Schauer, J. J., Sioutas, C., Ralph, C., Delfino, J. and Sioutash, C.: Associations of Primary and Secondary Organic Aerosols With Airway and Systemic Inflammation in an Elderly Panel Cohort, Source Epidemiol., 21181108(6), 892–902, 2010.

Donaldson, K., Stone, V., Seaton, A. and MacNee, W.: Ambient Particle Inhalation and the Cardiovascular System: Potential Mechanisms, Environ. Health Perspect., 109(4), 523–527, 2001.

Donaldson, K., Brown, D., Clouter, A., Duffin, R., Macnee, W., Renwick, L., Tran, L. and Stone, V.: The Pulmonary Toxicology of Ultrafine Particles, J. AEROSOL Med., 15(2), 213–220, 2002.

Dong, H. B., Zeng, L. M., Hu, M., Yu, Y. S., Zhang, Y. H., Slanina, J., Zheng, M., Wang, Z. F. and Jansen, R.: Technical Note: The application of an improved gas and aerosol collector for ambient air pollutants in China, Atmos. Chem. Phys., 12(21), 10519–10533, doi:10.5194/acp-12-10519-2012, 2012.

Eiguren-Fernandez, A., Kreisberg, N. and Hering, S.: An online monitor of the oxidative capacity of aerosols (o-MOCA), Atmos. Meas. Tech., 10(2), 633–644, doi:10.5194/amt-10-633-2017, 2017.

Fairfull-Smith, K. E. and Bottle, S. E.: The Synthesis and Physical Properties of Novel Polyaromatic Profluorescent Isoindoline Nitroxide Probes, European J. Org. Chem., 2008(32), 5391–5400, doi:10.1002/ejoc.200800597, 2008.

5 Fang, T., Verma, V., Guo, H., King, L. E., Edgerton, E. S. and Weber, R. J.: A semi-automated system for quantifying the oxidative potential of ambient particles in aqueous extracts using the dithiothreitol (DTT) assay: results from the Southeastern Center for Air Pollution and Epidemiology (SCAPE), Atmos. Meas. Tech. Discuss., 7(7), 7245–7279, doi:10.5194/amtd-7-7245-2014, 2014.

Fang, T., Verma, V., T Bates, J., Abrams, J., Klein, M., Strickland, J. M., Sarnat, E. S., Chang, H. H., Mulholland, A. J.,
10 Tolbert, E. P., Russell, G. A. and Weber, J. R.: Oxidative potential of ambient water-soluble PM2.5 in the southeastern United States: Contrasts in sources and health associations between ascorbic acid (AA) and dithiothreitol (DTT) assays, Atmos. Chem. Phys., 16(6), 3865–3879, doi:10.5194/acp-16-3865-2016, 2016.

Fuller, S. J., Wragg, F. P. H., Nutter, J. and Kalberer, M.: Comparison of on-line and off-line methods to quantify reactive oxygen species (ROS) in atmospheric aerosols, Atmos. Environ., 92, 97–103, doi:10.1016/j.atmosenv.2014.04.006, 2014.

15 Gao, D., Fang, T., Verma, V., Zeng, L. and Weber, R. J.: A method for measuring total aerosol oxidative potential (OP) with the dithiothreitol (DTT) assay and comparisons between an urban and roadside site of water-soluble and total OP, Atmos. Meas. Tech., 10(8), 2821–2835, doi:10.5194/amt-10-2821-2017, 2017.

Gussman, R. a., Kenny, L. C., Labickas, M. and Norton, P.: Design, Calibration, and Field Test of a Cyclone for PM 1 Ambient Air Sampling, Aerosol Sci. Technol., 36(3), 361–365, doi:10.1080/027868202753504461, 2002.

20 Harrison, R. M. and Yin, J.: Particulate matter in the atmosphere: Which particle properties are important for its effects on health?, Sci. Total Environ., 249(1–3), 85–101, doi:10.1016/S0048-9697(99)00513-6, 2000.

Hasson, A. S. and Paulson, S. E.: An investigation of the relationship between gas-phase and aerosol-borne hydroperoxides in urban air, J. Aerosol Sci., 34(4), 459–468, doi:10.1016/S0021-8502(03)00002-8, 2003.

