Simultaneous detection of atmospheric HONO and NO₂ utilizing an IBBCEAS system based on an iterative algorithm

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Abstract. We present an improved incoherent broadband cavity-enhanced absorption spectroscopy (IBBCEAS) system based on iterative retrieval algorithm for simultaneous detection of atmospheric nitrous acid (HONO) and nitrogen dioxide (NO₂). The conventional IBBCEAS retrieval algorithm depends on the absolute change of light intensity, which requires high stability of light source and stable transmission of light intensity of all optical components. The new algorithm has an iterative module to obtain the effective absorption optical path length, and then the concentrations of HONO and NO₂ are determined by differential optical absorption spectroscopy (DOAS) retrieval, so it is insensitive to the fluctuation of absolute light intensity. The robustness of the system is verified by simulating the influence of the relative change of light intensity on the spectral retrieval results. The effect of nitrogen purging in front of the cavity mirrors on shortening the actual cavity length was measured and corrected by using NO₂ gas samples. Allan deviation analysis was conducted to determine the system stability, indicating the detection limits (2σ) of HONO and NO₂ are 0.08 ppbv and 0.14 ppbv at an integration time of 60 s, respectively. Furthermore, Kalman filtering was used to improve the measurement precision of the system. The measurement precision of the system at an integration time of 3 s can be improved by 4.5-fold with applying Kalman filtering, which is equivalent to the measurement precision at an integration time of 60 s without applying Kalman filtering. The atmospheric HONO and NO₂ were observed by the IBBCEAS system based on iterative algorithm, and compared with those measured by conventional IBBCEAS.

1 Introduction

As HONO can absorb 300–400 nm solar radiation to form hydroxyl radical (OH) and nitric oxide (NO), it has been demonstrated that HONO contributes significantly to the OH budget during the daytime (Harrison et al., 1996). Recent studies have shown that the contribution of HONO to OH production plays an important role not only in the morning but also throughout the day (Spataro et al., 2013;Alicke, 2002). It is reported

that the contribution of HONO photolysis to OH production can reach 60% during the day (Michoud et al., 2012;Lu et al., 2013). However, the exact mechanisms leading to HONO formation are still under discussion. Existing gas-phase sources cannot explain the high concentration of HONO observed during the daytime (Zhou et al., 2002), indicating unknown sources of strong HONO during the day (Acker et al., 2006;Kleffmann, 2005). Therefore, fast and accurate measurement of HONO concentration is the premise of studying the atmospheric chemical behavior of HONO and its contribution to regional oxidation. However, the lifetime of HONO may be only a few minutes with a low concentration during the day, even a few hundred parts-per-trillion-by-volume (10^{-12}, pptv) (Laufs et al., 2017;Hou et al., 2016). Therefore, a rapid and accurate detection of HONO has become a challenge.

HONO detection methods are mainly classified into two categories: One is based on wet chemical techniques, and the other is based on spectroscopic methods. The wet chemical method mainly includes Denuder absorption coupling Ion Chromatography (Denuder-IC) (Neftel et al., 1996), Gas and Aerosol Collector (GAC) (Dong et al., 2012), Stripping Coil-Ion Chromatography (SC-IC) (Xue et al., 2019;Cheng et al., 2013), and Long Path Absorption Photometry (LOPAP) (Chen et al., 2014;Heland et al., 2001;Kleffmann et al., 2006). These kinds of methods have a lower detection limit and can reach several pptv. However, these methods need to be calibrated to obtain accurate HONO concentration. Furthermore, the maintenance is cumbersome and requires frequent replacement of the chemical solution. Spectroscopic methods are based on the Lambert-Beer law and quantify the concentration of HONO by measuring the its absorption spectrum in a specific wavelength region, which is not easily affected by chemical interferences. Spectroscopic methods can be divided into conventional absorption methods and cavity-enhanced methods (Fiedler et al., 2003). Conventional absorption methods mainly include Differential Optical Absorption Spectroscopy (DOAS) (Tsai et al., 2018;Qin et al., 2009), Fourier Transform Infrared Spectroscopy (FTIR) (Stockwell et al., 2014), and Infrared Quantum Cascade Laser (QCL) absorption spectroscopy (Cui et al., 2019). In these methods, the absorption spectrum of gas is obtained by the beam through multi-pass cells or long open path, the optical path length is a key factor that affects the sensitivity of the system. The cavity-enhanced methods are based on high-finesse optical cavity-enhanced absorption spectroscopy, mainly including Cavity Ring Down Spectroscopy (CRDS) (Wang and Zhang, 2000) and Incoherent Broadband Cavity-Enhanced Absorption Spectroscopy (IBBCEAS) (Jordan et al., 2019;Duan et al., 2018;Gherman et al., 2008;Nakashima and Sadanaga, 2017;Donaldson et al., 2013;Scharko et al., 2014;Wu et al., 2014). IBBCEAS methods have higher spatial resolution and are easier to deploy on different platforms. Compared to CRDS techniques based on a single wavelength, IBBCEAS can achieve simultaneous measurements of multi-gases by using a broadband light source. The IBBCEAS technology is such that the light beam is reflected back and forth in a high finesse optical cavity formed by two high-reflectivity mirrors. An optical cavity of several tens of centimeters can make the effective absorption path length reaching several kilometers, thereby improving the detection limit of the system. In recent years, IBBCEAS technology has been demonstrated to apply to HONO field measurements in remote (Duan et al., 2018;Tang et al., 2019) and urban regions (Crilley et al., 2019;Wu et al., 2014;Min et al., 2016;Nakashima and Sadanaga, 2017). Although a large number of HONO inter-

comparisons between different instruments have been carried out in previous studies (Crilley et al., 2019;Duan et al., 2018;Xue et al., 2019;Kleffmann et al., 2006;Stutz et al., 2010), deviations of HONO measurements between different methods still exist. In a recent field observation, the correlation between different instruments is high (r^2>0.97), unfortunately, the absolute concentration difference even reaches 39% (Crilley et al., 2019). The factors that affect the accuracy of measurement need to be discussed further.

Conventional IBBCEAS technology retrieves HONO concentration by measuring the absolute change of light intensity. It depends heavily on the stability of the instrument and is sensitive to environmental factors, such as temperature and vibration. Recently, some researchers have equated the optical cavity of IBBCEAS to multi-pass cell, and then the gas concentration is determined according to the DOAS retrieval algorithm (Herman et al., 2009;Hoch et al., 2014;Horbanski et al., 2019;Meinen et al., 2010;Platt et al., 2009;Thalman and Volkamer, 2010). Because DOAS retrieval uses the narrow-band differential absorption characteristics of trace gas to quantify the gas concentration, the IBBCEAS system based on DOAS retrieval is insensitive to the broadband change of light intensity. The key point of this technology is to determine the effective absorption path length, so as to determine the gas concentration by using DOAS retrieval. The most common method to correct the effective absorption path length uses the measured optical density (Hoch et al., 2014). However, in this way, as with conventional IBBCEAS, the retrieved gas concentration will be affected by fluctuations in the intensity of the light source. There are also attempts to calculate the effective absorption optical path with known absorption of O_4 gas, but the measurement accuracy is limited (Thalman and Volkamer, 2010;Herman et al., 2009). Recently, Horbanski et al. (2019) used an iterative method to calculate the effective absorption optical path, and developed a NO_2 instrument, indicating the effectiveness of the method (Horbanski et al., 2019). Compared with NO_2, HONO concentration ranges from pptv to several ppbv, and the spatial and temporal distribution is highly variable, which brings challenges to the accurate measurement of atmospheric HONO. Therefore, the measurement of HONO requires the IBBCEAS instrument to be of high stability, the application of iterative algorithm is helpful to improve the accuracy of HONO measurement.

This paper describes the improvement of the IBBCEAS system for the simultaneous detection of atmospheric HONO and NO_2. The concentration of HONO and NO_2 are determined by multiple iterations combined with the DOAS retrieval algorithm. The instrument can eliminate the influence of the broadband change of light intensity by using the iterative algorithm and prevent the instability of the light source and mechanical vibration. Kalman filtering technology is an effective post-processing technology for gas concentration measurement(Wu et al., 2010;Leleux et al., 2002). Kalman filtering is first applied in laser absorption spectroscopy to real-time measurement of CO_2 and NH_3 at parts-per-million-by-volume level (Leleux et al., 2002). In this work, we applied the Kalman filtering technique to realize trace gas concentration measurements, thus improving the measurement precision of the system. To our knowledge, the Kalman filtering technique is first used for HONO and NO_2 measurements. Our instrument's capability of making fast high-sensitivity measurements of HONO and NO_2 is of great significance for understanding the source of HONO and studying its role in atmospheric chemistry.

