

1 **Dynamic Infrared Gas Analysis from Longleaf Pine Fuelbeds Burned in a Wind**  
2 **Tunnel: Observation of Phenol in Pyrolysis and Combustion Phases**

3 Catherine A. Banach,<sup>1</sup> Olivia N. Williams,<sup>1</sup> Ashley M. Bradley,<sup>1</sup> Russell G. Tonkyn,<sup>1</sup>  
4 Joey Chong,<sup>2</sup> David R. Weise,<sup>2</sup> Tanya L. Myers,<sup>1</sup> Timothy J. Johnson<sup>1\*</sup>  
5 <sup>1</sup>Pacific Northwest National Laboratory, Richland, WA USA  
6 <sup>2</sup>USDA Forest Service, Pacific Southwest Research Station, Riverside, CA, USA  
7 \*Contact: Timothy J. Johnson, [timothy.johnson@pnnl.gov](mailto:timothy.johnson@pnnl.gov)

8  
9 **0. Abstract**

10 Pyrolysis is the first step in a series of chemical and physical processes that produce flammable  
11 organic gases from wildland fuels that can result in a wildland fire. We report results using a new  
12 time-resolved Fourier transform infrared method that correlates the measured FTIR spectrum to  
13 an infrared thermal image sequence enabling identification and quantification of gases within  
14 different phases of the fire process. The flame from burning fuel beds composed of pine needles  
15 (*Pinus palustris*) and mixtures of sparkleberry, fetterbush and inkberry plants was the natural heat  
16 source for pyrolysis. Extractive gas samples were analyzed and identified in both static and  
17 dynamic modes synchronized to thermal infrared imaging: A total of 29 gases were identified  
18 including small alkanes, alkenes, aldehydes, nitrogen compounds and aromatics, most previously  
19 measured by FTIR in wildland fires. This study presents one of the first identifications of phenol  
20 associated with both pre-combustion and combustion phases, using ca. 1 Hz temporal resolution.  
21 Preliminary results indicate ~2.5x greater phenol emission from sparkleberry and inkberry  
22 compared to fetterbush, with differing temporal profiles.

23 **Keywords:** Fourier transform infrared, time-resolved infrared, biomass burning, pyrolysis, phenol,  
24 benzene, naphthalene, *Pinus palustris*, *Lyonia lucida*, *Ilex glabra*, *Vaccinium arboreum*

25 **1. Introduction**

26 Wildland fire is an important component of many ecosystems and has been used by humans for  
27 several thousand years (Crutzen and Goldammer, 1993; Pyne, 1997; Scott et al., 2014). Many  
28 North American ecosystems have evolved as a result of persistent fire (Barbour and Billings,

29 2000). The importance of fire in pine forests worldwide including the southern U.S. is well-known  
30 (Agee, 2000; Christensen, 2000). In the U.S., prescribed burning is used on approximately 8  
31 million ha annually to accomplish a variety of forestry and agricultural objectives (Melvin, 2015);  
32 the impact of smoke from these fires has been studied for over 50 years (Chi et al., 1979; Biswell,  
33 1989; Ward and Hardy, 1991; Hardy et al., 2001;). In the southern U.S., forest management  
34 objectives include hazardous fuel reduction, site preparation, improved wildlife habitat, insect and  
35 disease control, enhanced appearance and perpetuation of fire dependent species and natural  
36 communities (Carter and Foster, 2004; Waldrop and Goodrick, 2012). The U.S. Department of  
37 Defense (DoD) uses prescribed burning on approximately 243,000 ha annually for many of these  
38 objectives as well as maintenance of critical training areas (Cohen et al., 2014). Many land  
39 managers rely on fire behavior models to calculate fire movement on the landscape, energy release,  
40 smoke plume development, dispersion and content (Bytnerowicz et al., 2009, Paton-Walsh et al.  
41 2014). However, few fire behavior models account for the plethora of chemical reactions involved  
42 in the fire. The heat transfer processes that take place in the fire environment are also only coarsely  
43 described. In order to improve the use of prescribed burning to accomplish refined objectives,  
44 more detailed description and modeling of the physical and chemical processes in fire are needed  
45 (Cohen et al., 2014).

46 The chemical phases of wildland fire, described as preheating, flaming, smoldering and glowing  
47 (Ward, 2001) are understood in a chemical sense, but only to varying degrees: While the chemical  
48 effluents of flaming and smoldering phases have been characterized for many ecosystems and fuel  
49 types at different scales (Ward and Radke, 1993), the physics and chemistry of the preheating  
50 (pyrolysis) phase, have fewer studies beyond the bench scale (e.g. Depew et al., 1972;  
51 Dimitrakopoulos, 2001; Susott, 1982; Tihay, 2010). To improve fire application models and to  
52 accomplish the desired fire effects and limit potential fugitive emissions, improved understanding  
53 is thus needed for many fundamental processes, particularly for pyrolysis and ignition in  
54 heterogeneous fuel beds of live and dead fuels that reflect the diversity of vegetation found  
55 worldwide. (Guérette et al. 2018).

56 Prior to oxidative combustion, biomass thermally decomposes in a heated environment. To study  
57 this decomposition, thermogravimetric analysis has been applied to a small set of plant species  
58 deemed to represent major wildland fuel types (e.g. Burgan and Susott, 1991; Susott, 1982). Others

59 have determined caloric content of southern fuels which is related to the composition of pyrolysis  
60 products (Hough, 1969; Behm et al., 2004). However, most such prior work used dried, ground  
61 fuel samples in either an inert or oxidizing environment subject to uniform heating and heat  
62 transfer, (Kibet et al. 2012) thereby eliminating the effects of moisture and heat transfer which are  
63 key fire behavior variables. While pyrolysis and combustion of wildland fuels are known to be  
64 complex processes (Zhou and Mahalingam, 2001), they are often modeled using simple  
65 approximations in the relevant computer codes using the dominant gases of H<sub>2</sub>, CO, CO<sub>2</sub> and CH<sub>4</sub>.  
66 Heat transfer in a wildland setting is less efficient than in thermogravimetric analysis: The amount  
67 and composition of pyrolyzed species produced depend strongly on heating rate and temperature  
68 and typically consists of oxidized small-molecule gases such as CO, CO<sub>2</sub>, or H<sub>2</sub>O, as well as non-  
69 oxidized or partially oxidized species such as H<sub>2</sub>, CH<sub>4</sub>, C<sub>x</sub>H<sub>y</sub>, C<sub>x</sub>H<sub>y</sub>O<sub>z</sub> as well as tars. The products  
70 of primary pyrolysis may react in the gas phase at elevated temperatures (i.e., secondary pyrolysis),  
71 which may affect the amount of tar remaining.