Hedayat, F., Stevanovic, S., Miljevic, B., Bottle, S. and Ristovski, Z. D. D.: Review-evaluating the molecular assays for
25 measuring the oxidative potential of particulate matter, Chem. Ind. Chem. Eng. Q., 21(1–2), 201–210, doi:10.2298/CICEQ140228031H, 2015.

Hering, S. V, Spielman, S. R. and Lewis, G. S.: Moderated Water Based Condensational Particle Growth in a Laminar Flow, Aerosol Sci. Technol., 48(4), 401–408, doi:10.1080/02786826.2014.881460, 2014.

Huang, W., Zhang, Y., Zhang, Y., Zeng, L., Dong, H., Huo, P., Fang, D. and Schauer, J. J.: Development of an automated
30 sampling-analysis system for simultaneous measurement of reactive oxygen species (ROS) in gas and particle phases: GAC-ROS, Atmos. Environ., 134, 18–26, doi:10.1016/j.atmosenv.2016.03.038, 2016a.

Huang, W., Zhang, Y., Zhang, Y., Fang, D. and Schauer, J. J.: Optimization of the Measurement of Particle-Bound Reactive Oxygen Species with 2′,7′-dichlorofluorescin (DCFH), Water. Air. Soil Pollut., 227(5), 164, doi:10.1007/s11270-016-2860-9, 2016b.

Hung, H.-F. and Wang, C.-S.: Experimental determination of reactive oxygen species in Taipei aerosols, J. Aerosol Sci., 32(10), 1201–1211, doi:10.1016/S0021-8502(01)00051-9, 2001.

Kao, M. C. and S., W. C.: Reactive Oxygen Species in Incense Smoke, Aerososl Air Qual. Res., 2(1), 61–69, 2002.

Kenny, L. C., Gussman, R. and Meyer, M.: Development of a Sharp-Cut Cyclone for Ambient Aerosol Monitoring

5   Applications, Aerosol Sci. Technol., 32(4), 338–358, doi:10.1080/027868200303669, 2000.

Khlystov, A.: The steam-jet aerosol collector, Atmos. Environ., 29(17), 2229–2234, doi:10.1016/1352-2310(95)00180-7, 1995.

Kidwell, C. B. and Ondov, J. M.: Development and Evaluation of a Prototype System for Collecting Sub-Hourly Ambient Aerosol for Chemical Analysis, Aerosol Sci. Technol., 2010.

10   King, L. E. and Weber, R. J.: Development and testing of an online method to measure ambient fine particulate reactive oxygen species (ROS) based on the 2',7'-dichlorofluorescin (DCFH) assay, Atmos. Meas. Tech., 6(7), 1647–1658, doi:10.5194/amt-6-1647-2013, 2013.

Koehler, K., Shapiro, J., Sameenoi, Y., Henry, C. and Volckens, J.: Laboratory evaluation of a microfluidic electrochemical sensor for aerosol oxidative load, Aerosol Sci. Technol., 48(5), 489–497, doi:10.1080/02786826.2014.891722, 2014.

15   Li, N., Sioutas, C., Cho, A., Schmitz, D., Misra, C., Sempf, J., Wang, M., Oberley, T., Froines, J. and Nel, A.: Ultrafine Particulate Pollutants Induce Oxidative Stress and Mitochondrial Damage, Environ. Health Perspect., 111(4), 455–460, doi:10.1289/ehp.6000, 2002.

Li, Q., Wyatt, A. and Kamens, R. M.: Oxidant generation and toxicity enhancement of aged-diesel exhaust, Atmos. Environ., 43(5), 1037–1042, doi:10.1016/j.atmosenv.2008.11.018, 2009.

20   Ma, Q., Cai, S., Wang, S., Zhao, B., Martin, R. V., Brauer, M., Cohen, A., Jiang, J., Zhou, W., Hao, J., Frostad, J., Forouzanfar, M. H. and Burnett, R. T.: Impacts of coal burning on ambient PM2.5 pollution in China, Atmos. Chem. Phys., 17(7), 4477–4491, doi:10.5194/acp-17-4477-2017, 2017.

Miljevic, B., Fairfull-Smith, K. E., Bottle, S. E. and Ristovski, Z. D.: The application of profluorescent nitroxides to detect reactive oxygen species derived from combustion-generated particulate matter: Cigarette smoke - A case study, Atmos.

