

# **A Novel Approach for Simple Statistical Analysis of High-Resolution Mass Spectra**

Yanjun Zhang<sup>1</sup>, Otso Peräkylä<sup>1</sup>, Chao Yan<sup>1</sup>, Liine Heikkinen<sup>1</sup>, Mikko Äijälä<sup>1</sup>, Kaspar R.  
Daellenbach<sup>1</sup>, Qiaozhi Zha<sup>1</sup>, Matthieu Riva<sup>1,2</sup>, Olga Garmash<sup>1</sup>, Heikki Junninen<sup>1,3</sup>, Pentti Paatero<sup>1</sup>,  
Douglas Worsnop<sup>1,4</sup>, and Mikael Ehn<sup>1</sup>

<sup>1</sup> Institute for Atmospheric and Earth System Research / Physics, Faculty of Science, University of  
Helsinki, Helsinki, 00140, Finland

<sup>2</sup> Univ Lyon, Université Claude Bernard Lyon 1, CNRS, IRCELYON, F-69626, Villeurbanne,  
France

<sup>3</sup> Institute of Physics, University of Tartu, Tartu, 50090, Estonia

<sup>4</sup> Aerodyne Research, Inc., Billerica, MA 01821, USA

12

*First author:* Yanjun Zhang & Otso Peräkylä

*Correspondence to:* Yanjun Zhang (yanjun.zhang@helsinki.fi) & Chao Yan (chao.yan@helsinki.fi)

15

16

**Abstract.** Recent advancements in atmospheric mass spectrometry provide huge amounts of new information, but at the same time present considerable challenges for the data analysts. High-resolution (HR) peak identification and separation can be effort- and time-consuming, yet still tricky and inaccurate due to the complexity of overlapping peaks, especially at larger mass-to-charge ratios. This study presents a simple and novel method, mass spectral binning combined with positive matrix factorization ('binPMF') to address these problems. Different from unit mass resolution (UMR) analysis or HR peak fitting, which represent the routine data analysis approaches for mass spectrometry datasets, binPMF divides the mass spectra into small bins and takes advantage of PMF's strength in separating different sources or processes based on different temporal patterns. In this study, we applied the novel approach to both ambient and synthetic datasets to evaluate its performance. It not only succeeded in separating overlapping ions, but was found to be sensitive to subtle variations as well. Being fast and reliable, binPMF has no requirement for a priori peak information and can save much time and effort from conventional HR peak fitting, while still utilizing nearly the full potential of HR mass spectra. In addition, we identify several future improvements and applications for binPMF, and believe it will become a powerful approach in the data analysis of mass spectra.

**Keywords.** (Atmospheric) mass spectrometry, binned positive matrix factorization (binPMF), high-resolution (HR) mass spectra, peak fitting, chemical ionization mass spectrometer (CIMS), highly oxygenated molecules (HOM)

## 1. Introduction

Volatile organic compounds (VOC) are emitted to the atmosphere both from biogenic and anthropogenic sources (Guenther et al., 1995; Wei et al., 2008). After oxidation, these gaseous species can partition to the particle phase and contribute to atmospheric organic aerosol (OA), a major component of tropospheric particulate matter (Zhang et al., 2007). The chemical components, both in particulate (OA) and gaseous phase (VOC and their oxidation products), play important roles in many atmospheric physical and chemical processes. They can deteriorate air quality causing adverse health effects, and aerosol particles can influence Earth's climate by altering the radiative balance, as well as decrease visibility (Stocker et al., 2013; Zhang et al., 2016; Pope III et al., 2009; Shiraiwa et al., 2017).

Recent instrumental advances in mass spectrometry have greatly enhanced our capability to investigate the chemical composition and evolution of aerosol particles and their precursors. The Aerodyne aerosol mass spectrometer (AMS) is widely applied in atmospheric research (Canagaratna et al., 2007), measuring the bulk composition and temporal behavior of the non-refractory aerosol, and has successfully identified different/unique OA sources utilizing factor analysis (Jimenez et al., 2009; Zhang et al., 2011). With the development of gas-phase chemical ionization mass spectrometry (CIMS) (Huey, 2007), and the commercially available ToF-CIMS (Bertram et al., 2011) and CI-API-TOF (chemical ionization atmospheric pressure interface time-of-flight mass spectrometer (Jokinen et al., 2012)), these instruments are becoming more popular in atmospheric chemistry research. Due to these new advances, the detection methods for aerosol precursor vapors and the understanding of their formation mechanisms have been greatly improved. For example, the discovery of highly oxygenated molecules (HOM) by the CI-API-TOF has led to increased knowledge regarding atmospheric oxidation pathways, with large implications on secondary organic aerosol (SOA) and new particle formation (Ehn et al., 2014; Jokinen et al., 2015; Kirkby et al., 2016; Yan et al., 2016). In particular, biogenic VOC such as monoterpenes ( $C_{10}H_{16}$ ), promptly produce HOM upon ozonolysis, e.g.  $C_{10}H_{16}O_7$  and  $C_{10}H_{16}O_9$ .

While a mass spectrum can contain large amounts of information representing the highly complex nature of the atmospheric sample, it also presents considerable challenges for the analysis and interpretation of the data. One example of such a challenge is the identification and separation of peaks with a similar but not identical masses. A single integer mass can contain tens of distinct ions, with mass-to-charge ratios ( $m/z$ ) close to each other. In all cases, specific spectral fitting techniques are needed to resolve the overlapping peaks at the same integer mass. Typically, a least squares fit is made to the spectrum, using analysis software such as Squirrel/PIKA (<http://cires1.colorado.edu/jimenez-group/ToFAMSResources/ToFSoftware/>), tofTools (Junninen et al., 2010) or Tofware (<https://www.tofwerk.com/software/tofware/>). However, these techniques require a pre-defined list of ions. This makes the analysis resource-intensive, and it can easily introduce subjective bias in determining the peak list.

Figure 1 depicts a concrete example, measured by a nitrate-based CI-API-TOF, where peak separation is not large enough to allow unambiguous fitting of all the ions, and the final result will depend on which ions the analyst chooses to include. As the  $m/z$  increases, the number of possible ions at a certain unit mass increases rapidly (Kroll et al., 2011; Stark et al., 2015). Too closely overlapping peaks will sometimes lead to ambiguously fitted peaks and arbitrarily resolved ions, resulting in unreliable separation of signals. Additionally, mass calibration errors can also affect correct peak

80 assignment or fitting. A few recent studies discuss in more detail the uncertainties of ion identification  
81 and separation in HR mass spectra (Stark et al., 2015;Corbin et al., 2015;Cubison and Jimenez, 2015).

82 Another typical analysis approach is to utilize only the unit mass resolution, or UMR data. As opposed  
83 to high resolution fitting, where the signals of individual ions are separated from the total measured  
84 signal, in UMR analysis all signals at a given integer mass is integrated and treated together. This  
85 approach is more straightforward and less subjective than HR fitting, but loses all possible high  
86 resolution details in the spectrum (see Figure 2).

87 Even with perfect high resolution peak fits, a spectrum typically contains information of hundreds, if  
88 not thousands, of ions, many of which come from similar sources. This wealth of data itself presents  
89 a challenge for data analysis. Factor analysis enables the reduction of data dimensions and can help  
90 to apportion the signals to factors. These factors may correspond to different sources or formation  
91 processes. Positive matrix factorization (PMF) (Paatero and Tapper, 1994) has been widely utilized  
92 in environmental sciences, applied to UMR and HR AMS data, succeeding in identifying multiple  
93 OA sources (Lanz et al., 2008;Ulbrich et al., 2009;Sun et al., 2011;Zhang et al., 2011). Compared to  
94 AMS data, PMF has been applied to CIMS data analysis much less frequently. To our knowledge,  
95 only Yan et al. (2016) and Massoli et al. (2018) have reported PMF analysis on nitrate-based CI-API-  
96 TOF, utilizing UMR and HR data, respectively.

97 UMR-PMF cannot utilize the full information content provided by HR mass spectrometers, but is  
98 more straightforward to apply. In contrast, accurate HR peak fitting can better preserve the  
99 information content of the raw data than UMR, and thus provide more information to PMF, resulting  
100 in more interpretable results. However, incorrectly fitted peaks can severely disturb the PMF  
101 modeling and the factor interpretation. In addition, mass spectra from iodide-adduct Tof-CIMS (Lee  
102 et al., 2014), often contain more peaks per mass than the  $\text{NO}_3^-$ -CI-API-TOF, making HR fitting much  
103 more complex (or in some cases, even unmanageable), severely limiting the potential of HR PMF.

104 In this study, a novel, yet simple and reliable, data analysis method, binned mass spectra combined  
105 with PMF (binPMF), is proposed to try to tackle the abovementioned problems in both HR and UMR  
106 PMF. Instead of using traditional UMR or HR fitting techniques to the mass spectra, we binned the  
107 mass spectra prior to PMF analysis (Figure 2). We applied binPMF to both ambient and synthetic  
108 datasets, succeeding in separating the key components of different sources/processes. Compared to  
109 UMR PMF, binPMF preserves more of the high resolution information content of the mass spectra,  
110 without the immense effort and subjectivity associated with high-resolution peak fitting. As a result,

111 this novel method can improve our understanding of sources/formation processes governing the  
112 particulate and gaseous phases in more detail and in a less subjective manner.

## 113 2. Methodology

114 We divided the mass spectra into narrow bins as presented in Figure 2, and carried out PMF analysis  
115 to extract more information from the dataset. Details on the data preparation (binning the mass  
116 spectra), and error estimation for the PMF input are discussed in the Sections 2.2 and 2.3. To test the  
117 performance of binPMF under different scenarios, we first constructed synthetic datasets, using a  
118 simple one-/two-mass system (Section 2.4.1). In the second step, we applied binPMF to an ambient  
119 dataset measured with a  $\text{NO}_3^-$ -CI-API-TOF in a boreal forest site located in Southern Finland (Section  
120 2.4.2).

### 121 2.1. Positive matrix factorization

122 The PMF model was developed by Paatero and Tapper (Paatero and Tapper, 1994) in the 1990s and  
123 has been widely applied in the analysis of various types of environmental data ever since (Zhang et  
124 al., 2017; Yan et al., 2016; Ulbrich et al., 2009; Song et al., 2007). By decomposing the observed  
125 dataset into different factors, PMF helps to simplify the complex data matrix and extract useful  
126 information contained within it. Compared to other common source apportionment tools, like  
127 chemical mass balance (CMB) (Schauer et al., 1996), PMF requires no prior knowledge of source  
128 information as essential input. Nevertheless, as a statistical method, PMF does require more data as  
129 input, which is typically not a problem for environmental mass spectrometry datasets. The main  
130 distinction of PMF from other factor analysis techniques is that PMF utilizes a least squares  
131 minimization scheme weighted with data uncertainties, and non-negative constraints, to minimize the  
132 ambiguity caused by rotation of the factors (Huang et al., 1999; Paatero and Tapper, 1994).

133 In PMF modelling, a measurement of chemical species is assumed to be a sum of contributions from  
134 several relatively fixed sources/processes. The measured data matrix is broken down to two smaller  
135 matrices, and a residual term as follows:

$$136 \quad X_{(m \times n)} = \text{TS}_{(m \times p)} \times \text{MS}_{(p \times n)} + R_{(m \times n)} \quad (1)$$

137 where  $X$  represents an  $m \times n$  data matrix of original measurement for species  $n$  (e.g.  $m/z$ ) at time point  
138  $m$ ,  $\text{TS}$  is the  $m \times p$  time series matrix of factor contributions,  $\text{MS}$  is the  $p \times n$  matrix of factor profiles  
139 or the factor mass spectra, and  $R$  is the residual between the modelled and the measured data.  $p$  is the  
140 number of factors, which needs to be determined based on the interpretability of the PMF results,

141 among other criteria. Thus, in PMF, the original data matrix is approximated in terms of  $p$  factors,  
142 each of which has a distinct mass spectrum and time series.

143 To find the solution, the PMF model utilizes uncertainty estimates for each element in the data matrix  
144  $X$ . These uncertainty estimates are used to weight the residuals ( $R$ ), in order to calculate the  $Q$  value  
145 as

$$146 \quad Q = \sum_{i=1}^m \sum_{j=1}^n \left( \frac{R_{ij}}{S_{ij}} \right)^2 \quad (2)$$

147 where  $S_{ij}$  is the estimated uncertainty of species/mass  $j$  at time point  $i$ , and  $R_{ij}$  is the residual of that  
148 mass at the same time.  $Q$  is then minimized iteratively to find the mathematically optimal solution.  
149 An expected  $Q$  value ( $Q_{exp}$ ) can be calculated as the number of non-down-weighted data values in  $X$   
150 minus the sum of elements in TS and MS. If the data follows the requirements of PMF, the solution  
151 with the correct number of factors should have a  $Q/Q_{exp}$  value near unity. When this is true, the  
152 residuals on average fall within the expected uncertainties for each time point and variable. More  
153 details about uncertainty estimation will be discussed in Section 2.3. The PMF analysis in this study  
154 was performed with the toolkit of Source Finder (SoFi, v6.3) (Canonaco et al., 2013) by multi-linear  
155 engine (ME-2) (Paatero, 1999). Masses with low signal-to-noise ratio ( $SNR < 0.2$ , see Section 2.3 on  
156 error estimation) were down-weighted by a factor of 10, and masses with  $0.2 < SNR < 2$  were down-  
157 weighted by a factor of 2, as suggested by Paatero and Hopke (Paatero and Hopke, 2003). The down-  
158 weighting effect was considered in the  $Q_{exp}$  calculation. In this study, PMF was operated in robust  
159 mode, where outliers ( $\left| \frac{R_{ij}}{S_{ij}} \right| > 4$ ) were dynamically down-weighted (Paatero, 1997).

