



Measurements of PM_{2.5} with PurpleAir under atmospheric conditions

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Abstract. The PurpleAir PA-II unit is a low-cost sensor for monitoring changes in the concentrations of Particulate Matter (PM) of various sizes. There are currently more than 9000 PA-II units worldwide; some of them are located in areas where no other reference air monitoring system is present. Previous studies have examined the performance of these PA-II units (or the sensor within them) in comparison to a co-located reference air monitoring system. However, because PA-II units are installed
10 by PurpleAir customers, the PA-II units are not co-located with a reference air monitoring system and, in many cases, are not near one. This study aimed to examine how PA-II units perform under atmospheric conditions when exposed to a variety of pollutants and PM_{2.5} concentrations. We were interested in knowing how accurate these PA-II units are when measuring PM_{2.5} concentrations with their sensitivity to concentration changes in comparison to the Environmental Protection Agency (EPA) Air Quality Monitoring Stations (AQMS) that are not co-located with them. For this study, we selected eight different locations,
15 where each location contains multiple PA-II units (minimum of seven per location, a total of 86 units) and at least one AQMS (total of 14). PM_{2.5} measurements from each PA-II unit were compared to those from the AQMS and other PA-II units in its area. The comparisons were made based on hourly and daily PM_{2.5} measurements. In most cases, the AQMS and PA-II units were found to be in good agreement; they measured similar values and followed similar trends, that is, when the PM_{2.5} values measured by the AQMS increased or decreased, so did those of the PA-II. In some high-pollution events, the PA-II measured
20 higher PM_{2.5} values compared to those measured by the AQMS. We found PA-II PM_{2.5} measurements to remain unaffected by changes in temperature or Relative Humidity (RH). Overall, the PA-II unit seems to be a promising tool for identifying relative changes in PM_{2.5} concentration with the potential to complement sparsely distributed monitoring stations and to aid in assessing and minimizing the public exposure to PM, particularly in areas lacking the presence of an AQMS.

25 1. Introduction

Atmospheric particulate matter (PM) with an aerodynamic diameter smaller than 2.5 μm (PM_{2.5}) is one of the leading contributors to the global burden of disease (GBD, Cohen et al., 2017; Forouzanfar et al., 2015; Lim et al., 2012). These particles are small enough to penetrate deep into the human lungs (Ling and van Eeden, 2009), where they have a negative impact on human health (Shiraiwa et al., 2017). Exposure to high PM_{2.5} concentrations was found to be correlated with the
30 daily number of hospitalizations and mortality cases (Schwartz et al., 1996; Klemm and Mason, 2000; Di et al., 2017). In the US, 3 %–5 % of annual deaths are attributed to PM_{2.5} (Cohen et al., 2017). Determining the pollution-level PM_{2.5} exposure can



be challenging as a limited number of in-situ instruments are available for monitoring ground-level $\text{PM}_{2.5}$ concentrations (Ford et al., 2019).

35 In the United States, the Environmental Protection Agency (EPA) monitors ambient $\text{PM}_{2.5}$ concentrations by using air quality monitoring stations (AQMSs). These stations use equipment that implements either a federal reference method or federal equivalent method (FRM and FEM, respectively; Clements et al., 2017). The FRM is a gravimetric measurement in which particles are collected on a filter and the difference in filter weight before and after exposure is used to determine the 24-h PM concentration (Watson et al., 2017). The FEM measures PM using optical, beta ray attenuation and trapped element oscillation

40 to provide hourly PM concentrations. A single FEM $\text{PM}_{2.5}$ sensor in each AQMS costs thousands of dollars. Further, the operation of these AQMSs requires trained personnel and significant infrastructure; they are subject to strict maintenance and calibration routines to ensure high-quality data and comparability between different locations (Castell et al., 2017). AQMSs generally have sparse geographic coverage and are located at fixed sites, mainly in large population centers; they are not present in smaller cities and underdeveloped regions. The high temporal and spatial resolution of $\text{PM}_{2.5}$ concentrations may

45 vary significantly within a region, therefore, $\text{PM}_{2.5}$ concentration values provided by a single AQMS site may not accurately represent the $\text{PM}_{2.5}$ concentrations present near people who are concerned about their possible health effects (Wang et al., 2015). These limitations create a growing need for air quality sensor networks that will produce both temporal and spatial high-resolution pollution maps that can be used to identify peak events across large areas (Morawska et al., 2018).

50 Recent advancements in technology and a rise in public awareness have led to an increase in the popularity of low-cost air-quality sensors that are relatively cheap and easy-to-use (Commodore et al., 2017; Woodall et al., 2017). Such sensors enable communities and individuals alike to obtain granular information on the spatial and temporal distribution of PM concentrations in their area (Gupta et al., 2018; Morawska et al., 2018), thereby enabling them to monitor local air quality conditions (Williams et al., 2018). Many types of low-cost air-quality sensors are available, and they vary in performance (Williams et al., 2018);

55 however, despite the proposed benefits of these sensors, their accuracy and precision remain unknown (Kuula et al., 2017). Data quality remains a major concern that hinders the widespread adoption of low-cost sensor technology. To assure data quality, it is important to test these sensors and compare them to FRM/FEM measurements under both laboratory and field conditions, particularly under atmospheric conditions with various air pollution levels in which the sensors are expected to operate (Kelly et al., 2017; Morawska et al., 2018). Testing these sensors at multiple locations will allow for exposure to

60 different atmospheric conditions and pollutant types (AQ-SPEC, 2018).

Among the limitations of low-cost sensors are environmental factors that affect the sensor's abilities. Some low-cost sensors have exhibited sensitivity to temperature and relative humidity (RH) (Clements et al., 2017). When working in the laboratory, these environmental conditions can be controlled; however, it is impossible to achieve such stability in the field under

65 atmospheric conditions. Therefore, additional measurements under a variety of ambient conditions are needed (Kelly et al.,



2017). In addition, some sensors have exhibited a drift in sensitivity over time (reduction of efficiency). The rate of drift over time is a crucial parameter in sensor characterization as it determines the interval of calibration as well as the overall useable lifetime of the sensor (Clements et al., 2017; Hagan et al., 2018).

70 The PA-II unit is a low-cost sensor sold by PurpleAir company. It is meant for outdoor usage and is the subject of our study. Each PA-II unit contains two Plantower particulate matter sensors (PMS5003 sensors) that provide real-time measurements of $PM_{1.0}$, $PM_{2.5}$, and PM_{10} . The usage of PA-II has grown rapidly in the last two years with the result that more than 9000 such sensors are in use across five continents, with the majority being operated in the US and Europe. PurpleAir provides live information on their website in the form of a color-coded air quality index (AQI) together with actual PM concentrations (PurpleAir, 2019). Several studies have already evaluated the PA-II unit or the sensors (PMS5003) it contains; however, in all such studies, the PA-II unit (or the PMS5003 sensor) was co-located with a reference unit. The AQ Sensor Performance Evaluation Center (AQ-SPEC) evaluated the performance of a PA-II unit using FEM sensors as reference under laboratory and field conditions in the Los Angeles area. Their evaluation showed a very good comparison between the two for both $PM_{2.5}$ and PM_{10} (AQ-SPEC, 2018). An additional comparison between three different PA-II sensors and a single FEM was performed for eight weeks between December 2016 and January 2017 at the South Coast Air Quality Management District Rubidoux Air Monitoring Station. Good correlation ($R^2 > 0.9$) was found between the three PA-II units and the FEM unit. However, although the PA-II unit follows diurnal and day-to-day fluctuations very well, it consistently overestimated the $PM_{2.5}$ concentrations measured by the FEM (Gupta et al., 2018). Sayahi et al. (2019) conducted a long-term comparison (320 days) between two PMS5003 sensors and both FRM and FEM units that were all co-located at Salt Lake City, Utah. One of their PMS5003 sensors overestimated the $PM_{2.5}$ concentration whereas the other measured similar values to those measured by the FEM. According to Gupta et al. (2018), the performance of PA-II compared against FEM units in a high-pollution environment ($PM_{2.5} > 100 \mu g m^{-3}$) is unknown and requires further evaluation. In addition, the sensitivity of the PA-II sensors to changes in RH, temperature, and other environmental parameters remains a topic of further investigation (Gupta et al. 2018). Answers to these questions are crucial if we are to assess the possibility of using measurement data from multiple PA-II units to properly represent the air quality of an area, thus allowing the residents to protect themselves when high pollution events occur.

95 This study aimed to examine how PA-II units perform under atmospheric conditions when exposed to a variety of pollutants and $PM_{2.5}$ concentrations. Comparison of PA-II units to $PM_{2.5}$ measurements taken by an AQMS that was not co-located with them are presented. Further, a comparison of PA-II units to other nearby PA-II units and their efficiency as a network of low-cost sensors are discussed.