2 System and principle

 2.1 Theory of IBBCEAS

The basic idea of the IBBCEAS system based on an iterative algorithm is to use a high finesse optical cavity to increase the effective absorption light path, thereby improving the detection sensitivity of the instrument, and using the DOAS retrieval algorithm to determine the gas concentration (Hoch et al., 2014;Herman et al., 2009;Meinen et al., 2010;Platt et al., 2009;Thalman et al., 2015;Thalman and Volkamer, 2010). For this

115   system, the cavity-enhance optical density $D_{CE}(\lambda)$ is defined as follows (Horbanski et al., 2019):

$$D_{CE}(\lambda) = \overline{L}_{eff}(\lambda) \cdot \left(\sum_i c_i \cdot \sigma_i(\lambda) + \varepsilon_b(\lambda) \right) \tag{1}$$

where c_i is the concentration of the gas species i, $\sigma_i(\lambda)$ is the absorption cross-section of the gas at wavelength λ, $\varepsilon_b(\lambda)$ is the broadband extinction caused by Rayleigh scattering and Mie scattering, and $\overline{L}_{eff}(\lambda)$ is an effective path length.

120   The DOAS evaluation determines the gas concentrations based on:

$$\ln\left(\frac{I_0(\lambda)}{I(\lambda)} \right) = \overline{L}_{eff} \cdot \left[\Sigma \sigma_i'(\lambda) c_i + \text{ polynomial } \right] \tag{2}$$

Here, $I_0(\lambda)$ is the intensity of light passing through the cavity without gas absorption, $I(\lambda)$ is the intensity of light passing through the cavity with gas absorption, $\sigma_i'(\lambda)$ is the differential part of the absorption cross section of the gas, and the polynomial term in Eq. (2) represents the broadband spectral structure in the

125   measurement spectrum other than the differential absorption part. In traditional DOAS, the optical path length $\overline{L}_{eff}$ is a constant as determined by the physical distance, whereas in IBBCEAS it is not a constant and has a dependence on the optical density (Platt et al., 2009).

The gas concentrations of IBBCEAS measurements can be obtained by using the DOAS evaluation. Here, the wavelength-dependent effective absorption optical path $\overline{L}_{eff}(\lambda)$ is calculated by an iterative algorithm.

130   The implementation of the algorithm is described in Section 3.3.

2.2 Optical Layout

The IBBCEAS system in this study was based on Duan et al. work (2018) [25]. The optical layout of the IBBCEAS system is shown in Fig. 1. The light source of the instrument uses a near-ultraviolet LED (LEDengin) with a center wavelength of 368 nm. In order to ensure the stability of the LED light intensity,

135   the LED is mounted to a Peltier device with heat sink, and the temperature of the LED is stabilized at 20°C ± 0.1°C via a thermistor temperature sensor and a PID controller. The light from the LED is coupled to a long 68 cm optical cavity through the achromatic lens (Edmund). The optical cavity is composed of two high-reflectivity mirrors and a perfluoroalkoxy alkanes (PFA) tube with an inner diameter of 22 mm. The high-

reflectivity mirror (Layertec) is installed in the adjustment frame at both ends of the optical cavity. The light
140   transmitted through the optical cavity is filtered by a band-pass filter (BG3, Newport) and then focused by
an off-axis parabolic mirror (Edmund), and finally coupled to one end of an optical fiber (600 μm, Ocean
Optics). The other end of the optical fiber is connected to a spectrometer (QE65000pro, Ocean Optics), and
the spectrometer is used to collect the corresponding spectral signal.

In order to prevent the reflectivity degradation during the measurement due to adsorptions of aerosol or
145   organic species onto the mirror surfaces, the surfaces of the two high-reflectivity mirrors are purged with
high-purity nitrogen (99.999%). The flow rates of the purge are controlled by two mass flow controllers at
0.1 standard liter per minute. The sampling tube of the instrument consists of a PFA tube with an outer
diameter of 6 mm. A 0.2-μm PTFE filter membrane is connected to the inlet of the sampling port to prevent
particles from entering the optical cavity. A diaphragm pump draws the ambient air into the instrument
150   through the sampling tube at a flow rate of 6 standard liters per minute. The ambient air enters the system
and is divided into two paths through a three-way PFA joint, one of which is discharged from the air outlet,
and the other airflow enters the cavity and at a flow rate of 1.2 standard liters per minute using a mass flow
controller. The use of this sampling gas path can ensure that the residence time of the atmosphere air in the
sampling tube can be shortened. It does so by increasing the total sampling flow rate while maintaining a
155   fractional constant flow rate of the cavity, thereby reducing the secondary generation and loss of HONO in
the sampling tube. The software control interface is programed with LabVIEW language to ensure the orderly
operation of the mass flow controller and spectrometer in the process of instrument operation.

3 Results and analysis

3.1 Determination of Mirror Reflectivity

160   Since the absorption optical path of the gas in the optical cavity is related to the reflectivity of mirrors, it is
necessary to determine the reflectivity of the high-reflectivity mirrors before the gas concentration is retrieved.
According to the method mentioned by Washenfelder et al. (2008)(Washenfelder et al., 2008), the Rayleigh
scattering difference between nitrogen and helium is used to determine the wavelength-dependent reflectivity
of the cavity mirrors ($R(\lambda)$):

165    $$R(\lambda) = 1 - d_0 \frac{\left(\frac{I_{N_2}(\lambda)}{I_{He}(\lambda)} \varepsilon_{Ray}^{N_2}(\lambda)\right) - \left(\varepsilon_{Ray}^{He}(\lambda)\right)}{1 - \left(\frac{I_{N_2}(\lambda)}{I_{He}(\lambda)}\right)} \qquad (3)$$

Here, d_0 is the cavity length, ε_{Ray} is the extinction caused by Rayleigh scattering, I_{N_2} and I_{He} are the light
intensity when the cavity is filled with nitrogen and helium, respectively.

The accuracy in determining the mirror reflectivity will affect the accuracy of subsequent gas concentration
measurements. Therefore, for achieving high measurement accuracy, it is necessary to stabilize the gas

170   temperature inside the optical cavity. Then the high purity nitrogen (99.999%) and high purity Helium (99.999%) are introduced into the optical cavity in turn, and the corresponding spectra are recorded after the spectrum become stable after gas filling. The wavelength-dependent mirror reflectivity is calculated by substitute the ratio of the recorded nitrogen spectrum intensity and helium spectrum intensity into the Eq. (3). The resulting dependence of the mirror reflectivity on wavelength is shown in Fig. 2. The red line is the

175   spectrum measured when the cavity is flushed with nitrogen; the black line is obtained when the cavity is filled with helium. It can be seen that the reflectivity of the mirrors is ~0.99980 at the wavelength of 368.2 nm.

3.2 Calibration of Effective Cavity Length

Considering the effect of mixing between the actual atmosphere gas in the cavity and the nitrogen purge gas

180   within a narrow space in front of the cavity mirrors during measurements, the effective cavity length becomes shorter than its physical length. The calibration experiments are needed to determine the effective cavity length (d_{eff}). We determine the value of d_{eff} by measuring the effective concentration of a NO_2 gas mixture, with and without mirror purge gas to the cavity mirrors. The NO_2 gas sample is made by mixing a cylinder gas with a nominal 10 ppm NO_2 and zero air in a Teflon (FEP) gas bag, which has low adsorption property.

185   Then the NO_2 gas mixture in the FEP gas bag is injected into the IBBCEAS system and measured by the iterative algorithm. The inlet flow rate of NO_2 was 1 standard liter per minute, and the purging flow rate at both ends of the optical cavity was 0.1 standard liter per minute. Open and close the purge gas intermittently during the operation of the instrument and record the spectrum at the corresponding time. The NO_2 concentration is retrieved according to the spectra measured during this period. Fig. 3 shows the change of

190   the appearing NO_2 concentration during purge on and off. The effective concentration of NO_2 become higher after purge off, and the concentration of NO_2 is returned to the previous value again after purge on. This indicates that the mirror reflectivity and NO_2 gas mixture before and after purge off remain unchanged, and the concentration of NO_2 are relatively stable during purge on and off. The average concentration of NO_2 measured at purge on is 62.64±0.32 ppb, the average concentration of NO_2 measured after purge off is

195   70.92±0.19 ppb. For the cavity with physical length (d_0) of 68 cm, d_{eff} can be calculated according to the equation (4). The calculated d_{eff} is 60.06 cm.

$$d_{eff} = d_0 \times \frac{[NO_2]_{\text{purge}\,on}}{[NO_2]_{\text{purge}\,off}} \qquad (4)$$

3.3 Determination of Effective Absorption Optical Path and Concentration Retrieval

For traditional DOAS fitting, the retrieval result is not affected as long as there is no change of narrow-band