160 One of the problems in any factorization analysis is rotational ambiguity, which is caused by an  
161 infinite number of similar solutions generated by PMF (Paatero et al., 2002; Henry, 1987). Generally,  
162 the non-negativity constraint alone is not sufficient for solution uniqueness. Rotating a certain  
163 solution and assessing the rotated results is one possible way to determine the most physically  
164 reasonable solution. Known source profiles or source contributions can also serve as constraints. In  
165 addition, if there is a sufficient number of time points when the contribution of a source is nearly  
166 zero, independent of other sources, rotational uniqueness of solutions can be achieved (Paatero et al.,  
167 2002). The same is true if specific variables in the profiles go to zero. Otherwise, the correct solution  
168 (correct rotation) may only be obtained by skillful use of rotational tools. Ambient measurement data  
169 can often contain zero values in most sources/processes, greatly reducing rotational ambiguity of the  
170 PMF results. The issue of rotational ambiguity is not explored in detail in this manuscript, as it is  
171 common to all PMF approaches, and the main purpose here is to illustrate the new methodology of

binPMF. All the solutions shown in this study were achieved without considering their rotational uniqueness. Finally, we note that in addition to rotational ambiguity, binPMF also inherits all other fundamental limitations and strengths of the underlying PMF method.

## 2.2. binPMF data matrix preparation

Instead of UMR or HR fitting of the mass spectra, the mass spectra were divided into small bins after mass calibration (Figure 2 and Figure S1 in the Supplement). Data were first linearly interpolated to a mass interval of 0.001 Th, and then divided into bins of 0.02 Th width. At an integer mass  $N$ , only the signals between  $N-0.2$  Th and  $N+0.3$  Th (“the signal region”) were binned to avoid unnecessary computation of masses without any signal. With the binning, there were 25 data points for each nominal mass, instead of only one signal in UMR or several fitted peaks in HR analysis. All the parameters mentioned above, e.g. bin width and signal region size, should be adjusted to suit the mass spectrometer and the data being analyzed. Further details on binning procedures are discussed in Section 3.3.

## 2.3. binPMF error matrix preparation

Beside the data matrix, an error matrix describing the expected uncertainty for each element in the data matrix is also required as input in PMF analysis. Here, the error matrix (Polissar et al., 1998) is estimated as

$$S_{ij} = \sigma_{ij} + \sigma_{noise} \quad (3)$$

where the uncertainty of mass  $j$  at time point  $i$ ,  $S_{ij}$ , is composed of analytical uncertainty  $\sigma_{ij}$  and instrument noise  $\sigma_{noise}$ .  $\sigma_{ij}$  is the uncertainty arising from counting statistics and is estimated as

$$\sigma_{ij} = a \times \frac{\sqrt{I}}{\sqrt{t_s}} \quad (4)$$

in which  $I$  is the signal intensity, in counts per second,  $t_s$  is the averaging time in seconds and  $a$  is an empirical parameter incorporated to include unaccounted uncertainties (Allan et al., 2003; Yan et al., 2016). In our study, we applied binPMF with CI-API-TOF data as an example, and the same  $a$  value of 1.28 was utilized as estimated previously from laboratory experiments in the work of Yan et al. (2016). The  $\sigma_{noise}$  is calculated as the median of the standard deviation of instrument noise, calculated from the bins between two nominal masses that should be least influenced by real signals (the noise region), i.e.  $N+0.5$  -  $N+0.8$  Th (see Figure S1 in the Supplement).

## 2.4. Data sources and description

201 This study utilized both ambient and synthetic datasets to test the performance of binPMF. The  
202 ambient data was collected at the SMEAR II station (Station for Measuring Ecosystem–Atmosphere  
203 Relations (Hari and Kulmala, 2005)) in the boreal forest in Hyytiälä, Southern Finland. Located in a  
204 rural forest area, the station has a wide range of continuous measurements of meteorology, aerosol  
205 and gas phase properties year-round. There are no strong anthropogenic sources close to the site, but  
206 two sawmills 5 km to southeast and the city of Tampere 60 km to the southwest. Detailed  
207 meteorological parameters and concentrations of trace gases during this campaign have been  
208 presented earlier (Zha et al., 2018). Before the application to ambient data, we constructed a simple  
209 synthetic dataset, to examine how well binPMF can separate overlapping ions under different  
210 conditions.

#### 211 **2.4.1. Synthetic dataset**

212 As a first test of the performance of binPMF, we generated a series of synthetic datasets based on two  
213 distinct sources. Each synthetic dataset  $Y$  was created by summing up the signals of the two sources.  
214 Each source consisted of a constant source profile (represented as the matrix  $MS$ ), and had a unique  
215 temporal behavior (represented as the matrix  $TS$ ). Each source was the multiplication of  $MS$  (mass  
216 spectra / source profile) and  $TS$  (time series). The two  $TS$  for the two sources were generated  
217 randomly and independently from each other, as shown in Figure S2 in the Supplement (correlation  
218 coefficient  $R = 0.375$ ). To avoid rotational ambiguity (See section 2.1) in these tests, we added zero  
219 values to the time series of the two sources, independently of each other.

220 As shown in Figure 3, each source profile ( $MS$ ) was generated to consist of either one or two separate  
221 peaks, covering either one or two unit masses, respectively. The peaks were generated as Gaussians  
222 of known width and centroid position. The peaks of the different sources/profiles were partially  
223 overlapping, with the exact overlap, i.e. the distance ( $m/z$  difference) between the overlapping peaks,  
224 being varied from one experiment to another.

225 Peaks in the synthetic  $MS$  profiles were first generated as perfect smooth peaks (fine  $m/z$  interval of  
226 0.00001 Th), with mass resolution of 5000 Th/Th. We define the resolution  $R$  of a peak as  $R =$   
227  $M/\Delta M$ , where  $M$  is the mass of the ion, and  $\Delta M$  is the full width at half maximum signal intensity,  
228 FWHM. As an example, with  $R = 5000$  Th/Th, an ion at  $m/z$  of 300 Th will have a FWHM of 0.06  
229 Th, corresponding to 200 ppm. Multiplying the source profiles and the time series, we generated an  
230 ideal data matrix. From this ideal matrix, we sampled with  $m/z$  interval of 0.015 Th to simulate the  
231 real measurement data. The interval selected was close to that typically used for the HTOF mass  
232 analyzer on a CI-APi-TOF. After the sampling, Gaussian distributed noise, both from background



random noise and signal dependent noise were added to make up the data matrix  $Y'$ , point by point, as shown in Eq. 6 below. The variance of the Gaussian distributed noise was estimated as one hundredth of the coefficient 'c', which is the average value of  $Y$ .

$$Y = TS \times MS \quad (5)$$

$$Y'_{ij} = Y_{ij} + \text{Gaussian}(0, 0.01 \times c) + \text{Gaussian}(0, 1) \times \sqrt{Y_{ij}} \quad (6)$$

Finally, random  $m/z$  shift within  $\pm 10$  ppm was added to simulate mass calibration error, spectrum by spectrum. This error, resulting from inaccurate conversion of the time-of-flight into mass-to-charge ratio, is one of the main causes of ambiguous or incorrect peak assignment or fitting. In our study, with the bin width of 0.02 Th and the mass calibration error of 10 ppm, a maximum of 15% of one bin's signal may incorrectly shift to the adjacent bin, for a mass at 300 Th ( $(10 \text{ ppm} \times 300 \text{ Th}) / 0.02 \text{ Th} \times 100 \% = 15\%$ ). The impact of this mass shift will effectively be smaller, due to the high temporal correlation of adjacent bins, as the signal from an ion will spread to several adjacent bins (the FWHM is  $\sim 3$  times the bin width). In the case of HR fitting of peaks, a 10 ppm mass calibration error may cause much more dramatic changes than merely shifting 15% of the signal. There is also no reason for ions from a given source to selectively end up at the same integer mass, meaning that the signal is likely to be shifted to another ion from a completely different source.

Twenty-one synthetic experiments were designed, varying the mass difference between peaks ( $m/z$  difference) and number of unit masses included in the MS, as shown in Table S1 and Figure 3. For experiments 1-10, each of the two source profiles consisted of one peak (A1 and B1), both located at the same unit mass (chosen to be 310 Th in this study), with varying separation of the peak centroids. In experiments 11-20, we added one more peak to each profile (peaks A2 and B2), in addition to peaks A1 and B1. The additional peaks were added at another unit mass (311 Th) and their  $m/z$  difference was fixed at 0.05 Th (161 ppm), while the position of peak B1 was varied as in experiments 1-10. For experiment 21, peaks A2 and B2 were added at two different masses (311 Th and 312 Th), corresponding to a  $m/z$  difference sufficiently large that there was no meaningful overlap between them. In the MS (i.e. mass spectra profiles), all peaks had the same intensity level initially. The variation of the peak intensity ratio comes from variations in the time series (Figure S2 in Supplement). The same time series for each of the two sources was used in all experiments 1-21.

With this approach of only using two masses, we purposefully provide a challenging dataset for binPMF, as in most real datasets there would be many more masses to help to constrain the final solutions. Nevertheless, as we will show, this simple synthetic dataset already provided a wealth of

264 useful information in the results attainable with binPMF, and provided a good comparison to the  
265 traditional HR fitting approach.

## 266 2.4.2. Ambient dataset

267 The ambient dataset was measured at ground level during the Influence of Biosphere-Atmosphere  
268 Interactions on the Reactive Nitrogen budget (IBAIRN) campaign (Zha et al, 2018) in September,  
269 2016. The measurements were conducted using a  $\text{NO}_3^-$ -CI-APi-TOF that has been described in detail  
270 elsewhere (Jokinen et al., 2012; Junninen et al., 2010; Yan et al., 2016). Here, the ambient gas-phase  
271 molecules clustered with the nitrate ion were measured with about 4000 Th/Th mass resolving power.  
272 Data from September 1<sup>st</sup> to 26<sup>th</sup>, averaged to 1-hour time resolution, in the mass range of 300 - 350  
273 Th (a typical monoterpene HOM “monomer” range, Ehn et al., 2014) were utilized for the binPMF  
274 analysis. A baseline subtraction was applied to the mass spectra, which caused some small signals  
275 next to large peaks to become negative. In our analysis, any  $m/z$  bin where the median signal was  
276 negative was excluded from the data matrix.

## 277 3. Results and discussion

### 278 3.1. Synthetic dataset

#### 279 3.1.1. Experiment settings

280 As introduced in Section 2.4.1, the synthetic datasets were constructed to assess the response of  
281 binPMF to varied  $m/z$  difference, peak intensity ratios, and number of masses included, as shown in  
282 Table S1 and Figure 3. The smaller the distance between the two peaks, the harder it is to accurately  
283 separate them with traditional HR peak identification and fitting. In our experiments, the  $m/z$   
284 difference was decreased stepwise from 0.050 Th (161 ppm) to 0.001 Th (3 ppm), in a system where  
285 the FWHM was roughly 200 ppm.

286 The analysis procedure of the synthetic dataset is briefly described here. In all cases, the parameter  
287 of interest is to see how well binPMF is able to deconvolve the adjacent peaks A1 and B1 at  $m/z$  310  
288 Th. First, binPMF was applied to the synthetic datasets, and factors profiles (mass spectra) were  
289 extracted. The optimal number of factors for the synthetic dataset is two, the same as the number of  
290 sources, so only the two-factor solution was studied with binPMF. The results of the diagnostic  
291 parameter  $Q/Q_{exp}$  for each experiment are included in Table S1. Gaussian fitting was then performed  
292 on the factor profiles to retrieve the locations of peaks A1 and B1, and thereby assess how well  
293 binPMF was able to retrieve the original peak positions.

294 In addition to applying binPMF to the synthetic datasets, traditional HR peak fitting was also  
295 conducted as comparison (by tofTools in our study). For the tofTools fits, we constrained the peak  
296 locations and widths to those originally used for generating the data (Table S1). Peaks fitted by  
297 tofTools and peaks fitted to the binPMF factors were compared, as well as the retrieved time series  
298 correlation with the original datasets. More details are presented and discussed in the following  
299 sections.

### 300 **3.1.2. Comparison of peak fitting**

301 We examined the performance of traditional HR fitting and binPMF by comparing their results to the  
302 original input data. In Figure 4, the shaded areas depict the original data, the dashed lines the  
303 traditional HR peak fitting result, and the solid lines the binPMF factors. Red and blue represent  
304 source/factor A and B, respectively. Panels a-d (in Figure 4) show four scenarios of peak fitting results  
305 from experiments 1, 5, 10 and 20, at the 79<sup>th</sup> time point, where the two peaks had similar signal  
306 intensities. When the two peaks were separated by 0.05 Th (Figure 4a), both methods captured the  
307 peak intensities quite well. However, as the  $m/z$  difference narrowed, the performance of both  
308 methods declined, with the HR fitting results deteriorating faster than those from binPMF. As  $m/z$   
309 difference reached 0.001 Th (3 ppm), the traditional HR fitting method completely failed to fit the  
310 two peaks (panels c and d), instead attributing all the signal to just one fitted mass. In panels e-h, the  
311 peak fitting results at the 21<sup>st</sup> time point are displayed, where the ratio of the two peaks was roughly  
312 1:6. Here, the traditional fitting method failed to extract the two peaks already at a  $m/z$  difference of  
313 0.01 Th (30 ppm), attributing all signal to Peak B1 (panels g and h). As shown in panels d and h,  
314 when a second set of peaks, separated by 0.05 Th, was introduced for the two sources in the datasets,  
315 binPMF was able to utilize the temporal behavior of peaks A2 and B2, performing much better, even  
316 in the extremely difficult cases when the  $m/z$  difference for the two peaks was only 0.001 Th (3 ppm).  
317 It is an inherent advantage of binPMF over traditional peak fitting methods that the temporal behavior  
318 and the correlations between different variables can be utilized.