25   Environ., 44(18), 2224–2230, doi:10.1016/j.atmosenv.2010.02.043, 2010.

Miljevic, B., Hedayat, F., Stevanovic, S., Fairfull-Smith, K. E., Bottle, S. E. and Ristovski, Z. D.: To sonicate or not to sonicate PM filters: Reactive Oxygen Species generation upon ultrasonic irradiation, Aerosol Sci. Technol., Accepted f(12), 1276–1284, doi:10.1080/02786826.2014.981330, 2014.

Morgan, T. E., Davis, D. a., Iwata, N., Tanner, J. a., Snyder, D., Ning, Z., Kam, W., Hsu, Y.-T. T., Winkler, J. W., Chen, J.-

30   C. C., Petasis, N. a., Baudry, M., Sioutas, C. and Finch, C. E.: Glutamatergic neurons in rodent models respond to nanoscale particulate urban air pollutants in vivo and in vitro, Environ. Health Perspect., 119(7), 1003–1009, doi:10.1289/ehp.1002973, 2011.

Mudway, I. S., Duggan, S. T., Venkataraman, C., Habib, G., Kelly, F. J. and Grigg, J.: Combustion of dried animal dung as biofuel results in the generation of highly redox active fine particulates, Part. Fibre Toxicol., 2(1), 6, doi:10.1186/1743-8977-

2-6, 2005.

Nel, A.: Air Pollution-Related Illness: Effects of Particles, Science (80-.)., 308(5723), 804–806, doi:10.1126/science.1108752, 2005.

Orsini, D. a., Rhoads, K., McElhoney, K., Schick, E., Koehler, D. and Hogrefe, O.: A Water Cyclone to Preserve Insoluble

5   Aerosols in Liquid Flow - An Interface to Flow Cytometry to Detect Airborne Nucleic Acid, Aerosol Sci. Technol., 42(5), 343–356, doi:10.1080/02786820802072881, 2008.

Orsini, D. A., Ma, Y., Sullivan, A., Sierau, B., Baumann, K. and Weber, R. J.: Refinements to the particle-into-liquid sampler (PILS) for ground and airborne measurements of water soluble aerosol composition, Atmos. Environ., 37(9), 1243–1259, doi:10.1016/S1352-2310(02)01015-4, 2003.

10   Pal, A. K., Bello, D., Budhlall, B., Rogers, E. and Milton, D. K.: Screening for Oxidative Stress Elicited by Engineered Nanomaterials: Evaluation of Acellular DCFH Assay, Dose. Response., 10(3), 308–330, doi:10.2203/dose-response.10-036.Pal, 2012.

Peltier, R. E., Weber, R. J. and Sullivan, A. P.: Investigating a Liquid-Based Method for Online Organic Carbon Detection in Atmospheric Particles, Aerosol Sci. Technol., 41(12), 1117–1127, doi:10.1080/02786820701777465, 2007.

15   Penttinen, P., Timonen, K. L., Tiittanen, P., Mirme, A., Ruuskanen, J. and Pekkanen, J.: Ultrafine particles in urban air and respiratory health among adult asthmatics, Eur. Respir. J., 17(3), 428–435, doi:10.1183/09031936.01.17304280, 2001.

Popovicheva, O., Persiantseva, N. M., Shonija, N. K., Demott, P., Koehler, K., Petters, M., Kreidenweis, S., Tishkova, V., Demirdjian, B. and Suzanne, J.: Water interaction with hydrophobic and hydrophilic soot particles, Phys. Chem. Chem. Phys., 10(17), 2332–2344, doi:10.1039/b718944n, 2008.

20   Posner, L. N. and Pandis, S. N.: Sources of ultrafine particles in the Eastern United States, Atmos. Environ., 111, 103–112, doi:10.1016/j.atmosenv.2015.03.033, 2015.

Puthussery, J. V., Zhang, C. and Verma, V.: Development and field testing of an online instrument for measuring the real-time oxidative potential of ambient particulate matter based on dithiothreitol assay, Atmos. Meas. Tech., 11(10), 5767–5780, doi:10.5194/amt-11-5767-2018, 2018.