2. Method

2.1. PurpleAir PA-II Unit Structure and Data



The PurpleAir PA-II unit has size of 85×125 mm. It contains two PMS5003 sensors (see two blue rectangles in Fig. 1A), a
100 BME280 environmental sensor, and an ESP8266 microcontroller. The BME280 sensor is used to monitor the units' inner
pressure, temperature, and humidity; the sensor measurements are not to be used for monitoring ambient conditions (PurpleAir,
personal communication, 2019). The ESP8266 microcontroller is used to communicate with both the two PMS5003 sensors
and with the PurpleAir server over Wi-Fi, thereby allowing the PM concentration to be presented live on the PurpleAir map
(<https://www.purpleair.com/map>). The PMS5003 sensors provide real-time measurements of $PM_{1.0}$, $PM_{2.5}$, and PM_{10}
105 concentrations; the sensors are based on the light scattering principle, and a photodiode detector converts the scattered light to
a voltage pulse. A fan draws the particles into the sensor and past the laser path (Fig. 1B) at a flow rate of 0.1 L/min. The
particle count is calculated by counting the pulses from the scattering signal and converting the number of pulses to a mass
concentration for six diameters between 0.3 and 10 μm using an algorithm for outdoor PM (CF_ATM - average particle
density). Each PMS5003 sensor has an effective measurement range for $PM_{2.5}$ concentration of 0–500 $\mu\text{g m}^{-3}$ with a resolution
110 of 1 $\mu\text{g m}^{-3}$, and the maximum standard $PM_{2.5}$ concentration is above 1000 $\mu\text{g m}^{-3}$. According to the manufacturer, each
PMS5003 sensor will work effectively in a temperature range of -10 °C to 60 °C and RH range of 0 %–99 % (Yong, 2016).

The microcontroller in the PA-II unit reads the $PM_{1.0}$, $PM_{2.5}$, and PM_{10} concentrations from the PMS5003 sensors every second;
it averages the concentration values across 20 s and displays the results using UTC time (PurpleAir, personal communication,
115 2019). The use of a dual PMS5003 sensor setup serves as an internal check for the PA-II unit's integrity. The
similarity/difference in the PM concentrations obtained from the two PMS5003 sensors (named as A and B) allows users to
evaluate the efficiency and validity of their PA-II unit. The two PMS5003 sensors, A and B, should agree with each other all
the time; failure to report the same value indicates that something is wrong with one of the sensors. PurpleAir does not calibrate
their devices; instead, before each PA-II unit is sent out to a customer, the company performs a comparison test with a dozen
120 other PA-II units to find and remove outliers from the shipment (PurpleAir, personal communication, 2019).

All the data regarding the PA-II units and their measurements was downloaded from the PurpleAir website. Information about
all the PA-II units was downloaded in a JSON formatted file. Each PA-II unit has a name (given by the owner), a unique ID
number (designated by the company for each sensor), the unit location (latitude and longitude), and a date on which the unit
125 was installed. We initially selected all the PA-II units that were active between January 1, 2017, and December 31, 2018 (UTC
time). For each selected PA-II unit, we downloaded an Excel file containing the measurement data in 20-s intervals for both
PMS5003 sensors (A and B). Because our focus was on $PM_{2.5}$ measurements, we calculated the $PM_{2.5}$ hourly average and
standard deviation (SD) based on the original measurement values and the daily average and standard deviation based on
hourly averages that we had calculated previously. Our final dataset included only days that had a minimum of 20 h of
130 measurements per day (80 % of the day). Only times which had a good agreement ($R^2 > 0.9$) of hourly $PM_{2.5}$ measurements
between the two PMS5003 sensors (A and B) were used.



2.2. PM_{2.5} Measurements from AQMS

Hourly measurements of PM_{2.5} (FRM/FEM Mass code - 88101 file) from all AQMSs collected by the EPA from January 1, 2017, to December 31, 2018, were selected from the EPA website (<https://aqs.epa.gov/api>). The location of each AQMS was provided in the same file. Each AQMS is identified by the combination of state code, county code, site number, and Parameter Occurrence Code (POC) number. The POC is used to represent cases in which more than one unit performs PM_{2.5} measurements at the same site. All timestamps were converted to UTC to match the PA-II measurement timestamps. The PM_{2.5} daily average and standard deviation were calculated based on the hourly PM_{2.5} measurements; only days with a minimum of 20 h of measurements per day (80 % of the day) were considered.

2.3. Identification of Locations for Analysis - Areas with Multiple PA-II units and at least one AQMS

By using the JSON file for the PA-II and the 88101 file for the AQMS, the distances between all units was calculated to identify locations with multiple PA-II units (a minimum of five units) and at least one AQMS. All the units in these locations needed to be active during the designated time period of January 1, 2017, to December 31, 2018. Eight different locations containing a total of 14 different AQMSs and 86 different PA-II units were identified: Pittsburgh, PA; Denver, CO; Berkeley-Oakland, CA; San Francisco, CA; Vallejo, CA; Ogden-South Ogden, UT; Lindon-Orem, UT; and Salt Lake City, UT. Fig. S1 shows a map with all the PA-II units and AQMSs at each location. Table 1 provides information on each of the eight locations with the names of the units, their location, first and last time of measurement, and the minimum and maximum PM_{2.5} hourly values.

In Pittsburgh, two AQMSs (42-3-8-3 and 42-3-1376-1) and eleven PA-II units (ID - 3723, 3981, 9016, 9026, 9038, 9096, 9878, 9880, 9892, 9896, and 9906) were used. In Denver, three AQMS (8-31-26-3, 8-31-27-3, and 8-31-28-3) and eight PA-II units (ID - 2249, 2267, 2269, 2719, 2900, 3924, 4022, and 7956) were used. In Berkeley-Oakland, three AQMSs (6-1-11-3, 6-1-12-3, and 6-1-13-3) and ten PA-II units (ID - 2574, 3082, 3854, 4335, 4506, 4795, 4825, 5414, 6410, and 10114) were used. San Francisco, Vallejo, Ogden-South Ogden, and Lindon-Orem all had a single AQMS (6-75-5-3, 6-95-4-4, 49-57-2-5, and 49-49-4001-5, respectively) but multiple PA-II units. San Francisco had nine PA-II units (ID - 1226, 2031, 2910, 3348, 3996, 4372, 4770, 5776, and 6344); Vallejo had 15 units (the maximum; ID - 1142, 1870, 1874, 1878, 1882, 2480, 2906, 3686, 3758, 3769, 3782, 3784, 3960, 4928, and 5127); Ogden-South Ogden had seven PA-II units (the minimum; ID - 465, 1104, 5178, 5454, 6604, 7858, and 7860); and Lindon-Orem had 12 PA-II units (ID - 5135, 5143, 5145, 5728, 5732, 5736, 5750, 5754, 5760, 6304, 6948, and 6986). Salt Lake City had two AQMSs at the same location (49-35-3006-4 and 49-35-3006-5, different POCs) and 14 PA-II units (ID - 884, 3388, 5014, 5460, 5742, 5802, 5990, 6078, 6356, 6360, 6434, 6608, 6622, and 10050).

2.4. Comparison between PA-II and AQMS



To evaluate the similarities and differences between the AQMS and the PA-II units, a set of calculations and comparisons was performed. First, graphs showing the distribution of $PM_{2.5}$ values were plotted. Second, a regression between the AQMS and each PA-II unit was made based on hourly and daily $PM_{2.5}$ measurements. From the regression, R-squared (R^2) and root mean square error (RMSE) values as well as the best fit information, including the slope and intercept, were obtained. We performed
170 different comparisons for both the entire study period and for specific events that we wanted to examine in greater detail.

2.5. Meteorological Information

Meteorological measurements including temperature, RH, and wind speed/direction were used from the EPA website (<https://www.epa.gov/outdoor-air-quality-data>). Only some AQMSs had these meteorological measurements: 42-3-1376-1 and
175 42-3-8-3 from Pittsburgh, 8-31-26-3 and 8-31-28-3 from Denver, 49-57-2-5 from Ogden-South Ogden, 49-49-4001-5 from Lindon-Orem, and 49-35-3006-4 from Salt Lake City.

Additional meteorological measurements such as temperature, RH, wind speed and gust, wind direction, and visibility of different meteorological stations were obtained from the Iowa Environmental Mesonet website
180 (<https://mesonet.agron.iastate.edu/request/download.phtml>). For meteorological information about the selected locations, the following meteorological stations were used: AGC-Pittsburgh/ Allegheny station in Pittsburgh, the Denver International Airport (DEN) station in Denver, the Ogden-Hinckley Muni (OGD) station in Utah, the Provo Muni (PVU) station in Ogden-South Ogden, the Salt Lake City International airport (SLC) station in Lindon-Orem, the California Oakland (OAK) station in Berkeley-Oakland and San Francisco, and the Napa County (APC) station in Vallejo.