319 Figure 5 shows an overview of all the results of peaks fitted with binPMF. Experiments 1-10 for the  
320 one-mass system are shown with green lines, and experiments 11-20 for the two-mass system in  
321 yellow. Mass accuracy was calculated as the difference between fitted peak center mass and the  
322 original mass, divided by the original mass, in ppm. When the  $m/z$  difference got smaller, the mass  
323 accuracy of peaks fitted to binPMF factors declined (Figures 5a and 5c). At a  $m/z$  difference of 0.01  
324 Th (32 ppm), the mass accuracy was  $-4 \pm 2$  ppm and  $7 \pm 2$  ppm for peaks A1 and B1, respectively. The  
325 uncertainties were estimated by repeating the analysis with 10 different random time series for the  
326 two sources (Brown et al., 2015). For comparison, this separation approximately corresponds to that

327 between  $\text{C}_{10}\text{H}_{16}\text{O}_7\cdot\text{NO}_3^-$  (310.0780 Th) and  $\text{C}_9\text{H}_{16}\text{N}_2\text{O}_6\cdot\text{NO}_3^-$  (310.0892 Th). With  $m/z$  difference  
328 decreasing, the position of peak A1 (the left red peak in Figure 3), as identified by binPMF, shifted  
329 gradually to the left, while peak B1 (the right blue peak) shifted to the right. When peaks A2 and B2  
330 were introduced to the sources, the mass accuracy improved ( $< 6$  ppm). The resolution of the peaks  
331 fitted to binPMF factor profiles stayed fairly constant, but had degraded compared to the original  
332 input data (5000 Th/Th), explained at least partially by the data binning (Figures 5b and 5d). Overall,  
333 binPMF performs relatively well in peak separation, with reasonable mass accuracy and peak  
334 resolution compared to the original datasets.

### 335 3.1.3. Correlation of time series

336 In addition to the peak positions, we also compared the temporal behavior of both the binPMF factors  
337 and the time series obtained through traditional fitting, to the original time series. When  $m/z$  difference  
338 was larger than 0.02 Th (65 ppm), both methods worked similarly well in reproducing the original  
339 time series (Figure 6). As the  $m/z$  difference decreased below 0.02 Th (65 ppm), correlations  
340 decreased rapidly (panels a and c), with that of the traditional method decreasing faster. However, as  
341 shown by the yellow lines, when peaks A2 and B2 were added to each source profile, the time series  
342 correlation coefficients between original data and peaks extracted by binPMF were close to unity in  
343 experiments 11-20. The coefficients stayed similar to that from the experiment with  $m/z$  difference of  
344 0.05 Th (161 ppm), which was the fixed  $m/z$  difference for the two new peaks added at 311 Th in  
345 experiments 11-20. This means that the separation of the factor time series was mainly driven by the  
346 additional, better separated peaks. Again, the traditional HR fitting method could not utilize the  
347 information at 311 Th, and therefore no improvement to the peak deconvolution at 310 Th was seen.  
348 In addition to the correlation analysis, also the assignment of absolute signal to peaks A1 and B1 was  
349 evaluated. This was done by a linear fit (through zero) to the data points retrieved by the different  
350 methods as a function of the original input data. The slopes of the fitted lines are plotted in Figures  
351 6b and 6d, and show that the signal was for the most part correctly attributed to within a few percent.  
352 The largest scatter in the determined slopes were observed for binPMF experiments with only one  
353 mass, at low peak separations.

### 354 3.1.4. Summary and discussion

355 Based on the results shown above, binPMF was found to be as capable of separating different peaks  
356 as traditional peak fitting techniques when the two peaks were separated by more than the mass  
357 calibration uncertainty (yet still in all cases by less than the FWHM of the peaks). As the  $m/z$   
358 difference of the two overlapping peaks decreased, the performance of the traditional method declined

359 faster than that of binPMF. This was shown for signal attribution of fitted peaks and time series  
360 correlation with original data. When masses with co-varied temporal behavior of the targeted  
361 overlapping peaks were introduced in the dataset, the performance of binPMF improved significantly.

362 The peak fitting principle of the traditional method and binPMF are very different. For example,  
363 tofTools fits peaks based on pre-determined instrument parameters (e.g. peak shape and peak width),  
364 as well as the peak location, either as a numeric value, or a chemical composition from which the  
365 location is calculated (Junninen et al., 2010). HR peak fitting by tofTools can be effective if the  
366 majority of the components (peaks) are known and provided in a peak list, which is valuable  
367 information for peak separation that was not provided to binPMF in this study. However, this  
368 information can be hard to achieve due to unknown numbers and/or identities of all the ions at a given  
369 mass, in combination with the limited mass resolving power of the mass spectrometer. HR peak fitting  
370 is also sensitive to mass calibration error, increasingly so when many ions in close proximity to each  
371 other need to be fit. On the contrary, in binPMF, peaks are separated based on the temporal variation  
372 of masses, which is an inherent advantage of PMF, though no information of the peaks is provided  
373 beforehand. To be more specific, a conceptual illustration is shown in Figure S3 in the Supplement.

374 The red peaks belong to Source A and the blue peaks to Source B. As mentioned before, the time  
375 series of sources A and B were totally independent and random. The shaded areas (the tails of the  
376 peaks), e.g. red shaded area in Figure S3a, contained masses that only had significant signal from  
377 peak A1 (left red peak). Similarly, the blue shaded area in Figure S3a was mostly from peak B1. The  
378 different temporal behaviors of the red and blue shaded areas helped the separation and correct  
379 attribution also in the regions with overlapped signals. When the  $m/z$  difference of peaks A1 and B1  
380 decreased, shown in Figure S3b, the two shaded areas also became smaller. This is the main reason  
381 why the fitted masses of binPMF had lower mass accuracy and lower correlation coefficients  
382 compared to the original data, as the  $m/z$  difference decreased.

383 When peaks A2 and B2 ( $m/z$  difference of 0.05 Th) were added in the dataset, peaks A1 and B1 were  
384 better separated and fitted by binPMF compared to the scenarios with only one mass. This is because,  
385 as shown in Figure S3c, the red and blue shaded areas became larger due to the addition of two more  
386 peaks. In this case, it was peaks A2 and B2 that dominated the separation of sources A and B. In  
387 experiment 21, three integer masses were included in the dataset. Though it was still equally difficult  
388 for the traditional HR method to separate and fit peaks A1 and B1 with  $m/z$  difference of 0.001 Th (3  
389 ppm), it was the easiest experiment for binPMF out of all the experiments because of the large  $m/z$   
390 difference of peaks A2 and B2 (1.000 Th, 3225 ppm). In experiment 21, the mass accuracies for peaks  
391 A1 and B1 were -3.2 ppm and 2.6 ppm, respectively, and the time series correlation with original data

392 was 1.000 and 0.999, respectively. In most real-world applications, individual sources typically  
393 contain multiple peaks, and the correlations of these can be utilized by binPMF.

394 We note once more that the results of binPMF and traditional HR peak fitting are not totally  
395 comparable. Information about the peaks, like the exact peak centroid position, peak width  
396 (resolution) and number of peaks, was provided to the traditional fitting method. For binPMF, no  
397 prior information about the peaks was given, except for the optimal number of factors, i.e. two.

### 398 **3.2. Ambient dataset**

399 With the success of binPMF for the synthetic datasets, we applied the new method to a real ambient  
400 dataset. Here we used data collected in September 2016, from Hyytiälä in Finland. The SMEAR II  
401 station is a forest site dominated by monoterpene ( $C_{10}H_{16}$ ) emissions (Hakola et al., 2006). Previous  
402 CI-API-TOF measurements of HOM at the site (Ehn et al., 2014; Yan et al., 2016) have presented  
403 bimodal distributions, with one mode corresponding to HOM monomers (range 300-400 Th) and the  
404 second to HOM dimers (450-650 Th). For testing the binPMF analysis on our ambient dataset, we  
405 selected the HOM monomer range of 300-350 Th. While the synthetic dataset primarily compared  
406 binPMF to traditional HR fitting analysis, in this section, we compare the binPMF results with that  
407 of traditional UMR-PMF, as employed by Yan et al. (2016). HR fitting was not performed for the  
408 ambient dataset, for all reasons mentioned in earlier sections, including the difficulty and efforts of  
409 producing a proper unambiguous peak list, as well the limitations of overlapping peaks.

410 As mentioned above, no prior knowledge was provided to PMF before the analysis. To determine the  
411 number of factors for further analysis, we conducted runs with two to eight factors. As the number of  
412 factors increased, more information could be extracted from the raw data. However, after the optimal  
413 number of factors, the additional factors may split the physically reasonable factors into meaningless  
414 fragments. There has been many studies on evaluations of PMF runs and selections of PMF factor  
415 number (Zhang et al., 2011; Craven et al., 2012). This is an inherent challenge in any PMF analysis,  
416 and not specific to binPMF, and therefore we do not put emphasis on this here. In this study, based  
417 on commonly used mathematical parameters and physical interpretation, we chose the seven-factor  
418 result, as presented below. Our main aim with this work is to present a ‘proof of concept’ for the  
419 binPMF methodology, and we will therefore not provide a detailed interpretation of all the factors  
420 (though several of the factors are easily validated based on earlier studies). The factor evolution from  
421 two to eight factors are briefly discussed below.

422 From two to six factors,  $Q/Q_{exp}$  showed a dramatic decrease from 6.5 to 2.7. Then for seven and eight  
423 factors,  $Q/Q_{exp}$  decreased to 2.3 and 2.0, respectively. The unexplained variation also declined from

14% to 8.8% going from two to six factors, then reached 8.0% for seven factors, and 7.6% for eight factors. The two-factor solution first split the data into a daytime factor and a nighttime factor, with very distinct mass spectral profiles. The daytime factor was characterized by signals at 307 Th, 311 Th, 323 Th, 339 Th and other odd masses, while the nighttime factor was dominated by 308 Th, 325 Th, 340 Th and 342 Th. The odd masses are typical signatures of day-time monoterpene-derived organonitrates at the site, while the even masses, and specific odd masses e.g. a radical at 325 Th, have been identified as monoterpene ozonolysis products (Ehn et al., 2014; Yan et al., 2016). As the number of factors increased, the daytime factor was further split into new daytime factors, with diurnal profiles having various peak times around noon or early afternoon. When the number of factors increased to seven, a clear sawtooth shape in the diurnal trend was resolved with marker masses at 308 Th, 324 Th, 325 Th, and 339 Th. Many of the profiles resolved in the seven-factor solution are similar to those found by Yan et al. (2016), and separating more factors did not yield new factors that we could interpret. Therefore, we opted to use this seven-factor result for the main discussion below, as it provided us with enough information to evaluate the binPMF method for this dataset.

Figure 7 shows the mass spectral profiles and factor time series for the 7-factor result, while Figure 8 displays the diurnal trends and factor contributions to the total signal. As shown in Figure 8a, the seven factors separated by binPMF consist of one nighttime factor (Factor 1), five daytime factors (Factors 2, 3, 4, 5 and 7) and a sawtooth-pattern factor (Factor 6). The same dataset was also analyzed by UMR-PMF, and the corresponding seven-factor results are also included in Figures 7 and 8 for comparison.

Overall, the results between UMR-PMF and binPMF are very similar. UMR-PMF also resolved one clear nighttime factor, and additionally six daytime factors. For the nighttime factor, both binPMF and UMR-PMF showed comparable temporal behavior, diurnal trend (peak at 17:00), mass spectral profiles (peaks at 340 Th, 308 Th, 325 Th, 342 Th) and factor contribution (~ 20%). This factor has been validated in both chamber and ambient studies to be formed from monoterpene ozonolysis (Ehn et al., 2014; Yan et al., 2016). As shown in Figure 7a, both methods also resolved similar, though not identical, mass spectral profiles for the other six factors, with mostly comparable time series (Figure 7b) and peak times in the diurnal trends (Figure 8a).

Despite the similarities, there also existed distinct differences between the results from binPMF and UMR-PMF. As the most distinctive dissimilarity, binPMF Factor 6 revealed a “contamination factor”. This factor was found to be related to automated instrument zeroing every three hours, giving rise to the distinct three-hour sawtooth pattern. The zeroing system introduced some additional

457 compounds into the sampling lines, and the semi-volatile nature of these compounds caused them to  
458 linger, and slowly decay, in the tubing even after the instrument had returned to sampling ambient  
459 air. binPMF accurately retrieved the 3-hour interval of the zero measurements. However, with the  
460 same mass range (300-350 Th), UMR-PMF failed to extract the contamination factor, regardless of  
461 the number of factors retrieved (up to 20 factor solutions were evaluated). Instead, these  
462 contamination signals were always mixed into the other factors. Factor 6 from UMR-PMF contributed  
463 almost twice as much as that estimated by binPMF due to the inaccurate factor separation (Figure  
464 8b). The time series of other factors, e.g. Factors 5 and 7 in UMR-PMF, were clearly influenced by  
465 Factor 6. Compared to UMR-PMF, binPMF thus showed a clear advantage in providing more  
466 information out of the data by being more sensitive to subtle variations.