200   structure, the DOAS system is insensitive to the variation of absolute light intensity. The optical absorption path length of a usual DOAS system is constant. However, in the IBBCEAS system, the effective optical absorption path length is strongly dependent on the wavelength, which is related to the wavelength-

dependency of the mirror reflectivity and intra-cavity absorption. For conventional IBBCEAS, the wavelength dependence problem is solved by converting the optical density to extinction absorption, but this method will seriously depend on the absolute stability of the light intensity. For IBBCEAS system based on iterative algorithm, the absorption optical path is not equal to the average optical path ($\bar{L}_0(\lambda)$), so $\bar{L}_0(\lambda)$ needs to be corrected. Previous studies have tried to calculate the effective absorption optical path by using the concentration of known gases, such as O_4. However, the measurement accuracy is limited when the relative variation of the wavelength dependence of the path length is corrected with a single wavelength, and the peak wavelength of O_4 in retrieval band is at 380nm and the absorption cross-section of O_4 at 380nm is three times weaker than that at 477 nm, the error may occur in retrieving the slant column density of O_4. According to the description of Horbanski et al. (2019), the scalar correction factor is not suitable for strong differential absorbers because their absorption band distortion can only be corrected by wavelength-resolved. Recently, Horbanski et al. (2019) proposed an iterative method to calculate the wavelength-dependent effective absorption optical path length. Through multiple iterative retrieval, the effective absorption optical path can be finally determined. Because the filter membrane is added to the front end of the sampling port of the system in this work, the influence of gas absorption and Rayleigh scattering is only considered in the retrieval of concentration. According to Platt et al. (2009), the cavity-enhanced $D_{CE}(\lambda)$ of the system can be defined as:

$$D_{CE}(\lambda) = \ln(\frac{I_{tot0}}{I_{tot}}) = \ln\left(\frac{\int_0^\infty I_{in0}(n)dn}{\int_0^\infty I_{in}(n)dn}\right) \tag{5}$$

I_{tot}, I_{tot0} is the light intensity detected when with and without gas absorption in the optical cavity, n indicates the n-th intra-cavity reflection, the total transmission can be considered as a sum of transmissions of all the individual intra-cavity sub beams of consecutive mirror reflections. By using Lambert-Beer law, Eq. (5) can be changed to:

$$D_{CE}(\lambda) = \ln\left(\frac{\int_0^\infty I_{in0}(n)dn}{\int_0^\infty I_{in}(n)dn}\right) = \ln\left(\frac{\int_0^\infty I_{in}(0)\cdot e^{-(1-R(\lambda)+\varepsilon_R(\lambda)d_0)n}dn}{\int_0^\infty I_{in}(0)\cdot e^{-(1-R(\lambda)+\varepsilon_R d_0+\sigma(\lambda)\cdot c\cdot d_0)n}dn}\right) = \ln\left(1+\frac{\sigma(\lambda)\cdot c\cdot d_0}{1-R(\lambda)+\varepsilon_R(\lambda)d_0}\right) \tag{6}$$

According to the description of Platt et al. (2009), the relationship between the effective optical path length ($\bar{L}_{eff}(\lambda)$) and the optical density ($D_{CE}(\lambda)$) is as follows:

$$\bar{L}_{eff}(\lambda) = \frac{D_{CE}(\lambda)}{\sigma(\lambda)\cdot c} \tag{7}$$

By substituting Eq. (6) into Eq. (7), The $\bar{L}_{eff}(\lambda)$ can be calculated:

$$\bar{L}_{eff}(\lambda) = \frac{d_0}{1-R(\lambda)+d_0\cdot\varepsilon_R(\lambda)}\cdot\frac{D_{CE}(\lambda)}{e^{D_{CE}(\lambda)}-1} \tag{8}$$

Therefore, if we know the optical density, we can correct the effective absorption optical path, and we can continuously approach the real effective absorption optical path by multiple iterative retrieval methods.

Horbanski et al. (2019) have made a detailed derivation; here is only a brief introduction, mainly divided into the following steps:

1. Assuming that the concentration of HONO and NO$_2$ is known, the optical density can be calculated according to Eq. (6).

2. Combined with the optical density and lens reflectivity calculated in (1), the effective absorption optical path is calculated according to Eq. (8).

3. Using the DOAS method to fit $D_{\text{CE, meas}}(\lambda)$ and $\bar{L}_{\text{eff}}(\lambda) \cdot \sigma_i(\lambda)$, new HONO and NO$_2$ concentration values are obtained.

4. The HONO and NO$_2$ concentrations obtained in Step 3 are replaced into Step 1 to recalculate the optical density, so the Step 1, Step 2, and Step 3 are repeated, until the change of HONO and NO$_2$ concentrations to their values of the previous iteration become less than an allowable tolerance range, a stop condition for iteration is that the concentration difference between two retrieval is less than the fit error. The final retrieval results of HONO and NO$_2$ concentrations are obtained. The retrieval steps are shown in Fig. 4.

All the data processing is based on DOASIS software (Kraus, 2006). The algorithm takes the high-resolution cross sections of HONO (Stutz et al., 2000), NO$_2$ (Voigt et al., 2002), and O$_4$ (Greenblatt et al., 1990) as the input, and then convolutes these high-resolution cross sections with the instrument functions of 0.49 nm FWHM. The fitting range of the spectrum is from 363 nm to 388 nm. Fig. 5 shows the change of effective optical path length in iterative algorithm retrieval, where L1, L2, L3, L4 and L5 are the corresponding absorption optical path from zero iteration to the fourth iteration, respectively. The effective absorption optical path length is finally converged as the number of iterations increases. Based on the effective absorption optical path length at the final iteration, the real concentration can be obtained according to the DOAS fitting. Fig. 6 shows the final HONO and NO$_2$ concentrations obtained by using the iterative algorithm mentioned above in the actual large spectrum retrieval. The HONO and NO$_2$ concentrations obtained by the final fitting are 0.78 ppbv and 29.18 ppbv, respectively.

3.4 Detection Limit and Uncertainty of the System

The Allan deviation is often used to calculate the sensitivity and stability of the system. Ideally, the detection sensitivity of the system can be improved by averaging successive measurements or by integrating the signal for longer time. However, actual measurement processes are affected by instrument drifts and other noise contributions. The system can only achieve an optimal detection sensitivity before slow-varying factors become dominant. Therefore, Allan deviation can be used to describe the overall performance and stability of the system. Zero air was introduced into the optical cavity at a flow rate of 0.1 standard liters per minute, and 10,000 spectra were continuously recorded. The integration time of each spectrum is 3 s, and each spectrum is fitted using the DOAS algorithm to obtain HONO and NO$_2$ concentrations. Then the Allan deviation is calculated according to Eq. (9):

$$\delta_A^2(\tau) = \frac{1}{2(m-1)} \sum_{k=1}^{m-1} \left(y_{k+1}(\tau) - y_k(\tau) \right)^2 \tag{9}$$

where m is the number of time series, and $y_k(\tau)$ is the average concentration during a time interval of τ. Fig. 7 shows the variation of Allan deviation of the system with integration time. Under the integration time of 60 s, the detection limits (2σ) of HONO and NO_2 are 0.08 ppbv and 0.14 ppbv, respectively. Table 1 shows the detection limits of HONO and NO_2 measured by different IBBCEAS instruments reported in literatures. Our system has higher detection sensitivity than most other instruments. The Allan deviation continuously decreases over the average time for several hours. This shows that the instrument is very stable in performances.

The uncertainty of the system may be determined by a Gaussian error propagation. The uncertainty of the system is mainly composed of contributed by the uncertainties in absorption cross sections of the spectral features, mirrors' reflectivity, effective cavity length, and temperature and pressure in the cavity. The uncertainty of the mirrors' reflectivity is 5%. The uncertainty of the effective cavity length is 3%. The uncertainty of the temperature and pressure in the cavity is 1%. The uncertainty of the fit retrieval is 4%. According to the literature, the uncertainty of the NO_2 and HONO absorbed cross-sections is 4% (Voigt et al., 2002) and 5% (Stutz et al., 2000), respectively. Therefore, the total uncertainty of the instrument is about 8.1% for NO_2 measurement and 8.7% for HONO measurement.