25   Ristovski, Z. D., Miljevic, B., Surawski, N. C., Morawska, L., Fong, K. M., Goh, F. and Yang, I. a.: Respiratory health effects of diesel particulate matter, Respirology, 17(2), 201–212, doi:10.1111/j.1440-1843.2011.02109.x, 2012.

Sa, A., Daher, N., Shafer, M. M., Schauer, J. J., Sioutas, C., Saffari, A., Daher, N., Shafer, M. M., Schauer, J. J. and Sioutas, C.: Global Perspective on the Oxidative Potential of Airborne Particulate Matter: A Synthesis of Research Findings, Environ. Sci. Technol., 48(13), 7576–7583, doi:10.1021/es500937x, 2014.

30   Sabaliauskas, K., Jeong, C. H., Yao, X., Jun, Y. S. and Evans, G.: Cluster analysis of roadside ultrafine particle size distributions, Atmos. Environ., 70, 64–74, doi:10.1016/j.atmosenv.2012.12.025, 2013.

Sameenoi, Y., Koehler, K., Shapiro, J., Boonsong, K., Sun, Y., Collett, J., Volckens, J. and Henry, C. S.: Microfluidic electrochemical sensor for on-line monitoring of aerosol oxidative activity., J. Am. Chem. Soc., 134(25), 10562–8, doi:10.1021/ja3031104, 2012.

Sameenoi, Y., Panymeesamer, P., Supalakorn, N., Koehler, K., Chailapakul, O., Henry, C. S. and Volckens, J.: Microfluidic paper-based analytical device for aerosol oxidative activity, Environ. Sci. Technol., 47(2), 932–940, doi:10.1021/es304662w, 2013.

Shi, T., Duffin, R., Borm, P. J. A., Li, H., Weishaupt, C. and Schins, R. P. F.: Hydroxyl-radical-dependent DNA damage by

5   ambient particulate matter from contrasting sampling locations, Environ. Res., 101(1), 18–24, doi:10.1016/j.envres.2005.09.005, 2006.

Shima, H., Koike, E., Shinohara, R. and Kobayashi, T.: Oxidative ability and toxicity of n-hexane insoluble fraction of diesel exhaust particles, Toxicol. Sci., 91(1), 218–226, doi:10.1093/toxsci/kfj119, 2006.

Silverman, D. T., Samanic, C. M., Lubin, J. H., Blair, A. E., Stewart, P. A., Vermeulen, R., Coble, J. B., Rothman, N., Schleiff,

10   P. L., Travis, W. D., Ziegler, R. G., Wacholder, S. and Attfield, M. D.: The Diesel Exhaust in Miners study: a nested case-control study of lung cancer and diesel exhaust, J. Natl. Cancer Inst., 104(11), 855–868, doi:10.1093/jnci/djs034, 2012.

Simon, P. K. and Dasgupta, P. K.: Wet Effluent Denuder Coupled Liquid/Ion Chromatography Systems: Annular and Parallel Plate Denuders, Anal. Chem., 65(9), 1134–1139, doi:10.1021/ac00057a006, 1993.

Simon, P. K. and Dasgupta, P. K.: Continuous Automated Measurement of the Soluble Fraction of Atmospheric Particulate

15   Matter, Anal. Chem., 67(1), 71–78, doi:10.1021/ac00097a012, 1995.

Simoneit, B. R. T., Kobayashi, M., Mochida, M., Kawamura, K., Lee, M., Lim, H. J., Turpin, B. J. and Komazaki, Y.: Composition and major sources of organic compounds of aerosol particulate matter sampled during the ACE-Asia campaign, J. Geophys. Res. D Atmos., 109(19), D19S10, doi:10.1029/2004JD004598, 2004.

Stevanovic, S., Ristovski, Z. D., Miljevic, B., Fairfull-Smith, K. E. and Bottle, S. E.: Application of profluorescent nitroxides

20   for measurements of oxidative capacity of combustion generated particles, Chem. Ind. Chem. Eng. Q., 18(4–2), 653–659, 2012a.

Stevanovic, S., Miljevic, B., Eaglesham, G. K., Bottle, S. E., Ristovski, Z. D. and Fairfull-Smith, K. E.: The use of a nitroxide probe in DMSO to capture free radicals in particulate pollution, European J. Org. Chem., (30), 5908–5912, doi:10.1002/ejoc.201200903, 2012b.