467 In addition to better resolving certain factors from the data, the binPMF mass spectral profiles will  
468 still contain more information than visible in Figure 7, due to the multiple bins at each unit mass. As  
469 an example, binPMF Factor 6 showed masses with clear negative mass defects, e.g. at 324 Th and  
470 339 Th (Figure 9). We identified many ions in this factor as different fluorinated carboxylic acids,  
471 common interference signals in negative ion CIMS, outgassing from e.g. Teflon tubing (Brown et al.,  
472 2015; Ehn et al., 2012; Heinritzi et al., 2016). The exact source of these products in our setup was not  
473 established, but it is not surprising that the additional valves, filters and/or tubing in the zeroing line  
474 could have caused this type of signal to be introduced to the instrument with the zero air. In general,  
475 this finding highlights the usefulness of the binPMF approach, where factor separation can be  
476 performed first, and the specific factor profiles can be utilized in interpreting the physical meaning  
477 of the different factors. This is in complete contrast to the more traditional approach, where all ions  
478 need to be identified first, and only then can HR PMF be attempted. As not all ions are going to be  
479 observable at all times, many ions may remain unidentified. For example, if peak identification would  
480 only have been done during periods when the HOM signals were high, as in the case shown in Figure  
481 9a, the fluorinated ion at 339 Th would not have been found (contributing only 0.45% to the total  
482 signal at this time point), even though it on average contributes nearly 10% of the signal at this mass  
483 over the entire campaign. binPMF, on the other hand, utilized the full dataset for the identification,  
484 and was able to separate several ions at 339 Th. By fitting gaussian signals to the factor profiles,  
485 similar to the synthetic data in section 3.1.2, we see that the two major peaks were fitted with decent  
486 resolution (Figure 9). Also the contamination factor (Factor 6) was clearly separated and fitted, and  
487 the resolution (3136 Th/Th) is slightly underestimated by the fit, as only one gaussian was fitted to  
488 each profile, yet there are clearly more than one ion at 339 Th in Factor 6. As shown in Figure 9c,  
489 there is also clear indication that Factors 3 and 5, which together make up as much signal at 339 Th



as the contamination Factor 6, mainly contain signals from another molecule ( $C_{10}H_{13}O_9$ ) than the dominant signals at this mass ( $C_{10}H_{15}NO_8$ ). However, further work will be needed to validate this. Factor 1 has marginal contribution to the signal at 339 Th (as seen in Figure 9b) throughout the campaign, and we expect it does not contain useful signal, as is suggested by the unreasonably high resolution, i.e. narrow peak width, of the fitted peak. The resolving power of the instrument was around 4000 Th/Th, and thus any apparent peak resolution above that will be unrealistic. However, as this factor contains signal at the outer edges of the main peaks at this  $m/z$ , it is possible that this factor relates to some instrumental variability affecting the peak shape. This is highly speculative, but such a phenomenon may be worth looking into in later studies utilizing binPMF. In summary, resolving multi-overlapping peaks by traditional methods is time-consuming and can be tricky and ambiguous. Here, binPMF greatly simplified this problem, by providing additional separation between the ions.

### 3.3. Future improvements and applications

The new technique for mass spectra analysis, binPMF, as presented above, shows clear promise in utilizing HR information while saving time and effort, as well as decreasing ambiguity related to conventional HR peak fitting. It is also more sensitive to subtle variation than standard UMR analysis. We consider this study a succesful proof-of-concept, and note that several future improvements and applications are still foreseeable. We list some of these below:

(1) **Varied bin width.** The full width at half maximum of an individual peak in a mass spectrometer is mass dependent, with peaks getting wider at higher masses. In binning the mass spectrum with a constant bin width, like in this study, the average number of bins per peak increases as a function of mass. To represent the peaks in a comparable manner, the bin width should thus be dependent on the mass. Varying the bin width as a function of the mass, and the mass resolution of the instrument, would enable a constant number of bins (e.g., seven) per peak. Too few bins per peak would mean that we may lose valuable information in the binning, and potentially risk introducing aliasing effects, while too many points per peak would lead to an unnecessarily high number of variables, without noticeable gain in information content. This would also result in high computational cost. If targeting 7 bins per peak, then the function for determining bin width based on  $m/z$  and resolution ( $R$ , which is mass-dependent) could be

$$\because R(m/z) = \frac{m/z}{\Delta m}$$

$$Bin\ width \times 7 = 2 \times \Delta m$$

522  
523  
524  
525  
526  
527  
528  
529  
530  
531  
532  
533  
534  
535  
536  
537  
538  
539  
540  
541  
542  
543  
544  
545  
546  
547  
548  
549  
550  
551  
552

$$\therefore \text{Bin width} = \frac{2}{7} \times \frac{m/z}{R(m/z)}$$

$\Delta m$  is the full peak width at half maximum signal intensity. If we consider an instrument with approximate constant resolution of 5000 Th/Th for masses from 200 Th to 600 Th, the bin width at 200 Th and 600 Th should be around 0.01 Th and 0.03 Th, respectively.

- (2) **Optimization of binning region.** Similarly to bin width, the binning region, i.e. the signal region ( $[N-0.2, N+0.3]$  in this study, introduced in section 2.2), should also be mass-dependent. Due to the widening of the peaks with increasing mass, the binning region should also get wider. In addition, the typical mass defect of measured ions typically varies with mass. This means that the binning regions should not necessarily be defined with respect to the integer masses, but to some chosen mass defect. Another approach would be to bin all the data, and remove the bins not meeting a certain criterion, such as one related to the signal to noise ratio in that bin, afterwards. In this case, there would be no need for a pre-defined mass defect or region width, and one could utilize the signals that do not fall within the expected regions.
- (3) **Error estimation.** Good error estimation is crucial to PMF calculation. Uncertainty in PMF analyses arises from three main causes, random errors in data values, rotational ambiguity, and modeling errors (Paatero et al., 2014). Variations in mass calibration are one example of a modeling error. It is common practice to increase uncertainty values  $S_{ij}$  specified for data values disturbed by modeling errors. This increase does not account for the mass calibration error in the sense that the effect of mass calibration variation would disappear. The increase simply balances residuals in different data values so that the best possible result may be obtained. In addition to the two error estimation terms discussed in section 2.3,  $\sigma_{ij}$  and  $\sigma_{noise}$ , a third form of error, to balance the mass calibration variation could also be considered for error estimation. Similarly, although generally rare and suggestive of some instrumental problem, if the peak shape or resolution shift over time, this would also require an improved error estimation in order to account for increased variability.
- (4) **Multi-peak fitting.** As discussed, peak identification is one of the most time-consuming and potentially ambiguous tasks in HR analysis, and with binPMF this may not always be a necessary task. However, as binPMF often resolves several peaks (chemical components) at each integer mass, peak identification can be made easier if peak identification is constrained to several binPMF factor profiles rather than just the initial HR spectrum. The optimal approach for this will be the target of a future study.

Most likely several other improvements to the approach will be identified in future studies, and simplicity of the analysis remains a critical consideration. We propose that binPMF is a good tool for initial exploration of new datasets, at which stage optimizing all parameters is not necessarily crucial, if the results can help guide further analysis directions. However, for maximizing the information content that can be extracted from a given data set, optimized routines are important.

#### 4. Conclusions

While recent advances in mass spectrometry have greatly enhanced our understanding of atmospheric chemistry, the increased information content in mass spectra also brings difficulties and challenges to the data analysis. Peak identification and separation can be challenging and ambiguous, as well as extremely time-consuming and involving large uncertainties. Constructing peak-lists, i.e. deciding which ions to fit to the mass spectra, and validating the results is becoming one of the most labor-intensive parts of the entire work. In this study, we propose a simple and reliable method, binPMF, to try to avoid many of these problems, while still being able to distinguish different chemical pathways/sources in the atmosphere.

Different from traditional analysis, binned positive matrix factorization (binPMF), divides the mass spectra into smaller bins, before applying PMF to distinguish different types of factors and behavior in the data. This method utilizes more available information than classical UMR-PMF, and requires no prior peak information as in the case of traditional HR-PMF. We applied binPMF successfully to both ambient and synthetic datasets to test its usefulness under different circumstances.

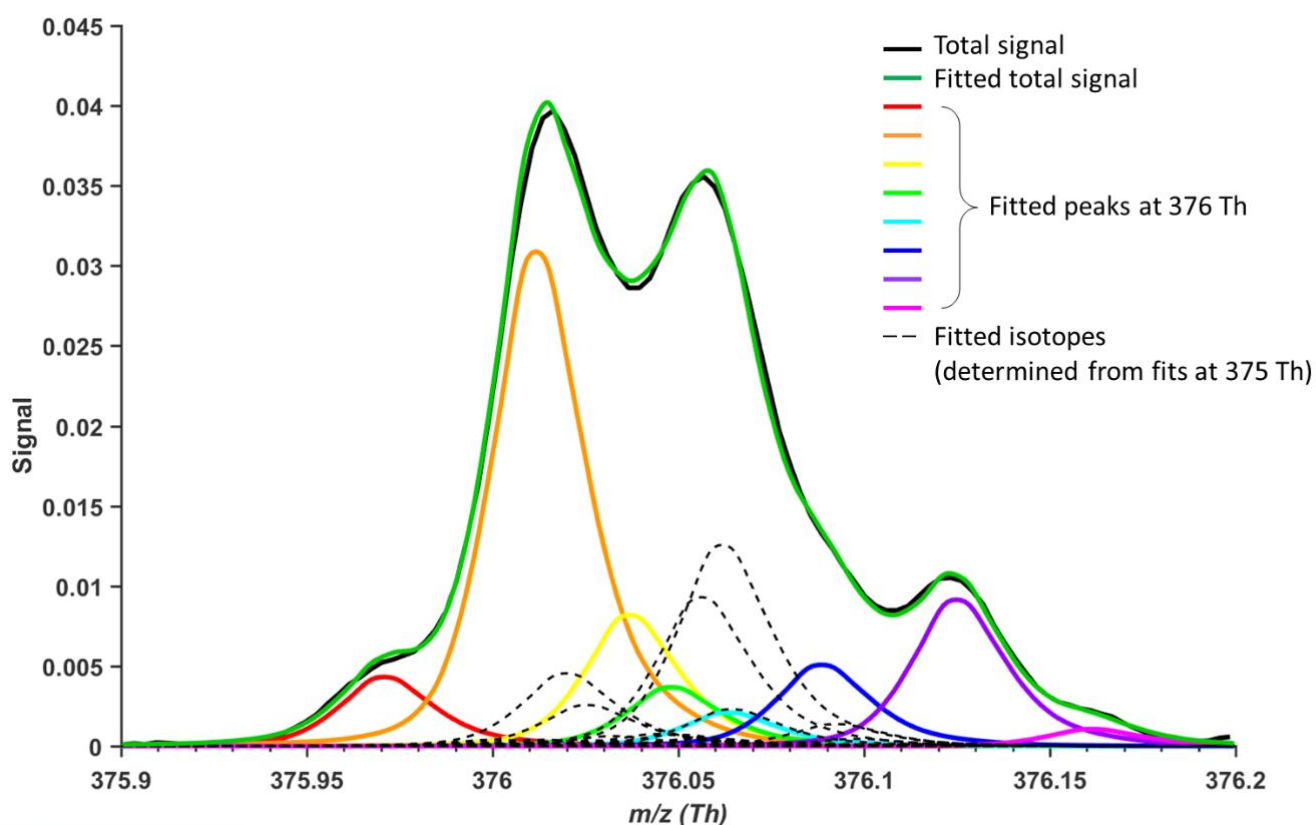
Traditional HR analysis fits peaks to each mass according to a pre-defined list, and is not able to utilize any information across masses or time. In our analysis of a simple synthetic data set with two overlapping ions at a single integer mass, we found that binPMF was able to separate the contributions of each ion even in cases where the HR analysis failed completely. This was the case for overlapping ions where binPMF had help in constraining the time series from another integer mass. When applied to an ambient dataset of HOM measured by a CI-API-TOF, binPMF identified more physically meaningful factors than UMR-PMF. Additionally, for factors where the two PMF approaches agreed, binPMF still contained more mass spectral information for ion identification, as compared to UMR-PMF.

We provide a proof-of-concept for the utility of binPMF, showing that it can outperform the two traditional analysis approaches, UMR and HR. We identify several future improvements and applications for binPMF, including an approach to greatly facilitate the time-consuming process of peak-list construction. We expect binPMF to become a powerful tool in the data exploration and analysis of mass spectra.