3.5 Effect of Light Intensity Fluctuation

In order to verify the insensitivity of IBBCEAS based on iterative algorithm to the broadband change of light intensity, we carried out the experiment of light intensity fluctuation. By adjusting the angle of the off-axis parabolic mirror to change the intensity of the spectrum, the original lamp spectrum is attenuated by 95%, 90%, 85%, 80% and 75% respectively. After adjusting the light intensity of the lamp spectrum every time, the actual atmospheric HONO concentration is measured. Finally, three methods are used to retrieve the measured atmospheric spectrum, the original lamp spectrum combined with iterative algorithm is used for concentration retrieval in method 1, the original lamp spectrum combined with conventional IBBCEAS retrieval algorithm is used for concentration retrieval in method 2, the lamp spectrum after each change combined with the conventional IBBCEAS retrieval algorithm is used for concentration retrieval in method 3. In order to ensure that the light intensity is not affected by the external environmental factors during the measurement process and to avoid the influence of the light intensity fluctuation caused by the external environmental factors on the measurement results, the measurement time of each cycle should be as short as possible. Fig.8 shows the retrieval results of three methods under different relative light source intensity. The results show that the retrieval results of method 2 are influenced greatly by the fluctuation of light intensity, while the difference between the retrieval results of methods 1 and 3 is relatively small. Fig. 9 shows the fitting results of method 2 and method 3 after light intensity change. When the light intensity changes, the RMS of the fitting residual increases by 2 times. Although the HONO concentration can also be fitted well after the light intensity changes in method 2, the attenuation signal of the lamp spectrum is considered to be caused by HONO absorption, which leads to the high measurement result. Method 3 uses the lamp spectrum after the change of the light intensity to participate in the retrieval, which ensures the absolute stability of the

light intensity. But the lamp spectrum after light intensity change cannot be obtained in time in actual measurement. Method 1 always uses the original lamp spectrum to participate in the retrieval, which shows that it is independent of the fluctuation of the light intensity of the light source.

3.6 Ambient Measurement and Compared with Conventional IBBCEAS

Atmospheric HONO and NO_2 observations were carried out in the suburbs of Hefei City, Anhui Province (31.89°N, 117.17°E), from the 29th of September to the 1st of October 2019. The IBBCEAS system based on iterative algorithm was placed in a room with approximately constant temperature controlled by air conditioner, and the sampling port was outside of the room. A conventional IBBCEAS (Duan et al., 2018) and the IBBCEAS system based on iterative algorithm were used to measure the concentrations of HONO and NO_2, respectively. The conventional IBBCEAS instrument was developed and reported by our research group previously [25]. The length of the air sampling tubes for both instruments is about 2 m. The time series of HONO and NO_2 concentrations measured by the two instruments is shown in Fig. 10. The integration time was 1 min. The highest HONO concentration of 3.12 ppbv appeared on the evening of September 29th, whereas the average value of HONO during the measurement period was 0.96 ppbv. The average concentration of NO_2 was 15.45 ppbv, and the maximum value was 49.55 ppbv. Fig. 11 shows the correlation between the measurement results of the conventional IBBCEAS and the new IBBCEAS system based on iterative algorithm. The correlation coefficients (R^2) of HONO and NO_2 results obtained by the two instruments are 0.94 and 0.99, respectively. The differences between two IBBCEAS systems are 1% and 7% for HONO and NO_2 measurements, respectively, which is within the measurement uncertainty of instruments discussed in Section 3.4. Although there are many inter-comparisons about HONO measurements, the differences between different instruments have always existed. Crilley et al. (2019) reported that the IBBCEAS instrument and wet chemical method have good consistency in field observations, but the absolute concentration difference is 12%–39% (Crilley et al., 2019). The cause of the difference was not clearly identified, so more experiments on HONO comparison are needed in the future.

3.7 Applications of Kalman Filtering to measurement results

Allan deviation describes the relationship between the integration time of the system and the stability of the system. According to the results of the Allan deviation, the optimum integration time of the system can be obtained, so that the system achieves the best detection sensitivity. Kalman filtering technology can further improve the measurement precision of the system(Leleux et al., 2002;Fang et al., 2017;Wu et al., 2010). Compared with simple moving average, Kalman filtering can deal with "lag" effect and abnormal peak value(Leleux et al., 2002). The basic idea of Kalman filtering is to obtain the predicted state of the present time based on the state of the previous time, and the predicted state of the present time is fused with the observed state measured by the sensor to obtain the estimation of the current state (Wu et al., 2010;Leleux et al., 2002). It can be expressed as:

$$\hat{x}_k = \hat{x}_k^- + K_k \left(z_k - \hat{x}_k^- \right) \tag{10}$$

Here, $\hat{x}_k$ is the predicted value of time sequence k, $\hat{x}_k^-$ is the predicted value of time sequence k-1, K_k is the Kalman gain, and z_k is the measurement value of time sequence k. In this work, the variance of previous 10 values of concentration measurements are used in the Karman filtering. Fig.12 shows the measurements of HONO and NO_2 concentrations under zero air conditions, together with the application of the Kalman filtering. The measurement fluctuation at the integration time of 3 s was 0.33 ppbv and 0.18 ppbv for the NO_2 and HONO. After applying Kalman filtering, the effect of the fluctuation is reduced and the measurement precision at the integration time of 3 s was 0.04 ppbv and 0.07 ppbv for the NO_2 and HONO. The results show that the measurement precision improved by a factor of 4.5. This result is comparable to the measurement precision at the integration time of 60 s. Therefore, using Kalman filtering technology can enhance measurement precision and reducing measurement noise. Fig.13 shows the measurement of HONO and NO_2 concentrations with and without Kalman filtering under ambient conditions. In order to capture the sharp change of NO_2 concentration, the filter's gain parameter is set to 40. The filtering results follow the change of the measured concentration, effectively reduce the influence of noise on the concentration results, and improve the measurement precision of the system.

4 Conclusions

Here, we have developed an IBBCEAS system based on an iterative algorithm for simultaneous measuring of atmospheric HONO and NO_2. The effective absorption optical path length is obtained by the iterative algorithm. The concentrations of HONO and NO_2 are then determined by DOAS retrieval. The iterative algorithm is insensitive to the broadband change of light intensity and has good robustness. The reflectivity of the high-reflectivity mirror is characterized by the difference in observed Rayleigh scattering between nitrogen and helium. The reflectivity of the mirror is measured to be 0.99980 at the wavelength of 368.2 nm. The effect of the cavity-mirror-protecting nitrogen purge on the effective cavity length is calibrated using a stable NO_2 concentration. The detection sensitivity of the system is analyzed by Allen deviation analysis. The detection limit of the system is 0.08 ppbv (2σ value) and 0.14 ppbv (2σ value) for HONO and NO_2 at an integration time of 60 s, respectively. The IBBCEAS based on iterative algorithm is in good agreement with the conventional IBBCEAS system when applied to a 'proof-of-concept" atmospheric measurement for more than two days. The total uncertainty of the system is about 8.1% for NO_2 measurement and 8.7% for HONO measurement.

We also utilized a Kalman filtering technique to improve the measurement precision of the IBBCEAS system. This helps to realize high-precision measurements of atmospheric HONO and NO_2. After applying Kalman filtering technology, the measurement precision at an integration time of 3 s can reach the corresponding precision under the integration time of 60 s without Kalman filtering. The system has a good application prospect for follow-up researches of atmospheric HONO under several different platforms, such as vehicle, balloon, and airborne platforms.

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Reference

Acker, K., Möller, D., Wieprecht, W., Meixner, F. X., Bohn, B., Gilge, S., Plass-Dülmer, C., and Berresheim, H.: Strong daytime production of OH from HNO2 at a rural mountain site, Geophys. Res. Lett., 33, 10.1029/2005gl024643, 2006.

Alicke, B.: Impact of nitrous acid photolysis on the total hydroxyl radical budget during the Limitation of Oxidant Production/Pianura Padana Produzione di Ozono study in Milan, J. Geophys. Res., 107, 10.1029/2000jd000075, 2002.

Chen, L., Hou, S.-Q., Wang, W.-G., Tong, S.-R., Pei, K.-M., and Ge, M.-F.: Development of a Home-Made Long Path Absorption Photometer for the Sensitive Detection of Nitrous Acid, Acta Phys.-Chim. Sin., 30, 1408-1415, 10.3866/pku.whxb201406032, 2014.

Cheng, P., Cheng, Y., Lu, K., Su, H., Yang, Q., Zou, Y., Zhao, Y., Dong, H., Zeng, L., and Zhang, Y.: An online monitoring system for atmospheric nitrous acid (HONO) based on stripping coil and ion chromatography, J. Environ. Sci., 25, 895-907, 10.1016/s1001-0742(12)60251-4, 2013.

Crilley, L. R., Kramer, L. J., Ouyang, B., Duan, J., Zhang, W., Tong, S., Ge, M., Tang, K., Qin, M., Xie, P., Shaw, M. D., Lewis, A. C., Mehra, A., Bannan, T. J., Worrall, S. D., Priestley, M., Bacak, A., Coe, H., Allan, J., Percival, C. J., Popoola, O. A. M., Jones, R. L., and Bloss, W. J.: Intercomparison of nitrous acid (HONO) measurement techniques in a megacity (Beijing), Atmos. Meas. Tech., 12, 6449-6463, 10.5194/amt-12-6449-2019, 2019.