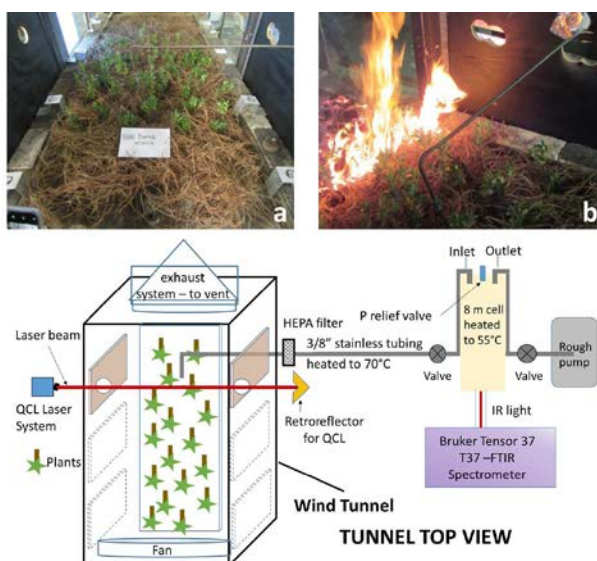
72 This work is part of a larger project to measure and model pyrolysis gases from common wildland  
73 fuels found on DoD installations in the southern United States (Weise et al. 2018). The project  
74 includes bench-level, laboratory-scale and field plot burns; integrating the results of the field and  
75 laboratory measurements with the modeling results to identify potential improvements that can  
76 enhance understanding of pyrolysis and ignition in wildland fuels. During the course of the project  
77 Fourier transform infrared (FTIR) methods have been used on several occasions to non-intrusively  
78 measure the composition and concentration of the pyrolysis gases including the gases liberated by:  
79 i) heating single leaf samples from several common southern fuels using different heating modes  
80 in a pyrolyzer and in a simple flat-flame burner system, (Amini et al., 2019; Safdari et al., 2020)  
81 ii) heating shrubs in prescribed burns at Ft. Jackson, South Carolina, (Scharko et al., 2019a, b) and  
82 iii) heating nursery plants with flames from longleaf pine needle fuel beds inside a wind tunnel  
83 (Aminfar et al., 2019). In order to achieve the goal that the results be applicable to prescribed  
84 burns, a key focus has been linking the bench scale, wind tunnel and field data to the models using  
85 realistic values and identities for the pyrolysis gases. Chemical analysis of the foliage and results  
86 of the bench scale tests so far suggest that describing wood pyrolysis may not be suitable for foliage  
87 fuels (live and dead) (Jolly et al., 2012; Jolly et al., 2016; Matt et al., 2020). To date, pyrolysis and  
88 ignition of wildland fuels have typically been based on results for only cellulose or wood (e.g.  
89 Varhegyi et al., 1994; Di Blasi, 2008). In this paper a small wind tunnel was used to bridge the

90 bench-scale studies to the field-scale pyrolysis measurements using a subset of the plant species  
91 from the bench-scale tests. The wind tunnel measurements were set to emulate the larger scale  
92 FTIR experiments using canister samples in 0.1 ha prescribed burns at Ft. Jackson in May 2018.  
93 The specific goal is to provide better temporal / flame phase resolution than provided by larger  
94 field studies such as the Ft. Jackson burns or from large chamber facilities such as the Fire Sciences  
95 Laboratory (FSL) in Montana (Yokelson et al., 1996, Burling et al., 2010); the FSL has long path  
96 optical cell coupled to an FTIR as well as many other powerful analytical methods such as proton-  
97 transfer mass spectrometry (Christian et al., 2004; Warneke et al., 2011; Yokelson et al., 2013)  
98 and has made first detections for dozens of chemical species and pioneered the science of biomass  
99 burning in many ways. But because the sampling platform is 4 m above the floor, there is mixing  
100 of gases from different phases such as volatilization and pyrolysis. The combustion and smoldering  
101 phases are typically easier to differentiate, primarily via the intrinsic diagnostic of the modified  
102 combustion efficiency (Ward and Hao, 1991), a measure that is not independent of the composition  
103 of smoke (Weise et al 2020). Similar ambiguities as to the nature of the phase of the fire also  
104 applies to extractive methods whereby a sampling device attempts to capture pre-combustion  
105 phase gases. Such sampling systems, typically connected to a field canister are effective but are  
106 subject to vagaries of sniffer gas inlet placement, i.e. proximity to the pyrolyzing plant. (Scharko  
107 et al. 2019a,b). Here we describe use of an FTIR with a probe to temporally isolate, identify and  
108 quantify some of the early-stage/pyrolysis gases from burns at a mid-scale laboratory facility.  
109 Experiments were conducted at the Riverside Fire Lab (RFL) in a wind tunnel using fuel beds  
110 composed of longleaf pine needles and the live plants fetterbush, inkberry, sparkleberry and  
111 blueberry. Multiple methods were used such as quantum-cascade lasers (Phillips et al., 2020), gas  
112 chromatography-mass spectrometry as well as FTIR with the overall objectives of: i) using careful  
113 chemometric extraction from the acquired data to see what pyrolysis species can be identified by  
114 the techniques; ii) using the various methods to determine the degree of oxidation or combustion,  
115 i.e. pyrolysis characterization; iii) making first attempts to quantify the rates of evolution of  
116 pyrolysis products for certain plant species; and ideally; iv) determining if differences exist  
117 between the pyrolysis emissions / temporal profiles for different plant species. We take advantage  
118 of the high resolution and time-resolved capabilities offered by IR spectroscopy and couple these  
119 to the flame/solid fuel temperature diagnostics of an IR camera to analyze the emissions from a  
120 series of RFL burns.

121 **2. Materials and Methods**

122 *2.1 Wind Tunnel and Experimental Configuration*

123 As part of the project a total of 88 laboratory scale burns were conducted at the USDA Forest  
124 Service Pacific Southwest Research Station in Riverside, California; this paper reports on the 21  
125 burns from November 2018. The Riverside laboratory includes a wind tunnel ca. 3 m long and 1  
126 m wide which was set up to simulate a forest floor of litter and live plants. Fuel beds composed of  
127 1 kg of dry longleaf pine needles and various combinations of inkberry (*Ilex glabra* (L.) A. Gray),  
128 fetterbush (*Lyonia lucida* (Lam.) K. Koch), sparkleberry (*Vaccinium arboreum* L.) and blueberry  
129 (*V. darrowii* Camp) were burned under either “no wind” or 1 m s<sup>-1</sup> wind conditions. Fuel moisture  
130 content and mass loading, ambient temperature and relative humidity in the tunnel were varied  
131 between experiments; fuel beds were ignited with a line fire which propagated the length of the  
132 fuel bed as seen in Fig. 1b. Multiple techniques were used to study the fire characteristics as well  
133 as the gas effluents: thermocouples, Schmidt-Boelter flux sensor, nadir thermal IR camera and  
134 background-oriented Schlieren photography (Aminfar et al 2019) to estimate heat transfer / air  
135 flow around the plants, canister samples analyzed by GC/FID, quantum cascade (QC) infrared  
136 laser spectroscopy, (Phillips et al., 2020) as well as broadband Fourier transform infrared (FTIR)  
137 spectroscopy. A schematic overview of the experimental setup is seen in Figure 1c.



138  
139 **Figure 1.** a) Overhead view of wind tunnel down its length with longleaf pine needles and interspersed  
140 inkberry plants; b) flame front progressing down the wind tunnel with FTIR extraction tube visible; c)  
141 cartoon (top view) of experimental layout with laser and FTIR systems.