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2.6. AQI Calculations

The AQI is used for the reporting air quality levels. It allows the public to know how clean the air is and indicates the health effects a person may experience within a few hours or days of breathing unhealthy air. The AQI has six categories, each of which corresponds to a different level of health concern (EPA, 2014): Good (0–50, green), Moderate (51–100, yellow),
190 Unhealthy for Sensitive Groups (101–150, orange), Unhealthy (151–200, red), Very Unhealthy (201–300, purple), and Hazardous (301–500, maroon) (see Table S1). In our study, we calculated the AQI for $PM_{2.5}$ daily average as follows:

$$AQI = \frac{(\text{measured } PM_{2.5} - PM_{min})(AQI_{max} - AQI_{min})}{(PM_{max} - PM_{min})} + AQI_{min} \quad (1)$$

where the measured $PM_{2.5}$ is the daily average $PM_{2.5}$ value, PM_{max} and PM_{min} are respectively the maximum and minimum concentration of the AQI color category for the measured $PM_{2.5}$, AQI_{max} is the maximum AQI value for a color category that
195 corresponds to the measured $PM_{2.5}$, and AQI_{min} is the minimum AQI value for a color category that corresponds to the measured $PM_{2.5}$. Table S1 lists the different values and categories of PM_{max} , PM_{min} , AQI_{max} , and AQI_{min} .

3. Result and Discussion



3.1. Hourly and Daily PM_{2.5} Comparisons of AQMS and PA-II units.

200 This study examined measurements for a two-year period from January 1, 2017, to December 31, 2018, resulting in ample overlapping measurement times between the PA-II units and the different AQMSs. The number of concurrent hourly measurements in each comparison varies per location. Overall, the number of concurrent hourly measurements ranged from 1017 to 13975 h with an average of 6652 ± 2822 h per comparison. Other than the Lindon-Orem area where the local AQMS was active only from November 2017, measurements from January 2017 were available in all the other areas. Most of the PA-
205 II units became active only at the end of 2017. The distance between the different AQMSs and PA-II units ranged from 0.01 km to 13 km with an average of 4.2 ± 2.4 km. Table 2 lists the exact distance and number of PM_{2.5} hourly measurements used in comparisons of each AQMS and PA-II unit. Based on the overlap times, we identified and examined the distribution of daily PM_{2.5} values measured by the PA-II units and AQMS for each location and also performed additional comparisons between the units in these locations.

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3.1.1 Distribution of Daily PM_{2.5} Values

Fig. 2 shows the distribution of daily PM_{2.5} values for each unit at each of the eight locations. Overall, the daily PM_{2.5} values obtained from both the AQMS and the PA-II units seem to follow similar trends. When the AQMS values increase/decrease, the PA-II values also increase/decrease. The PA-II unit measurements of daily PM_{2.5} values start at $0 \mu\text{g m}^{-3}$, and the AQMS
215 can measure negative values owing to its calibration process. In some cases (locations and times), the AQMS measured higher PM_{2.5} daily values compared to the PA-II units, as seen during April–July 2018 in Berkeley-Oakland (Fig. 2C), Lindon-Orem (Fig. 2G), and Salt Lake City (Fig. 2H). However, regardless of the PM_{2.5} concentration, PA-II units usually measured higher values compared to those measured by the AQMS (see July and August 2018 in Pittsburgh, Fig. 2A). This overestimating of PM values by the PA-II units (or PMS sensors) compared to FRM and FEM units has also been observed previously (Kelly et
220 al., 2017; AQ-SPEC, 2018; Gupta et al., 2018; Sayahi et al., 2019) when the two were co-located.

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3.1.2 Linear Regression Tests

To evaluate the overall trends of the PA-II units compared to the AQMS, we performed a series of regression tests for each site. As in previous works (Gupta et al., 2018; Sayahi et al., 2019) and as commonly used (Clements et al., 2017), these
225 comparisons were performed using linear regression. Each AQMS was compared to all the PA-II units in its area based on hourly PM_{2.5} measurements. Table 2 lists R², RMSE values, and the slope and intercept of the linear fit. In general, the linear regression results were mixed. The total R² values for the hourly PM_{2.5} measurements ranged from 0.1 to 0.91 with an average of 0.63 ± 0.17 , which is relatively high. The RMSE values ranged from 3.89 to $13.13 \mu\text{g m}^{-3}$ with an average of $7.73 \pm 2.05 \mu\text{g m}^{-3}$. The slope ranged from 0.03 to 3.12, but was mostly around 1, with an average of 1.15 ± 0.35 .

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In some locations such as Denver (Table 2B) and Vallejo (Table 2F), high correlation values were found between the local AQMS and the PA-II units. Denver had three AQMSs; each comparison had a high R² value in the range of 0.53 to 0.91



(average of 0.72 ± 0.1 for all three AQMSs), average RMSE of $5.65 \pm 0.89 \mu\text{g m}^{-3}$, and average slope of 1.4 ± 0.18 . Vallejo had one AQMS with fifteen PA-II units; the R^2 values ranged from 0.55 to 0.91 with an average of 0.79 ± 0.13 . The RMSE values in Vallejo were higher than those in Denver, with an average of $8.95 \pm 1.28 \mu\text{g m}^{-3}$ but with lower average slope of 1.27 ± 0.11 . These high correlation values and relatively low RMSE indicate that although the PA-II units and the AQMS are not co-located, they still tend to behave in a similar way. At the other locations, except for Ogden-South Ogden, more than 75 % of the comparisons had high correlation values (>0.5) and only a few with low R^2 value. Several PA-II units had low R^2 values when compared to an AQMS, as in the case of unit 5414 in Berkeley-Oakland and unit 6344 in San Francisco. These two units also had low correlation values compared to the other PA-II units in their region (data not shown). We noticed that unit 6344 was exposed to very high $\text{PM}_{2.5}$ concentrations (up to $250 \mu\text{g m}^{-3}$ for a duration of 3 h) on May 13, 2018. We suspect that this exposure might have affected the instrument efficiency, as was suggested by Sayahi et al. (2019), and therefore, its measurements differ substantially from those of the AQMS. Another exception was Ogden-South Ogden, as all of the comparisons had very low R^2 values (ranging from 0.11 to 0.36 with an average of 0.28 ± 0.1) and high RMSE values (ranging from 8.27 to $10.6 \mu\text{g m}^{-3}$). However, when the PA-II units were compared to each other (and not to the AQMS), they showed high correlation values ranging from 0.83 to 0.98 with an average of 0.92 ± 0.05 (Fig. S2). These low correlation values and high RMSE values for the PA-II and AQMS comparisons were most likely caused by specific events and the location of each of the units, as explained below.

A comparison based only on hourly $\text{PM}_{2.5}$ values lower than $40 \mu\text{g m}^{-3}$, as performed by Sayahi et al. (2019), did not improve the hourly correlation values, as shown in Table S2. Around 88 % of the comparisons had lower correlation values compared to the case when all $\text{PM}_{2.5}$ concentrations were used; the R^2 values ranged from 0.04 to 0.9 with an average of 0.57 ± 0.16 . Some locations such as Pittsburgh (Table S2A) showed no change in their correlation values for $\text{PM}_{2.5} < 40 \mu\text{g m}^{-3}$ comparisons whereas others such as Ogden-South Ogden (Table S2F) and Lindon-Orem (Table S2G) showed improved correlation values. Unlike the correlation values, the RMSE values in the comparison of $\text{PM}_{2.5} < 40 \mu\text{g m}^{-3}$ improved in 93 % of the cases, resulting in lower RMSE values compared to those found when all $\text{PM}_{2.5}$ values were used. The RMSE values ranged from 2.89 to $12.96 \mu\text{g m}^{-3}$ with an average of $6.83 \pm 1.54 \mu\text{g m}^{-3}$.

Comparisons based on the $\text{PM}_{2.5}$ daily values improved the results (Table S3). The numbers of concurrent $\text{PM}_{2.5}$ daily measurements ranged from 18 to 574 days, with an average of 270 ± 119 days per comparison. The correlation values ranged from 0.17 to 0.97 with an average of 0.78 ± 0.15 . Further, the RMSE values had a wide range of 2.1– $12.8 \mu\text{g m}^{-3}$ with an average of $4.98 \pm 1.77 \mu\text{g m}^{-3}$. Overall, 95 % of the comparisons had a higher R^2 and 98 % of the comparisons had lower RMSE values compared to the hourly comparison. Even Ogden-South Ogden, which did not show an improvement in previous comparisons, exhibited better results (Table S3F). The average correlation values in Ogden-South Ogden improved from 0.28 ± 0.1 in the hourly comparison to 0.53 ± 0.12 in the daily comparison. The RMSE values also improved; they decreased from an average of $9.51 \pm 0.83 \mu\text{g m}^{-3}$ in the hourly comparisons to $6.95 \pm 0.46 \mu\text{g m}^{-3}$ in the daily comparisons.



3.2. Comparison of High Pollution Events

270 Different meteorological conditions such as wind direction or speed as well as pollution type (traffic, industrial, wildfire, fireworks, etc.) or source (local vs. regional) may affect the comparison between the AQMS and the PA-II units. We aimed to determine how the PA-II units behave in a high-pollution event when the daily $PM_{2.5}$ concentration exceeds the EPA daily regulation of $35 \mu\text{g m}^{-3}$. Therefore, we decided to investigate specific events with high $PM_{2.5}$ concentrations in different time frames under different atmospheric conditions.