Cui, X.-j., Yu, R., Chen, W.-d., Zhang, Z.-r., Pang, T., Sun, P.-s., Xia, H., Wu, B., and Dong, F.-z.: Development of a Quantum Cascade Laser-Based Sensor for Environmental HONO Monitoring in the Mid-Infrared at 8 μm, J. Lightwave Technol., 37, 2784-2791, 10.1109/jlt.2018.2876672, 2019.

Donaldson, M. A., Berke, A. E., and Raff, J. D.: Uptake of Gas Phase Nitrous Acid onto Boundary Layer Soil Surfaces, Environ. Sci. Technol., 48, 375-383, 10.1021/es404156a, 2013.

Dong, H. B., Zeng, L. M., Hu, M., Wu, Y. S., Zhang, Y. H., Slanina, J., Zheng, M., Wang, Z. F., and Jansen, R.: Technical Note: The application of an improved gas and aerosol collector for ambient air pollutants in China, Atmos. Chem. Phys., 12, 10519-10533, 10.5194/acp-12-10519-2012, 2012.

Duan, J., Qin, M., Ouyang, B., Fang, W., Li, X., Lu, K., Tang, K., Liang, S., Meng, F., Hu, Z., Xie, P., Liu, W., and Häsler, R.: Development of an incoherent broadband cavity-enhanced absorption spectrometer for in situ measurements of HONO and NO$_2$, Atmos. Meas. Tech., 11, 4531-4543, 10.5194/amt-11-4531-2018, 2018.

Fang, B., Zhao, W., Xu, X., Zhou, J., Ma, X., Wang, S., Zhang, W., Venables, D. S., and Chen, W.: Portable broadband cavity-enhanced spectrometer utilizing Kalman filtering: application to real-time, in situ monitoring of glyoxal and nitrogen dioxide, Opt. Express, 25, 26910-26922, 10.1364/OE.25.026910, 2017.

410 Fiedler, S. E., Hese, A., and Ruth, A. A.: Incoherent broad-band cavity-enhanced absorption spectroscopy, Chem. Phys. Lett., 371, 284-294, 10.1016/s0009-2614(03)00263-x, 2003.

Gherman, T., Venables, D. S., Vaughan, S., Orphal, J., and Ruth, A. A.: Incoherent broadband cavity-enhanced absorption spectroscopy in the near-ultraviolet: Application to HONO and NO2, Environ. Sci. Technol., 42, 890-895, 10.1021/es0716913, 2008.

415 Greenblatt, G. D., Orlando, J. J., Burkholder, J. B., and Ravishankara, A. R.: Absorption measurements of oxgen between 330nm and 1140nm, J. Geophys. Res.-Atmos., 95, 18577-18582, 10.1029/JD095iD11p18577, 1990.

Harrison, R. M., Peak, J. D., and Collins, G. M.: Tropospheric cycle of nitrous acid, J. Geophys. Res.-Atmos., 101, 14429-14439, 10.1029/96jd00341, 1996.

420 Heland, J., Kleffmann, J., Kurtenbach, R., and Wiesen, P.: A new instrument to measure gaseous nitrous acid (HONO) in the atmosphere, Environ. Sci. Technol., 35, 3207-3212, 10.1021/es000303t, 2001.

Herman, J. R., Thalman, R. M., Gao, W., and Volkamer, R. M.: Light emitting diode cavity enhanced differential optical absorption spectroscopy (LED-CE-DOAS): a novel technique for monitoring atmospheric trace gases, 7462, 74620H, 10.1117/12.826737, 2009.

425 Hoch, D. J., Buxmann, J., Sihler, H., Pöhler, D., Zetzsch, C., and Platt, U.: An instrument for measurements of BrO with LED-based Cavity-Enhanced Differential Optical Absorption Spectroscopy, Atmos. Meas. Tech., 7, 199-214, 10.5194/amt-7-199-2014, 2014.

Horbanski, M., Pöhler, D., Lampel, J., and Platt, U.: The ICAD (iterative cavity-enhanced DOAS) method, Atmos. Meas. Tech., 12, 3365-3381, 10.5194/amt-12-3365-2019, 2019.

430 Hou, S., Tong, S., Ge, M., and An, J.: Comparison of atmospheric nitrous acid during severe haze and clean periods in Beijing, China, Atmos. Environ., 124, 199-206, 10.1016/j.atmosenv.2015.06.023, 2016.

Jordan, N., Ye, C. Z., Ghosh, S., Washenfelder, R. A., Brown, S. S., and Osthoff, H. D.: A broadband cavity-enhanced spectrometer for atmospheric trace gas measurements and Rayleigh scattering cross sections in the cyan region (470–540 nm), Atmos. Meas. Tech., 12, 1277-1293, 10.5194/amt-12-1277-2019, 435 2019.

Kleffmann, J.: Daytime formation of nitrous acid: A major source of OH radicals in a forest, Geophys. Res. Lett., 32, 10.1029/2005gl022524, 2005.

Kleffmann, J., Loerzer, J. C., Wiesen, P., Kern, C., Trick, S., Volkamer, R., Rodenas, M., and Wirtz, K.: Intercomparison of the DOAS and LOPAP techniques for the detection of nitrous acid (HONO), Atmos. 440 Environ., 40, 3640-3652, 10.1016/j.atmosenv.2006.03.027, 2006.

Kraus, S.: DOASIS: A framework design for DOAS, 2006.

Laufs, S., Cazaunau, M., Stella, P., Kurtenbach, R., Cellier, P., Mellouki, A., Loubet, B., and Kleffmann, J.: Diurnal fluxes of HONO above a crop rotation, Atmos. Chem. Phys., 17, 6907-6923, 10.5194/acp-17-6907-2017, 2017.

445 Leleux, D. P., Claps, R., Chen, W., Tittel, F. K., and Harman, T. L.: Applications of Kalman filtering to real-time trace gas concentration measurements, Appl. Phys. B-Lasers Opt., 74, 85-93, 10.1007/s003400100751, 2002.

Lu, K. D., Hofzumahaus, A., Holland, F., Bohn, B., Brauers, T., Fuchs, H., Hu, M., Haeseler, R., Kita, K., Kondo, Y., Li, X., Lou, S. R., Oebel, A., Shao, M., Zeng, L. M., Wahner, A., Zhu, T., Zhang, Y. H., and

450 Rohrer, F.: Missing OH source in a suburban environment near Beijing: observed and modelled OH and HO2 concentrations in summer 2006, Atmos. Chem. Phys., 13, 1057-1080, 10.5194/acp-13-1057-2013, 2013.

Meinen, J., Thieser, J., Platt, U., and Leisner, T.: Technical Note: Using a high finesse optical resonator to provide a long light path for differential optical absorption spectroscopy: CE-DOAS, Atmos. Chem. Phys., 10, 3901-3914, 10.5194/acp-10-3901-2010, 2010.

455 Michoud, V., Kukui, A., Camredon, M., Colomb, A., Borbon, A., Miet, K., Aumont, B., Beekmann, M., Durand-Jolibois, R., Perrier, S., Zapf, P., Siour, G., Ait-Helal, W., Locoge, N., Sauvage, S., Afif, C., Gros, V., Furger, M., Ancellet, G., and Doussin, J. F.: Radical budget analysis in a suburban European site during the MEGAPOLI summer field campaign, Atmos. Chem. Phys., 12, 11951-11974, 10.5194/acp-12-11951-2012, 2012.

460 Min, K. E., Washenfelder, R. A., Dubé, W. P., Langford, A. O., Edwards, P. M., Zarzana, K. J., Stutz, J., Lu, K., Rohrer, F., Zhang, Y., and Brown, S. S.: A broadband cavity enhanced absorption spectrometer for aircraft measurements of glyoxal, methylglyoxal, nitrous acid, nitrogen dioxide, and water vapor, Atmos. Meas. Tech., 9, 423-440, 10.5194/amt-9-423-2016, 2016.

Nakashima, Y., and Sadanaga, Y.: Validation of in situ Measurements of Atmospheric Nitrous Acid Using

465 Incoherent Broadband Cavity-enhanced Absorption Spectroscopy, Anal. Sci., 33, 519-524, 10.2116/analsci.33.519, 2017.

Neftel, A., Blatter, A., Hesterberg, R., and Staffelbach, T.: Measurements of concentration gradients of HNO2 and HNO3 over a semi-natural ecosystem, Atmos. Environ., 30, 3017-3025, 10.1016/1352-2310(96)00011-8, 1996.

470 Platt, U., Meinen, J., Poehler, D., and Leisner, T.: Broadband Cavity Enhanced Differential Optical Absorption Spectroscopy (CE-DOAS) - applicability and corrections, Atmos. Meas. Tech., 2, 713-723, 10.5194/amt-2-713-2009, 2009.

Qin, M., Xie, P., Su, H., Gu, J., Peng, F., Li, S., Zeng, L., Liu, J., Liu, W., and Zhang, Y.: An observational study of the HONO–NO2 coupling at an urban site in Guangzhou City, South China, Atmos. Environ., 43,

475 5731-5742, 10.1016/j.atmosenv.2009.08.017, 2009.