586    **Acknowledgements**

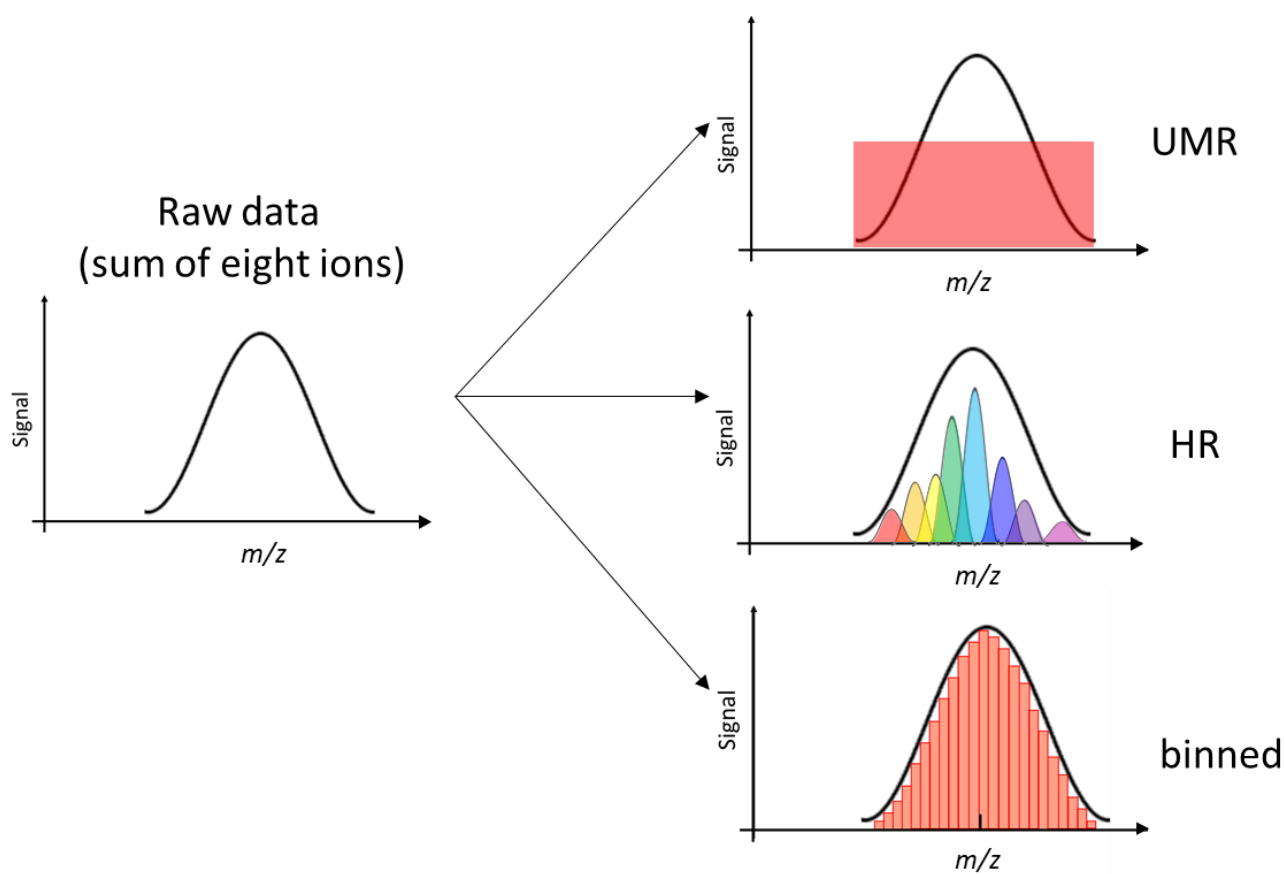
587    This research was supported by the European Research Council (Grant 638703-COALA), the  
588    Academy of Finland (grants 317380 and 320094), and the Vilho, Yrjö and Kalle Väisälä Foundation.  
589    K.R.D. acknowledges support by the Swiss National Science postdoc mobility grant  
590    P2EZP2\_181599. We thank the tofTools team for providing tools for mass spectrometry data  
591    analysis. The personnel of the Hyytiälä forestry field station are acknowledged for help during field  
592    measurements.

593



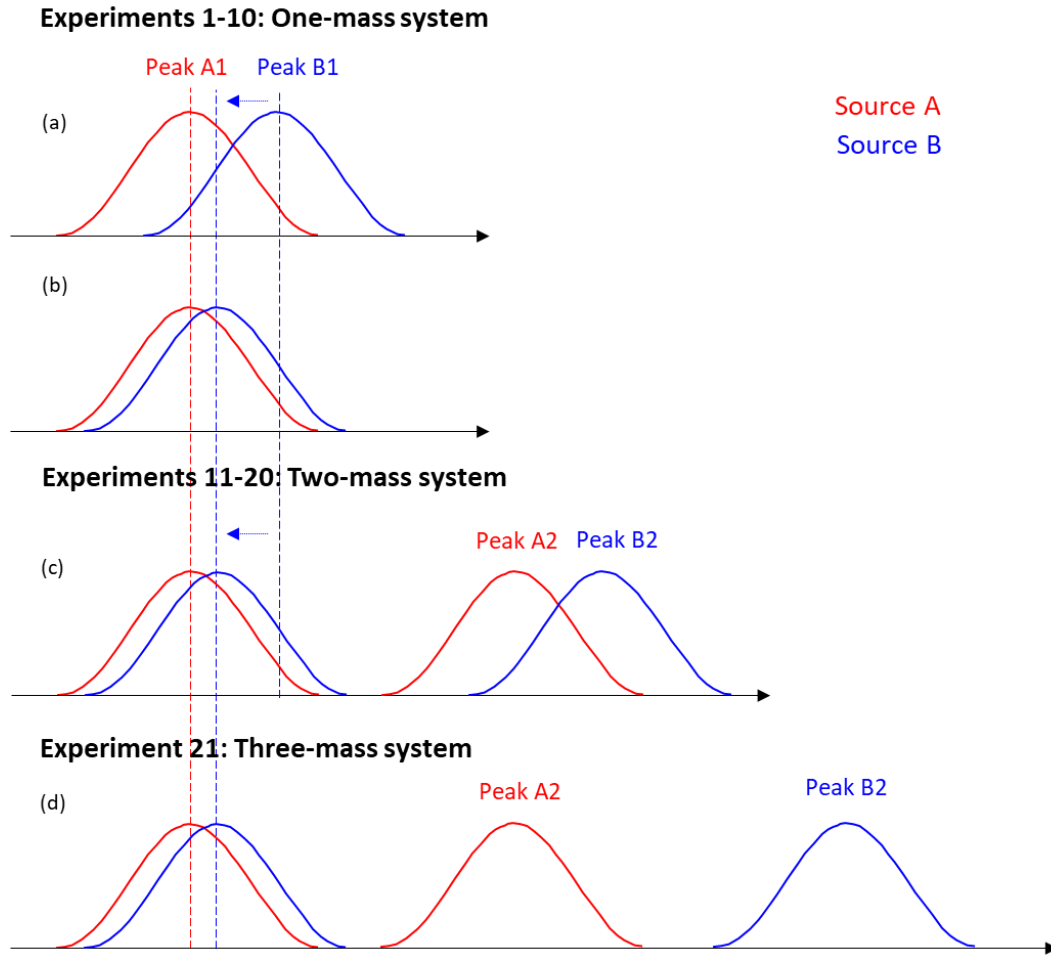
594

595 Figure 1. Example of traditional HR peak fitting. Potential peak fitting at  $m/z$  376 Th (10-h average)  
 596 in an atmospheric simulation chamber during a monoterpene ozonolysis experiment, utilizing a  
 597 nitrate-based CI-APi-TOF (resolving power of 13000 Th/Th). Even a minor shift in the mass axis  
 598 calibration could cause the signals of especially the yellow, green and blue peaks to change  
 599 dramatically. Similarly, adding or removing an ion would alter the amount of signal attributed to the  
 600 other fitted peaks.



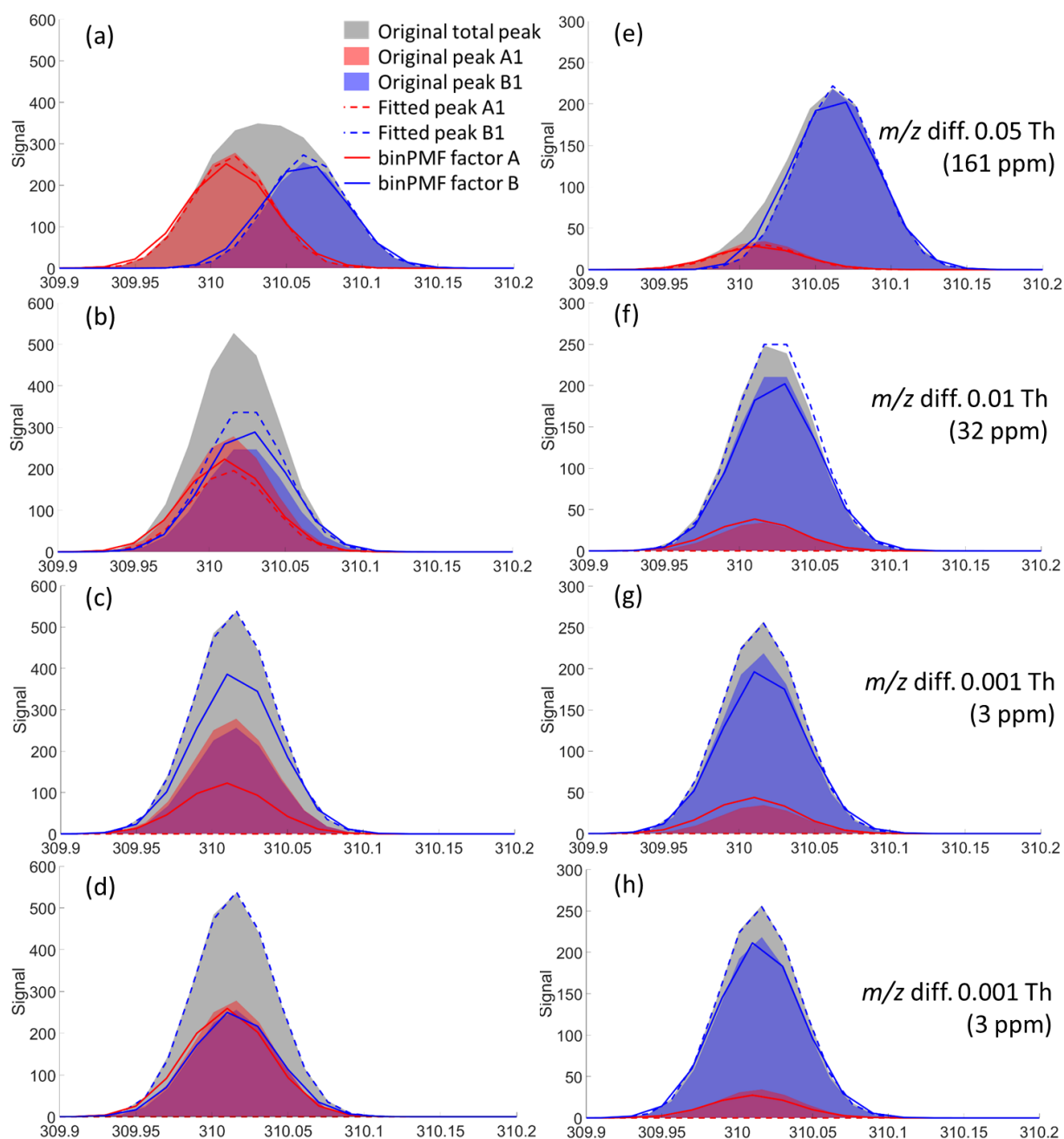
601

602 Figure 2. Conceptual comparison of traditional methods (UMR and HR) and binned mass spectra for  
 603 PMF analysis. The raw data signal is shown in the left and contains eight ions. By UMR analysis, the  
 604 information of the eight ions is totally lost. Using an analyst-determined peak-list, HR analysis  
 605 attempts to separate signals at this mass by fitting selected ions. By binning the spectra, we utilize the  
 606 HR information without any a priori information required.



607

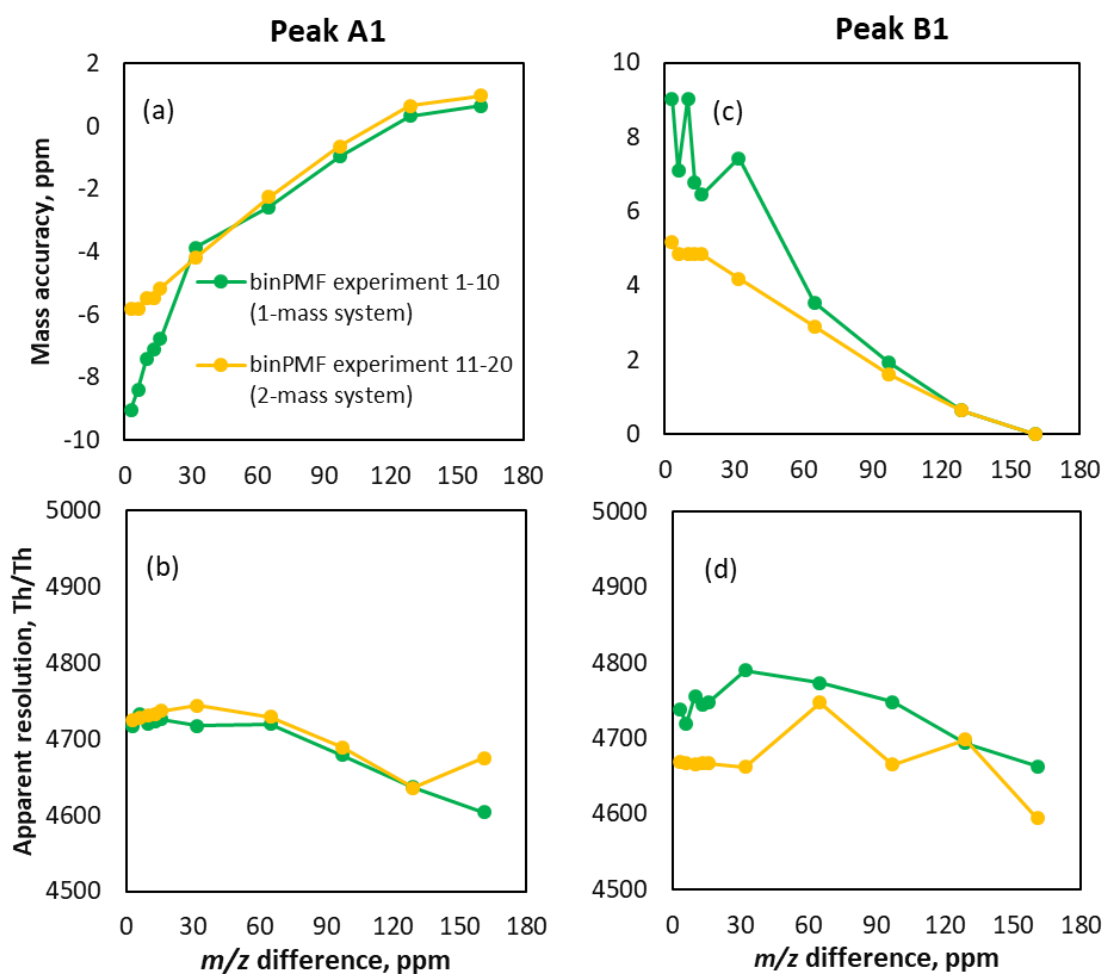
608 Figure 3. Conceptual schematic diagram for the synthetic datasets. Panels a and b describe  
 609 experiments 1-10 in the one-mass system, panel c is experiments 11-20 in the two-mass system. Panel  
 610 d shows experiment 21, with peaks A2 and B2 at separate integer masses (see text for details).