275 3.2.1. Fireworks in Ogden- South Ogden

In Ogden-South Ogden, major differences were observed in the $PM_{2.5}$ values measured during July 2018 (Fig. 3) by the PA-II units and the single AQMS. During this month, we noticed that the AQMS measured very high hourly $PM_{2.5}$ values (with peaks over $400 \mu\text{g m}^{-3}$), whereas none of the PA-II units exceeded $20 \mu\text{g m}^{-3}$. The regression test results for this month also showed low R^2 values with an average of 0.03 ± 0.01 . The location of the units (Fig. S1F), pollution type during this event, and meteorological conditions at the time revealed the cause of these differences. The increase in $PM_{2.5}$ was due to 4th of July fireworks (correlated to July 5, UTC time) that caused an increase in AQMS hourly $PM_{2.5}$ values $> 100 \mu\text{g m}^{-3}$ for a duration of 5 h. The AQMS was located downwind from the main fireworks event (Friendship Park, south of the AQMS) whereas all the PA-II units were far from any fireworks in a residential area on the slopes of Mt. Ogden. Local regulations did not allow the use of fireworks in a residential area (east of road 203; Ogden City Fire Department, 2019) where most of the PA-II units are located. Wind direction information obtained from the local metrological station (see Methods) revealed that the wind was blowing from the fireworks location toward the AQMS but was not reaching the PA-II units. Therefore, the PA-II units could not detect this increase. A similar result was seen in the previous year in July 2017 when only one PA-II unit was active (see Fig. 2F). We also noticed that on July 9, one of the PA-II units (ID 6604) measured high $PM_{2.5}$ values (up to $135 \mu\text{g m}^{-3}$) whereas all the other units measured much lower $PM_{2.5}$ values. This high concentration was measured during only one hour (23:00 UTC time); therefore, we suspected that this increase was caused by a local source near this specific unit, such as a small-scale fire, lawn mower, or barbeque.

In both cases, the presence of the PA-II sensors significantly benefited the areas' residents by allowing them to make informed decisions. In the case of the fireworks, if the residents were to base their actions solely on the AQMS data, they would assume that the air quality is unhealthy when actually it is not. If the wind direction was to change and blow from the fireworks toward the residential area, the AQMS data would not prepare the residents at all. In the second case, the localized pollution was



identified by the PA-II unit; the AQMS did not measure any changes owing to its location. Overall, the probability of any event being identified by a single AQMS is significantly lower than that of it being identified using multiple PA-II sensors.

300 The remaining days included both low-pollution days (July 1–5 and after July 9) and elevated-pollution days (July 7–8). During these days, the PA-II sensors and the AQMS exhibited similar trends, identified the same changes in $PM_{2.5}$ concentrations, and measured similar values. A repeat of the regression tests for only these days (without the fireworks and local event data) resulted in a significant improvement in correlation values; specifically, the average R^2 value increased to 0.69 ± 0.03 .

305 3.2.2. Inversion in Utah

In Utah, all three locations- Ogden-South Ogden, Lindon-Orem, and Salt Lake City-followed similar daily $PM_{2.5}$ trends during December 4-13, 2018 (Fig. 4). The entire area was affected by an inversion for several days (December 3–13) that increased the daily $PM_{2.5}$ values up to $67.2 \pm 4.17 \mu\text{g m}^{-3}$ and reduced the visibility to almost zero (see photos in Williams, 2019). Overall, at each of these three locations, the values measured by the PA-II units increased at the same time and followed a similar trend
310 to the AQMS measurements. However, whereas all the PA-II units measured similar $PM_{2.5}$ values, the AQMS measured lower $PM_{2.5}$ concentrations. $PM_{2.5}$ values only decreased after precipitation occurred on December 13. The linear regression for each area shows good correlation. In Ogden-South Ogden, Salt Lake City, and Lindon-Orem, the average R^2 was 0.93 ± 0.01 , 0.98 ± 0.01 for both AQMSs, and 0.96 ± 0.01 , respectively. Overall, at each of these three locations, the PA-II units measured similar values, but these seemed to be overestimated when compared to the AQMS measurements.

315

3.2.3. Wildfire in California

The three locations in California- Vallejo, Berkeley-Oakland, and San Francisco are relatively close to each other and were affected by a large wildfire that occurred in November 2018. According to the California Statewide Wildfire Recovery Resources (2019), the wildfire started on November 8 at Butte County (north of Vallejo) owing to a combination of strong
320 winds and very dry conditions. A southwesterly wind transferred the wildfire smoke from Butte County toward Vallejo, Berkeley-Oakland, and San Francisco. Very high daily $PM_{2.5}$ values ($>200 \mu\text{g m}^{-3}$) were measured from November 9 to 21 (Fig. 5). During this period, the area had stable meteorological conditions, with low wind speed, that reduced visibility down to 1.6 km (1 mile). The high daily $PM_{2.5}$ values decreased only after precipitation started on November 21. Overall, at each of the three locations, the values measured by the PA-II units increased at the same time and followed a similar trend to the
325 AQMS measurements. Regression test results of each area also show very similar results to each other. In Vallejo, the average R^2 was 0.97 ± 0.01 , and in Berkeley-Oakland, where there are three AQMSs, two of them had an average R^2 of 0.95 ± 0.04 and the third had average R^2 of 0.94 ± 0.03 . In both Vallejo (nine PA-II units) and Berkeley-Oakland (six PA-II units), the average daily $PM_{2.5}$ values of the PA-II units were higher than those measured by the AQMS (Fig. 5A-B). There was no active AQMS at San Francisco during these days, and therefore, only the PA-II units are shown in Fig. 5C. Out of the eight PA-II



330 units located in Berkeley-Oakland (Fig. 5B), two PA-II units (5414 and 10114) measured lower daily $PM_{2.5}$ values compared to the other PA-II units and even compared to the local AQMS.

Using AQI maps is another good way to see the spatial and temporal changes in $PM_{2.5}$ measurements; it is also important as the public's behavior is based on the interpretation of the AQI values. We calculated the AQI values for both the PA-II units and the AQMS of all three areas; these calculations were based on the daily $PM_{2.5}$ values (see Methods). We drew maps of all three areas for each day (Fig 6) that show the locations of the AQMS and PA-II units; the locations on the maps are color-coded based on the AQI value at that location on that day. Examining these maps shows us how, as the wildfire and smoke progressed, the air quality worsened. On November 6, before the wildfire started, the AQI for the entire area was moderate. As the fire progressed, the air quality changed from unhealthy on November 11 to very unhealthy on November 16; the air quality became good again only on November 22. Overall, the AQMS and PA-II units in these areas reported similar values and followed similar trends; AQI values differed between the AQMS and PA-II units on a few days are a result of the differences in the $PM_{2.5}$ values used in the calculation. Having multiple PA-II units in each area allows us to track air quality changes with higher resolution, as multiple sensors provide more data than a single AQMS. In the case of the San Francisco area where no AQMS was active, the PA-II units are the only source of data for providing the residents with crucial information about the air quality in their region.

3.3. Factors That May Impact PA-II Performance

Meteorological conditions such as wind direction and speed, pollutant type, and pollution source are some of the factors that might affect the performance of the PA-II units. It is therefore important to also evaluate and consider additional factors such as other meteorological conditions and underlying technology used when comparing the behavior and measurements of the PA-II units and the AQMS.

3.3.1. Temperature and RH

The sensitivity of the PA-II unit to changes in temperature and RH remains unknown (Gupta et al., 2018). We can assume that changes in temperature or RH may affect the performance of the PA-II unit especially under atmospheric conditions as they cannot be controlled. Jayaratne et al. (2018) tested an older version of the PMS unit (PMS1003) and reported such an effect. Most low-cost sensors have no heater or dryer at their inlet to remove water from the sample before measuring the particles; therefore, deliquescent or hygroscopic growth of particles, mainly under high RH conditions (>75 %), can lead to higher reported PM concentrations (Jayaratne et al., 2018). According to Rai et al. (2017), most low-cost sensors show some sensitivity to RH conditions but not to temperature. It is therefore important to evaluate whether the PA-II unit will be affected by changes in temperature or RH. To do so, we used temperature and RH measurements from the nearest available meteorological stations (see Methods for station information) and, in some cases, additional measurements from the AQMS (e.g., in Pittsburgh, Denver, Ogden-South Ogden, Lindon-Orem, and Salt Lake City).