142 Because the probe was inserted directly in the flame above the plants (Fig. 1b) and because  
 143 Teflon® melts at ca. 327 °C, the flame gas samples were pumped into the cell / FTIR instrument  
 144 via a stainless steel tube that was heated for those sections outside the wind tunnel. This has proven  
 145 effective at preventing adhesion for nearly all gases except amines (Scharko 2019a). Gas from  
 146 further probes was pumped into canisters for offline analysis using gas chromatography. Sixty-six  
 147 or seventy-four live plants were distributed within the longleaf pine needles in ceramic holders.  
 148 Figure 1a shows the configuration of the fuel bed with instrumentation for *in situ* analysis. Plant  
 149 species were prepared on site and samples of dry and live fuel were clipped to determine fuel  
 150 moisture content prior to each burn set. The experiments were set under varying fuel bed and  
 151 environmental conditions as summarized by Table 1 for the 21 experiments presented in this paper.

152 **Table 1.** Summary of burn schedule for November 2018 studies including burn number, date and time, fuel  
 153 description, acquisition method and spectral resolution used for wind tunnel experiments under 1 m s<sup>-1</sup> imposed air  
 154 flow. Geometric mean flame spread rate = 0.01 m s<sup>-1</sup>. The FTIR acquisition methods are described in the text.

155

<b>Burn number</b>	<b>Date (2018)</b>	<b>Local ignition time (PDT)</b>	<b>Local finish time (PDT)</b>	<b>Plant species</b>	<b>Acquisition method</b>	<b>Resolution (cm<sup>-1</sup>)</b>
76	30-Oct	11:48:01	11:52:00	inkberry	static	0.6
77	30-Oct	14:19:10	14:23:37	fetterbush	dynamic	2.0
78	30-Oct	15:12:30	15:16:33	sparkleberry	static	0.6
79	30-Oct	16:17:00	16:21:10	inkberry	dynamic	2.0
80	31-Oct	9:32:00	9:35:45	sparkleberry	static	0.6
81	31-Oct	10:35:00	10:38:52	fetterbush	dynamic	1.0
82	31-Oct	11:30:30	11:35:15	sparkleberry	static	0.6
83	31-Oct	13:19:00	13:22:58	inkberry	dynamic	1.0
84	31-Oct	14:12:15	14:16:30	fetterbush	static	0.6
85	31-Oct	15:30:30	15:34:24	fetterbush	dynamic	2.0
86	1-Nov	9:30:00	9:33:02	sparkleberry	dynamic	1.0
87	1-Nov	10:40:00	10:42:49	inkberry	dynamic	1.0
88	1-Nov	11:40:00	11:42:59	fetterbush	static	0.6
89	1-Nov	13:35:00	13:38:48	inkberry	static	0.6
90	1-Nov	14:45:00	14:49:47	sparkleberry	static	0.6
92	2-Nov	9:30:00	9:34:05	inkberry	dynamic	0.6
93	2-Nov	10:41:15	10:45:44	fetterbush	dynamic	1.0
94	2-Nov	11:28:15	11:32:28	sparkleberry	static	0.6
95	2-Nov	13:42:45	13:46:17	sparkleberry	static	0.6
97	2-Nov	15:38:38	15:41:40	sparkleberry	dynamic	0.6

156

157 *2.2 Instrumentation*

158 Gases were extracted from the burns via 3/8" stainless steel tubing, HEPA filtered to eliminate tar  
159 and char contamination and pumped into an 8-meter White cell (Bruker A136, 2.2 liter volume)<sup>1</sup>  
160 housed inside a Bruker Tensor 37 spectrometer (Figure 1c). The extractive probe was placed  
161 directly above a plant as close as possible to the foliage. To prevent analyte/tar condensation, both  
162 transfer tubing and the gas cell were heated to ~55 °C using heating tape/voltage regulator and a  
163 cell heating shroud, respectively. A thermocouple was suspended into the White cell to record the  
164 gas temperature for subsequent spectral analysis, with pressure gauge mounted atop the cell. Prior  
165 to data collection, the White cell was aligned using the FTIR's Ge/CaF<sub>2</sub> beamsplitter and tungsten  
166 lamp source. Once aligned, these were replaced with a Ge/KBr beamsplitter and mid-IR globar  
167 source, along with a mercury cadmium telluride detector, configuring the Tensor 37 to record  
168 spectral data from 7500 to 500 cm<sup>-1</sup>.

169 The FTIR system was tested for leaks, followed by a gas cell path length calibration using purified  
170 isopropyl alcohol (IPA - Sigma Aldrich 99.5%). Ten spectra with IPA pressures between 0.6 and  
171 10.5 Torr (0.8 -14.0 hPa) were recorded to 0.1 Torr accuracy using an MKS KF15 pressure  
172 transducer. The integrated area of the 3515-3290 cm<sup>-1</sup> spectral domain (Bruker OPUS 5.5  
173 software) along with recorded temperatures and pressures were used to create a Beer-Lambert plot  
174 (Scharko et al. 2019a). Using the integrals from the ten recorded spectra, the cell path length was  
175 determined to be 6.5 ± 0.2 m. Prior to spectral analysis (section 2.4), the infrared spectra were  
176 calibrated on the wavelength axis using a series of 30 water rotational-vibrational lines from the  
177 PNNL gas-phase database (Sharpe et al. 2002; Williams et al., 2013). FTIR interferograms were  
178 acquired using double-sided, forward-backward acquisition; these were apodized using a  
179 Blackman-Harris 3-Term function and phase corrected with Mertz's method prior to Fourier  
180 transformation. For both acquisition modes (static / dynamic), a single I<sub>0</sub> reference spectrum at the  
181 appropriate resolution was collected by flowing ambient gas into the cell at the start of each day  
182 to form the single (static) or multiple (dynamic) decadic absorbance spectra using Beer's law: -  
183 log<sub>10</sub>(I/I<sub>0</sub>). Acquiring such a blank or zero I<sub>0</sub> spectrum effectively accounts for any trace VOC  
184 emissions from the White cell, wind tunnel, tubing etc.

---

<sup>1</sup> The use of trade or firm names in this publication is for reader information and does not imply endorsement by the U.S. Department of Agriculture of any product or service.

### 185 2.3 Infrared spectral acquisition

186 Two data acquisition modes were used to analyze the burn gases: an extractive (or static) mode  
187 and a dynamic mode. In the extractive mode the gas flowing through the White cell was isolated  
188 for analysis; the inlet/outlet valves were simultaneously closed such that the emitted gases were  
189 isolated in the cell at a desired pressure, ca. 740-700 Torr (990-930 hPa) for high pressures, and  
190 430-400 Torr (570 - 530 hPa) for lower pressure measurements. The valves were closed just prior  
191 to the flame front reaching the probe, attempting to capture pre-combustion phases including  
192 evaporation and pyrolysis. The goal of the extractive mode was to obtain a higher fidelity  
193 “snapshot” for a given point in time of the burn; data were averaged longer at higher spectral  
194 resolution, allowing for detection of more gaseous species with higher sensitivity. (Scharcko et al.,  
195 2019a). The dynamic mode measurements recorded fewer scans at lower resolution to capture  
196 changing chemical identities/composition corresponding to different fire phases (pyrolysis,  
197 flaming combustion, smoldering combustion), achieving temporal resolutions of ca. 1 Hz.

