Measurement of NO\textsubscript{x} and NO\textsubscript{y} with a thermal dissociation cavity ring-down spectrometer (TD-CRDS): instrument characterisation and first deployment

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Abstract. We present a newly constructed, two-channel thermal dissociation cavity ring-down spectrometer (TD-CRDS) for the measurement of NO\textsubscript{x} (NO + NO\textsubscript{2}), NO\textsubscript{y} (NO\textsubscript{3} + HNO\textsubscript{3} + RO\textsubscript{2}NO\textsubscript{2} + N\textsubscript{2}O\textsubscript{5}\textsuperscript{L}} etc.), NO\textsubscript{z} (NO\textsubscript{3}−NO\textsubscript{x}) and particulate nitrate (pNit). NO\textsubscript{x}-containing trace gases are detected as NO\textsubscript{2} by the CRDS at 405 nm following sampling through inlets at ambient temperature (NO\textsubscript{3}) or at 850 °C (NO\textsubscript{x}). In both cases, O\textsubscript{3} was added to the air sample directly upstream of the cavities to convert NO (either ambient or formed in the 850 °C oven) to NO\textsubscript{2}. An activated carbon denuder was used to remove gas-phase components of NO\textsubscript{x} when sampling pNit. Detection limits, defined as the 2σ precision for 1 min averaging, are 40 pptv for both NO\textsubscript{x} and NO\textsubscript{y}. The total measurement uncertainties (at 50% relative humidity, RH) in the NO\textsubscript{x} and NO\textsubscript{y} channels are 11%+10 ppbv and 16%+14 ppbv for NO\textsubscript{2} respectively. Togammograms of various trace gases of the NO\textsubscript{x} family confirm stoichiometric conversion to NO\textsubscript{2} (and/or NO) at the oven temperature and rule out significant interferences from NH\textsubscript{3} detection (< 2%) or radical recombination reactions under ambient conditions. While fulfilling the requirement of high particle transmission (> 80% between 30 and 400 nm) and essentially complete removal of reactive nitrogen under dry conditions (> 99%), the denuder suffered from NO\textsubscript{3} breakthrough and memory effects (i.e. release of stored NO\textsubscript{3}) under humid conditions, which may potentially bias measurements of particle nitrate.

Summertime NO\textsubscript{x} measurements obtained from a ship sailing through the Red Sea, Indian Ocean and Arabian Gulf (NO\textsubscript{3} levels from <20 to 25 ppbv) were in excellent agreement with those taken by a chemiluminescence detector of NO and NO\textsubscript{2}. A data set obtained locally under vastly different conditions (urban location in winter) revealed large diel variations in the NO\textsubscript{x} to NO\textsubscript{y} ratio which could be attributed to the impact of local emissions by road traffic.

1 Introduction

1.1 Atmospheric NO\textsubscript{x} and NO\textsubscript{y}

Total reactive nitrogen NO\textsubscript{y} (= NO\textsubscript{x} + NO\textsubscript{2}) consists of nitrogen oxide, NO; nitrogen dioxide, NO\textsubscript{2} (NO + NO\textsubscript{2} = NO\textsubscript{x}); and their reservoir species, NO\textsubscript{3} (NO\textsubscript{3} + 2N\textsubscript{2}O\textsubscript{5} + HONO + RONO\textsubscript{2} + RO\textsubscript{2}NO\textsubscript{2} + XONO\textsubscript{2} + XNO\textsubscript{2} + pNit), where X is a halogen atom. HCN and NH\textsubscript{3} are generally not considered to be components of NO\textsubscript{y} (Logan, 1983).

The formation of both peroxy nitrates (PNs; RO\textsubscript{2}NO\textsubscript{2}) and alkyl nitrates (ANs; RONO\textsubscript{2}) requires the presence of organic peroxy radicals (RO\textsubscript{2}), which are formed by processes such as the reaction of OH radicals with volatile organic compounds (VOCs) and oxygen (Reaction R1). RO\textsubscript{2} radicals subsequently react with NO\textsubscript{2} or NO to form peroxy nitrates (PNs; RO\textsubscript{2}NO\textsubscript{2}) or alkyl nitrates (ANs; RONO\textsubscript{2}, Re-
actions R2 and R3). Reaction (R3) competes with the formation of an alkoxy radical (RO) and the oxidation of NO to NO2 (Reaction R4), which consumes the dominant fraction of RO2. The branching ratio between these two pathways depends on atmospheric conditions such as pressure and temperature and on the structure and length of the organic backbone (Lightfoot et al., 1992). HNO3 is produced mainly via the reaction of NO2 with OH (Reaction R5).

\[
\begin{align*}
\text{OH} + \text{RH} + \text{O}_2 & \rightarrow \text{RO}_2 + \text{H}_2\text{O} \quad \text{(R1)} \\
\text{RO}_2 + \text{NO}_2 + \text{M} & \rightarrow \text{RO}_2\text{NO}_2 + \text{M} \quad \text{(R2)} \\
\text{RO}_2 + \text{NO} + \text{M} & \rightarrow \text{RONO}_2 + \text{M} \quad \text{(R3)} \\
\text{RO}_2 + \text{NO} & \rightarrow \text{RO} + \text{NO}_2 \quad \text{(R4)} \\
\text{OH} + \text{NO}_2 + \text{M} & \rightarrow \text{HNO}_3 + \text{M} \quad \text{(R5)}
\end{align*}
\]

The lifetimes of peroxynitrates in the low troposphere are mainly governed by the temperature. PNs with an additional acyl group (PANs), such as peroxyacetyl nitrate (PAN), are generally more stable than PNs without an acyl group (e.g. pernitric acid, HO2NO2), which are observed only in cold regions (Slusher et al., 2002). Thus, of the peroxynitrates, only PANs are considered able to act as transportable reservoirs for NOx. At higher altitudes in the troposphere (above ca. 7 km) photolysis becomes the most important loss process for PAN, while the reaction with OH is negligible throughout the troposphere (Talukdar et al., 1995).

The absence of photolysis reactions and low levels of the OH radical at night-time provide alternative pathways for the formation of NOx species. NO2 is oxidised by O3 to produce the nitrate radical NO3, which exists in thermal equilibrium with N2O5 (Reactions R6 and R7). The reaction of NO3 with hydrocarbons represents a night-time source of alkyl nitrates (Reaction R8), and N2O5 can be hydrolysed on aqueous aerosol, resulting in the formation of HNO3 (Reaction R9) and ClNO2 (Reaction R10) if particulate chloride is available (Finlayson-Pitts et al., 1989).

\[
\begin{align*}
\text{NO}_2 + \text{O}_3 & \rightarrow \text{NO}_3 + \text{O}_2 \quad \text{(R6)} \\
\text{NO}_3 + \text{N}_2\text{O}_5 & \rightarrow \text{N}_2\text{O}_5 \quad \text{(R7)} \\
\text{NO}_3 + \text{R}=\text{R}(+\text{O}_2) & \rightarrow \text{RONO}_2 \quad \text{(R8)} \\
\text{N}_2\text{O}_5 + \text{H}_2\text{O} & \rightarrow 2\text{HNO}_3 \quad \text{(R9)} \\
\text{N}_2\text{O}_5 + \text{Cl}^- & \rightarrow \text{ClNO}_2 + \text{NO}_3^- \quad \text{(R10)}
\end{align*}
\]

Nitric acid formation via the reaction of NO2 and OH (Reaction R5), followed by wet or dry deposition of HNO3, is considered to be the dominant daytime loss process for atmospheric NOx (Roberts, 1990), although the reduction of NOx may result in an increasingly important role for organic nitrates (e.g. in the USA; Romer Present et al., 2020). As some organic nitrates have longer lifetimes than HNO3, the atmospheric transport of NOx to remote locations would lead to a more even distribution of NOx, instead of hotspots in polluted regions close to emission sources. Atmospheric removal processes for ANs include oxidation by OH or O3 (which may lead to a loss of the nitrate functionality), deposition to the Earth’s surface, and photolysis. Additionally, partitioning into the aerosol phase is possible for large and multifunctional ANs (Perring et al., 2013). Alkyl nitrates possessing no further functionality (e.g. double bonds or hydroxyl groups) can be unreactive and have long lifetimes (Talukdar et al., 1997). On the global average, RONO2 has a lifetime of close to 3 h (2.6–3 h) with ~30% being lost by hydrolysis (Zare et al., 2018).

The formation of NOx in the lower atmosphere reduces the NOx lifetime, and the partitioning of NOx into NOy and NO2 can provide information about the chemical history of an air mass (Day et al., 2002; Wild et al., 2014). In regions impacted by biogenic emissions, the sources and sinks of ANs account for a large fraction of NOx lost both during the day and night and, thus, control the lifetime of NOx (Romer et al., 2016; Sobanski et al., 2017).

Laboratory experiments have shown that particulate nitrates (pNits) are formed at high yields in the atmospheric degradation of terpenoids in the presence of NOx and play an important role in the formation and growth of secondary organic aerosol (SOA; Ng et al., 2017; Ammann et al., 2019). This has been confirmed in field studies, which provide evidence for the partitioning of organic nitrate to the aerosol phase both during day- and night-time (Rollins et al., 2012; Fry et al., 2013; Palm et al., 2017) with the formation of highly functionalised molecules and large contributions (up to 25%) of particulate organic nitrates to the total aerosol mass (Xu et al., 2015; Lee et al., 2016; Huang et al., 2019).