The hourly temperature measurements from the meteorological stations were compared with the hourly $PM_{2.5}$ measurements from each PA-II unit (86 units in total) using linear regression. The regression resulted in very low R^2 values that ranged from 1×10^{-9} to 0.07 with an average of 0.02 ± 0.02 . Similar results were found when the AQMS temperature measurements were used (52 units in total, Table S4); the R^2 values ranged from 6×10^{-5} to 0.13 with an average of 0.04 ± 0.03 . For the RH, two different comparisons were made: a comparison using all RH values and a comparison for only those cases in which the RH value was higher than 75 %. When using RH data from the meteorological stations and for the entire RH range, very low R^2 values were found. The correlations values ranged from 7.5×10^{-7} to 0.1 with an average of 0.02 ± 0.03 . Comparison results obtained using RH measurements from the AQMS were similar (Table S4); the R^2 values ranged from 1.01×10^{-5} to 0.17 with an average of 0.05 ± 0.04 . Even when only $RH > 75 \%$ was tested, the R^2 values ranged from 1.6×10^{-7} to 0.1 with an average of 0.01 ± 0.01 for RH measurements from the meteorological station. Similar values were also found for RH measured by the AQMS; R^2 values ranged from 5.5×10^{-6} to 0.18 with an average of 0.02 ± 0.04 . Similar results have been reported previously as well. For example, Sayahi et al. (2019) found very low correlation values between measurements from the PMS5003 sensor and the temperature/RH under atmospheric conditions. Holstius et al. (2014) found a negligible effect of temperature or RH on measurements performed using low-cost sensors under ambient conditions. However, several studies that used old PMS units, such as PMS1003 that was used in PA-I or PMS3003 that was never used in any PA units, found that these sensors were affected by RH (Kelly et al., 2017; Jayaratne et al., 2018; Zheng et al., 2018). AQ-SPEC (2018) tested the PA-II unit in a laboratory setting under different temperature and RH conditions and found that most temperature and RH combinations had a minimal effect on the PA-II's precision. Our findings for PA-II units in the field under atmospheric conditions are in agreement with those of the AQ-SPEC (2018).

3.3.2. Technology, Maintenance, and Placement

There are many differences between PA-II and AQMS units that can influence the comparison results, including the underlying technology and the manner in which units are placed. The $PM_{2.5}$ sensors in the AQMS perform gravimetric measurements using the mass of the particle; by contrast, the PA-II unit uses a laser particle counter to count electric pulses generated as particles cross through a laser beam. Another difference is the physical location of the units; whereas AQMSs are meticulously positioned in an open area, the location of a PA-II sensor is determined by its owner. Although PurpleAir recommends positioning the PA-II sensor in an open area, ultimately, it is the owner's decision. In practice, most of the PA-II units are located in residential areas with low-rise housing. Further, the height at which the sensor is located could affect the measurements. Whereas the height of the AQMS inlet is regulated and kept constant at each location, the owner of a PA-II unit can freely place it near the ground or higher up. The location of the PA-II units in residential areas can provide both an advantage and a disadvantage. For example, as in the case of Ogden-South Ogden, a single unit might be exposed to more localized PM sources such as a barbeque, lawn mower, or car, making it report different results compared with other units in its area. Maintenance and calibration are other possible causes of differences between the two. The $PM_{2.5}$ sensors in the AQMS have strict rules for the monthly evaluation of sensor performance, including through flow calibration or calibration based on



400 minimum value threshold (which, in some cases, causes the recording of negative PM values). By contrast, PA-II units do not have any quality control other than that done by the company for each sensor before shipment to the customer (PurpleAir personal communication, 2019).

3.3.3. Distance and Number of Comparisons Between the Units

405 Other factors that could affect the comparisons with the AQMS are the distances between the units or the number of observations. Previous studies obtained good results when comparing between the PA-II unit or PMS5003 sensor and the FRM and FEM units when the two units were co-located. The AQ-SPEC (2018) recently released a report comparing PA-II units to two FEM instruments under laboratory and field conditions. They found good correlations for hourly and daily values of both PM_{2.5} and PM₁₀ under field conditions with higher correlation values for PM_{2.5} compared to those for PM₁₀. Gupta et al. (2018) compared three PA-II units in California to a single FEM unit and obtained good correlation values ($R^2 > 0.9$). Sayahi et al. (2019) co-located reference air monitors (tapered element oscillating microbalance, TEOM), and FRM unit, next to a PMS5003 (used in the PA-II unit) in Salt Lake City. The PMS5003 PM_{2.5} measurements correlated well with the hourly TEOM measurements ($R^2 > 0.87$) and with the daily FRM measurements ($R^2 > 0.88$). In our study, we did not position the PA-II units. Further, in most cases, the AQMS and the PA-II units were not located at the same place; therefore, they might have been exposed to different particle types and concentrations. Some might claim that not having the PA-II and FRM units co-located, as was done in previous studies, might diminish the accuracy of the comparison between these units. Although lower correlation values were in fact observed in our study, as we were using PA-II units in their natural locations, this was expected. Further, as we saw that the correlation values are not much lower than those in the co-located cases described in previous studies, they are still statistically significant. Because the AQMS and the PA-II units were not co-located, we wanted to verify whether the distance between the AQMS and the PA-II units affected the R^2 values. We compared the R^2 values that we previously calculated for the hourly PM_{2.5} measurements with the corresponding distances between the PA-II units and AQMS (Fig. S3A). There was no correlation between the two, and similar results were found when the RMSE values were tested (Fig. S3B). The number of observations used for the comparison was also tested; comparing the same R^2 from the measurements with the number of observations revealed no effect of the number of observations on R^2 or RMSE values (Fig. S3C-D).

3.4. Next Steps with PA-II units

425 Ford et al. (2019) suggested the use of PA-II units as a network installed by residents in an in North Colorado. This seems like a good solution for locations that are lacking FRM or FEM units as multiple sensors can provide more data. However, it is important to consider the limitations of the PA-II unit. The PA-II unit needs to be monitored for changes in unit behavior. We recommend PurpleAir to monitor the measurements of the PA-II units, identify units that behave differently from other surrounding units or units whose internal sensors (A and B) report different values, flag them on the online map, and communicate instructions to the unit owners on how to clean the unit. The manufacturer of the PMS5003 sensor that is used in the PA-II units noted that it has a lifetime of ~3 years (Yong, 2016). None of the current units have been active for that long;



therefore, the efficiency of PA-II units over such a long period remains unknown and should be evaluated. It is possible that, after this duration, they will lose their efficiency (a behavior known as drift) and will become outliers.

435 4. Conclusions

PA-II units are becoming a common low-cost tool to monitor changes in the concentrations of PMs of various sizes. Previous studies have examined the performance of these PA-II units (or the sensor in them) by comparing them with a co-located EPA AQMS. However, PA-II units are not co-located in practice, and some of them are placed in areas where there is no reference air monitor system. This study aimed to examine the behavior of PA-II units under atmospheric conditions when exposed to a variety of pollutants and different PM_{2.5} concentrations. For this purpose, we used PA-II units that have already been active for some time irrespective of where they might be. Eight locations with multiple PA-II units and at least a single AQMS were identified. Each PA-II unit was compared to the AQMS and to other PA-II units in its surrounding area based on hourly or daily PM_{2.5} measurements. Overall, the PA-II units behaved in a similar way to the other PA-II units at their locations. We found that even though some PA-II units overestimated or underestimated at times, the AQMS and PA-II units were mostly in agreement and measured similar PM_{2.5} concentrations. PA-II was also found to not be affected by temperature or RH. We think that the PA-II unit is a promising tool for measuring PM_{2.5} concentrations and identifying relative concentration changes. Further, through the use of AQI, the current air quality can be successfully conveyed to the public. The PA-II unit has the potential to complement sparsely distributed monitoring stations, particularly in areas lacking a nearby AQMS.

450 **Data availability.** All data can be provided by the authors upon request.

Competing interests. The authors declare that they have no conflict of interest.

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Table legends

Table 1. Information on each of the eight locations with the names of the AQMS and PA-II units, their location (latitude and longitude), first and last time of measurement, minimum, and maximum $PM_{2.5}$ hourly values. AQMS ID represented by the numbers of State-County-Site-POC for each unit.

Table 2. Comparison between each AQMS and the different PA-II units per location (A-G) for average hourly $PM_{2.5}$ measurements. Distance and number of observations (hours) are provided for each comparison along with linear regression result such as R^2 , RMSE values, and the slope and intercept of the linear fit. Bold R^2 values represent values larger than 0.5.

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Figure legends

Figure 1. (A) Picture from the bottom of the PA-II unit containing two PMS5003 sensors (in blue). (B) Schematic of a single PMS5003 sensor. A fan draws the particles through the inflow (rounded holes) at the lower level of the sensor. The particles travel to the upper part of the sensor where they come out through the air flow holes and then pass through the laser path, causing the beam to scatter. Finally, the particles exit from the fan.

Figure 2. Distribution of daily $PM_{2.5}$ measurements from the AQMS and PA-II units in each of the eight areas: (A) Pittsburgh; (B) Denver; (C) Berkeley-Oakland; (D) San Francisco; (E) Vallejo; (F) Ogden-South Ogden; (G) Lindon-Orem, and (H) Salt Lake City. Measurements from AQMS are represented by the green lines and the PA-II units are indicated by purple lines. The numbers are the units' ID numbers.