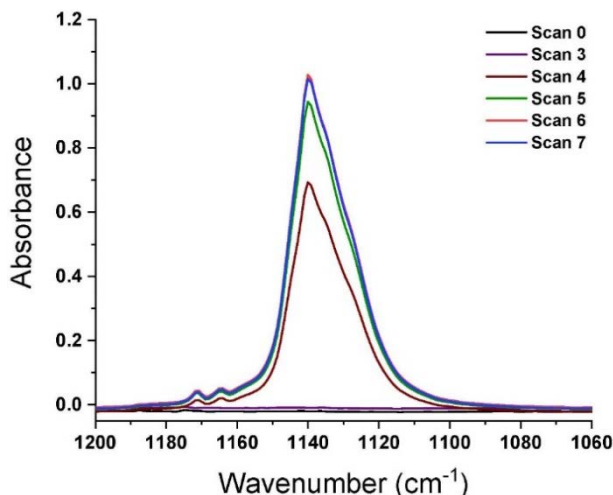
198 Of the 21 burns, 10 were recorded using the static method (Table 1). The static experiment spectra  
199 were recorded using the spectrometer’s  $0.6\text{ cm}^{-1}$  full resolution with an optical aperture of 2 mm.  
200 Interferograms were double sided with forward-backward acquisition. Due to the higher resolution  
201 and lower light throughput, multiple scans were averaged for a 30-min acquisition time, resulting  
202 in vastly improved signal/noise ratios (SNR). For analysis of such complicated gas-phase mixtures,  
203 infrared spectral resolutions of  $1\text{ cm}^{-1}$  or better have been demonstrated to be advantageous  
204 (Burling et al., 2010, Akagi et al., 2014, Scharcko et al., 2019a). While one goal was to isolate gases  
205 to include only the pyrolysis and pre-combustion phases, one vagary of the technique involved the  
206 timed closing of the valves relative to arrival of the flame front approaching the inlet. If the valves  
207 were shut too early, the captured emissions would consist of only (warmed) ambient gas before  
208 onset of thermal degradation of the solid fuel, as opposed to the desired pyrolysis phase.  
209 Conversely, if shut too late, flaming, or possibly even smoldering conditions would be sampled.

210 The second method was the dynamic mode whereby the OPUS software was used to continuously  
211 collect interferograms throughout the duration of the burn, capturing the chemical compositions  
212 associated with different phases, e.g. volatilization, heating, pyrolysis, flaming or smoldering  
213 combustion. Fourier transformation of the interferograms occurred after the burns to yield faster  
214 acquisition times. The dynamic acquisition mode was used in combination with thermal IR video



215 imaging recorded from above the flame bed, to help synchronize spectral acquisition to the various  
216 burn phases for a total of 11 burns. Instead of averaging for 30 minutes, the dynamic method  
217 allowed for 40-80 continuous interferometer scans (differing on the duration of the burn), and  
218 yielded a spectrum every 1.5 s for data taken at  $1.0\text{ cm}^{-1}$  resolution, every 0.79 s for data at  $2.0\text{ cm}^{-1}$   
219  $^1$  resolution, and every 2.5 s for  $0.6\text{ cm}^{-1}$  resolution spectra. Data acquisition began as the flame  
220 front encroached upon the extractive probe and continued until the flame had passed. Due to the  
221 faster acquisition rate these spectra are significantly noisier than the data collected using the  
222 extractive method. To compare results from the static and dynamic modes, fires 87 and 89 will be  
223 presented. The 2 m length fuel beds for both experiments 87 and 89 consisted of 1 kg longleaf pine  
224 needles with interspersed inkberry plants.

225 For time synchronization it was necessary to quantify the time lag from the time the emissions  
226 entered the extractive probe to midpoint in their flow through the White cell. A flow rate test was  
227 thus conducted using freon gas, CFC-11 (trichlorofluoromethane) which is comparable in  
228 molecular weight to the heavier gases detected by the FTIR. Figure 2 shows such a test of CFC-  
229 11 being introduced with spectra recorded every 0.79 s using  $2.0\text{ cm}^{-1}$  resolution. The time from  
230 introduction of the freon at the extractive probe ( $t=0$ , scan 0) to first appearance in scan 4 (maroon  
231 trace) was 3.2 seconds. The freon spectra had maximized at scan 6 (red trace) for a total  $\Delta t = 4.8$   
232 s lag from the probe to the instrument. With this information, FTIR time stamped data were then  
233 adjusted to reflect the 4.8 s delay which was used when correlating the spectral data to the visual  
234 and thermal IR video images.



235  
236 **Figure 2:** Dynamic spectra recording the introduction of CFC-11 from extractive probe to gas cell. Scan 0 (not  
237 shown) represents start of spectral acquisition/freon release near probe. Spectra were produced every 0.79 s.  
238 First freon observation occurs with scan 4; maximal absorbance of CFC-11 and stabilization occurs at scan 6.

## 239 2.4 Spectral Analysis

240 A combination of software was used for the post-acquisition spectral analysis and confirmation of  
241 the species observed during the campaign. The MALT5 software (Griffith, 2016) utilizing both  
242 HITRAN line-by-line data (Gordon et al., 2017) as well as the PNNL 50 °C gas-phase reference  
243 spectra (Kochanov, 2019; Johnson et al., 2006, 2010) as input libraries was used to identify and  
244 quantify vapor-phase chemicals in the spectra. Spectra were compiled into parameter files and  
245 analyzed by the MALT5 software using parameters including pressure, temperature, pathlength,  
246 resolution, as well as estimated initial values for chemical mixing ratios. The software generates a  
247 spectrum to simulate the measured spectrum, adjusting mixing ratios until the residual between  
248 the simulated and measured spectra is minimized. To confirm the species were actually present,  
249 each spectrum generated by MALT5 was input to OPUS and subtracted from the measured  
250 spectrum; the target compound was purposefully omitted from the subtraction process to visually  
251 inspect if the omitted compound was in fact present (see e.g. Figure 5).

## 252 3. Results and Discussion

### 253 3.1 Analysis of Static Spectra

254 Ten spectra were recorded from different burns using the static mode with the gas cell valves  
255 closed simultaneously; gases were sampled prior to arrival of the flame front. A total of 29  
256 compounds were detected and confirmed using MALT5 and OPUS 5.5. Along with CO, CO<sub>2</sub> and  
257 nitrogen compounds, the gas-phase species are largely lightweight hydrocarbons (HCs), volatile  
258 organic compounds (VOCs) and oxygenated volatile organic compounds (OVOCs). Table 2  
259 provides a summary of all compounds observed during the static measurements and is broken  
260 down into subcategories of chemical classes by rows labeled a-e, with ambient gases such as CO  
261 and CO<sub>2</sub> in group a, alkanes and alkenes in group b, alcohols, aldehydes and carboxylic acids in  
262 group c, aromatic species in group d, and N-bearing compounds in group e. The benefits of the *in*  
263 *situ* laboratory static measurements were controlled gas sample collection with FTIR analysis and  
264 longer scan times for increased SNRs at higher spectral resolution. Valves were shut before the  
265 flame front arrived, allowing for minimal mixing of air and flame gases near the extractive probe.  
266 In this manner the targeted pyrolysis phase was likely to be sampled with a greater mole fraction  
267 rather than that of the combustion phase. The gases listed in Table 2 have previously been  
268 observed in smoke in either field or laboratory settings, and some of them have been linked to

269 pyrolysis (Scharko, 2019a,b; Burling et al., 2010, 2011; Christian et al., 2003, 2004; Gilman et  
270 al., 2015; Goode et al., 1999, 2000; Hatch et al., 2017; Selimovic et al., 2018; Stockwell et al.,  
271 2014; Yokelson et al., 1996, 1997; Akagi et al., 2013, 2014; Alves et al., 2010; Hurst et al., 1994a,  
272 b; Karl et al., 2007; Paton-Walsh et al., 2010). Compounds associated with the pyrolysis phase  
273 and observed in several of the static measurements include acetic acid, ethene (C<sub>2</sub>H<sub>4</sub>), allene, 1,3-  
274 butadiene, acetaldehyde, formic acid, formaldehyde, acrolein, benzene, furan, furaldehyde,  
275 naphthalene and phenol.