1.2 Detection of NOx

Methods for the detection of NO and NO2 include chemiluminescence (CLD), differential optical absorption spectroscopy (DOAS), laser-induced fluorescence (LIF) and cavity ring-down spectroscopy (CRDS). A description and intercomparison of these methods is given in Fuchs et al. (2010), and we restrict the following discussion to an outline of the basic principles. The CLD method detects NO by chemiluminescent emission in its reaction with O3; detection of ambient NO2 by CLD follows its catalytic or photolytic conversion to NO. The best CLD devices have detection limits for NO and NO2 in the single-digit parts per trillion by volume (pptv) range (Hosaynali Beygi et al., 2011; Reed et al., 2016; Tadic et al., 2020). Detection of NO2 via LIF involves photoexcitation in its visible absorption band at wavelengths > 400 nm and detection of fluorescent emission at wavelengths > 600 nm, with detection limits of the order of parts per trillion by volume achieved for an integration time of a few seconds (Day et al., 2002; Javed et al., 2019). The structured spectrum of NO2 between ≈ 400 and 600 nm is used to detect light absorption by ambient NO2 by DOAS, using either broadband light sources (long-path DOAS, with a path length of more than a few kilometres) or natural sun-
light (Platt et al., 1979; Leser et al., 2003; Pohler et al., 2010; Merten et al., 2011).

The CRDS detection method for NO$_2$ also utilises its visible absorption spectrum, with high sensitivity being reached by achieving very long path lengths for optical extinction in an optical resonator (see Sect. 2.1). Limits of detection for NO$_2$ with CRDS of < 20 pptv with a 1 s integration time have been reported (Wild et al., 2014). NO can be detected as NO$_2$ following its oxidation by O$_3$ (Reaction R6; Fuchs et al., 2009).

### 1.3 Detection of NO$_y$

The first NO$_x$ measurements were based on the conversion of all reactive nitrogen trace gases (apart from NO) to NO on catalytic metal surfaces of gold at $\sim$ 300–320 °C or of molybdenum oxide (MoO$_x$) at $\sim$ 350–400 °C (Fahey et al., 1985; Williams et al., 1998), with subsequent CLD detection of NO. Au converters were designed to exclude particulate nitrates, whereas MoO$_x$ set-ups aimed at a response towards pNit (Williams et al., 1998). In recent years, the thermal decomposition of NO$_2$ to NO$_2$ has been employed to detect NO$_2$, using inlets held at temperatures high enough (> 650–700 °C) to thermally dissociate the most strongly bound reactive nitrogen trace gas, HNO$_3$, to NO$_2$ (Day et al., 2002; Rosen et al., 2004; Wooldridge et al., 2010; Perring et al., 2013; Wild et al., 2014) and/or using multiple inlets at intermediate temperatures (Paul et al., 2009; Paul and Osthoff, 2010; Sadanaga et al., 2016; Sobanski et al., 2016; Thieser et al., 2016). Following thermal decomposition, the NO$_2$ product can be detected using LIF (Day et al., 2002, 2003; Rosen et al., 2004; Murphy et al., 2006; Wooldridge et al., 2010) or cavity-enhanced absorption spectroscopy (Paul et al., 2009; Wild et al., 2014; Sadanaga et al., 2016; Sobanski et al., 2016; Thieser et al., 2016). These techniques are impacted to various degrees by secondary reactions at high temperatures, including the loss of NO$_2$ via recombination with $\alpha$-carbonyl peroxy radicals or reaction with OES$^3$ atoms (formed by the thermolysis of ambient O$_3$) and the generation of extra NO$_2$ from the oxidation of NO via reactions with peroxy radicals (Day et al., 2002; Sobanski et al., 2016; Thieser et al., 2016; Womack et al., 2017). Measures to reduce potential measurement artefacts and avoid excessive data correction include operation at low pressures (Day et al., 2002; Womack et al., 2017) and the addition of surfaces to scavenge peroxy radicals (Sobanski et al., 2016). Nonetheless, data correction may still be necessary and may involve laboratory characterisation and chemical simulation of the chemical reactions within the heated inlet (Sobanski et al., 2016; Thieser et al., 2016).

In this paper we present a two-channel TD-CRDS instrument for the detection of NO$_x$, NO$_y$, NO$_2$ and pNit that overcomes these limitations. Compared with the set-ups described by Thieser et al. (2016), the following changes were implemented: (1) the addition of O$_3$ for NO$_x$ detection, (2) a higher oven temperature (to detect HNO$_3$) and located directly at the front of the inlet, and (3) the use of a charcoal denuder for separate measurement of pNit and gas-phase NO$_2$. The addition of O$_3$ (after the TD inlet) ensures that we detect NO as well as NO$_2$ and, thus, removes bias caused by factors such as the pyrolysis of O$_3$ and reactions of O($^3$P) which reduce NO$_2$ to NO.

## 2 Experimental

Our TD-CRDS instrument consists of two identically constructed cavities to monitor NO$_2$ at 405 nm which are largely unchanged compared to those described by Thieser et al. (2016). In the present set-up, the two cavities are connected to three different inlets. One cavity monitors NO$_2$ via an inlet at ambient temperature, and the second samples air via one of two heated inlets (one equipped with a denuder, see below), thereby monitoring either NO$_x$ or particle nitrate. A schematic diagram (not to scale) of the instrument is given in Fig. 1.

### 2.1 CRDS operation principals

The optical resonator consists of two mirrors (1 m radius of curvature) with a nominal 0.999965 reflectivity at 405 nm (Advanced Thin Films), which are mounted 70 cm apart. The cavity volumes are defined by Teflon (FEP) coated DURAN glass tubes with an inner diameter of 10 mm. Under normal operating conditions each cavity samples 3.0 L (STP) min$^{-1}$ (slpm; standard litres per minute) of ambient air (where STP refers to 0 °C and 1013 hPa). Additional purge flows (0.14 slpm dry synthetic air) are introduced directly in front of each mirror to prevent surface degradation by atmospheric trace gases. The cavities are operated at pressures of 540 to 580 Torr (1 Torr = 1.333 hPa), resulting in a residence time of $\sim$ 1.2 s. A 405 nm laser light (square-wave modulated at 1666 Hz and a 50 % duty cycle) is provided by a laser diode (LASER COMPONENTS), the emission of which is coupled into an optical fibre with a Y piece for splitting into both cavities. Temperature and current control of the laser diode are achieved by a Thorlabs ITC502 control unit. The laser emission spectrum is monitored continuously by coupling scattered light from one of the cavity mirrors into a 3648 pixel CCD (charge-coupled device) spectrograph (OMT, $\sim$ 0.1 nm resolution).

The intensity of light exiting the cavity is measured with a photomultiplier (Hamamatsu Photonics), with ring-up and ring-down profiles recorded by a digital oscilloscope (PicoScope 3000). NO$_2$ mixing ratios are derived from the decay constant ($k$ or $k_0$) describing the exponential decrease in light intensity after the laser has been switched off:

$$[\text{NO}_2] = \frac{l}{d} \cdot \frac{1}{\sigma_c} (k - k_0),$$

where

$$d = \frac{\int_0^t (1 - \sigma_c) \, dt}{\int_0^t \sigma_c \, dt}.$$
where $c$ is the speed of light, $\sigma$ is the effective absorption cross section of NO$_2$ over the emission spectrum of the laser (Vandaele et al., 2002), and $k$ and $k_0$ are the decay constant with and without NO$_2$ present in the cavity respectively. Thus, $k_0$ is defined by the mirror reflectivity and light scattering by the dry, synthetic air.

The ratio $l/d$ accounts for the difference between the physical length of the cavity ($l$) and the effective optical path length ($d$) in which NO$_2$ is present as well as for dilution effects. $d$ is shorter than $l$ due to the purge flows of zero air in front of the mirrors, and a value of $l/d = 0.98 \pm 0.01$ TSI was determined by adding a constant flow of NO$_2$ and varying the purge-gas flow rate (Schuster et al., 2009; Thieser et al., 2016). $k_0$ is typically determined every 5 min (for 1 min) by overflowing the inlets with zero air from a commercial zero air generator (CAP 180, Fuhr GmbH) attached to a source of compressed ambient air. PTFE filters (47 mm diameter, 2 µm pore size) prevent particles from entering the cavities. The filter’s efficiency, tested with laboratory air containing $1.8 \times 10^3$ particles cm$^{-3}$ and a CPC (TSI 3025 A), was $> 98\%$.

Raw data sets (i.e. ring-down constants) undergo a few basic corrections before further analysis:

1. $k_0$ is interpolated onto the $k$ time grid. The first three data points after switching from sampling to zeroing are discarded in order to enable the stabilisation of the zero signal. The remaining data points of each zero cycle are averaged. Finally, a linear interpolation between the averaged $k_0$ values is performed, allowing for the subtraction of $k_0$ for each individual data point.