Figure 3. Hourly $PM_{2.5}$ measurements at Ogden-South Ogden in UT during July 1-11, 2018 (UTC time). Measurements from the AMQS unit are represented in green and those from the PA-II units, in different shades of purple. Each number represent the ID of the unit. Error bars represent the standard deviation values for each hour on each of the PA-II units. Note that local PA-II unit 465 was not active during this time.

Figure 4. Hourly measurements of $PM_{2.5}$ at (A) Ogden-South Ogden, (B) Lindon-Orem, and (C) Salt Lake City during December 1-14 2018 (UTC time). An increase in average daily $PM_{2.5}$ values was observed from December 4-13. The AMQS unit is represented by the different green lines and the PA-II units, by the different purple lines. Each number represents the ID of the unit. Bars represent the standard deviation values per day. Several PA-II units were not operating during these times.

Figure 5. Hourly measurements of $PM_{2.5}$ at (A) Vallejo, (B) Berkeley-Oakland (B), and (C) San Francisco during the November 2018 wildfire (UTC time). An increase in average daily $PM_{2.5}$ values was observed during November 9–20. The AMQS unit is represented by the different green lines and the PA-II units, by the different purple lines. Each number represent the ID of the unit. Bars represent the standard deviation values per day.

Figure 6. Spatial and temporal changes of AQI in California at Berkeley-Oakland, San Francisco, and Vallejo during November 8-22, 2018. Squares represent AQMS and circles, PA-II units. The colors of units represent the different AQI values.

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Table 1. Information on each of the eight locations with the names of the AQMS and PA-II units, their location (latitude and longitude), first and last time of measurement, minimum, and maximum PM_{2.5} hourly values. AQMS ID represented by the numbers of State-County-Site-POC for each unit.

Location	Unit Type	ID of Each Unit (PA-II - sensor A)	Latitude	Longitude	PA-II Unit label	First day of observation	Last day of observation	Minimum PM _{2.5} hourly average (µg/m ³)	Maximum PM _{2.5} hourly average (µg/m ³)	Number of observations (hours)
1. Pittsburgh, PA	AQMS	42-3-8-3 *	40.465	-79.961		1-Jan-17	31-Dec-18	-2	109	17302
		42-3-1376-1 &	40.437	-79.864		1-Jan-17	31-Dec-18	-2	67	16690
	PurpleAir sensor ID	3723	40.448	-79.916	Point Breeze	14-Oct-17	28-Sep-18	0.1	86.66	3438
		3981	40.438	-79.956	CMU CAPS PPA 010	20-Nov-17	8-Oct-18	0.19	80.84	2143
		9016	40.421	-79.914	Parkview Blvd-Summerset at Frick Park	27-May-18	31-Dec-18	0.36	79.69	2885
		9026	40.478	-79.93	Jancey St Morningside	22-Apr-18	31-Dec-18	0.13	47.59	3412
		9038	40.445	-79.915	Pillars In Squirrel Hill North	7-May-18	31-Dec-18	0.11	193.46	4250
		9906	40.436	-79.908	Frick Environmental Center - Squirrel Hill	9-May-18	31-Dec-18	0.01	55.4	3499
		9878	40.45	-79.911	juniata ct	6-May-18	31-Dec-18	0.09	49.39	4424
		9880	40.473	-79.914	HP Winterton St	6-May-18	26-Dec-18	0.09	192.89	3957
		9892	40.43	-79.918	Nicholson St	2-May-18	31-Dec-18	0	116.85	5729
		9896	40.441	-79.896	EastEndAve1	2-May-18	31-Dec-18	0.12	137.42	3378
		9906	40.43	-79.954	South Oakland	9-May-18	31-Dec-18	0.03	174.57	5613
2. Denver, CO	AQMS	8-31-26-3 \$	39.779	-105.005		1-Jan-17	31-Dec-18	0	76.5	16850
		8-31-27-3 #	39.732	-105.015		1-Jan-17	31-Dec-18	0.2	73.1	17259
		8-31-28-3 #	39.786	-104.989		1-Jan-17	24-Dec-18	0.3	75.1	16651
	PurpleAir sensor ID	2249	39.783	-104.96	The GrowHaus	5-Dec-17	31-Dec-18	0.08	121.99	9331
		2267	39.779	-105.006	La Casa	22-Aug-17	27-Feb-18	0.08	132.44	2156
		2269	39.781	-104.956	Swansea (DEH)	4-Aug-17	18-Jun-18	0.04	170.64	7128
		2719	39.755	-104.966	26th and Williams	12-Aug-17	1-Nov-18	0.1	155.27	6411
		2900	39.753	-105.041	West Denver PA-II	18-Aug-17	31-Dec-18	0.04	152.73	11980
		3924	39.779	-105.005	APCD La Casa	16-Nov-17	31-Dec-18	0.05	81.58	9004
		4022	39.708	-104.981	Wash Park West	8-Nov-17	31-Dec-18	0.04	80.46	9968



3. Berkeley -Oakland, CA	AQMS	7956	39.786	-104.989	Globeville	27-Feb-18	31-Dec-18	0.11	78.48	7326
		6-1-11-3 *	37.815	-122.282		1-Jan-17	31-Dec-18	-10	210	17210
		6-1-12-3 *	37.794	-122.263		1-Jan-17	31-Dec-18	-3	218	17283
		6-1-13-3 *	37.865	-122.303		1-Jan-17	31-Dec-18	-7	393	16882
	PurpleAir sensor ID	2574	37.901	-122.286	Berkeley Park and Coventry, Kensington, CA, USA	19-Sep-17	31-Dec-18	0.05	281.12	10476
		3082	37.906	-122.302	El Cerrito - Rust - Ohlone Greenway	6-Sep-17	16-Nov-18	0.04	291.35	10176
		3854	37.862	-122.247	Claremont Blvd	17-Oct-17	7-Oct-18	0.03	87.47	8007
		4335	37.81	-122.298	West Oakland, Oakland, CA	30-Nov-17	31-Dec-18	0.09	239.01	9471
		4506	37.875	-122.271	North Berkeley	3-Dec-17	31-Dec-18	0.05	307.69	9434
		4795	37.797	-122.216	Lodestar	6-Dec-17	19-Jun-18	0.12	58.92	4125
		4825	37.7637	-122.233	Northwood	22-Dec-17	31-Dec-18	0.08	272.35	8733
5414	37.8295	-122.248	Piedmont Ave	17-Nov-18	31-Dec-18	0.02	211.77	1048		
6410	37.858	-122.284	San Pablo Park / The Derby	15-Mar-18	31-Dec-18	0.03	297	6911		
10114	37.8	-122.249	CCEEB - Park & E. 19th	30-May-18	31-Dec-18	0.11	137.76	3927		
4. San Francisco, CA	AQMS	6-75-5-3 *	37.766	-122.399		1-Jan-17	31-Dec-18	-10	190	16309
	PurpleAir sensor ID	1226	37.768	-122.402	Volta Charging	17-Oct-17	31-Dec-18	0.08	263.78	10417
		2031	37.733	-122.424	St Mary's Park	15-Sep-17	31-Dec-18	0.07	265.08	11295
		2910	37.778	-122.408	tactrix rooftop	18-Sep-17	31-Dec-18	0.12	282.73	10861
		3348	37.787	-122.445	Lower Pacific Heights	13-Nov-17	23-Dec-18	0.1	180.53	3937
		3996	37.789	-122.391	South Beach	11-Nov-17	1-Oct-18	0.1	79.63	7783
		4372	37.754	-122.412	The Mission- Clean air is hip	5-Jan-18	31-Dec-18	0.09	250.54	7951
		4770	37.787	-122.417	930 Post	21-Dec-17	31-Dec-18	0.22	250.18	8883
		5776	37.745	-122.421	La Lengua Air Station Alpha	5-Jan-18	23-Dec-18	0	275.6	8033
		6344	37.759	-122.403	Kansas Gulch	28-Jan-18	17-Jun-18	0.11	252.71	3384
5. Vallejo, CA	AQMS	6-95-4-4 *	38.1	-122.24		1-Jan-17	31-Dec-18	-10	435	16630
	PurpleAir sensor ID	1142	38.104	-122.258	Carolina Street	14-Apr-17	9-Oct-18	0.06	457.06	9893
		1870	38.111	-122.243	Amador St @ Stutz Alley	17-Jul-17	31-Dec-18	0.04	468.49	12646
		1874	38.067	-122.22	Glen Cove Ridge	15-Jul-17	31-Dec-18	0.05	292.45	12460
		1878	38.086	-122.245	Winchester Hill	20-Jul-17	3-May-18	0.08	384.76	6432
		1882	38.078	-122.23	Navone St.	19-Jul-17	26-Dec-18	0.05	339.3	11406
		2480	38.122	-122.233	Howard Ave	17-Aug-17	31-Dec-18	0.06	477.83	11908
		2906	38.074	-122.24	Sandy Beach	10-Dec-17	31-Dec-18	0.04	303.5	9245
		3686	38.074	-122.231	Carquinez One	6-Dec-17	23-Aug-18	0.08	256.68	7143
		3758	38.114	-122.259	Buckles St	11-Nov-17	22-Aug-18	0.12	92.45	6774