276 As seen in Table 2, ammonia gas (NH<sub>3</sub>) was also detected at fairly low mixing ratios in the  
277 laboratory scale experiments, which had previously not been detected in the Ft. Jackson field  
278 study: The lack of NH<sub>3</sub> detection in those studies was ascribed to the known adsorptivity of the  
279 compound as it may have adhered to either the transfer canister walls, the extractive probe, or the  
280 White Cell, all at ambient temperatures as used in those studies (Scharko et al., 2019; Roscioli et  
281 al., 2015; Stockwell et al., 2014; Yokelson et al., 2003; Neuman et al., 1999). Adhesion losses  
282 were minimized in the present experiments by a) measuring the gas parcel directly without storage  
283 and b) heating transfer lines and gas cell to ~55 °C.

284 **Table 2.** Mixing ratio of chemicals from spectra collected using the static acquisition method. Burns are labeled by  
285 number and plant species. Mixing ratios are reported in ppm (with the exception of H<sub>2</sub>O and CO<sub>2</sub> reported as  
286 percents) and categorized by (a) background ambient compounds, (b) simple hydrocarbons, (c) oxygenated organic  
287 compounds, (d) aromatics and furans, and (e) N-bearing species.

	Burn 76	Burn 78	Burn 80	Burn 82	Burn 84	Burn 88	Burn 89	Burn 90	Burn 94	Burn 95	
	inkberry	sparkleberry	sparkleberry	sparkleberry	fetterbush	fetterbush	inkberry	sparkleberry	sparkleberry	sparkleberry	
<b>a</b>	% H <sub>2</sub> O	1.24	1.05	3.23	2.03	3.54	3.08	3.46	1.82	6.21	4.10
	% CO <sub>2</sub>	0.06	0.09	2.06	0.48	1.51	2.06	1.36	0.34	4.60	2.08
	CO	1.45	3.90	808	192	1089	1057	391	160	7506	2651
	N <sub>2</sub> O	0.35	0.34	1.21	0.50	1.28	1.79	0.44	0.41	3.22	1.78
<b>b</b>	CH <sub>4</sub>	2.27	2.21	45.3	10.7	54.5	50.3	15.4	11.3	682	198
	C <sub>2</sub> H <sub>2</sub>	0.01	0.06	23.8	4.52	23.4	23.2	8.82	5.62	351	96.5
	C <sub>2</sub> H <sub>4</sub>	0.07	0.05	29.3	7.05	39.9	39.3	9.66	6.52	452	133
	C <sub>2</sub> H <sub>6</sub>			0.83				2.76	4.0E-04	24.2	6.29
	C <sub>3</sub> H <sub>6</sub>			4.02	0.99	5.55	5.48	0.75	0.77	61.3	18.1
	allene	0.17		0.64	0.29	1.12	1.21	0.25	0.12	8.69	2.30
	1,3-butadiene			1.63	0.37	1.98	2.07	0.26	0.43	28.1	7.57
	isobutene			0.75		0.74	0.52			3.16	1.07
	isoprene			1.78	0.39	1.72	1.43	0.31	0.32	11.7	4.22
<b>c</b>	CH <sub>3</sub> OH	0.89	0.24	6.81	1.53	6.92	9.44	1.66	0.93	42.3	18.0
	C <sub>2</sub> H <sub>5</sub> OH	1.37									
	acetic acid	0.07		5.93	3.55	13.4	13.8	11.0	2.49	13.4	9.62
	formic acid			15.9	5.14	32.35	35.3	9.20	3.64	130	73.6
	acetaldehyde			5.87	1.69	7.62	8.65	1.51	0.94	73.6	22.6
	acrolein			2.59	1.29	3.99	4.35	0.98	0.00	26.0	9.53
	crotonaldehyde			1.51	0.54			0.73	0.17	9.97	5.64
	formaldehyde		0.08	13.6	4.31	21.3	22.5	5.41	3.33	114	52.8
<b>d</b>	benzene			4.08	2.23	5.19	4.24	1.93	1.48	61.3	18.6
	furan			0.75		0.39	0.54			3.07	1.16
	furfural			0.65	0.06				0.13	3.34	1.24
	naphthalene			4.48	1.06	3.60	4.80	3.40	0.82	14.6	1.42
	phenol			0.90	0.30	1.36	1.63	1.75	0.37	2.19	1.63
<b>e</b>	NH <sub>3</sub>	0.10	0.29	0.19	1.29	1.79	0.88	1.08	0.41	0.58	0.57
	HCN			5.84	2.19	8.25	6.94	3.36	1.69	64.2	21.0
	HNCO			1.89	0.67	2.61	2.94	1.27	0.70	5.37	1.98
	HONO		0.11	9.40	2.53	9.72	12.7	8.92	1.75	26.9	11.3

288

289

290 When comparing the RFL laboratory scale experiments to the 2018 Ft. Jackson field scale  
291 experiments (Table 3), it is evident that field scale values via the static mode are greater than those  
292 of the laboratory, even though the laboratory experiment attempted to replicate Ft. Jackson fuel  
293 beds and scenarios. In most cases, a comparison of compounds found in the RFL laboratory burns  
294 and the Ft. Jackson 2018 field burns finds Ft. Jackson mixing ratios approximately 4 to 10 times  
295 greater than those of the RFL 2018 tunnel data. Field scale measurements typically yield more  
296 emissions than experiments conducted in the laboratory due to larger fuel quantities (Yokelson,  
297 2013; Scharko, 2019b, Weise et al. 2015). However, while the mixing ratios may differ or be  
298 larger/smaller, the information describing the composition of the mixture is relative in nature and  
299 is contained in log-ratios of the various gases. Analysis of the data as compositional data (Aitchison  
300 1986), however, is beyond the scope of the present paper. Table 3 displays the minimum and  
301 maximum mixing ratio values in ppm for five compounds from the Ft. Jackson studies presented  
302 in Scharko et al. (2019a) versus the present RFL laboratory results. Of the five species compared,  
303 acetaldehyde, acrolein, and allene all follow the trend of having Ft. Jackson results being