2. Depending on conditions of flow, the pressure and the inlet set-up (see Sect. 2.2 and 2.3), changes in flow resistance between the zeroing and sampling periods result in slight changes in the cavity pressure. The resulting change in Rayleigh scattering of the 405 nm light owing to a pressure change of 6.5 Torr was found to be equivalent to a change of ca. 300 pptv in the NO$_2$ mixing ratio,
which is in accordance with earlier experiments using previous versions of this instrumental set-up (Thieser et al., 2016). We have also used an alternative set-up in which the inlet is overflowed with zero air added close to the tip of the inlet (downstream of the oven); this reduces the pressure difference but has the disadvantage that hot air is blown out of the instrument when zeroing, which may interfere with co-located inlets. The addition of zero air upstream of the quartz inlets would remove this problem, but it would also increase the complexity of the inlet and potentially result in the loss of sticky molecules such as HNO₃.

3. A further correction is associated with the difference in the Rayleigh scattering coefficient between dry air (during zeroing) and humid air (whilst taking ambient measurements). This effect was corrected using the H₂O scattering cross sections reported by Thieser et al. (2016), leading, for example, to a correction of 116 pptv at 70 % RH and 20 °C.

2.2 Detection of NOₓ

In order to measure NO (which does not absorb at 405 nm), it is converted to NO₂ by reaction with an excess of ozone (O₃); the O₃ was generated by passing zero air over a Hg Pen-Ray lamp emitting at 185 nm, which was housed inside a glass vessel at ca. 980 Torr pressure. The gas stream containing O₃ is split up equally by critical orifices and directed into two identical reaction volumes made of 88 cm long PFA tubing (1/2 inch outer diameter, residence time 1.05 s). The concentration of O₃ (monitored by a commercial monitor, Model 202, 2B Technologies) was optimised in laboratory experiments in which the efficiency of the conversion of NO to NO₂ was varied by changing the flow of air over the Pen-Ray lamp. The maximum conversion of NO₂ (corresponding to 96 % of the NO in the gas bottle) was observed when the flow over the Pen-Ray lamp was between 60 and 80 cm³ (STP) min⁻¹ (hereafter sccm), which resulted in 19 ppmv O₃ in the reaction volumes. This result could be confirmed by numerical simulation (see Table S1 in the Supplement) of the reactions involved in the formation and loss of NO₂ when NO reacts with O₃. According to the simulation, the maximum conversion of NO to NO₂ during the 1.05 s residence time occurs between ca. 12 and 20 ppmv O₃. The conversion efficiency decreases at higher O₃ concentrations due to the formation of N₂O₅ and NO₃. The results from the experiments to determine the optimum parameters for O₃ generation are summarised in Fig. S1.

For NO₂, the performance of the instrument was first described by Thieser et al. (2016), who reported a measurement uncertainty of 6 % + (20 pptv × RH/100) that was dominated by uncertainty in the effective cross section of NO₂ and the wavelength stability of the laser diode. The NO₃ detection limit of 40 pptv (2σ, 1 min average) for the present instrument (laboratory conditions) was derived from an Allan variance analysis and is worse than that reported by Thieser et al. (2016) (6 pptv at 40 s) due to degradation of the mirror reflectivity. Corrections applied to take humidity and pressure changes into account are discussed in Sect. 2.1. The total uncertainty in NOₓ will depend on the uncertainty in the conversion of both gaseous and particulate nitrate to NO₂ and, thus, depends on the individual components of NOₓ in the air sampled. For purely gaseous NOₓ, the major problem is likely to be related to the loss of sticky molecules at the inlet, and we choose to quote a “worst case” uncertainty of 15 %.

2.3 Thermal dissociation inlets: detection of NOᵧ

The thermal dissociation inlets used to dissociate NOₓ to NO₂ are quartz tubes housed in commercial furnaces (Carbolite, MTF 10/15/130). The oven temperature was regulated with a custom-made electronic module, which enabled spatial separation between the heating elements and insulation and the control electronics. The distance between the heated section of the quartz tubes and the point at which air was taken into the inlet was kept short (ca. 30 cm) in order to minimise losses of trace gases with a high affinity for surfaces, especially HNO₃ (Neuman et al., 1999). Experiments characterising the thermal conversion of various trace gases to NO₂ are described in Sect. 3.1. An electronic, PTFE three-way valve (Neptune Research, Inc., type 648T032, orifice diameter 4 mm) under software control switches between the two heated inlets, one of which is equipped with a denuder. Memory effects on the valves’ surfaces were not observed for NO₂. Bypassing the valve under normal sampling conditions led to a 0.6 Torr pressure change. The sampling flow through both heated inlets is 3.0 slpm. When sampling ambient air via the denuder, we expect to remove all gas-phase NOᵧ components and, thus, measure only particulate nitrate (pNit). Experiments to characterise the transmission of the denuder for particles and various trace gases are presented in Sect. 3.3.

2.4 Active carbon denuder

The active carbon denuder (Dynamic AQS) has a honeycomb structure with 225 quadratic channels (1 mm × 1 mm) of 10 cm length in a cylindrical form (diameter 3 cm) which is housed inside an aluminium casing with 1/2 inch connections (see Fig. 1). The geometric surface area of the denuder is ∼ 45 cm². Assuming a specific surface area for activated carbon of 1000 m² g⁻¹ (Atsuko et al., 1996), we calculate a BET (Brunauer–Emmett–Teller, Brunauer et al., 1938) surface area of the order of 10⁸ cm².

2.5 Chemicals

A stock, liquid sample of PAN in n-tridecane (> 98 %, Alfa Aesar) was synthesised according to the procedures described by Gaffney et al. (1984) and Talukdar et al. (1995).
Samples of lower concentration (as used in the experiments described below) were produced by diluting the original sample with additional n-tridecane. Acetone (> 99 %), isopropyl nitrate (> 98.0 %) and (R)-(+) -limonene (97 %) were obtained from Sigma-Aldrich. An ammonia permeation source (324 ng min⁻¹) was supplied by VICI Metronics. Methanol (> 99.9 %) was acquired from Merck, isoprene (98 %, stabilised) from Acros Organics, and ethanol from Martin and Werner Mundo oHG. Both nitric acid (65 %) and β-pinene (pure) were obtained from Carl Roth. N₂O₅ crystals were synthesised according to Davidson et al. (1978) by reacting NO (5 %) with excess O₃ in a glass reactor. O₃ was produced via electrical discharge through O₂ using a commercial ozone generator (Ozomat Com, Anseros). The crystals were trapped and stored at −78 °C in dry ice and ethanol.

3 Results and discussion

3.1 Trace gas thermograms

The fractional conversion of NO₂ to NO in the TD inlets was investigated in a series of experiments in which constant flows of various NO₂ trace gases were passed through the heated inlet (bypassing the denuder) while the temperature was varied and NO₂ was monitored. NO₂ impurity levels were determined either via the simultaneous operation of the NO₃ channel or via the NO₃ channel mixing ratio at room temperature, before and after heating the inlet. By inserting a thermocouple into the middle part of the heated section under normal sampling conditions, we were able to show that the temperature of the gas was ≈ 80 °C lower than that indicated by the oven’s internal temperature sensor in the 200–300 °C temperature range and about 40 °C lower at a set temperature of 600 °C (see Fig. S3a). We were unable to measure the temperatures of the gas stream at oven temperatures above about 600 °C, and we refer only to the temperature indicated by the internal sensor of the oven throughout the paper.

In the following, we show that the thermograms (plots of fractional dissociation of NO₂ to NO₂ versus temperature) which we measure with this instrument are broader and are shifted in temperature compared with other examples found in the literature, including those from this laboratory. As this instrument is built for the measurement of NO₃ and is not intended for separate measurement of species such as PNs, ANs and HNO₃, overlap of the individual thermograms does not represent a problem.

3.1.1 PAN

A stream of 200 sccm synthetic air was used to elute a constant supply of gaseous PAN from its solution (held at a constant temperature of 0 °C in a glass vessel) into the CRDS inlet. The thermogram is presented in Fig. 2, and the absolute NO₂ mixing ratios are depicted in Fig. S2a. In this experiment the maximum amount of NO₂ observed (at temperature > 400 °C) was 2.2 ppbv. At temperatures < 100 °C, there was no measurable thermal decomposition of PAN to NO₂. Increasing the temperature from 100 °C to 300 °C resulted in a sharp increase in NO₂ which flattened off at temperatures > 380 °C. We conclude that PAN is stoichiometrically converted to NO₂ at temperatures above 400 °C in our oven. The steepest part of the isotherm at ~ 200 °C, i.e. 50 % conversion of PAN to NO₂, is therefore shifted by ca. 80 °C compared with those reported in the literature by Wild et al. (2014), Thieser et al. (2016) and Sobanski et al. (2016). This is a consistent feature of our TD ovens and is related to the short time available for thermal decomposition (see below) and a significantly lower gas temperature than indicated by the oven’s internal temperature sensor.

3.1.2 Isopropyl nitrate

A 10 L stainless-steel canister containing 10.3 ppmv of isopropyl nitrate (iPN) at a pressure of 4 bar N₂ was prepared using a freshly vacuum-distilled liquid sample utilising standard manometric methods. NO₃ impurities were ~ 4.7 ppbv, although we note that diluted iPN stored in stainless-steel canisters for periods of several weeks degrades to form NO₂ and HNO₃.
The thermogram is displayed in Fig. 2, and the absolute concentrations are shown in Fig. S2b. Based on the mixing ratio of iPN in the canister and the dilution flows, 10.7 ppbv represents (101 ± 11) % conversion. The shaded area around the expected iPN mixing ratio in Fig. S2b signifies the uncertainty of this value based on the propagation of the errors during the manometric and dilution procedures (2 % for flow rates, 5 % for pressures measured with digital pressure gauges and 10 % for the last dilution step using the analogue pressure gauge of the canister).