611

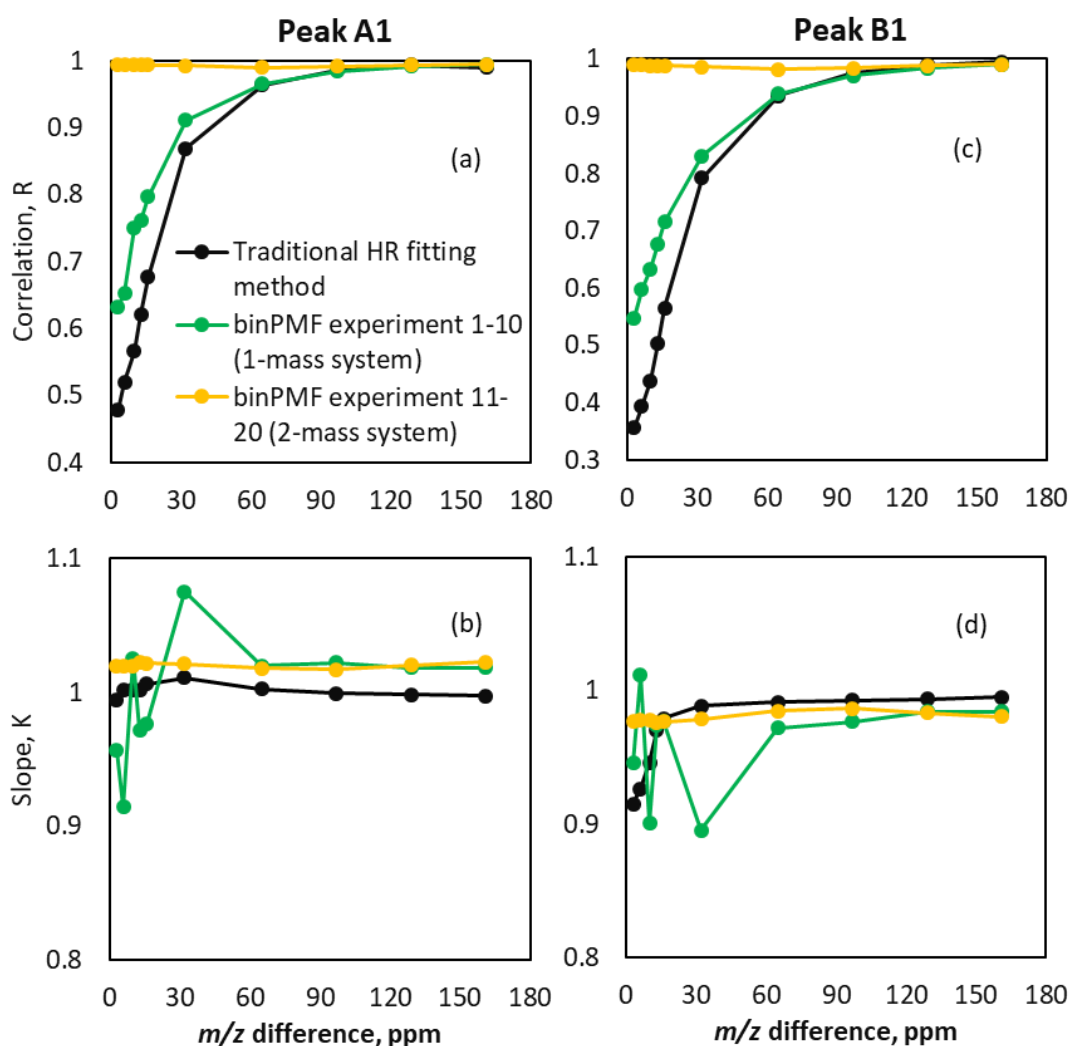
612 Figure 4. Peak separation results by a traditional HR fitting method (dashed lines) and binPMF (solid  
 613 lines), at the 79<sup>th</sup> time point (panels a-d) and at the 21<sup>st</sup> time point (e-h) for experiment numbers 1 (a,  
 614 e), 5 (b, f), 10 (c, g), and 20 (d, h). The signal intensity ratio of peaks A1 and B1 were about 1:1 and  
 615 1:6, respectively, at the 79<sup>th</sup> and the 21<sup>st</sup> time points. Panels a-c and e-g are for the one-mass system,  
 616 while panels d and h are for the two-mass system.





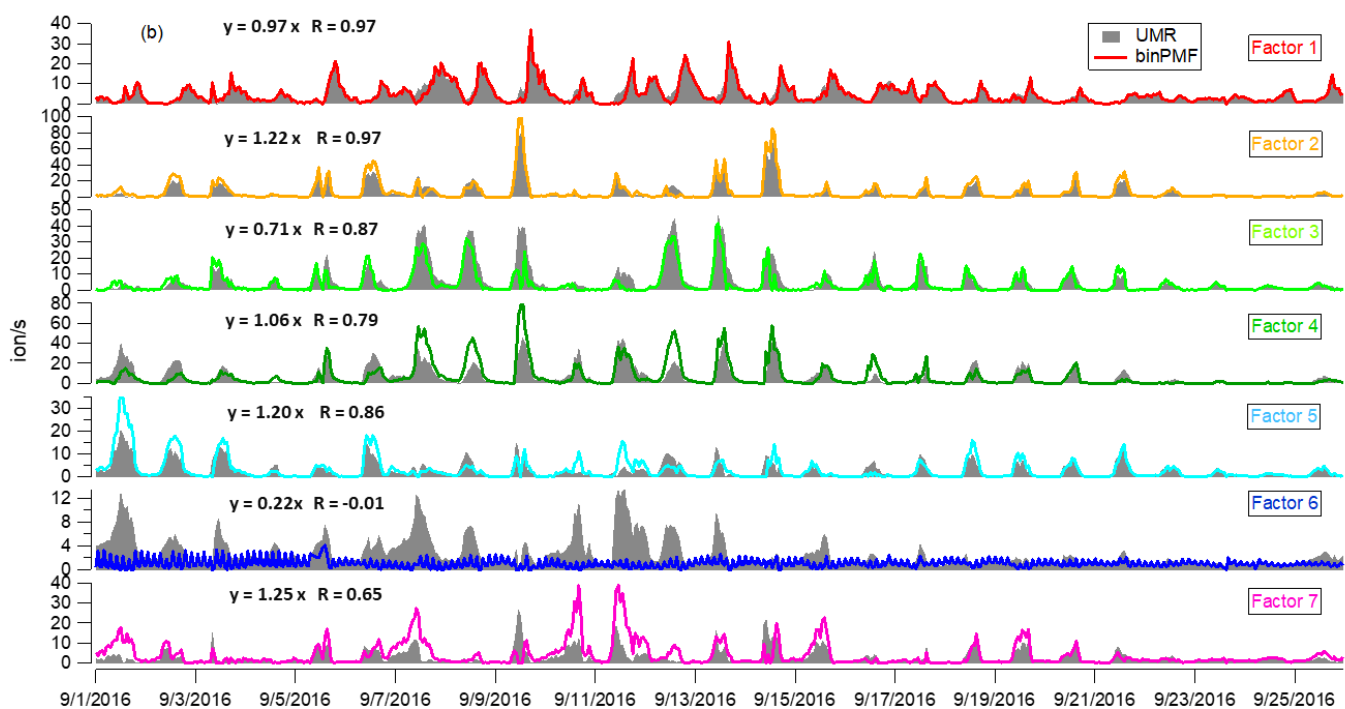
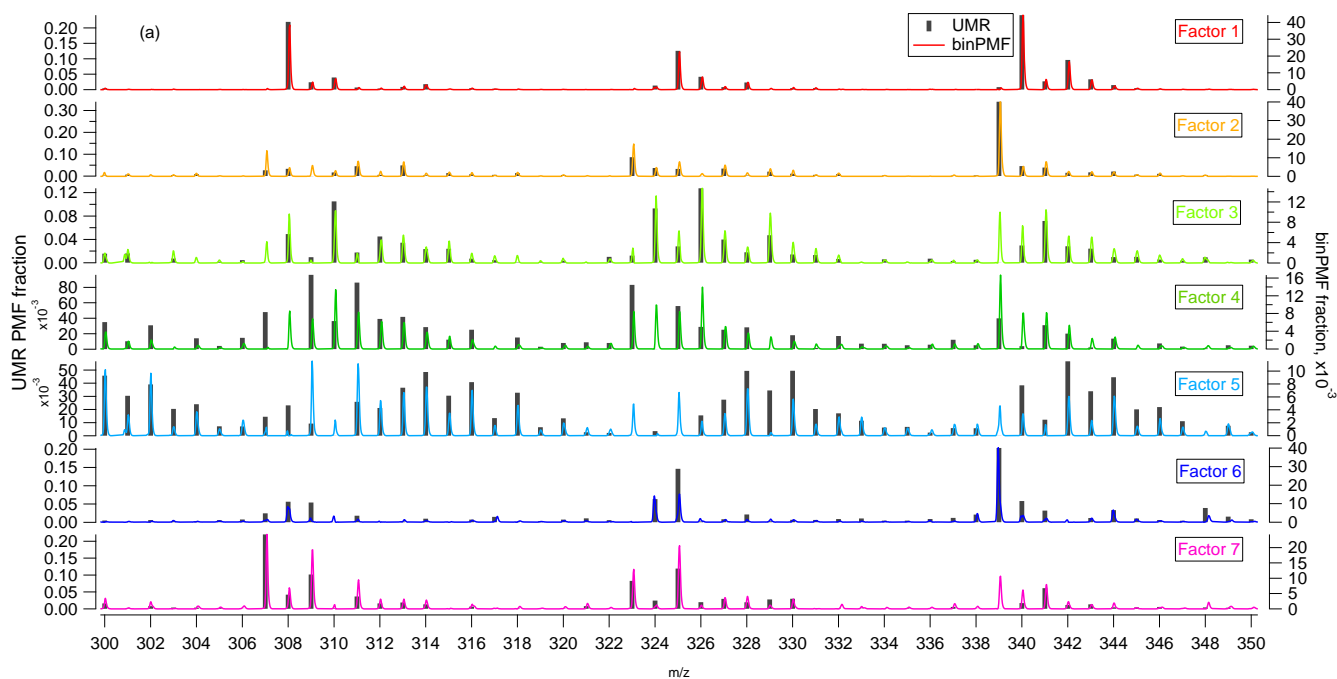
617

618 Figure 5. Characteristics of peaks fitted to binPMF factors. Panel a and b show results for peak A1,  
 619 and c and d for peak B1. In panels a and c, the mass accuracy of peaks resolved by binPMF are  
 620 compared to the original data. Panels b and d depict the resolution of the two fitted peaks. The original  
 621 resolution of the input data was 5000 Th/Th.



622

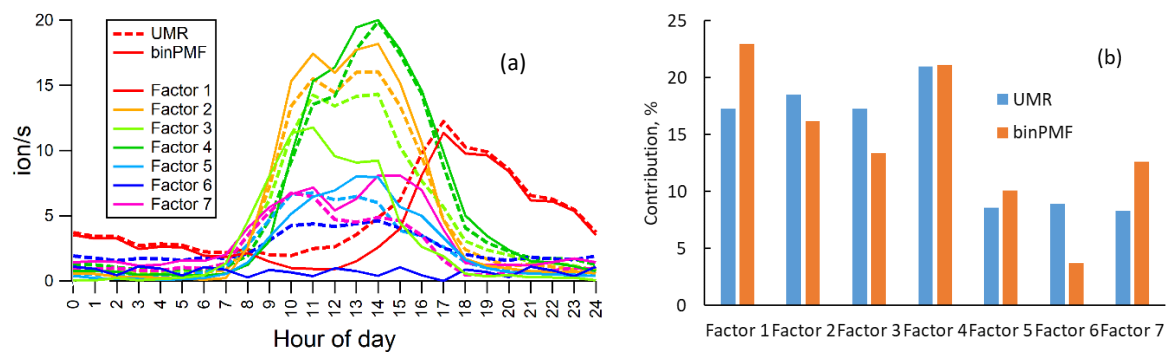
623 Figure 6. Comparison of time series of binPMF and HR fitting. Panel a and b show results for peak  
 624 A1, and c and d for peak B1. Correlation of time series (panels a and c) retrieved by binPMF (green  
 625 lines for experiments 1-10, yellow for 11-20) and traditional HR fitting (black lines) compared to  
 626 original input data. Panels b and d depict the slope K of the linear fit  $y = k \times x$ , where y is the signal  
 627 retrieved from the synthetic data by either binPMF or the HR fitting, and x is the original input signals.