304 significantly higher than the RFL studies by a factor of ~4. Naphthalene, a polycyclic aromatic  
 305 hydrocarbon (PAH) was the only exception to this trend, having comparable mixing ratio values  
 306 in the two studies. This anomaly could be attributed to one of naphthalene's pyrolysis formation  
 307 route as suggested by Fairburn et al., where a single ringed aromatic compound undergoes a Diels-  
 308 Alder reaction of an alkene (Fairburn et al., 1990; Liu et al., 2017). Of the four compounds  
 309 compared, naphthalene is the only one to be derived from a secondary reaction, whereas  
 310 acetaldehyde and acrolein are derived directly from the pyrolysis of cellulose (Stein et al., 1983),  
 311 while allene is a compound known to be a precursor of aromatic compounds and soot (Frenklach  
 312 et al., 1983, 1988). As noted, most compounds detected in the RFL laboratory studies yielded ~4  
 313 to 10x lower mixing ratios compared to the field scale studies at Ft. Jackson. Along with  
 314 naphthalene, however, acetic acid, formaldehyde, isoprene and isobutene were also found to have  
 315 comparable mixing ratios to those reported in the Ft. Jackson studies. This could be due to the  
 316 four compounds being products of secondary reactions, or fragmentation, of species such as lignin,  
 317 xylan and glucomannan (Collard and Blin, 2014). It should be noted that of the five novel  
 318 compounds detected in Scharko et al. (2019b), only four were detected in these laboratory scale  
 319 experiments. Methyl nitrite was not observed (Table 3). This is attributed to the field experiment  
 320 being on the Ft. Jackson base, where there is known to be unexploded ordinance (Scharko et al.,  
 321 2019b) or possibly due to lower concentration levels that are below the detection limits of the  
 322 present laboratory-scale experiment.

323 **Table 3.** Calculated minimum and maximum mixing ratios (ppm) for the 10 canister measurements taken at the  
 324 Ft. Jackson field measurements (Scharko et al., 2019) along with the minimum and maximum mixing ratios  
 325 (ppm) for the 10 static measurements during the RFL laboratory experiment of acetaldehyde, acrolein, allene,  
 326 methyl nitrite and naphthalene.

Target Compound	Ft. Jackson mixing ratio (ppm)		RFL mixing ratio (ppm)	
	min	max	min	max
acetaldehyde	34.5	264.8	0.9	73.6
acrolein	14.7	125.7	1.0	26.0
allene	2.2	37.8	0.1	8.7
methyl nitrite	2.3	21.0	-	-
naphthalene	1.4	19.9	0.9	14.6

327  
 328 It is clear that the static method as deployed was not perfect at either isolating strictly the pyrolysis  
 329 phase gases or capturing extremely high fractions of combustion gases. The method relied heavily

330 on valves being closed prior to the flame front, typically using visual cues as opposed to using  
331 other techniques, e.g. thermal IR detection. While not readily visible to the human eye, radiant  
332 and convective heating (as determined by Background-Oriented Schlieren measurements -  
333 Aminfar et al 2019) occurred well in advance of the flame front, suggesting this as a possible  
334 alternate visualization of pyrolysis gas release (Aminfar 2019). In any case, there is a narrow  
335 temporal window for the pre-combustion phase making the valve-close time extremely important.  
336 For example, spectra from burns 76 and 78 show largely the detection of only ambient air  
337 compounds indicating the valves were closed too early. Conversely, in other samples there is  
338 clearly some mixing of both upstream and downstream air before the gas enters the extractive  
339 probe. Despite the shortcomings of the static method, most of the attempts to obtain pre-  
340 combustion gases were successful as evidenced in part by the chemical composition of the isolated  
341 gases.

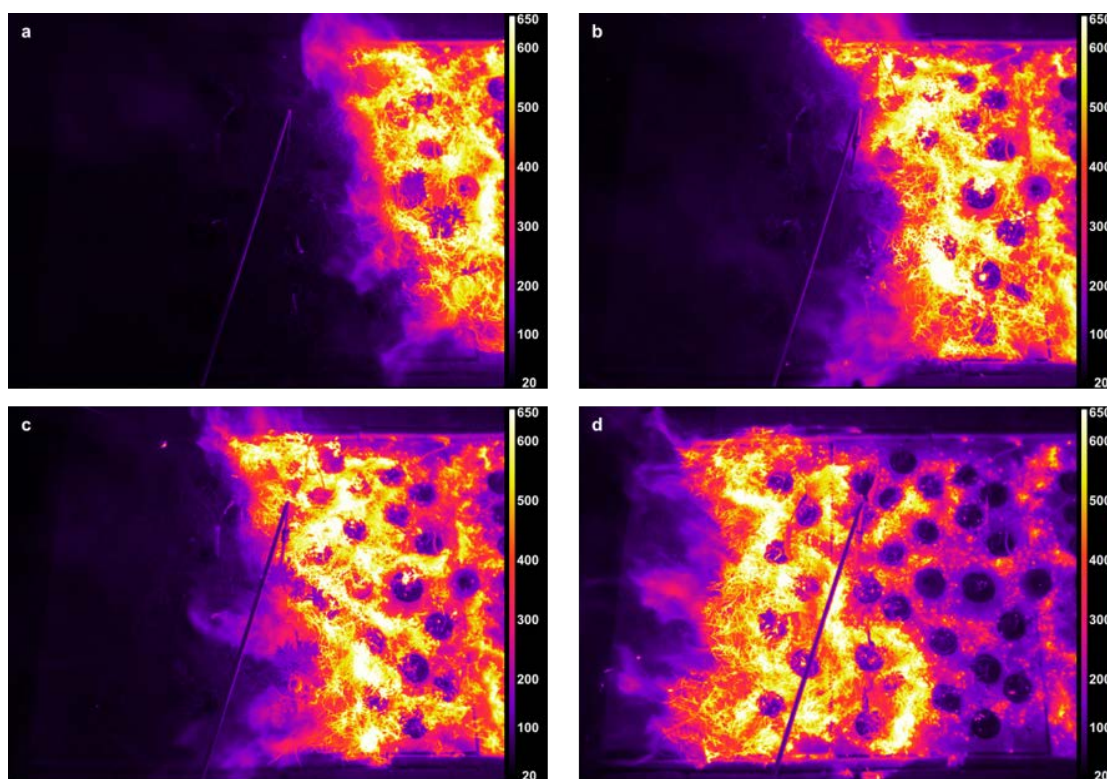
342

### 343 *3.2 Spectral-Thermal Correlation to Isolate Pyrolysis Phase*

344 Dynamic IR data and visual image acquisition proved advantageous to resolve the different phases  
345 of the experiments (e.g. pyrolysis, flaming combustion, smoldering combustion). This was  
346 important since the modified combustion efficiency (MCE) is a function of the gas composition  
347 and is not unique to phase, i.e. the same value of MCE results if the same relative amounts of CO  
348 and CO<sub>2</sub> are observed, whether in the pyrolysis, flaming or smoldering combustion phases. MCE,  
349 defined as  $\Delta\text{CO}_2/(\Delta\text{CO} + \Delta\text{CO}_2)$ , has many times been used to distinguish phases of combustion,  
350 namely flaming vs. smoldering although Ward and Radke (1993) recommended combustion  
351 efficiency as the preferred descriptor of the combustion system. MCE has not been used to identify  
352 pyrolysis nor should it be for the non-uniqueness described previously. Recent studies have  
353 introduced more sophisticated techniques to analyze smoke emissions data with compositional  
354 data methods (Weise et al. 2020). However, since primary and secondary pyrolysis occurs both  
355 prior to and after the onset of combustion or oxidation, methods such as the MCE are not  
356 appropriate. We were not able to use the metric suggested by Sekimoto et al. (2018), namely high  
357 temperature vs. low temperature pyrolysis as determined from the acetylene-to-furan ratio owing  
358 to weak furan signals in the present study due to shortened scan times. The analysis was further  
359 exacerbated because furan's strongest vibrational band, the  $\nu_{19}$  vibrational band near 745 cm<sup>-1</sup>,

360 corresponding to the C-H out-of-plane bend (Shimanouchi, 1972), was obscured by saturated  
361 carbon dioxide lines and thus MALT was not able to generate a satisfactory fit for this  
362 microwindow.