Between 550 and 850 °C, we observe a weak increase in NO2 from 10.7 to 11.2 ppbv, which is likely due to small amounts of HNO3 in the sample. For iPN, the temperature at 50 % conversion is 50 °C higher than those reported by Thieser et al. (2016) and Sobanski et al. (2016). Wild et al. (2014) employed a gaseous mixture of different alkyl nitrates and also observed an initial increase in NO2 (up to 80 % conversion) for temperatures < 300 °C, followed by a slower increase up to 800 °C. The alkyl nitrates thermogram of Wild et al. (2014) has been included into Fig. S2b to illustrate this behaviour and to facilitate direct comparison.

### 3.1.3 HNO3

A custom-made permeation source was used to provide a constant, known flow of HNO3 (with ~8 % NOx impurity) to the TD-CRDS inlet. The permeation source consisted of a length (~1 m) of PFA tubing immersed in 66 % HNO3 solution held at 50 °C through which 100 sccm of dry, zero air was passed. The concentration of HNO3 and, thus, its permeation rate, \((1.62 ± 0.2) \times 10^{-4} \text{ sccm}\), was derived by measuring the optical extinction of HNO3 at 185 nm using the absorption cross section of Dulitz et al. (2018). The uncertainty is related to the uncertainty in the absorption cross section and the reproducibility of the output. The HNO3 thermogram (Figs. 2 and S2c) has a plateau at temperatures above \(\approx 800 °C\). In the plateau region of Fig. 2, the HNO3 mixing ratio measured is \(13.0 ± 0.8 \text{ ppbv}\), which (within combined uncertainties) is in agreement with the expected value \((15.2 ± 1.98 \text{ ppbv})\) calculated from the permeation rate and uncertainty in the dilution factor. We cannot rule out some loss of HNO3 in the tubing connecting the permeation source to the TD-CRDS, although previous studies have shown that irreversible losses are \(~ 5 %\) or less under dry conditions (Neuman et al., 1999). We note that inlet loss of HNO3 is minimised under ambient sampling conditions, as only a short section (~20 cm) of the quartz tubing at ambient temperature is upstream of the heated section in which HNO3 is converted to NO2. Therefore, our observations are in accord with previous studies that found the complete conversion of HNO3 to NO2 in similar set-ups (Day et al., 2002; Di Carlo et al., 2013; Wild et al., 2014; Womack et al., 2017).

### 3.1.4 N2O5

A sample of N2O5 was prepared by flowing 80 sccm of synthetic air over N2O5 crystals, kept at ~70 °C, with further dilution with 20 slpm synthetic air. An 8.5 cm long nylon tube was used to reduce HNO3 impurity. Two distinct dissociation steps can be observed in Fig. 2. The first step, between 50 and 185 °C (in which NO2 increases to 4.2 ppbv, see Fig. S2d), is due to the dissociation of N2O5 to NO2 + NO3. In the second step, in which NO3 is dissociated to NO3 between 450 and 800 °C, the NO2 mixing ratio was 9.2 ppbv. As the amount of N2O5 derived from the first dissociation step was in accord with simultaneous measurements of N2O5 using a further TD-CRDS set-up (Sobanski et al., 2016), we conclude that some HNO3 was present in the sample, which was presumably the result of N2O5 hydrolysis. Our thermogram is similar to that reported by Womack et al. (2017), who also observed two steps – the first with a plateau at \(T > 100 °C\) and the second at \(T > 650 °C\). The shift in temperature (100–150 °C) compared with our results is rationalised in Sect. 3.1.8.

### 3.1.5 HONO

Gaseous HONO was produced by flowing HCl in air (22 ppbv; relative humidity, RH, ca. 50 %), over a bed of continuously stirred sodium nitrate crystals (Wollenhaupt et al., 2000). In our set-up, the thermal dissociation of HONO to NO starts at ~400 °C and reaches a plateau (6.2 ppbv) between ca. 800 and 850 °C (Fig. 2). We did not have access to independent instrumentation to characterise the concentrations of HONO and potential impurities generated using this method. Previous investigations have reported that HONO thermally dissociates between 450 and 650 °C (Perez et al., 2007) and between 200 and 700 °C (Wild et al., 2014). The reasons for such large divergence in the positions and widths of the thermograms may be partially related to the presence of impurities in the HONO samples used, although the details of the samples used to thermally dissociated HONO also play an important role as described in sect. 3.1.8.

### 3.1.6 CINO2

CINO2 was generated by passing Cl2 (33 ppbv in air) over sodium nitrate at room temperature. The thermogram, depicted in Fig. 2, has two steps – one with an apparent plateau at ~500 °C and a second at ~800 °C. The lower temperature plateau in which CINO2 dissociates to NO2 corresponds to that reported previously (Thaler et al., 2011; Sobanski et al., 2016; Thieser et al., 2016). The observation of further NO2 formation at higher temperature is consistent with the observations of Wild et al. (2014). We hypothesise, that the second dissociation step might be associated with the presence of CINO which dissociates to NO and would therefore not have been detected by instruments that monitor NO2 rather
than NO$_3$). Even at temperatures $> 850 \degree C$, we still see an increase in the NO$_2$ signal. However, as CINO is not considered to be an important atmospheric trace gas, this has no repercussions for the deployment of the instrument.

3.1.7 NH$_4$NO$_3$ and NaN$_3$O$_2$ particles

NH$_4$NO$_3$ and NaN$_3$O$_2$ particles were generated from an aqueous solution (ca. 1 g in 500 mL deionised water) using an atomiser (TSI 3076). The particles were dried prior to size selection (DMA, TSI 3080) and were diluted in a total flow of 6 slpm synthetic air which was split between a condensation particle counter (CPC) and the heated TD-CRDS inlet (after further dilution). The relative thermogram for NH$_4$NO$_3$ is displayed in Fig. 2 and, similar to HNO$_3$, displays a plateau region at temperatures above 830 \degree C. The shift in the thermogram when comparing HNO$_3$ and NH$_4$NO$_3$ (which we expect to detect in a two-step process in which NH$_4$NO$_3$ first decomposes to HNO$_3$) may be related to the time required to fully thermally decompose particles (e.g. of 200 nm diameter) containing several million molecules. Particle numbers (in cm$^{-3}$) detected by the CPC were converted to molar mixing ratios via the diameters and densities of the dry particles (1.72 g cm$^{-3}$ for NH$_4$NO$_3$ and 2.26 g cm$^{-3}$ for NaN$_3$O$_2$). The fraction of NH$_4$NO$_3$ detected as NO$_3$ following passage through the oven is illustrated in Fig. S5, which indicates values between $\approx 60\%$ and $120\%$ depending on particle size. The total uncertainty in the concentration was estimated as $41\%$, which includes $10\%$ uncertainty in the particle diameter (based on measured size distributions of latex calibration particles), $20\%$ uncertainty in the particle number (including the error in the multiple charge correction) and $10\%$ uncertainty in the density, due to possible differences between single crystal and bulk density. As the particle mass scales to the third order with the particle diameter, the correction for double-charged particles introduces a large uncertainty in the calculated mixing ratios, with the effect being largest in the size range between 100 and 150 nm, which probably explains the lower NH$_4$NO$_3$ detection efficiencies in this range. We consider the data obtained at 200 nm to be the most reliable and conclude that, similar to other TD instruments (Womack et al., 2017; Garner et al., 2020), our instrument also detects NH$_4$NO$_3$ particles as NO$_2$ with close to $100\%$ efficiency. In contrast, our experiments using NaN$_3$O$_2$ resulted in much smaller NO$_x$ concentrations despite identical experimental conditions in back-to-back experiments and produced detection efficiencies of close to $25\%$. While the inefficient detection of NaN$_3$O$_2$ is consistent with a previous reports suggesting that NaN$_3$O$_2$ would not be detected in TD inlets (Womack et al., 2017), it contrasts strongly with the very recent result of Garner et al. (2020), who observed quantitative conversion of NaN$_3$O$_2$ to NO$_2$ at 600 \degree C. This difference may be related to residence times in the heated section of the inlet.