25   Stevanovic, S., Miljevic, B., Madl, P., Clifford, S. and Ristovski, Z.: Characterisation of a Commercially Available Thermodenuder and Diffusion Drier for Ultrafine Particles Losses, AEROSOL AIR Qual. Res., 15(1), 357, doi:10.4209/aaqr.2013.12.0355, 2015.

Stevanovic, S., Vaughan, A., Hedayat, F., Salimi, F., Rahman, M. M., Zare, A., Brown, R. A. J., Brown, R. A. J., Wang, H., Zhang, Z., Wang, X., Bottle, S. E., Yang, I. A. and Ristovski, Z. D.: Oxidative potential of gas phase combustion emissions -

30   An underestimated and potentially harmful component of air pollution from combustion processes, Atmos. Environ., 158, 227–235, doi:10.1016/j.atmosenv.2017.03.041, 2017.

Stoeger, T., Takenaka, S., Frankenberger, B., Ritter, B., Karg, E., Maier, K., Schulz, H. and Schmid, O.: Deducing in vivo toxicity of combustion-derived nanoparticles from a cell-Free oxidative potency assay and Metabolic activation of organic compounds, Environ. Health Perspect., 117(1), 54–60, doi:10.1289/ehp.11370, 2009.

Takeuchi, M., Ullah, S. M. R. and Dasgupta, P. K.: Continuous Collection of Soluble Atmospheric Particles with a Wetted Hydrophilic Filter, Anal. Chem. [H.W.Wilson - GS], 77(24), 8031, 2005.

Venkatachari, P. and Hopke, P. K.: Development and Laboratory Testing of an Automated Monitor for the Measurement of Atmospheric Particle-Bound Reactive Oxygen Species (ROS), Aerosol Sci. Technol., 42(8), 629–635, doi:10.1080/02786820802227345, 2008.

Venkatachari, P., Hopke, P. K., Brune, W. H., Ren, X., Lesher, R., Mao, J. and Mitchell, M.: Characterization of Wintertime Reactive Oxygen Species Concentrations in Flushing, New York, Aerosol Sci. Technol., 41(2), 97–111, doi:10.1080/02786820601116004, 2007.

Verma, V., Rico-Martinez, R., Kotra, N., King, L., Liu, J., Snell, T. W. and Weber, R. J.: Contribution of water-soluble and insoluble components and their hydrophobic/hydrophilic subfractions to the reactive oxygen species-generating potential of fine ambient aerosols, Environ. Sci. Technol., 46(20), 11384–11392, doi:10.1021/es302484r, 2012.

Wang, Y., Hopke, P. K., Sun, L., Chalupa, D. C. and Utell, M. J.: Laboratory and field testing of an automated atmospheric particle-bound reactive oxygen species sampling-analysis system, J. Toxicol., 2011, 419476–419479, doi:10.1155/2011/419476, 2011.

Weber, R. J., Orsini, D., Daun, Y., Lee, Y.-N. N., Klotz, P. J. and Brechtel, F.: A Particle-into-Liquid Collector for Rapid Measurement of Aerosol Bulk Chemical Composition, Aerosol Sci. Technol., 35(3), 718–727, doi:10.1080/02786820152546761, 2001.

Wragg, F. P. H., Fuller, S. J., Freshwater, R., Green, D. C., Kelly, F. J. and Kalberer, M.: An automated online instrument to quantify aerosol-bound reactive oxygen species (ROS) for ambient measurement and health-relevant aerosol studies, Atmos. Meas. Tech., 9(10), 4891–4900, doi:10.5194/amt-9-4891-2016, 2016.

Zhao, J. Y. and Hopke, P. K.: Concentration of Reactive Oxygen Species (ROS) in Mainstream and Sidestream Cigarette Smoke, Aerosol Sci. Technol., 46(2), 191–197, doi:10.1080/02786826.2011.617795, 2012.

Zhou, J., Bruns, E. A., Zotter, P., Stefenelli, G., Prévôt, A. S. H., Baltensperger, U., EI-Haddad, I. and Dommen, J.: Development, characterization and first deployment of an improved online reactive oxygen species analyzer, Atmos. Meas. Tech. Discuss., 1–27, doi:10.5194/amt-2017-161, 2017.