Scharko, N. K., Berke, A. E., and Raff, J. D.: Release of nitrous acid and nitrogen dioxide from nitrate photolysis in acidic aqueous solutions, Environ Sci Technol, 48, 11991-12001, 10.1021/es503088x, 2014.

Spataro, F., Ianniello, A., Esposito, G., Allegrini, I., Zhu, T., and Hu, M.: Occurrence of atmospheric nitrous acid in the urban area of Beijing (China), Sci. Total Environ., 447, 210-224, 10.1016/j.scitotenv.2012.12.065,
480   2013.

Stockwell, C. E., Yokelson, R. J., Kreidenweis, S. M., Robinson, A. L., DeMott, P. J., Sullivan, R. C., Reardon, J., Ryan, K. C., Griffith, D. W. T., and Stevens, L.: Trace gas emissions from combustion of peat, crop residue, domestic biofuels, grasses, and other fuels: configuration and Fourier transform infrared (FTIR) component of the fourth Fire Lab at Missoula Experiment (FLAME-4), Atmos. Chem. Phys., 14, 9727-9754,
485   10.5194/acp-14-9727-2014, 2014.

Stutz, J., Kim, E. S., Platt, U., Bruno, P., Perrino, C., and Febo, A.: UV-visible absorption cross sections of nitrous acid, J. Geophys. Res.-Atmos., 105, 14585-14592, 10.1029/2000jd900003, 2000.

Stutz, J., Oh, H.-J., Whitlow, S. I., Anderson, C., Dibb, J. E., Flynn, J. H., Rappenglück, B., and Lefer, B.: Simultaneous DOAS and mist-chamber IC measurements of HONO in Houston, TX, Atmos. Environ., 44,
490   4090-4098, 10.1016/j.atmosenv.2009.02.003, 2010.

Tang, K., Qin, M., Duan, J., Fang, W., Meng, F., Liang, S., Xie, P., Liu, J., Liu, W., Xue, C., and Mu, Y.: A dual dynamic chamber system based on IBBCEAS for measuring fluxes of nitrous acid in agricultural fields in the North China Plain, Atmos. Environ., 196, 10-19, 10.1016/j.atmosenv.2018.09.059, 2019.

Thalman, R., and Volkamer, R.: Inherent calibration of a blue LED-CE-DOAS instrument to measure iodine
495   oxide, glyoxal, methyl glyoxal, nitrogen dioxide, water vapour and aerosol extinction in open cavity mode, Atmos. Meas. Tech., 3, 1797-1814, 10.5194/amt-3-1797-2010, 2010.

Thalman, R., Baeza-Romero, M. T., Ball, S. M., Borrás, E., Daniels, M. J. S., Goodall, I. C. A., Henry, S. B., Karl, T., Keutsch, F. N., Kim, S., Mak, J., Monks, P. S., Muñoz, A., Orlando, J., Peppe, S., Rickard, A. R., Ródenas, M., Sánchez, P., Seco, R., Su, L., Tyndall, G., Vázquez, M., Vera, T., Waxman, E., and Volkamer,
500   R.: Instrument intercomparison of glyoxal, methyl glyoxal and NO2 under simulated atmospheric conditions, Atmos. Meas. Tech., 8, 1835-1862, 10.5194/amt-8-1835-2015, 2015.

Tsai, C., Spolaor, M., Colosimo, S. F., Pikelnaya, O., Cheung, R., Williams, E., Gilman, J. B., Lerner, B. M., Zamora, R. J., Warneke, C., Roberts, J. M., Ahmadov, R., de Gouw, J., Bates, T., Quinn, P. K., and Stutz, J.: Nitrous acid formation in a snow-free wintertime polluted rural area, Atmos. Chem. Phys., 18, 1977-1996,
505   10.5194/acp-18-1977-2018, 2018.

Voigt, S., Orphal, J., and Burrows, J. P.: The temperature and pressure dependence of the absorption cross-sections of NO2 in the 250-800 nm region measured by Fourier-transform spectroscopy, J. Photochem. Photobiol. A-Chem., 149, 1-7, 10.1016/s1010-6030(01)00650-5, 2002.

Wang, L. M., and Zhang, J. S.: Detection of nitrous acid by cavity ring down spectroscopy, Environ. Sci.
510   Technol., 34, 4221-4227, 10.1021/es0011055, 2000.

Washenfelder, R. A., Langford, A. O., Fuchs, H., and Brown, S. S.: Measurement of glyoxal using an incoherent broadband cavity enhanced absorption spectrometer, Atmos. Chem. Phys., 8, 7779-7793, 10.5194/acp-8-7779-2008, 2008.

515   Wu, T., Chen, W., Kerstel, E., Fertein, E., and Brückner, D.: Kalman Filtering Real-Time Measurements of H2O Isotopologue Ratios by Laser Absorption Spectroscopy at 2.73 µm, Opt. Lett., 35, 634-636, 2010.

Wu, T., Zha, Q., Chen, W., Xu, Z., Wang, T., and He, X.: Development and deployment of a cavity enhanced UV-LED spectrometer for measurements of atmospheric HONO and NO2 in Hong Kong, Atmos. Environ., 95, 544-551, 2014.

Xue, C., Ye, C., Ma, Z., Liu, P., Zhang, Y., Zhang, C., Tang, K., Zhang, W., Zhao, X., Wang, Y., Song, M.,
520   Liu, J., Duan, J., Qin, M., Tong, S., Ge, M., and Mu, Y.: Development of stripping coil-ion chromatograph method and intercomparison with CEAS and LOPAP to measure atmospheric HONO, Sci. Total Environ., 646, 187-195, 10.1016/j.scitotenv.2018.07.244, 2019.

Zhou, X., Civerolo, K., Dai, H., Huang, G., Schwab, J., and Demerjian, K.: Summertime nitrous acid chemistry in the atmospheric boundary layer at a rural site in New York State, J. Geophys. Res.-Atmos., 107,
525   ACH 13-11-ACH 13-11, 10.1029/2001jd001539, 2002.

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Table 1. The detection limits of HONO and NO$_2$ measured by various IBBCEAS instruments.

Time resolution (s)	HONO detection limit (ppbv, 2σ)	NO$_2$ detection limit (ppbv, 2σ)	References
20	4	14	Gherman et al. (2008)
120	0.6	2	Wu et al. (2014)
900	1.5	1.5	Donaldson et al. (2014)
600	0.6	0.6	Scharko et al. (2014)
5	0.35	0.08	Min et al. (2016)
300	0.4	Not mentioned	Naksshima and Sadanaga. (2017)
30	0.18	0.34	Duan et al. (2018)
300	0.11	0.42	Jordan and Osthoff. (2020)
60	0.08	0.14	This work

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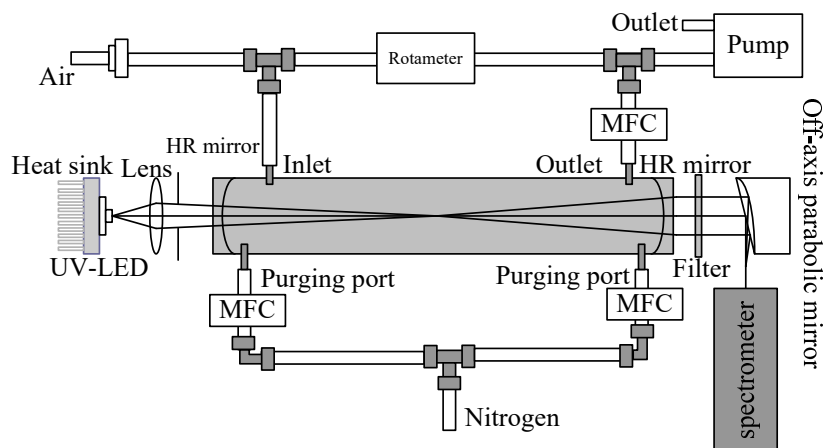
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Figure 1. The optical layout of the IBBCEAS system. MFC: mass flow controller; HR mirror: high-reflectivity mirror.

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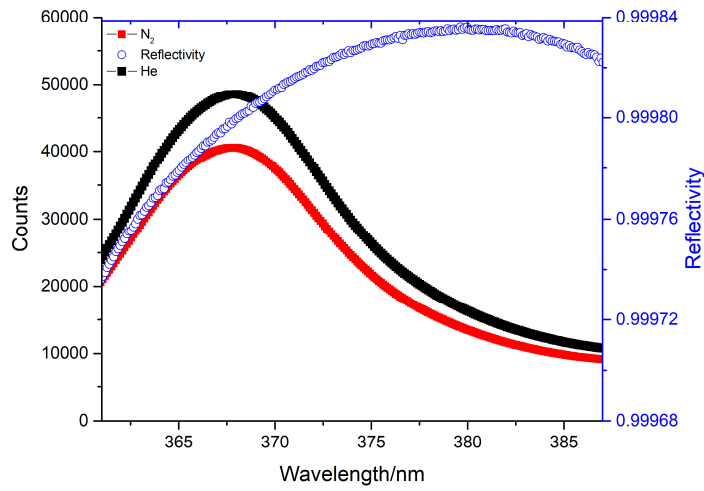
Figure 2. Dependences of transmission intensity when cavity is filled with N_2 and He gas, respectively, and the calculated cavity mirror reflectivity.