3.1.8 Summary of thermograms

The thermograms obtained by the present instrument deviate from others reported in the literature, with the temperatures required for $50\%$ dissociation generally being higher by, for example, 80 \degree C for PAN, 50 \degree C for iPN and 150 \degree C for HNO$_3$ respectively (Day et al., 2002; Wild et al., 2014; Sobanski et al., 2016; Thieser et al., 2016; Womack et al., 2017). This lack of agreement with other set-ups is not unexpected, as the degree of dissociation of a trace gas at any temperature depends not only on the temperature but also on the time over which the molecule is exposed to that temperature (Womack et al., 2017). To illustrate this, based on rate coefficients (related to bond-dissociation energies, BDEs) for the thermal dissociation of PAN (Bridge et al., 1991), iPN (Barker et al., 1977), HNO$_3$ (Glinzer and Troe, 1974), N$_2$O$_3$ (Ammann et al., 2019), CINO$_2$ (Baulch et al., 1981) and HONO (Tsang and Herron, 1991), we calculated the theoretical $50\%$ conversion temperature for each molecule as a function of residence time inside the oven (see Fig. S5b). At short residence times the dependence on temperature is very steep (especially for large BDEs) which partially explains the differences between our short heated section inlet and longer ones. However, in practise, we do not know the precise average temperature of the gas at the centre of the oven nor can we characterise the axial and radial gradients in temperature in the quartz tubes; thus, the calculations of fractional dissociation (or complete thermograms) based on bond-dissociation energies are only a rough guide at best. We note that the use of different flows, oven diameters and operational pressures will strongly affect heat transfer from the oven walls to the gas; therefore, reporting the temperature of the external oven wall (as done here and in all reports in the literature) to some extent precludes comparison between different set-ups. The width of the thermograms (i.e. the temperature difference between e.g. $10\%$ and $90\%$ dissociation) will also depend on details of axial and radial temperature gradients in the tubing located within the oven as well as in the downstream section of tubing, which represents a transition regime between oven and room temperature. The impact of temperature gradients inside the quartz tube was explored by calculating the HNO$_3$ thermogram using an Arrhenius expression for its thermal dissociation and the gas residence time within the quartz tube. We initially assumed that all HNO$_3$ molecules experience the same temperature and then compared this to the situation in which $20\%$ of the HNO$_3$ molecules experience an 80 \degree C lower temperature and $20\%$ experience an 80 \degree C higher temperature. The resultant thermograms are displayed in Fig. S5c and indicate that the presence of temperature gradients results in an increase in the width of the thermogram from 250 to 350 \degree C.

The thermograms we report here serve only to determine the temperature needed to ensure the maximum conversion of each trace gas to NO$_2$. This is achieved in the present set-up with a temperature of 850 \degree C. Where possible, we have veri-
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fied that operation at the plateau of the thermogram resulted in quantitative conversion of the traces gases and particles studied, with one exception – NaNO$_3$ particles. We further note that, in an instrument designed only to measure NO$_x$, there is no need to ensure separation (in temperature) of the thermograms for different classes of molecules.

3.1.9 Detection of NH$_3$

As described previously (Wild et al., 2014; Womack et al., 2017) ammonia represents a potential interference in NO$_x$ measurements. In order to quantify this interference, we measured NO$_2$ formation in air containing 131 ppbv NH$_3$ delivered by a calibrated permeation source (VICI METRON-ICS, permeation rate 324 ng min$^{-1}$ at 45°C). The results are summarised in Fig. 2. In NH$_3$–air mixtures, we observe a small NO$_2$ signal, increasing slowly at first and then (from ≈ 700°C) rapidly with temperature; the amount of NO$_2$ observed at 850°C corresponds to a fractional conversion of NH$_3$ to NO$_2$ of 0.006 ± 0.002. This result is in broad agreement with Wild et al. (2014), who found a conversion efficiency of < 0.01 at 700°C. In experiments with NH$_3$ in zero air with relative humidities of 17%, 31% and 53%, we were unable to observe the conversion of NH$_3$ to NO$_2$, which is again consistent with the humidity-related suppression of NO$_2$ formation observed by Wild et al. (2014).

In additional experiments, we investigated the potential influence of ozone on the NH$_3$ to NO$_2$ conversion efficiency in zero air containing O$_3$. The addition of O$_3$ results in a significant increase in NO$_2$ with a linear dependence on the O$_3$ mixing ratio (Fig. 3) and up to 11.4% conversion of NH$_3$ to NO$_2$ at 200 ppbv O$_3$. This was not reduced measurably by the addition of water vapour to the air–O$_3$ mixture. In further experiments, we spiked air with the headspace of various organic liquids (acetone, methanol, ethanol, β-pinene, limonene and isoprene). The gas-phase mixing ratios of the organic trace gases were unknown, but the formation of NO$_2$ was suppressed or completely stopped in each case. A more quantitative investigation was carried out using a known concentration (1 ppbv gas bottle) of isoprene. We found that the addition of 30 ppbv isoprene to zero air (containing 330 ppbv O$_3$) did not significantly reduce the NH$_3$-to-NO$_2$ conversion efficiency under dry conditions, but it decreased it by a factor of 2 when the RH was increased to 50%.

A tentative chemical mechanism, based partially on Womack et al. (2017), to explain the formation of NO$_2$ from NH$_3$ and O$_3$ at high temperatures and the processes that suppress it is given in Reactions (R11) to (R15). In this scheme, the oxidation of NH$_3$ is initiated and propagated by O($^3$P), which is formed from the thermal dissociation of O$_3$ (Peukert et al., 2013). This leads to the formation of NO and HNO (Reaction R13a and R13b), both of which can be oxidised to NO$_2$ (Reactions R14 and R15). Forward and reverse rate coefficients for Reaction (R11) indicate that O$_3$ is converted almost stoichiometrically to O($^3$P) in the

$\approx 10$ ms reaction time in the heated inlet. The rate constants (at 1123 K) for the subsequent reactions involving O($^3$P) are as follows: $k_{12} = 4.4 \times 10^{-13}$ cm$^3$ molec.$^{-1}$ s$^{-1}$, $k_{13a} = 8.3 \times 10^{-12}$ cm$^3$ molec.$^{-1}$ s$^{-1}$ and $k_{13b} = 7.5 \times 10^{-11}$ cm$^3$ molec.$^{-1}$ s$^{-1}$ (Cohen and Westberg, 1991). Reaction (R12) converts 0.3% of the initial NH$_3$ molecules to NH$_2$ within 10 ms (at 100 ppbv O$_3$ and 1123 K).

\[
\begin{align*}
O_3 + M & \rightarrow O(^3P) + O_2 + M & (R11) \\
O(^3P) + NH_3 & \rightarrow NH_2 + OH & (R12) \\
NH_2 + O(^3P) & \rightarrow NO + H_2 & (R13a) \\
& \rightarrow HNO + H & (R13b) \\
HNO + O(^3P)/OH/H & \rightarrow NO + OH/H_2O/H_2 & (R14) \\
NO + O(^3P) + M & \rightarrow NO_2 + M & (R15)
\end{align*}
\]

The experimental results obtained in zero air indicate that reactions involving O($^3$P) from O$_3$ thermolysis can result in the conversion of NH$_3$ to NO and NO$_2$. These results could, however, not be reproduced when adding NH$_3$ to ambient air sampled from outside of the building. In this case, the addition of NH$_3$ (at 50–60 ppbv O$_3$) did not result in a measurable increase in NO$_2$, which was in accord with the observations of Womack et al. (2017). The scavenging of NH$_2$ radicals and O($^3$P) by both volatile organic compounds and H$_2$O provides a likely explanation for this. Womack et al. (2017) also found that the addition of 100 ppbv CO can reduce the conversion of NH$_3$ to NO$_x$.

In summary, our experiments indicate that the conversion of NH$_3$ to NO$_2$ is suppressed in ambient air samples, or in synthetic air with added VOCs and water. The ambient air used in these experiments was from an urban and polluted environment (typical NO$_x$ levels between 10 and 50 ppbv; see

Sect. 4.2). As high levels of atmospheric NH₃ are associated with agricultural activity (Langford et al., 1992; Schlesinger and Hartley, 1992) and are often accompanied by high NOₓ and VOC levels, the NH₃ interference under these conditions is most likely to be small compared with ambient NOₓ levels. Long-term measurements of NH₃ have additionally found a positive correlation between NH₃ concentrations and ambient temperature (Yamamoto et al., 1988; Wang et al., 2015; Yao and Zhang, 2016), with the latter promoting the presence of high levels of biogenic VOCs, such as isoprene (Tingey et al., 1979), which would also help to minimise the NH₃-related interference.

3.2 Bias caused by secondary reactions in the TD ovens

Thermal dissociation techniques coupled to CRD systems for the measurement of organic nitrates suffer bias to different degrees owing to reactions of organic peroxy radicals with NO and NO₂ (Sobanski et al., 2016; Thieser et al., 2016). According to previous studies (Day et al., 2002; Rosen et al., 2004; Thieser et al., 2016), in experiments using iPN at an oven temperature of 450 °C, an overestimation of ANs in the presence of NO is caused by reactions of the initially formed alkoxy radical, C₃H₂O, which results in the formation of both HO₂ and CH₃O₂ (Reactions R16–R21).

\[
\begin{align*}
C_3H_7NO_2 + M & \rightarrow C_3H_7O + NO_2 + M \quad \text{(R16)} \\
C_3H_7O + O & \rightarrow CH_3C(O)CH_3 + HO_2 \quad \text{(R17)} \\
C_3H_7O + M & \rightarrow CH_3 + CH_3CHO + M \quad \text{(R18)} \\
CH_3 + O_2 + M & \rightarrow CH_3O_2 + M \quad \text{(R19)} \\
CH_3O_2 + NO & \rightarrow CH_3O + NO_2 \quad \text{(R20)} \\
HO_2 + NO & \rightarrow HO + NO_2 \quad \text{(R21)} \\
CH_3C(O)O_2NO_2 + M & \rightarrow CH_3C(O)O_2 + NO_2 + M \quad \text{(R22a)} \\
CH_3C(O)O_2 + NO_2 + M & \rightarrow CH_3C(O)O_2NO_2 + M \quad \text{(R22b)} \\
CH_3C(O)O_2 + M & \rightarrow CH_3CO + O_2 + M \quad \text{(R23a)} \\
CH_3C(O)O_2 & \rightarrow CH_2C(O)OOH \quad \text{(R23b)} \\
OH + NO_2 + M & \rightarrow HNO_3 + M \quad \text{(R24)}
\end{align*}
\]

In order to investigate the potential bias in the measurement of ANs under the present experimental conditions, a set of experiments was conducted in which NO (up to 12 ppbv) was added to various amounts of iPN. The NO mixing ratio was determined by modulating the addition of O₃. The results (Fig. 4a) show that NO₂ derived from the thermal decomposition of iPN increases with the amount of NO added and results in overestimation (factor of ~1.6) at 12 ppbv NO, which is consistent with the observations by Thieser et al. (2016). This disappears when 19 ppmv ozone is added in front of the cavity so that NO₂ rather than NO₃ is measured (blue data points). This is readily explained by the compensation of the additional NO₂ formed via reactions of NO with RO₂ by an equal loss in NO, which is only detected when introducing O₃. This is illustrated graphically in Fig. S4.