363 The pre-flame arrival gases were identified by either of two methods: The first method involved a  
364 simple time subtraction of 4.8 s from the recording of the infrared spectrum time stamp, and  
365 associating that time to the corresponding visual and thermal infrared video images (Fig. 3). This  
366 provided a relatively accurate verification that the gases being investigated were emitted prior to  
367 the onset of combustion as seen in Table 4. The second method used the FTIR spectra directly:  
368 demarcations for the flame front were denoted by the maximal value obtained for both CO and  
369 CO<sub>2</sub> concentrations, i.e. greatest fraction of gas from the combustion phase. From this value the  
370 FTIR scans were selected for pyrolysis corresponding to the ~10 seconds before arrival (~0.1 m  
371 distance) of the flame front.



372

373 **Figure 3.** Burn 87 inkberry on longleaf pine needle fuel bed – FLIR thermal imaging for burn progression.  
374 (a) frame corresponding to FTIR scan 5 signaling the pre-combustion phase, (b) frame corresponding to  
375 FTIR scan 16, flame front nearing sample probe, (c) frame corresponding to FTIR scan 21, inkberry bush  
376 consumed by flame, (d) frame corresponding to FTIR scan 44, flame front has passed the probe. Dark  
377 circles are ceramic plant holders. The temperature scale is seen on the right axis.

378 The FTIR time-resolved scans (including derived chemical mixing ratios) synchronized to the RFL  
379 time-stamped thermal IR temperature images provide insight into the chemical composition of  
380 each burn. As an example, Table 4 pairs data from the two systems for Burn 87. FTIR scan number,  
381 FTIR time stamp, RFL FLIR recorded temperature near the extractive probe, and a selection of  
382 chemical concentrations are shown. The table demonstrates that spectral data for FTIR scans 0-8  
383 saw no significant detections above ambient levels as corroborated by the FLIR images displaying  
384 temperatures ranging from 40 to 80 °C (see Table 4 and Figure 3a); the extractive probe is still in  
385 the low temperature region. The gradual increase in mixing ratios for most compounds (excluding  
386 ammonia, which is primarily a smoldering gas) begins after FTIR scan 9. The magenta and orange  
387 colored domains seen in Figure 3b indicate the encroaching flame front and a rise in thermal  
388 temperatures. The frames corresponding to FTIR scans 16-19 display IR temperatures between  
389 175 and 220 °C. In this temperature range compounds associated with the pyrolysis phase such as  
390 acetaldehyde, acetic acid and allene (shown in Table 4) are not only manifest in the IR spectra, but  
391 their mixing ratios rise rapidly. Shortly thereafter the greatest mixing ratios of CO<sub>2</sub> occur at scans  
392 20 through 22, indicating the flaming stage; this is corroborated by thermal IR video of the inkberry  
393 plant beginning to be fully consumed in flames (Fig 3c). As the flame front progressed down the  
394 tunnel, temperatures near the plant holder began to drop with the onset of the smoldering phase as  
395 indicated by lower mixing ratios as well as the thermal IR visual, seen in Fig. 3d. [We note in  
396 Figure 3 that the temperature directly near/above the holders is much cooler due to minimal duff  
397 cover and the plants being green.] The video stopped recording at scan 48, when the flame reached  
398 the end of the fuel bed although the FTIR continued to collect interferograms to monitor  
399 smoldering from the fire.

400

401 **Table 4:** Burn 87 inkberry amongst pine needle fuel bed FTIR scan summary synchronized to FLIR temperature data.  
402 Scan number, FTIR time stamp, along with FLIR video emissions temperature at extractive probe accounting for time  
403 delay and mixing ratios from carbon dioxide (CO<sub>2</sub>), carbon monoxide (CO), ethene (C<sub>2</sub>H<sub>4</sub>), acetic acid (CH<sub>3</sub>COOH),  
404 formaldehyde (HCHO), acetaldehyde (CH<sub>3</sub>CHO), and phenol (C<sub>6</sub>H<sub>6</sub>O).