		3769	38.081	-122.215	Old Glen Cove	14-Oct-17	31-Dec-18	0.08	287.42	10426
		3782	38.12	-122.241	El Camino Real/Valle Vista	29-Nov-17	1-Oct-18	0.05	85.32	7401
		3784	38.098	-122.26	Little Old Lady By The River	20-Oct-17	31-Dec-18	0.05	278.12	10148
		3960	38.141	-122.26	211 Sonora pass rd	18-Jan-18	24-Oct-18	0.03	227.05	6508
		4928	38.09	-122.239	1300 Block Lemon	1-Dec-17	31-Dec-18	0.03	296.62	8253
		5127	38.108	-122.256	Vallejo	2-Dec-17	31-Dec-18	0.08	243.68	9406
6. Ogden - South Ogden, UT	AQMS	49-57-2-5 **	41.21	-111.98		4-Jan-17	31-Dec-18	-10	790.3	13574
	PurpleAir sensor ID	465	41.185	-111.935	Beus Park	1-Jan-17	30-Nov-17	0.06	83.72	8002
		1104	41.179	-111.946	University Village - Weber State University	31-Jan-18	31-Dec-18	0.07	96.44	7712
		5178	41.216	-111.931	Taylor Canyon	9-Dec-17	31-Dec-18	0.57	110.39	7797
		5454	41.192	-111.942	WSU Marriott Health	4-Apr-18	31-Dec-18	0.06	64.49	5124
		6604	41.185	-111.938	Bobwhite Ct	1-Feb-18	31-Dec-18	0	135.89	5687
		7858	41.195	-111.947	WSU Public Safety Building	5-Apr-18	31-Dec-18	0	104.43	6135
		7860	41.193	-111.943	WSU Stewart Library	4-Apr-18	31-Dec-18	0	95.09	6248
7. Lindon - Orem, UT	AQMS	49-49-4001-5 **	40.341	-111.714		8-Nov-17	31-Dec-18	0.1	204	9984
	PurpleAir sensor ID	5135	40.324	-111.715	Orem Bonneville Park powered by UTOPIA Fiber	10-Jan-18	31-Dec-18	0.08	165.33	8420
		5143	40.315	-111.667	Orem Foothill Park powered by UTOPIA Fiber	19-Dec-17	19-Oct-18	0.01	53.67	3949
		5145	40.308	-111.705	Orem 600N 400W powered by UTOPIA Fiber	18-Jan-18	17-Jun-18	0.12	46.11	3471
		5728	40.314	-111.697	Orem Fire Department #2 powered by UTOPIA Fiber	19-Jan-18	31-Dec-18	0	166.86	7273
		5732	40.308	-111.73	Orem Public Works powered by UTOPIA Fiber	18-Jan-18	31-Dec-18	0	192.96	7440
		5736	40.299	-111.705	Orem 400W 75N powered by UTOPIA Fiber	18-Jan-18	31-Dec-18	0	183.61	7315
		5750	40.302	-111.712	Orem Geneva Park powered by UTOPIA Fiber	21-Jan-18	19-Oct-18	0	187.95	6253
		5754	40.317	-111.677	Orem Orchard Elementary powered by UTOPIA Fiber	19-Jan-18	19-Oct-18	0	102.15	6089
		5760	40.31	-111.713	Orem Junior High powered by UTOPIA Fiber	23-Oct-18	31-Dec-18	0.09	72.01	1628
		6304	40.308	-111.689	Orem Sharon Park powered by UTOPIA Fiber	1-Feb-18	31-Dec-18	0	130.03	7882
		6948	40.338	-111.694	Lindon City - Murdock Canal Trail	21-Mar-18	24-Aug-18	0.17	100.75	2987
6986	40.34	-111.718	Lindon City Center	20-Mar-18	31-Dec-18	0	156.33	4954		
8. Sal	AQMS	49-35-3006-4 **	40.74	-111.87		1-Jan-17	31-Dec-18	0	87.5	16529



	49-35-3006-5 **	40.74	-111.87		1-Jan-17	31-Dec-18	-10	89.1	17030
PurpleAir sensor ID	884	40.777	-111.895	Quince and Apricot	15-Feb-17	31-Dec-18	0	114.33	14426
	3388	40.733	-111.822	Montessori Community School	20-Oct-17	31-Dec-18	0.03	79.22	10248
	5014	40.771	-111.9	KSL Triad	28-Nov-17	31-Dec-18	0.08	123.65	9520
	5460	40.728	-111.861	1027 Hollywood	14-Jan-18	31-Dec-18	0.08	125.71	8097
	5742	40.734	-111.846	Wasatch Hollow	7-Jan-18	31-Dec-18	0	156.33	6815
	5802	40.71	-111.832	Yuma View	7-Jan-18	14-May-18	0.03	47.05	2610
	5990	40.72	-111.82	Lynwood	5-Jul-18	31-Dec-18	0.04	187.18	4013
	6078	40.764	-111.86	Victory Park	29-Jan-18	25-Jul-18	0.03	39.9	1231
	6356	40.774	-111.883	Cobble Knoll	29-Jan-18	31-Dec-18	0	116.68	8050
	6360	40.767	-111.867	Capitol Hill Construction	26-Jan-18	31-Dec-18	0.03	92.98	8105
	6434	40.696	-111.877	3450 South 500 East	26-Jan-18	31-Dec-18	0.08	157.06	7875
	6608	40.774	-111.851	4th AveCat	5-Mar-18	31-Dec-18	0	243.54	7179
	6622	40.744	-111.876	Tracy Aviary	25-Feb-18	31-Dec-18	0	128.51	6468
	10050	40.749	-111.912	Utah Paperbox	2-May-18	31-Dec-18	0.15	150.72	5771

AQMS Sensor Type - * Met One BAM-1020 Mass Monitor; ** Thermo Scientific Model 5030; & Thermo Scientific 5014i; \$ Teledyne T640; # GRIMM EDM Model 180



Table 2. Comparison between each AQMS and the different PA-II units per location (A-G) for average hourly PM_{2.5} measurements. Distance and number of observations (hours) are provided for each comparison along with linear regression result such as R², RMSE values, and the slope and intercept of the linear fit. Bold R² values represent values larger than 0.5.

A. Pittsburgh			PurpleAir sensor ID										
			3723	3981	9016	9026	9038	9096	9878	9880	9892	9896	9906
EPA AQMS ID	42-3-1376-1	Distance (km)	4.58	7.79	4.66	7.24	4.44	3.73	4.24	5.79	4.65	2.79	7.72
		Obs (h)	3394	2116	2861	3380	4207	3470	4379	3913	5672	3352	5558
		R ²	0.51	0.43	0.54	0.53	0.57	0.61	0.59	0.51	0.57	0.54	0.49
		RMSE	8.04	8.72	7.35	6.42	7.22	6.49	6.35	7.50	6.90	7.17	7.63
		Slop	0.99	0.86	1.10	0.99	1.16	1.16	1.09	1.06	1.16	1.12	1.06
		Intercept	4.29	6.18	3.61	4.23	2.37	2.01	2.72	3.05	2.33	3.93	3.43
	42-3-8-3	Distance (km)	4.26	3.09	6.32	2.96	4.48	5.54	4.55	4.07	5.34	6.1	4
		Obs (h)	3207	2035	2737	3186	4026	3301	4132	3677	5418	3128	5300
		R ²	0.52	0.51	0.46	0.58	0.56	0.49	0.57	0.50	0.52	0.46	0.53
		RMSE	8.04	8.12	8.08	6.21	7.39	7.48	6.60	7.63	7.37	7.91	7.37
		Slop	1.20	1.18	1.11	1.09	1.26	1.13	1.16	1.16	1.20	1.10	1.21
		Intercept	2.91	3.25	0.34	0.14	-1.76	-0.54	-1.03	-0.99	-0.69	1.20	-0.84

B. Denver			PurpleAir sensor ID							
			2249	2267	2269	2719	2900	3924	4022	7956
EPA AQMS ID	8-31-26-3	Distance (km)	3.89	0.08	4.25	4.34	4.23	0.01	8.19	1.57
		Obs (h)	9130	2144	7060	6336	11763	8807	9765	7151
		R ²	0.76	0.91	0.81	0.81	0.80	0.81	0.73	0.75
		RMSE	4.80	4.26	5.01	5.11	4.79	4.51	5.15	4.36
		Slop	1.41	1.70	1.52	1.51	1.50	1.54	1.37	1.40
		Intercept	-1.25	-2.25	-2.21	-2.41	-1.61	-1.77	-1.45	-0.74
	8-31-27-3	Distance (km)	7.34	5.26	7.49	4.93	3.19	5.33	3.95	6.39
		Obs (h)	8708	2145	6859	6319	11338	8407	9337	6907
		R ²	0.67	0.83	0.74	0.75	0.73	0.70	0.70	0.68
		RMSE	5.64	6.04	5.91	5.91	5.60	5.75	5.45	4.85
		Slop	1.37	1.64	1.51	1.49	1.47	1.47	1.38	1.33
		Intercept	-1.80	-2.73	-2.92	-3.27	-2.24	-2.24	-2.43	-1.01