We also explored the potential for bias caused by the recombination of CH₃C(O)O₂ and NO₂ (Reaction R22b), following the thermal decomposition of PAN (Reaction R22a). Thieser et al. (2016) reported that PAN was underestimated by a factor 0.45 when adding 10 ppbv NO₂ to an air sample containing PAN at 200 °C. This behaviour was not apparent at 450 °C, which is related to the decomposition (Reaction R23a) or isomerisation (Reaction R23b) of CH₃C(O)O₂ at the higher temperature. The results of a similar experiment with our 850 °C inlet are presented in Fig. 4b. In this plot, the measured NO₂ relative to the input PAN is plotted versus the mixing ratio of added NO₂. For PAN concentrations from 1.5 to 2.2 ppbv, no effect was observed for NO₂ concentrations.
of up to 10 ppbv. This is consistent with the reaction scheme presented by Thieser et al. (2016) at 450 °C.

A potential source of bias when measuring HNO₃ includes its reformation via the reaction of OH and NO₂ (Reaction R24). Compared with the RO and RO₂ radicals formed in the thermal dissociation of PN₃s and AN₃s, OH exhibits a higher affinity for surfaces and is likely to be efficiently removed at the oven wall. Day et al. (2002) estimated that wall losses are the dominant OH sink and that the resulting underestimation of HNO₃ would be < 2% for NOₓ levels < 5 ppbv. At our oven temperature, the diffusion coefficient for OH (D_{OH}) can be calculated according to Tang et al. (2014):

\[ D_{OH} (1123 \text{ K}) = D_{OH} (296 \text{ K}) \cdot \left( \frac{296}{T} \right)^{1.75}. \]  

Using an average of the literature values for D_{OH} at room temperature from Ivanov et al. (2007), 165 Torr cm² s⁻¹, and Bertram et al. (2001), 192 Torr cm² s⁻¹, a value of D_{OH} (1123 K) = 1841 Torr cm² s⁻¹ was derived. The maximum rate constant for OH wall loss (assuming laminar flow) can subsequently be approximated according to Zasypkin et al. (1997):

\[ k_{wall} = \frac{D_{OH} \cdot 3.66}{\pi^2 \cdot p}. \]  

With the radius of the oven quartz tube r (0.45 cm) and the pressure p (760 Torr), the maximum value of k_{wall} is 44 s⁻¹. The first-order loss rate coefficient for the reaction of OH with NO₂ is given by k_{(OH+NO₂)} [NO₂], where k_{(OH+NO₂)} is the rate coefficient for the reaction between OH and NO₂ at 1123 K ~ 5 × 10⁻¹⁴ cm³ molec⁻¹ s⁻¹ (Ammann et al., 2019) and [NO₂] = 6.5 × 10¹⁰ molec cm⁻³ (the concentration of 10 ppb NO₂ at the pressure and temperature of the oven). The first-order loss rate of OH via reaction with NO₂ is then 3 × 10⁻³ s⁻¹. Clearly, the efficiency of uptake of OH to the wall would have to be very low in order to reduce the maximum value of 44 s⁻¹ to values that are comparable to the reaction with NO₂, which is very unlikely. We conclude that reformation of HNO₃ via Reaction (R24) will not bias measurements of HNO₃ with the present set-up.

3.3 Denuder characterisation

The efficiency of the removal of NOₓ trace gases and the transmission of submicron particles was determined in a series of experiments, which are described below.

3.3.1 Transmission of ammonium nitrate particles (10–414 nm)

In order to characterise the transmission of the denuder for particles of different diameter, a constant flow of particles was generated by passing 3.3 slpm of nitrogen through an atomiser (TSI 3076) containing an aqueous solution of ammonium nitrate. The flow rate (3.3 slpm) was matched to the typical sampling flow through the denuder. A total of 0.28 slpm of the flow was sampled into a scanning mobility particle sizer (SMPS)/CPC system (TSI 3080 and TSI 3025A) to measure the number density and size distribution (10–414 nm) of the ammonium nitrate particles. The flow was delivered either directly to the SMPS/CPC via straight metal tubing (length 27 cm, inner diameter 0.9 cm), for which we assume 100% particle transmission, or via the denuder. Thus, the ratio of the particle numbers in each size bin represents the size-dependent denuder transmission. As shown in Fig. 5, the transmission of the denuder is > 80% for particles between 30 and 400 nm in diameter. As expected, some diffusive loss is observed for particles < 20 nm in diameter and loss due to impact and/or settling is observed for particles > 300 nm. The particle transmission T as a function of particle diameter D can be represented by the following empirical expression:

\[ T(\%) = \frac{D(\text{nm}) - 5.79}{0.035 + 0.010 \cdot (D - 5.79) + 1.78 \cdot 10^{-6} \cdot (D - 5.79)^2} \]  

The particle transmission through the denuder channels was also calculated using the Particle Loss Calculator (PLC) developed by von der Weiden et al. (2009). The results of this calculation are plotted in Fig. 5. The observed loss of particles smaller than 40 nm is not replicated by the PLC, which was developed for cylindrical piping and not for the square honeycomb shape of the denuder; thus, it does not consider losses due to impact at the finite surface area which the gas and particle flow is exposed to at the entrance to the honeycomb. The PLC does a better job at predicting a reduction in transmission for the largest particles that we measured, indicating a transmission of 74% at 1 μm and 45% at 2 μm. Therefore, in certain environments, nitrate associated with coarse-mode particles represents a potential bias for TD-CRDS measurements of NO₃.

3.3.2 Efficiency of removal of NOₓ trace gases

The efficiency of the removal of trace gases in the denuder under typical flow conditions (3.3 slpm) was investigated for NO, NO₂, PAN, iPN, HONO, N₂O₅, CINO₂ and HNO₃ as representative NOₓ species. The efficiency of removal of each trace gas (generally present at 5–40 ppbv) was determined by measuring its relative concentration when flowing through the denuder (pNit channel) and when bypassing the denuder (NOₓ channel). The results (Fig. 6) indicate that, in dry air, all of these trace gases were removed with an efficiency of close to 100%. However, when the main dilution flow was humidified, RH-dependent breakthrough of NO was observed, with only 60% stripped from the gas phase at an RH close to 100%. HONO was removed with 85% efficiency at an RH of 46%, and CINO₂ was removed...
Figure 5. Transmission of ammonium nitrate particles through the denuder inlet. Relative transmissions are derived by dividing the number size distribution when sampling through the denuder by a size distribution obtained without the denuder. Error bars are based on the standard deviation of three consecutive measurements with and without the denuder. An aerosol flow of 3.3 slpm was directed through the denuder (diameter 3 cm, see Sect. 2.4), and a DMA subsequently sampled 0.3 slpm from the stream exiting the denuder. The plot also includes a fit of the experimental (solid, black line) data and a theoretical transmission distribution computed with the Particle Loss Calculator (PLC).

Figure 6. Removal efficiency of the denuder for various NO\textsubscript{x} trace gases as a function of RH. The inset is identical to the main graph. See Table S2 for the exact values and information on the error determination.

with a 75% efficiency at an RH of 60%. In contrast, humidification only had a marginal effect on the scrubbing efficiency for NO\textsubscript{2}, iPN and HNO\textsubscript{3}, for which an efficiency of \geq 95% was observed. The precise values that the removal efficiencies in Fig. 6 were determined from are listed in Table S2.