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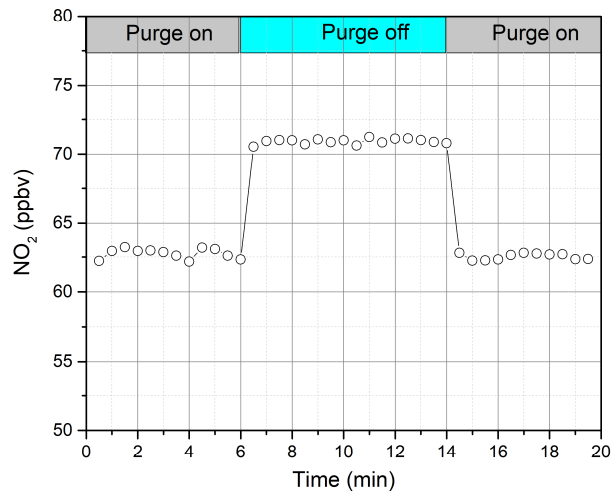
Figure 3. The change of effective NO_2 concentrations during purge on and off. The effective cavity length is calculated according to the retrieved concentrations when purging is opened and closed.

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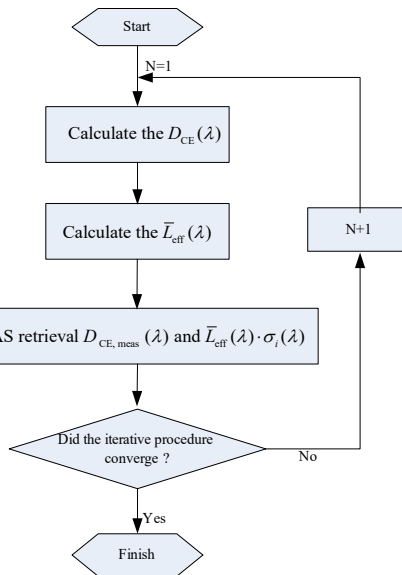
Figure 4. Flowchart of an iterative retrieval scheme for IBBCEAS.
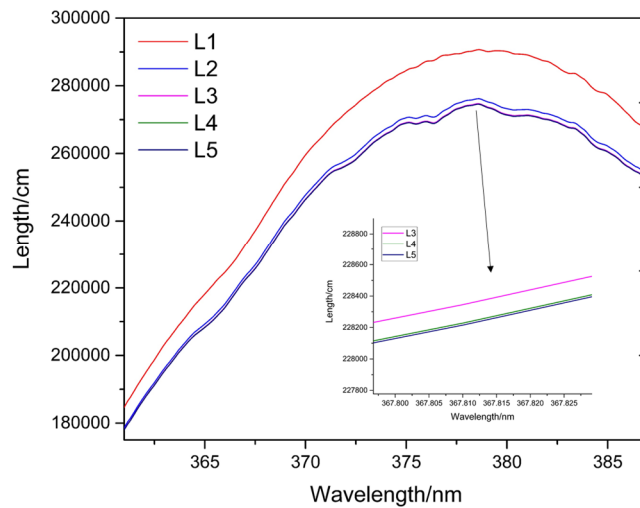
Figure 5. Example of the effective absorption optical path lengths calculated during consecutive iterations (L1, L2, L3, L4 and L5) of the iterative retrieval algorithm. As the number of iterations increases, the effective absorption optical path gradually converges.

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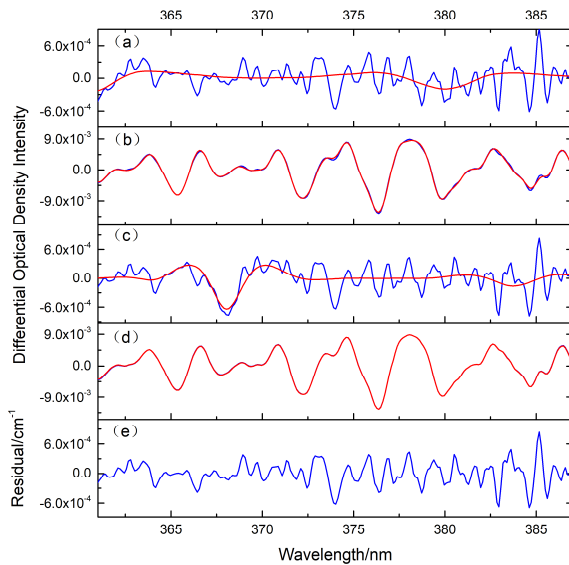
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Figure 6. Examples of spectral retrieval of O_4, NO_2 and HONO in the measured atmosphere. (a) The blue line shows the absorption spectrum of O_4, and the red line shows the fitting spectrum; (b) the blue line shows the absorption spectrum of NO_2, the red line shows the fitting spectrum, and the retrieval NO_2 concentration is 29.18 ppb; (c) the blue line shows the absorption spectrum of HONO, the red line shows the fitting spectrum of HONO, and the retrieval HONO concentration is 0.78 ppb; (d) the blue line shows the differential optical density line, and the red line shows the fitting differential optical density; (e) the blue line shows the fitting residuals.

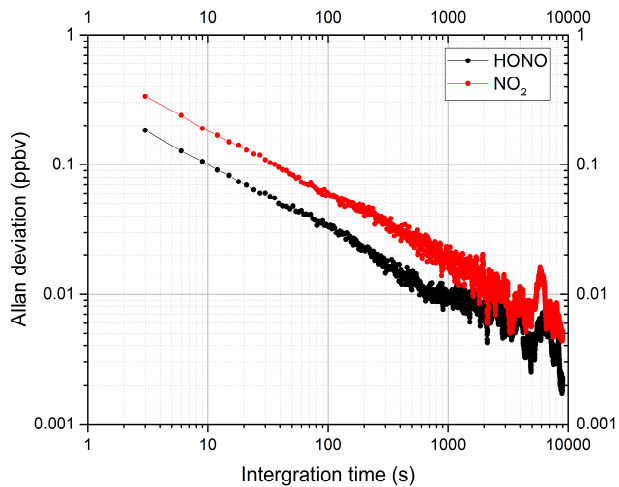



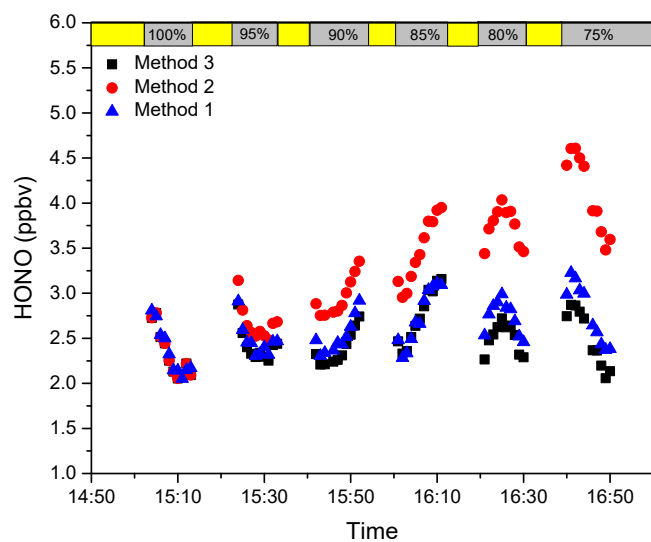
Figure 7. Allan deviation plots for measurements of HONO and NO_2. The Allan deviations of HONO and NO_2 are 0.08ppbv and 0.14ppbv at an integration time of 60 s, respectively.