8-31-28-3	Distance (km)	2.49	1.66	2.87	3.99	5.78	1.59	8.68	0.03
	Obs (h)	8750	2145	6970	5956	11380	8444	9382	6866
	R ²	0.61	0.78	0.65	0.62	0.59	0.59	0.53	0.66
	RMSE	6.15	6.83	6.80	7.29	6.97	6.67	6.75	5.00
	Slop	1.11	1.72	1.40	1.35	1.19	1.15	1.04	1.07
	Intercept	-0.86	-4.34	-2.72	-2.94	-1.07	-0.79	-0.76	-0.03

C. Berkeley -Oakland		PurpleAir sensor ID										
		2574	3082	3854	4335	4506	4795	4825	5414	6410	10114	
EPA AQMS ID	6-1-11-3	Distance (km)	9.56	10.33	6.13	1.45	6.79	6.13	7.14	2.67	4.83	3.36
		Obs (h)	10448	10147	7988	9459	9422	4117	8725	1046	6905	3924
		R ²	0.76	0.69	0.36	0.86	0.79	0.43	0.85	0.38	0.82	0.65
		RMSE	12.05	11.55	8.08	8.55	11.82	7.72	10.40	13.13	12.16	10.93
		Slop	1.21	1.21	0.68	1.19	1.25	0.71	1.36	0.40	1.30	0.60
		Intercept	-4.17	-3.22	0.68	-3.10	-4.23	-0.99	-4.69	2.92	-2.67	6.08
	6-1-12-3	Distance (km)	12.08	12.99	7.76	3.51	9.09	4.16	4.26	3.3	7.41	1.4
		Obs (h)	10323	10026	7943	9324	9287	4091	8592	1042	6790	3898
		R ²	0.84	0.78	0.57	0.87	0.87	0.59	0.90	0.39	0.88	0.70
		RMSE	9.95	9.75	6.60	8.17	9.11	6.56	8.60	13.11	9.86	10.26
		Slop	1.28	1.30	0.96	1.22	1.34	0.95	1.42	0.41	1.36	0.62
		Intercept	-5.62	-5.08	-3.24	-3.60	-5.63	-3.63	-5.75	3.01	-3.90	5.26
	6-1-13-3	Distance (km)	4.26	4.64	4.93	6.12	3.04	10.68	12.8	6.41	1.78	8.64
		Obs (h)	10181	9912	7825	9167	9114	4036	8444	1017	6675	3733
		R ²	0.79	0.71	0.35	0.81	0.83	0.53	0.82	0.41	0.85	0.63
		RMSE	11.45	11.40	8.18	9.97	10.57	6.96	11.52	12.70	11.38	11.49
		Slop	1.22	1.20	0.67	1.15	1.28	0.92	1.32	0.43	1.30	0.57
		Intercept	-0.79	0.68	2.83	0.60	-1.17	-0.61	-0.82	1.79	-0.02	7.21

D. San Francisco		PurpleAir sensor ID									
		1226	2031	2910	3348	3996	4372	4770	5776	6344	
EPA AQMS ID	6-75-23	Distance (km)	0.35	4.23	1.6	4.65	2.65	1.75	2.86	3.03	0.85
		Obs (h)	9290	10157	9725	3546	7558	6954	7867	7024	3223



		R ²	0.63	0.65	0.65	0.64	0.58	0.53	0.55	0.42	0.19
		RMSE	6.71	7.58	7.82	7.30	7.34	7.16	7.15	6.97	7.41
		Slop	1.03	1.07	1.11	1.01	0.98	0.97	0.96	0.76	0.53
		Intercept	0.72	0.74	1.45	0.55	2.18	1.27	3.43	1.61	2.55

E. Vallejo		PurpleAir sensor ID															
		1142	1870	1874	1878	1882	2480	2906	3686	3758	3769	3782	3784	3960	4928	5127	
EPA AQMS ID	6-95-4-4	Distance (km)	1.78	1.07	4.27	1.96	2.8	2.21	3.22	3.23	2.24	3.16	2	1.96	4.73	1.35	1.64
		Obs (h)	9525	11824	11647	6257	10654	11085	8440	6791	6432	9612	7044	9340	6224	7459	8600
		R ²	0.76	0.91	0.83	0.76	0.86	0.89	0.88	0.56	0.70	0.86	0.57	0.89	0.55	0.91	0.89
		RMSE	10.78	7.96	10.60	11.14	9.78	8.40	9.65	9.29	6.79	9.43	7.51	8.33	7.74	8.73	8.11
		Slop	1.47	1.32	1.22	1.27	1.24	1.25	1.39	1.29	1.26	1.31	0.96	1.27	1.16	1.33	1.24
		Intercept	-5.25	-1.97	-1.77	-2.47	-2.26	-2.77	-2.69	-2.26	-1.40	-2.13	0.19	-2.60	-1.41	-1.32	-2.09

F. Ogden- South Ogden		PurpleAir sensor ID							
		465	1104	5178	5454	6604	7858	7860	
EPA AQMS ID	49-57-2-5	Distance (km)	4.15	3.95	3.92	3.26	3.95	2.72	3.15
		Obs (h)	5127	7679	6944	5105	5662	6106	6219
		R ²	0.11	0.36	0.34	0.16	0.30	0.36	0.36
		RMSE	9.08	9.15	9.27	8.27	10.51	10.60	9.68
		Slop	0.21	0.68	0.64	0.36	0.62	0.73	0.66
		Intercept	4.65	2.27	3.80	3.54	2.71	2.86	2.68

G. Lindon - Orem		PurpleAir sensor ID												
		5135	5143	5145	5728	5732	5736	5750	5754	5760	6304	6948	6986	
EPA AQMS ID	49-49-4001-5	Distance (km)	1.91	4.93	3.75	3.36	3.92	4.81	4.43	4.12	3.48	4.27	1.74	0.4
		Obs (h)	8388	3911	3465	7242	7408	7283	6224	6060	1626	7850	2963	4925
		R ²	0.22	0.50	0.20	0.49	0.50	0.43	0.55	0.51	0.66	0.48	0.58	0.52
		RMSE	0.41	4.04	4.43	8.97	9.16	8.56	7.86	6.72	8.60	7.86	4.61	8.97
		Slop	0.03	0.75	0.71	1.19	1.27	1.11	1.15	0.95	3.12	1.05	0.59	1.25



		Intercept	0.09	-0.41	0.77	0.54	0.92	0.75	0.13	-0.01	-3.54	0.31	0.59	1.09
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H. Salt Lake City		PurpleAir sensor ID														
		884	3388	5014	5460	5742	5802	5990	6078	6356	6360	6434	6608	6622	10050	
EPA AQMS ID	49-35-3006-4	Distance (km)	4.95	4.29	4.47	1.34	2.21	4.49	4.78	3.20	4.33	3.42	4.48	4.58	0.87	3.67
		Obs (hours)	13524	9283	8570	7450	6074	2126	3926	1200	7766	7766	7541	6944	6241	5614
		R ²	0.72	0.72	0.78	0.81	0.72	0.37	0.70	0.40	0.77	0.77	0.73	0.63	0.80	0.77
		RMSE	6.14	6.45	6.85	5.00	6.37	3.89	7.81	4.10	5.70	5.39	7.32	7.17	5.43	6.94
		Slop	1.36	1.24	1.51	1.41	1.31	0.79	1.34	0.78	1.40	1.33	1.58	1.22	1.43	1.58
		Intercept	-2.45	-2.21	-2.18	-3.06	-2.15	-0.41	-1.87	-0.38	-3.13	-2.75	-1.73	-2.62	-2.94	-1.74
	49-35-3006-5	Distance (km)	4.95	4.29	4.47	1.34	2.21	4.49	4.78	3.20	4.33	3.42	4.48	4.58	0.87	3.67
		Obs (h)	13975	10158	9431	8022	6748	2570	3981	1224	7982	8037	7808	7142	6421	5736
		R ²	0.68	0.67	0.72	0.70	0.64	0.37	0.65	0.20	0.67	0.68	0.66	0.55	0.72	0.71
		RMSE	6.83	7.11	7.98	6.09	7.01	5.18	8.47	4.72	6.77	6.31	8.00	7.76	6.20	7.74
		Slop	1.31	1.18	1.46	1.39	1.30	1.06	1.35	0.50	1.37	1.31	1.58	1.17	1.42	1.58
		Intercept	-1.10	-0.86	-0.45	-1.69	-1.01	-0.67	-0.50	1.39	-1.77	-1.51	-0.45	-1.29	-1.80	-0.20