628

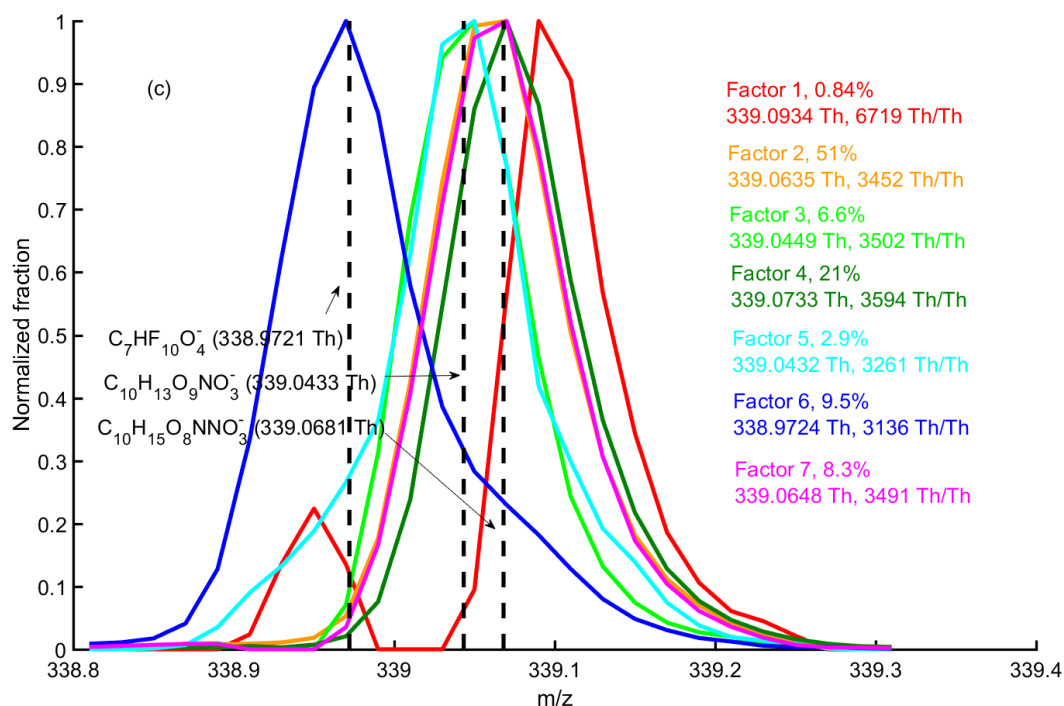
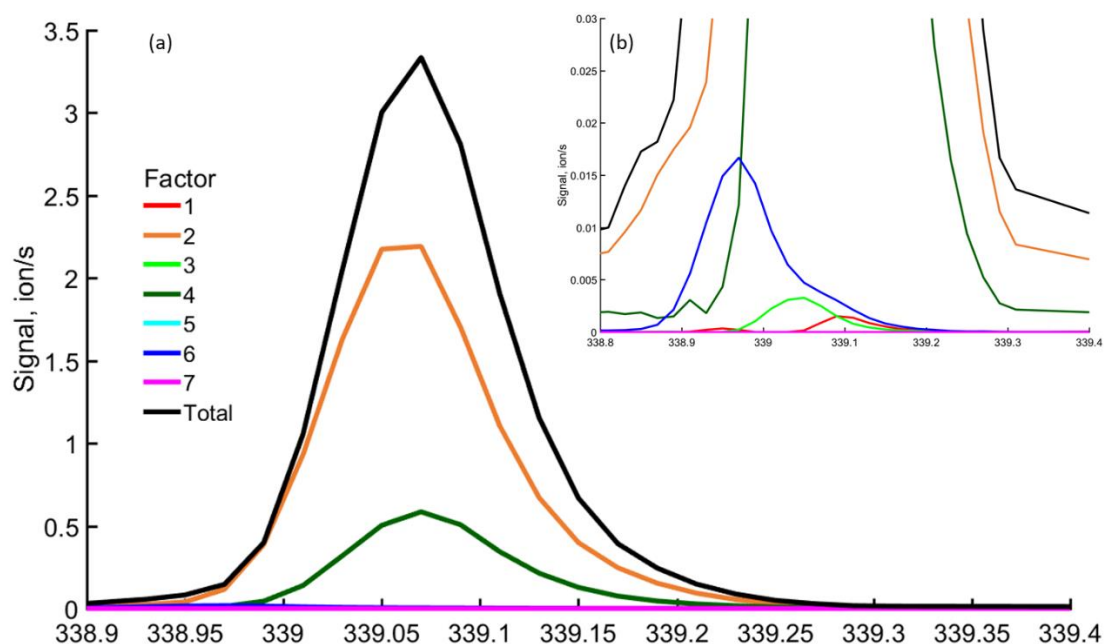
629

630 Figure 7. Comparison of binPMF and UMR-PMF for factor mass spectral profiles (panel a) and time  
 631 series (panel b). The equations in each panel describe how signals from binPMF (y) compare with  
 632 the UMR-PMF solution (x). R is the correlation coefficient between the time series.



633

634 Figure 8. Comparison of binPMF and UMR-PMF for (a) diurnal trend and (b) factor contribution



635

636

637 Figure 9. binPMF factor profiles at  $m/z$  339 Th at 12:00 on September 9<sup>th</sup>. Panels a and b show the  
 638 absolute concentrations of each factor, while in panel c, the factor profiles are normalized to the same  
 639 maximum peak heights. The fitted peak location (Th) and the apparent resolution (Th/Th) for each  
 640 factor is given in panel c,. The contribution of different factors to the integer  $m/z$  339 Th is shown as  
 641 a percentage. Three potential chemical compositions were marked with black vertical dashed lines.

## 642    **References**

- 643    Allan, J. D., Jimenez, J. L., Williams, P. I., Alfarra, M. R., Bower, K. N., Jayne, J. T., Coe, H., and Worsnop,  
644    D. R.: Quantitative sampling using an Aerodyne aerosol mass spectrometer I. Techniques of data  
645    interpretation and error analysis, *Journal of Geophysical Research: Atmospheres*, 108, 2003.
- 646    Bertram, T. H., Kimmel, J. R., Crisp, T. A., Ryder, O. S., Yatavelli, R. L. N., Thornton, J. A., Cubison, M. J.,  
647    Gonin, M., and Worsnop, D. R.: A field-deployable, chemical ionization time-of-flight mass spectrometer,  
648    *Atmos. Meas. Tech.*, 4, 1471-1479, 10.5194/amt-4-1471-2011, 2011.
- 649    Brown, S. G., Eberly, S., Paatero, P., and Norris, G. A.: Methods for estimating uncertainty in PMF  
650    solutions: Examples with ambient air and water quality data and guidance on reporting PMF results, *Science*  
651    *of The Total Environment*, 518-519, 626-635, <https://doi.org/10.1016/j.scitotenv.2015.01.022>, 2015.
- 652    Canagaratna, M., Jayne, J., Jimenez, J., Allan, J., Alfarra, M., Zhang, Q., Onasch, T., Drewnick, F., Coe, H.,  
653    and Middlebrook, A.: Chemical and microphysical characterization of ambient aerosols with the aerodyne  
654    aerosol mass spectrometer, *Mass Spectrometry Reviews*, 26, 185-222, 2007.
- 655    Canonaco, F., Crippa, M., Slowik, J., Baltensperger, U., and Prévôt, A.: SoFi, an IGOR-based interface for  
656    the efficient use of the generalized multilinear engine (ME-2) for the source apportionment: ME-2  
657    application to aerosol mass spectrometer data, *Atmospheric Measurement Techniques*, 6, 3649, 2013.
- 658    Corbin, J. C., Othman, A., Allan, J. D., Worsnop, D. R., Haskins, J. D., Sierau, B., Lohmann, U., and  
659    Mensah, A. A.: Peak-fitting and integration imprecision in the Aerodyne aerosol mass spectrometer: effects  
660    of mass accuracy on location-constrained fits, *Atmos. Meas. Tech.*, 8, 4615-4636, 10.5194/amt-8-4615-2015,  
661    2015.
- 662    Craven, J. S., Yee, L. D., Ng, N. L., Canagaratna, M. R., Loza, C. L., Schilling, K. A., Yatavelli, R. L. N.,  
663    Thornton, J. A., Ziemann, P. J., Flagan, R. C., and Seinfeld, J. H.: Analysis of secondary organic aerosol  
664    formation and aging using positive matrix factorization of high-resolution aerosol mass spectra: application  
665    to the dodecane low-NO<sub>x</sub> system, *Atmos. Chem. Phys.*, 12, 11795-11817, 10.5194/acp-12-  
666    11795-2012, 2012.
- 667    Cubison, M. J., and Jimenez, J. L.: Statistical precision of the intensities retrieved from constrained fitting of  
668    overlapping peaks in high-resolution mass spectra, *Atmos. Meas. Tech.*, 8, 2333-2345, 10.5194/amt-8-2333-  
669    2015, 2015.
- 670    Ehn, M., Kleist, E., Junninen, H., Petäjä, T., Lönn, G., Schobesberger, S., Dal Maso, M., Trimborn, A.,  
671    Kulmala, M., Worsnop, D. R., Wahner, A., Wildt, J., and Mentel, T. F.: Gas phase formation of extremely  
672    oxidized pinene reaction products in chamber and ambient air, *Atmos. Chem. Phys.*, 12, 5113-5127,  
673    10.5194/acp-12-5113-2012, 2012.
- 674    Ehn, M., Thornton, J. A., Kleist, E., Sipila, M., Junninen, H., Pullinen, I., Springer, M., Rubach, F.,  
675    Tillmann, R., Lee, B., Lopez-Hilfiker, F., Andres, S., Acir, I.-H., Rissanen, M., Jokinen, T., Schobesberger,  
676    S., Kangasluoma, J., Kontkanen, J., Nieminen, T., Kurten, T., Nielsen, L. B., Jorgensen, S., Kjaergaard, H.  
677    G., Canagaratna, M., Dal Maso, M., Berndt, T., Petaja, T., Wahner, A., Kerminen, V.-M., Kulmala, M.,  
678    Worsnop, D. R., Wildt, J., and Mentel, T. F.: A large source of low-volatility secondary organic aerosol,  
679    *Nature*, 506, 476-+, 10.1038/nature13032, 2014.
- 680    Guenther, A., Hewitt, C. N., Erickson, D., Fall, R., Geron, C., Graedel, T., Harley, P., Klinger, L., Lerdau,  
681    M., McKay, W. A., Pierce, T., Scholes, B., Steinbrecher, R., Tallamraju, R., Taylor, J., and Zimmerman, P.:  
682    A GLOBAL-MODEL OF NATURAL VOLATILE ORGANIC-COMPOUND EMISSIONS, *Journal of*  
683    *Geophysical Research-Atmospheres*, 100, 8873-8892, 10.1029/94jd02950, 1995.
- 684    Hakola, H., Tarvainen, V., Bäck, J., Ranta, H., Bonn, B., Rinne, J., and Kulmala, M.: Seasonal variation of  
685    mono- and sesquiterpene emission rates of Scots pine, *Biogeosciences*, 3, 93-101, 10.5194/bg-3-93-2006,  
686    2006.
- 687    Hari, P., and Kulmala, M.: Station for Measuring Ecosystem–Atmosphere Relations (SMEAR II), *Boreal*  
688    *Environment Research*, 10, 315-322, 2005.
- 689    Heinritzi, M., Simon, M., Steiner, G., Wagner, A. C., Kürten, A., Hansel, A., and Curtius, J.:  
690    Characterization of the mass-dependent transmission efficiency of a CIMS, *Atmos. Meas. Tech.*, 9, 1449-  
691    1460, 10.5194/amt-9-1449-2016, 2016.
- 692    Henry, R. C.: Current factor analysis receptor models are ill-posed, *Atmospheric Environment* (1967), 21,  
693    1815-1820, [https://doi.org/10.1016/0004-6981\(87\)90122-3](https://doi.org/10.1016/0004-6981(87)90122-3), 1987.

694 Huang, S., Rahn, K. A., and Arimoto, R.: Testing and optimizing two factor-analysis techniques on aerosol  
695 at Narragansett, Rhode Island, *Atmospheric Environment*, 33, 2169-2185, [https://doi.org/10.1016/S1352-](https://doi.org/10.1016/S1352-2310(98)00324-0)  
696 2310(98)00324-0, 1999.

697 Huey, L. G.: Measurement of trace atmospheric species by chemical ionization mass spectrometry:  
698 Speciation of reactive nitrogen and future directions, *Mass Spectrometry Reviews*, 26, 166-184,  
699 10.1002/mas.20118, 2007.