In further experiments, we examined the potential for release of NO\textsubscript{y} that had previously been stored in the activated carbon substrate of the denuder. In these experiments, in which either NO\textsubscript{x} or NO\textsubscript{y} was continuously monitored, the denuder was exposed to a flow of 9.5 ppm iPN in dry nitrogen for 90 min during which 2.30 \times 10^{17} molecules of iPN were stripped from the gas phase and deposited onto the denuder. This exposure is equivalent to a month-long exposure to 20 ppb of iPN. The air passing through the denuder was subsequently humidified to an RH of 65%. The results (Fig. 7a) indicate a high initial (11:40–11:50 UTC) rate of release of NO\textsubscript{x} under humid conditions (resulting in a maximum mixing ratio of 2 ppbv at 11:45 UTC). At 11:50 UTC, humidification of the air was stopped and the rate of release of NO\textsubscript{x} dropped gradually towards zero. During a second period, in which the air was again humidified (from 11:50 UTC onwards), NO\textsubscript{x} was released from the denuder, albeit at a lower rate than during the first humidification. From 12:25 UTC onwards, the oven behind the denuder was heated to 850°C so that NO\textsubscript{y} was added. No significant increase in NO\textsubscript{2} was observed, indicating that the trace gas(es) species released from the denuder surface upon humidification are predominantly NO\textsubscript{x}. During this experiment, 2.55 \times 10^{15} molecules of NO\textsubscript{x} desorbed from the denuder, indicating that the major fraction of iPN molecules remained stored on the denuder surface upon humidification.

Similar denuder exposure experiments were performed with HNO\textsubscript{3} and NO\textsubscript{2}. For HNO\textsubscript{3}, no evidence for desorption of NO\textsubscript{x} or NO\textsubscript{y} during exposure to humidified air was observed, whereas NO\textsubscript{2} exhibited a similar behaviour to iPN (Fig. 7b). After loading the denuder with 5 sccm from a 0.831 ppm NO\textsubscript{2} gas bottle for 4.8 d (a total of 7.60 \times 10^{17} molecules were deposited as derived from the flow rate, the exposure time and the gas bottle mixing ratio), the effect of passing humidified air through the denuder was to release NO\textsubscript{x}, which was observed at concentrations up to \approx 39 ppbv. While the relative humidity was kept constant at close to 100%, the NO\textsubscript{x} released decreased over time so that after 30 min, 3.2 ppbv of NO\textsubscript{x} could still be detected. By switching the O\textsubscript{3} source off (at \approx 10:45 UTC), the NO\textsubscript{2} measured went to \approx 0, indicating that predominantly NO was released (and not NO\textsubscript{2}). Integrating these results over time yielded a value of 1.63 \times 10^{16} molecules desorbed NO from the denuder surface in humid air. Qualitatively similar results, i.e. humidity-induced formation and release of NO\textsubscript{x} from the denuder, were observed when the denuder was exposed to variable levels of NO\textsubscript{x} (i.e. up to 20 ppbv) under dry conditions for periods of weeks.

Clearly, adsorption of water molecules onto the denuder surface can initiate and/or catalyse chemical transformation at the surface that convert stored NO\textsubscript{x} into forms which can desorb and be detected as NO\textsubscript{x}. To further our understanding
of the underlying processes that occur upon humidification, a series of experiments were conducted to examine the adsorption of water on the denuder. In these experiments, the denuder was first dried by exposing it to dry air for several hours until the relative humidity of the air exiting the denuder was close to zero. Subsequently, humidified air was passed through the denuder and the RH of air exiting it was continuously monitored. The results of an experiment in which the air was humidified to 68% are shown in Fig. 8a. After 77 min of exposure to this humidity, the RH of the air exiting the denuder acquired a maximum value of 64%. After switching back to dry synthetic air (at 09:37 UTC), ~60 minutes passed before the RH dropped to values close to zero. In this period, the RH did not decrease monotonically, with the rate of change of relative humidity exiting the denuder revealing a number of discrete steps. Figure 8b plots the derivatives (dRH/dt) of the drying phases of a series of experiments in which the initial RH varied between 47% and 75%. A similar pattern emerges for each experiment with the greatest desorption rates occurring at the beginning of the drying phase followed by a minimum in the desorption rate and a second maximum (at ~15% relative humidity). This behaviour is a clear indication that H2O is bound to more than one chemically or physically distinct surface sites on the activated carbon.

By measuring the change in the RH of the air flowing into and out of the denuder, we can derive an equilibrium adsorption isotherm for H2O at the active carbon surface. An example is given in Fig. 9 where it is also compared to a literature isotherm for the adsorption of water vapour on activated carbon fibre (Kim et al., 2008). The data of Kim et al. (2008) have been scaled by matching the number of adsorbed water molecules at an RH of 65.9% to our observed value at an RH of 67.2%. The exposure of carbonaceous surfaces to inert gases at high temperatures (~2000°C) reduces the capacity for water uptake, whereas functionalising the carbon surface with oxygen-containing groups (e.g. from HNO3) enhances the water adsorption capacity (Dubinin et al., 1982; Barton and Koresh, 1983; Liu et al., 2017). Thus, in our experiments, the uptake of gas-phase NOx is expected to generate oxygenated sites on our denuder surface, which, in turn, will influence water uptake and subsequent further trace gas accommodation.

The chemistry leading to the formation of gas-phase NOx from NOy trace gases adsorbed at the denuder surface under humid conditions cannot be elucidated in detail with our experimental set-up. However, a strong humidity dependence in the heterogeneous generation of HONO and NO from NO2 adsorbed on soot particles has been reported (Kalberer et al., 1999; Kleffmann et al., 1999). Formation of HONO from NOx has also been observed on wet aerosol and ground surfaces in field studies (Lammel and Perner, 1988; Notholt et al., 1992). Previous investigations report the adsorption of NO2 on activated carbon at ambient or close to ambient temperatures (30–50°C), followed by its reduction to NO, with the simultaneous oxidation of the carbon surface (Shirahama et al., 2002; Zhang et al., 2008; Gao et al., 2011). These results are consistent with our observation of processes such as the conversion of NO2 to NO at the denuder surface under humid conditions. In our experiments, we observed that NO2 was converted to NO (rather than HONO) at the denuder surface. It is possible that the initial step is the formation of HONO (e.g. by the surface-catalysed hydrolysis of NO2 or iPN) which undergoes further reduction on the surface to NO. The release of HONO and NO has also been observed from soil, after the nitrification of NH3/NO3− or the reduction of NOx (Su et al., 2011; Pilegaard, 2013; Meusel et al., 2018). Oswald et al. (2013) found compara-
Figure 8. (a) RH of humidified zero air after passing through the denuder. The initial RH was determined by bypassing the denuder before and after the experiment. Zero air was humidified by flowing a fraction of the stream through deionised water stored in a glass vessel. The time at which the experiment was conducted is given on the x axis. Until ca. 09:35 UTC, zero air with constant humidity (RH ca. 68 %) was sent through the denuder. Afterwards the denuder was exposed to dry zero air. (b) Derivative of the measured RH during the drying period. The step during the drying phase occurs in a higher RH area when starting the drying from a larger RH value.

Figure 9. Number of adsorbed water molecules onto the denuder surface at equilibrium versus RH. The red line represents (scaled) results from a study on activated carbon fibre (Kim et al., 2008).

3.4 Comparison of the TD-CRDS with existing methods for NO\textsubscript{x} and NO\textsubscript{y} and instruments

In this section, our instrument’s performance is compared to other methods for NO\textsubscript{x} and NO\textsubscript{y} detection. We consider only instruments that measure NO\textsubscript{x} and/or NO\textsubscript{y} and not those that measure individual trace gases from each family.

NO\textsubscript{x} has traditionally been measured by two-channel CLD instruments, which have one channel for NO and one channel for NO\textsubscript{2} respectively: NO is detected by chemiluminescence from its reaction with O\textsubscript{3}, and NO\textsubscript{2} is converted to NO on a molybdenum converter or by photolysis at wavelengths close to 390 nm. Examples of this type of instrument as well as the reported levels of detection (LODs) and overall uncertainties are listed in Table 1. In Table 1, we also list the LOD and uncertainty of a recently developed CRDS set-up which is...
similar in principal of operation to the one described here (Fuchs et al., 2009).

NO\textsubscript{y} has frequently been measured by CLD instruments with Au and/or MoO-coated thermal dissociation inlets that reduce NO\textsubscript{z} to NO, which is detected as described above for CLD-NO\textsubscript{x} instrument. Although such instruments have very low detection limits, they have been shown to be vulnerable to degradation of the NO\textsubscript{y} conversion efficiency and suffer from interferences by HCN, CH\textsubscript{3}CN and NH\textsubscript{3} (Kliner et al., 1997) as well as loss of NO\textsubscript{y} in the inlet (Zenker et al., 1998; Parrish et al., 2004). Table 1 summarises the LODs and uncertainties reported by other NO\textsubscript{y} instruments.

Table 1 indicates that the total uncertainty of the present instrument is comparable to those reported for both NO\textsubscript{x} and NO\textsubscript{y}. However, our present detection limit for both NO\textsubscript{x} and NO\textsubscript{y} is worse than that reported (for NO\textsubscript{2}) for the same instrument in 2016 (Thieser et al., 2016), which is a result of mirror degradation since that study.

4 Application of the instrument in field experiments

4.1 NO\textsubscript{x} intercomparison and pNit measurements during the AQABA campaign

The first deployment of the instrument was during the AQABA (Air Quality and climate change in the Arabian Basin) ship campaign in summer 2017. From 31 July to 2 September, the ship “Kommander Iona” followed a route from southern France via the Mediterranean Sea, the Suez Channel, the Red Sea, the Arabian Sea and the Arabian Gulf to Kuwait and back. The instrument was located in a container in front of the ship, with the inlet ovens located in an aluminium box on the roof of the container. The (unheated) tips of quartz inlet tubes protruded about 15 cm from the side of the aluminium box. Here, we compare the NO\textsubscript{x} and pNit measurements with other measurements of these parameters made during the campaign.