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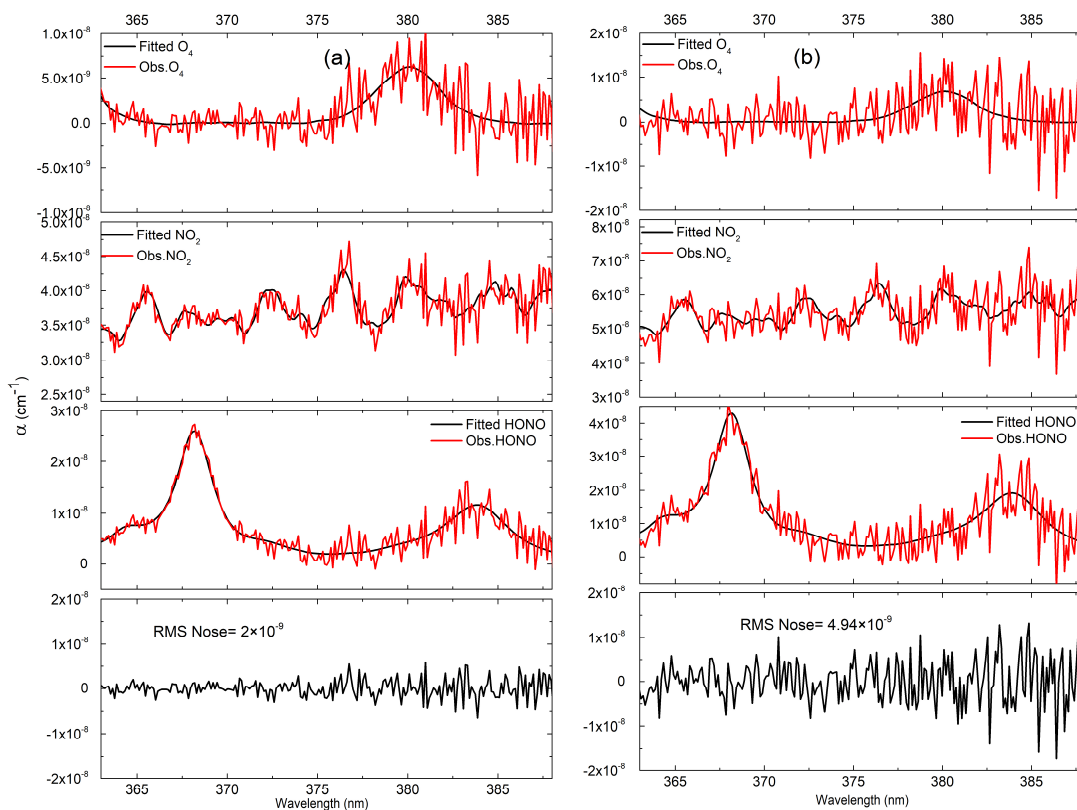
Figure 8. Retrieval results of three methods under different relative light source intensity. The yellow area represents the measuring time of the lamp spectrum and the gray area represents the atmospheric measurement after changing the relative light intensity of the light source.

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Figure 9. Spectral fitting results of (a) method 3 and (b) method 2 after light intensity change. The RMS of fit residuals in method 2 is 4.94×10⁻⁹, The RMS of fit residuals in method 3 is 2×10⁻⁹.
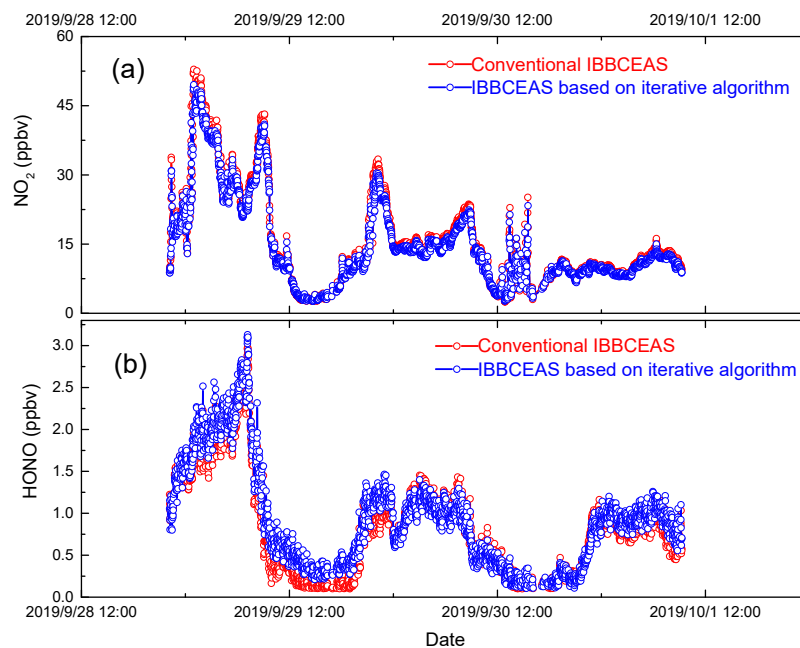
Figure 10. The time series of (a) NO$_2$ and (b) HONO concentrations measured by the two instruments. All data are 1 min averages. The red data point is the result of retrieval using iterative algorithm, and the blue data point is the result of retrieval using conventional IBBCEAS algorithm.
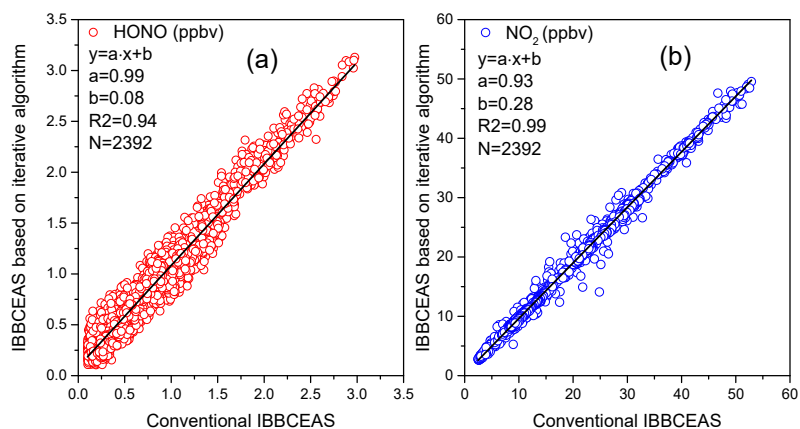
Figure 11. The correlation between the measurement results of the two IBBCEAS systems. (a) Correlation between the HONO concentrations determined by two IBBCEAS systems; (b) correlation between the NO$_2$ concentrations determined by two IBBCEAS systems.
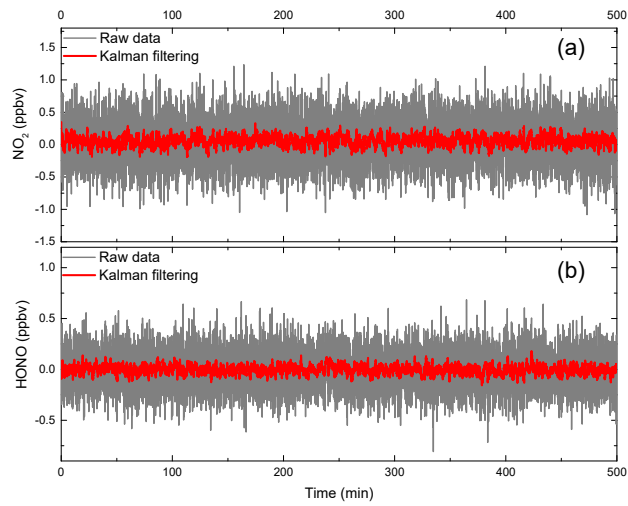
Figure 12. The measurement results of (a) HONO and (b) NO$_2$ with and without Kalman filtering under zero air conditions. The red line is the concentration result after Kalman filtering technology processing, gray line is the result of original concentration result.

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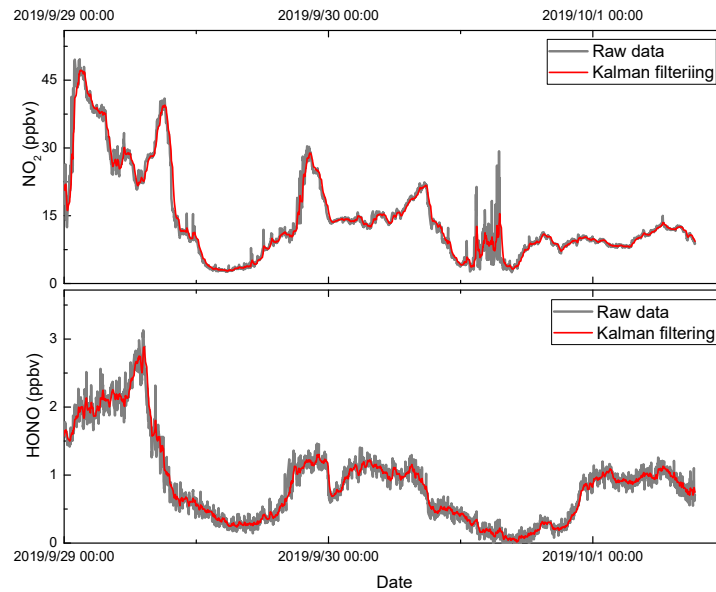
640 **Figure 13.** The measurement results of HONO and NO$_2$ with and without Kalman filtering under ambient conditions. The red line is the concentration result after Kalman filtering technology processing, gray line is the result of original concentration result.