700 Jimenez, J. L., Canagaratna, M. R., Donahue, N. M., Prevot, A. S. H., Zhang, Q., Kroll, J. H., DeCarlo, P. F.,  
701 Allan, J. D., Coe, H., Ng, N. L., Aiken, A. C., Docherty, K. S., Ulbrich, I. M., Grieshop, A. P., Robinson, A.  
702 L., Duplissy, J., Smith, J. D., Wilson, K. R., Lanz, V. A., Hueglin, C., Sun, Y. L., Tian, J., Laaksonen, A.,  
703 Raatikainen, T., Rautiainen, J., Vaattovaara, P., Ehn, M., Kulmala, M., Tomlinson, J. M., Collins, D. R.,  
704 Cubison, M. J., Dunlea, J., Huffman, J. A., Onasch, T. B., Alfarra, M. R., Williams, P. I., Bower, K., Kondo,  
705 Y., Schneider, J., Drewnick, F., Borrmann, S., Weimer, S., Demerjian, K., Salcedo, D., Cottrell, L., Griffin,  
706 R., Takami, A., Miyoshi, T., Hatakeyama, S., Shimono, A., Sun, J. Y., Zhang, Y. M., Dzepina, K., Kimmel,  
707 J. R., Sueper, D., Jayne, J. T., Herndon, S. C., Trimborn, A. M., Williams, L. R., Wood, E. C., Middlebrook,  
708 A. M., Kolb, C. E., Baltensperger, U., and Worsnop, D. R.: Evolution of Organic Aerosols in the  
709 Atmosphere, *Science*, 326, 1525-1529, 10.1126/science.1180353, 2009.

710 Jokinen, T., Sipilä, M., Junninen, H., Ehn, M., Lönn, G., Hakala, J., Petäjä, T., Mauldin Iii, R. L., Kulmala,  
711 M., and Worsnop, D. R.: Atmospheric sulphuric acid and neutral cluster measurements using CI-API-TOF,  
712 *Atmos. Chem. Phys.*, 12, 4117-4125, 10.5194/acp-12-4117-2012, 2012.

713 Jokinen, T., Berndt, T., Makkonen, R., Kerminen, V.-M., Junninen, H., Paasonen, P., Stratmann, F.,  
714 Herrmann, H., Guenther, A. B., Worsnop, D. R., Kulmala, M., Ehn, M., and Sipila, M.: Production of  
715 extremely low volatile organic compounds from biogenic emissions: Measured yields and atmospheric  
716 implications, *Proceedings of the National Academy of Sciences of the United States of America*, 112, 7123-  
717 7128, 10.1073/pnas.1423977112, 2015.

718 Junninen, H., Ehn, M., Petäjä, T., Luosujärvi, L., Kotiaho, T., Kostianinen, R., Rohner, U., Gonin, M., Fuhrer,  
719 K., Kulmala, M., and Worsnop, D. R.: A high-resolution mass spectrometer to measure atmospheric ion  
720 composition, *Atmos. Meas. Tech.*, 3, 1039-1053, 10.5194/amt-3-1039-2010, 2010.

721 Kirkby, J., Duplissy, J., Sengupta, K., Frege, C., Gordon, H., Williamson, C., Heinritzi, M., Simon, M., Yan,  
722 C., Almeida, J., Troestl, J., Nieminen, T., Ortega, I. K., Wagner, R., Adamov, A., Amorim, A., Bernhammer,  
723 A.-K., Bianchi, F., Breitenlechner, M., Brilke, S., Chen, X., Craven, J., Dias, A., Ehrhart, S., Flagan, R. C.,  
724 Franchin, A., Fuchs, C., Guida, R., Hakala, J., Hoyle, C. R., Jokinen, T., Junninen, H., Kangasluoma, J.,  
725 Kim, J., Krapf, M., Kuerten, A., Laaksonen, A., Lehtipalo, K., Makhmutov, V., Mathot, S., Molteni, U.,  
726 Onnela, A., Peraekylae, O., Piel, F., Petaejae, T., Praplan, A. P., Pringle, K., Rap, A., Richards, N. A. D.,  
727 Riipinen, I., Rissanen, M. P., Rondo, L., Sarnela, N., Schobesberger, S., Scott, C. E., Seinfeld, J. H., Sipilae,  
728 M., Steiner, G., Stozhkov, Y., Stratmann, F., Tome, A., Virtanen, A., Vogel, A. L., Wagner, A. C., Wagner,  
729 P. E., Weingartner, E., Wimmer, D., Winkler, P. M., Ye, P., Zhang, X., Hansel, A., Dommen, J., Donahue,  
730 N. M., Worsnop, D. R., Baltensperger, U., Kulmala, M., Carslaw, K. S., and Curtius, J.: Ion-induced  
731 nucleation of pure biogenic particles, *Nature*, 533, 521-+, 10.1038/nature17953, 2016.

732 Kroll, J. H., Donahue, N. M., Jimenez, J. L., Kessler, S. H., Canagaratna, M. R., Wilson, K. R., Altieri, K. E.,  
733 Mazzoleni, L. R., Wozniak, A. S., and Bluhm, H.: Carbon oxidation state as a metric for describing the  
734 chemistry of atmospheric organic aerosol, *Nature Chemistry*, 3, 133, 2011.

735 Lanz, V. A., Alfarra, M. R., Baltensperger, U., Buchmann, B., Hueglin, C., Szidat, S., Wehrli, M. N.,  
736 Wacker, L., Weimer, S., Caseiro, A., Puxbaum, H., and Prevot, A. S. H.: Source Attribution of Submicron  
737 Organic Aerosols during Wintertime Inversions by Advanced Factor Analysis of Aerosol Mass Spectra,  
738 *Environmental Science & Technology*, 42, 214-220, 10.1021/es0707207, 2008.

739 Lee, B. H., Lopez-Hilfiker, F. D., Mohr, C., Kurtén, T., Worsnop, D. R., and Thornton, J. A.: An Iodide-  
740 Adduct High-Resolution Time-of-Flight Chemical-Ionization Mass Spectrometer: Application to  
741 Atmospheric Inorganic and Organic Compounds, *Environmental Science & Technology*, 48, 6309-6317,  
742 10.1021/es500362a, 2014.

743 Massoli, P., Stark, H., Canagaratna, M. R., Krechmer, J. E., Xu, L., Ng, N. L., Mauldin, R. L., Yan, C.,  
744 Kimmel, J., Misztal, P. K., Jimenez, J. L., Jayne, J. T., and Worsnop, D. R.: Ambient Measurements of  
745 Highly Oxidized Gas-Phase Molecules during the Southern Oxidant and Aerosol Study (SOAS) 2013, *ACS*  
746 *Earth and Space Chemistry*, 10.1021/acsearthspacechem.8b00028, 2018.

747 Paatero, P., and Tapper, U.: Positive matrix factorization: A non-negative factor model with optimal  
748 utilization of error estimates of data values, *Environmetrics*, 5, 111-126, 1994.

Paatero, P.: Least squares formulation of robust non-negative factor analysis, *Chemometrics and Intelligent Laboratory Systems*, 37, 23-35, [https://doi.org/10.1016/S0169-7439\(96\)00044-5](https://doi.org/10.1016/S0169-7439(96)00044-5), 1997.

Paatero, P.: The Multilinear Engine—A Table-Driven, Least Squares Program for Solving Multilinear Problems, Including the n-Way Parallel Factor Analysis Model, *Journal of Computational and Graphical Statistics*, 8, 854-888, 10.1080/10618600.1999.10474853, 1999.

Paatero, P., Hopke, P. K., Song, X.-H., and Ramadan, Z.: Understanding and controlling rotations in factor analytic models, *Chemometrics and Intelligent Laboratory Systems*, 60, 253-264, [https://doi.org/10.1016/S0169-7439\(01\)00200-3](https://doi.org/10.1016/S0169-7439(01)00200-3), 2002.

Paatero, P., and Hopke, P. K.: Discarding or downweighting high-noise variables in factor analytic models, *Analytica Chimica Acta*, 490, 277-289, [https://doi.org/10.1016/S0003-2670\(02\)01643-4](https://doi.org/10.1016/S0003-2670(02)01643-4), 2003.

Paatero, P., Eberly, S., Brown, S. G., and Norris, G. A.: Methods for estimating uncertainty in factor analytic solutions, *Atmos. Meas. Tech.*, 7, 781-797, doi:10.5194/amt-7-781-2014, 2014.

Polissar, A. V., Hopke, P. K., Paatero, P., Malm, W. C., and Sisler, J. F.: Atmospheric aerosol over Alaska: 2. Elemental composition and sources, *Journal of Geophysical Research: Atmospheres*, 103, 19045-19057, 1998.

Pope III, C. A., Ezzati, M., and Dockery, D. W.: Fine-particulate air pollution and life expectancy in the United States, *New England Journal of Medicine*, 360, 376-386, 2009.

Schauer, J. J., Rogge, W. F., Hildemann, L. M., Mazurek, M. A., Cass, G. R., and Simoneit, B. R.: Source apportionment of airborne particulate matter using organic compounds as tracers, *Atmospheric Environment*, 30, 3837-3855, 1996.

Shiraiwa, M., Ueda, K., Pozzer, A., Lammel, G., Kampf, C. J., Fushimi, A., Enami, S., Arangio, A. M., Fröhlich-Nowoisky, J., Fujitani, Y., Furuyama, A., Lakey, P. S. J., Lelieveld, J., Lucas, K., Morino, Y., Pöschl, U., Takahama, S., Takami, A., Tong, H., Weber, B., Yoshino, A., and Sato, K.: Aerosol Health Effects from Molecular to Global Scales, *Environmental Science & Technology*, 51, 13545-13567, 10.1021/acs.est.7b04417, 2017.

Song, Y., Shao, M., Liu, Y., Lu, S., Kuster, W., Goldan, P., and Xie, S.: Source apportionment of ambient volatile organic compounds in Beijing, *Environmental science & technology*, 41, 4348-4353, 2007.

Stark, H., Yatavelli, R. L. N., Thompson, S. L., Kimmel, J. R., Cubison, M. J., Chhabra, P. S., Canagaratna, M. R., Jayne, J. T., Worsnop, D. R., and Jimenez, J. L.: Methods to extract molecular and bulk chemical information from series of complex mass spectra with limited mass resolution, *International Journal of Mass Spectrometry*, 389, 26-38, <https://doi.org/10.1016/j.ijms.2015.08.011>, 2015.

Stocker, T., Qin, D., Plattner, G., Tignor, M., Allen, S., Boschung, J., Nauels, A., Xia, Y., Bex, V., and Midgley, P.: IPCC, 2013: Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change, 1535 pp, in, Cambridge Univ. Press, Cambridge, UK, and New York, 2013.

Sun, Y.-L., Zhang, Q., Schwab, J., Demerjian, K., Chen, W.-N., Bae, M.-S., Hung, H.-M., Hogrefe, O., Frank, B., and Rattigan, O.: Characterization of the sources and processes of organic and inorganic aerosols in New York city with a high-resolution time-of-flight aerosol mass spectrometer, *Atmospheric Chemistry and Physics*, 11, 1581-1602, 2011.

Ulbrich, I. M., Canagaratna, M. R., Zhang, Q., Worsnop, D. R., and Jimenez, J. L.: Interpretation of organic components from Positive Matrix Factorization of aerosol mass spectrometric data, *Atmos. Chem. Phys.*, 9, 2891-2918, 10.5194/acp-9-2891-2009, 2009.

Wei, W., Wang, S., Chatani, S., Klimont, Z., Cofala, J., and Hao, J.: Emission and speciation of non-methane volatile organic compounds from anthropogenic sources in China, *Atmospheric Environment*, 42, 4976-4988, <https://doi.org/10.1016/j.atmosenv.2008.02.044>, 2008.

Yan, C., Nie, W., Aijala, M., Rissanen, M. P., Canagaratna, M. R., Massoli, P., Junninen, H., Jokinen, T., Sarnela, N., Hame, S. A. K., Schobesberger, S., Canonaco, F., Yao, L., Prevot, A. S. H., Petaja, T., Kulmala, M., Sipila, M., Worsnop, D. R., and Ehn, M.: Source characterization of highly oxidized multifunctional compounds in a boreal forest environment using positive matrix factorization, *Atmospheric Chemistry and Physics*, 16, 12715-12731, 10.5194/acp-16-12715-2016, 2016.

Zha, Q., Yan, C., Junninen, H., Riva, M., Sarnela, N., Aalto, J., Quéléver, L., Schallhart, S., Dada, L., Heikkinen, L., Peräkylä, O., Zou, J., Rose, C., Wang, Y., Mammarella, I., Katul, G., Vesala, T., Worsnop, D. R., Kulmala, M., Petäjä, T., Bianchi, F., and Ehn, M.: Vertical characterization of highly oxygenated molecules (HOMs) below and above a boreal forest canopy, *Atmos. Chem. Phys.*, 18, 17437-17450, 10.5194/acp-18-17437-2018, 2018.



804 Zhang, Q., Jimenez, J. L., Canagaratna, M., Allan, J., Coe, H., Ulbrich, I., Alfarra, M., Takami, A.,  
805 Middlebrook, A., and Sun, Y.: Ubiquity and dominance of oxygenated species in organic aerosols in  
806 anthropogenically-influenced Northern Hemisphere midlatitudes, *Geophysical Research Letters*, 34, 2007.  
807 Zhang, Q., Jimenez, J. L., Canagaratna, M. R., Ulbrich, I. M., Ng, N. L., Worsnop, D. R., and Sun, Y.:  
808 Understanding atmospheric organic aerosols via factor analysis of aerosol mass spectrometry: a review,  
809 *Analytical and Bioanalytical Chemistry*, 401, 3045-3067, 10.1007/s00216-011-5355-y, 2011.  
810 Zhang, Y., Lin, Y., Cai, J., Liu, Y., Hong, L., Qin, M., Zhao, Y., Ma, J., Wang, X., and Zhu, T.: Atmospheric  
811 PAHs in North China: spatial distribution and sources, *Science of the Total Environment*, 565, 994-1000,  
812 2016.  
813 Zhang, Y., Cai, J., Wang, S., He, K., and Zheng, M.: Review of receptor-based source apportionment  
814 research of fine particulate matter and its challenges in China, *Science of the Total Environment*, 586, 917-  
815 929, 2017.

816

817

818