During AQABA, NO\textsubscript{x} levels ranged from a few parts per trillion by volume (maritime background) up to several tens of parts per billion by volume in heavily polluted air masses in shipping lanes or in harbours. In Fig. 10a, we compare NO\textsubscript{x} measured with the TD-CRDS with the results of a chemiluminescence detector (CLD 790 SR, ECO PHYSICS; Tadic et al., 2020), which measured NO and NO\textsubscript{2}. The data points represent 1 min averages for the entire campaign, excluding air masses that were contaminated by the ship’s exhaust. Additionally, periods with very high NO\textsubscript{x} variability were not included, with a data point being discarded whenever the differences in mean values exceeded 2 ppbv for consecutive data points.

A bivariate fit to the data sets (York, 1966), which incorporates total uncertainties for both instruments (CLD: 8.6 %; TD-CRDS: 11 % + 20 pptv × RH/100) resulted in a slope of 0.996 ± 0.003 and an intercept of −1.3 pptv. This very good agreement serves to underline the general applicability of the
Figure 10a shows a 10 h time frame with pNit measurements from the denuder channel of the TD-CRDS. Unfortunately, the TD oven of the denuder channel broke down very early in the campaign and was not operational afterwards. The data from the pNit channel are presented along with the NO\textsubscript{x} and NO\textsubscript{y} measurements as well as particulate nitrate mass concentrations measured by an aerosol mass spectrometer (Aerodyne HR-ToF-AMS; DeCarlo et al., 2006; Brooks et al., 2020). The night-to-day transition is indicated via the NO\textsubscript{2} photolysis rates \(J_{\text{NO}_2}\) derived from a spectral radiometer (Metcon GmbH). The relative humidity was \(>80\%\) throughout the period shown. During the two periods when, apart from some short spikes, NO\textsubscript{x} was very low (21:10–23:45 and 02:20–03:40 UTC), TD-CRDS data indicate the presence of 300–400 pptv of pNit, which would then constitute \(\sim80\%\) NO\textsubscript{y}. Such mixing ratios of particulate nitrate are not commensurate with those measured by the AMS, which, on average, are a factor of 6–8 lower. As the AMS does not detect particles larger than \(\sim600\) nm with high efficiency (Drewnick et al., 2005), the difference could potentially indicate that a significant fraction of the particulate nitrate is associated with coarse-mode aerosol. In the lower panel of Fig. 10b, we plot the coarse-mode aerosol mass concentration determined from measurements of an optical particle counter (OPC) that measures particles between 0.2 and 20 \(\mu\)m. In the two low-NO\textsubscript{x} periods outlined above, the OPC-derived aerosol mass concentrations were between 3 and 5 \(\mu\)g m\(^{-3}\). If 10% of this coarse-mode aerosol mass concentration was nitrate, which is a typical value in the Mediterranean (Koulouri et al., 2008; Calzolai et al., 2015; Malaguti et al., 2015), this would account for 100–200 pptv of the pNit observed by the TD-CRDS and not by the AMS. However, the time profile of pNit measured by the TD-CRDS is not consistent with those of either the OPC or the AMS and rather resembles the NO\textsubscript{x} mixing ratios. This strongly suggests that the large difference between pNit reported by the TD-CRDS and the AMS does not result from the non-detection or detection of supermicron particulate nitrate by the AMS but instead results from the denuder artefacts described in Sect. 3.3.2. This short case study serves to highlight the potential positive bias in denuder-based TD-CRDS measurements of pNit under humid field conditions.

4.2 Ambient NO\textsubscript{x} and NO\textsubscript{y} measurements in an urban environment (Mainz, Germany)

NO\textsubscript{x} and NO\textsubscript{y} mixing ratios were obtained in air sampled outside the Max Planck Institute for Chemistry (MPIC). The MPIC (49\(^{\circ}\)59′27.5″N 8\(^{\circ}\)13′44.4″E) is located on the outskirts of Mainz but within 200 m of two busy two- and four-lane roads and within 500 m of additional university buildings as well as commercial and residential areas. The city of Mainz (217 000 inhabitants) is located in the densely populated Rhine–Main area along with the cities of Frankfurt (753 000 inhabitants) and Wiesbaden (278 000 inhabitants);
the air in this region is strongly influenced by local pollution. The sampling location was on the top floor of a three-story building (ca. 12 m above ground level). Air was subsampled to the inlets of the instrument from a ~1 m long 0.5 inch outer diameter PFA tube that was connected to a membrane pump and flow controller to generate a 20 slpm bypass flow. Aerosol transmission was probably < 100% in these measurements.

Figure 11a summarises the 8 d of measurement (data coverage 82%) as a time series for NO$_x$, NO$_y$, NO$_z$, (10 min averages) wind speed (1 h averages) and the NO$_z$/NO$_y$ ratio. The NO$_x$ and NO$_y$ mixing ratios were highly variable throughout this period, with NO$_x$ mixing ratios between 0.7 and 148.3 ppbv (mean and median values of 22.1 and 6.9 ppbv respectively). Traffic-related morning rush-hour peaks in NO$_x$ were observed on all weekdays (14, 16, 17 and 20 January) between 05:00 and 10:00 UTC. The morning NO$_x$ peak is reduced or absent on the weekends (18 and 19 January). NO$_x$ levels stayed above 50 ppbv for nearly a full day from 18:00 UTC on 16 January until 18:00 UTC on 17 January, which coincides with constantly low wind speeds and the sampling of air masses that were predominantly local, and thus highly polluted. NO$_z$ mixing ratios were usually between 0.5 and 2.5 ppbv (minimum < LOD, maximum 3.1 ppbv, mean 1.0 ppbv and median 0.9 ppbv), with NO$_z$/NO$_y$ ratios below 0.5. These values indicate that the air masses had been impacted by recent (local) NO$_x$ emissions.

The NO$_z$/NO$_y$ ratio can be used as an indicator of the degree of chemical processing of an air mass. In Fig. 11b, a median diel profile (including all measurement days) for the NO$_z$/NO$_y$ ratio from the ambient measurement is shown. The diel profile displays two distinct minima in NO$_z$/NO$_y$ during the morning and evening rush hours, where NO$_z$ only makes up 5%–10% of the total NO$_y$. This fraction increases up to 15% during midday and up to 25% during nighttime, when emissions of NO$_x$ are reduced. The diel profiles of NO$_z$/NO$_y$ are strongly influenced by fresh emissions of NO$_x$. As the measurement location is strongly influenced by traffic, there is a decrease in NO$_x$ (and an increase in NO$_z$/NO$_y$) at night-time. Night-time increases in NO$_x$ (13–14, 15–16, 18–19 and 19–20 January 2020) may also be partially caused by the formation of N$_2$O$_5$ as previously observed (Schuster et al., 2009) and which would have been favoured by the low night-time temperatures (< 10°C) in winter.

These measurements serve to illustrate the applicability of our TD-CRDS over a wide range of NO$_x$ and NO$_y$ concentrations under realistic field conditions and in the investigation of processes that transform NO$_x$ into its gas- and particle-phase reservoirs.

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5 Conclusions

We report on the development, characterisation and first deployment of a TD-CRDS instrument for the measurement of NO$_x$, NO$_y$, NO$_z$ and pNit. Our laboratory experiments suggest that the different gas-phase NO$_x$ species investigated
(PAN, iPN, N₂O₅, HONO, ClNO₂ and HNO₃) are converted with near-stoichiometric efficiency to NOₓ at a nominal oven temperature of 850 °C. NH₄NO₃ particles of 200 nm in diameter are also detected quantitatively as NOₓ, whereas the efficiency of detection of NaNO₃ particles of similar diameter was closer to 25 %. The efficiency of the detection of coarse-mode particles will be further reduced by their lower transmission through the denuder.

The potential for NH₃ to bias NOₓ measurements was assessed and was found to be insignificant in ambient air or synthetic air containing VOCs and water. The conversion to NO₂ (by reaction with O₃) of atmospheric NO, as well as NO formed in the heated inlet, circumvents bias resulting from O₃ thermolysis (leading to an NO₂ overestimation) and secondary processes, which are initiated by the thermal dissociation of organic nitrates.

For our activated carbon denuder, we observed > 90 % transmission for ammonium nitrate particles with diameters between 40 and 400 nm. Under humid conditions, the denuder suffered from the direct breakthrough of NO and the re-release of previously stored iPN and NO₂ in the form of NO, indicating a potential bias of pNit measurements using this technique and potentially limiting its deployment to low-NOₓ and low-NOy environments. When using comparable denuders, we recommend regular checks with humidified zero air to characterise potential breakthrough. Our experiments demonstrated that the release of NOₓ from the denuder exposed to humid zero air for several hours can decrease to values below 1 ppbv, which, in a first approximation, could be treated as an offset. Cycling between multiple denuders would help to reduce the size of any bias.

The performance of the instrument under field conditions was demonstrated by measurements in Mainz, Germany, and during the AQABA ship campaign. NOₓ measurements with the new instrument were in good agreement with those from an established, independent CLD-based instrument.

References


Competing interests. The authors declare that they have no conflict of interest.

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