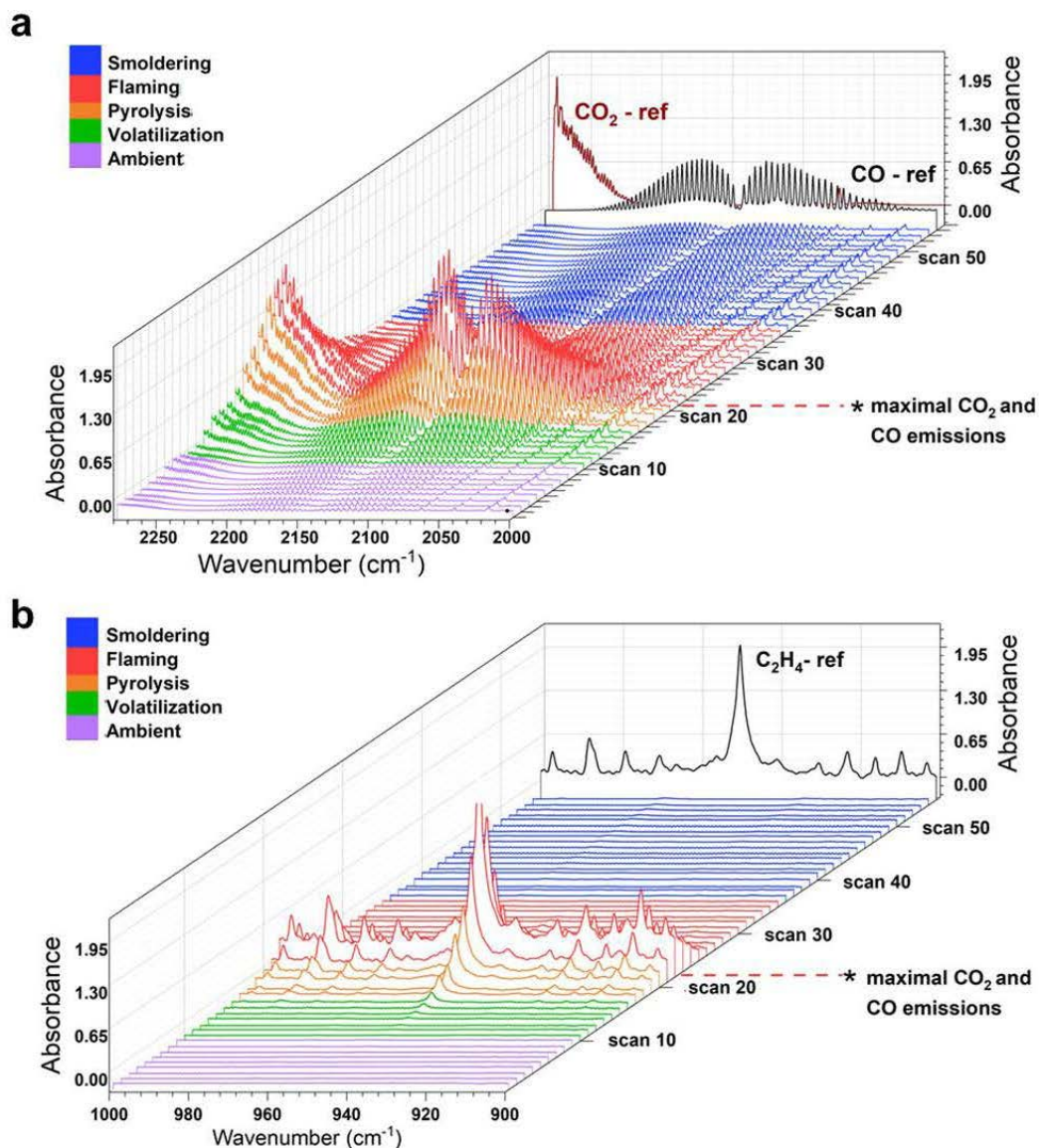


	FTIR Scan Number	FTIR time stamp	4.8 s Delayed	FLIR		CO <sub>2</sub> (ppm)	CO (ppm)	C <sub>2</sub> H <sub>4</sub> (ppm)	CH <sub>3</sub> COOH (ppm)	HCHO (ppm)	CH <sub>3</sub> CHO (ppm)	C <sub>6</sub> H <sub>6</sub> O (ppm)
			FLIR Video Time Stamp	temperature at inlet (°C)								
<b>Ambient</b>	scan 0	10:41:23.69	10:41:18.89	41.4	1548	49.5	2.4	0.0	0.6	-	-	
	scan 1	10:41:25.20	10:41:20.40	44.2	1912	74.5	1.9	1.7	0.9	-	-	
	scan 2	10:41:26.70	10:41:21.90	51.0	1562	63.7	2.3	1.1	0.3	-	-	
	scan 3	10:41:28.21	10:41:23.41	54.9	1290	58.4	1.1	0.5	0.5	-	-	
	scan 4	10:41:29.71	10:41:24.91	53.4	1882	63.1	1.2	1.2	0.8	-	-	
	scan 5	10:41:31.21	10:41:26.41	63.2	1946	57.0	1.5	1.1	0.4	-	-	
	scan 6	10:41:32.72	10:41:27.92	72.8	3811	102	4.7	1.8	1.2	-	-	
	scan 7	10:41:34.22	10:41:29.42	77.6	4722	138	2.7	1.9	1.5	-	-	
scan 8	10:41:35.73	10:41:30.93	86.2	3553	109	0.9	2.2	1.4	-	-		
<b>Volatilization + Pyrolysis</b>	scan 9	10:41:37.23	10:41:32.43	116.9	2957	97	1.0	2.4	0.7	-	-	
	scan 10	10:41:38.73	10:41:33.93	132.1	3360	138	1.8	2.8	1.2	1.4	-	
	scan 11	10:41:40.24	10:41:35.44	115.7	7476	246	4.2	2.6	3.6	0.0	-	
	scan 12	10:41:41.74	10:41:36.94	121.3	10274	291	6.2	9.0	4.8	4.2	0.7	
	scan 13	10:41:43.25	10:41:38.45	183.3	10890	391	11.4	11.5	7.1	1.1	1.6	
	scan 14	10:41:44.75	10:41:39.95	159.4	11833	635	23.5	14.5	13.5	3.3	1.5	
	scan 15	10:41:46.25	10:41:41.45	163.8	16080	1214	48.5	16.8	28.7	7.2	1.5	
	scan 16	10:41:47.76	10:41:42.96	176.9	25757	2217	95.3	16.3	54.4	17.6	1.0	
	scan 17	10:41:49.26	10:41:44.46	181.3	31856	2915	129	14.4	70.5	24.0	1.4	
	scan 18	10:41:50.77	10:41:45.97	220.1	41291	2878	260	13.1	92.6	34.5	1.4	
	scan 19	10:41:52.27	10:41:47.47	219.0	61166	8228	435	12.3	121	44.3	2.1	
<b>Flaming Combustion</b>	scan 20	10:41:53.77	10:41:48.97	296.1	79332	11354	747	13.1	178	80.7	2.9	
	scan 21	10:41:55.28	10:41:50.48	261.7	54381	9729	1167	17.6	255	140	4.7	
	scan 22	10:41:56.78	10:41:51.98	456.1	64077	12954	1025	15.3	185	103	6.0	
	scan 23	10:41:58.29	10:41:53.49	429.9	41495	8620	530	15.4	123	63.3	6.3	
	scan 24	10:41:59.79	10:41:54.99	516.5	25879	3453	257	15.5	74.5	33.9	6.8	
	scan 25	10:42:01.29	10:41:56.49	514.5	15965	3110	116	15.1	45.7	19.6	6.8	
	scan 26	10:42:02.80	10:41:58.00	460.1	11819	2416	53.2	15.6	35.2	12.0	5.6	
	scan 27	10:42:04.30	10:41:59.50	453.7	8566	1875	36.6	14.7	27.4	6.9	5.2	
	scan 28	10:42:05.81	10:42:01.01	448.0	5795	1320	14.5	13.2	21.8	3.0	5.6	
	scan 29	10:42:07.31	10:42:02.51	440.2	5235	1302	11.0	14.6	20.0	5.1	4.5	
	scan 30	10:42:08.81	10:42:04.01	484.7	3626	916	5.8	15.1	15.0	5.2	4.2	
	scan 31	10:42:10.32	10:42:05.52	470.4	2368	570	3.3	11.5	11.1	2.7	3.8	
	scan 32	10:42:11.82	10:42:07.02	497.5	1636	377	1.6	10.8	9.4	0.3	3.7	
	scan 33	10:42:13.33	10:42:08.53	477.4	1684	399	1.0	9.2	8.8	2.1	3.4	
	scan 34	10:42:14.83	10:42:10.03	450.1	1986	519	0.9	10.4	9.5	-1.5	3.0	
<b>Smoldering Combustion</b>	scan 35	10:42:16.33	10:42:11.53	397.9	1968	518	0.7	9.9	9.3	2.8	3.2	
	scan 36	10:42:17.84	10:42:13.04	410.2	1901	495	1.4	9.6	8.8	0.9	2.8	
	scan 37	10:42:19.34	10:42:14.54	401.6	1936	516	1.5	9.2	9.4	0.0	2.7	
	scan 38	10:42:20.85	10:42:16.05	358.3	1935	513	1.6	9.0	9.5	1.1	2.8	
	scan 39	10:42:22.35	10:42:17.55	341.8	1753	439	1.8	9.7	8.9	1.1	2.3	
	scan 40	10:42:23.85	10:42:19.05	320.6	1438	345	1.4	10.4	8.3	-0.5	2.9	
	scan 41	10:42:25.36	10:42:20.56	305.9	1224	277	-0.1	10.9	7.4	1.4	2.8	
	scan 42	10:42:26.86	10:42:22.06	295.0	1377	324	1.4	11.6	8.2	-2.0	2.4	
	scan 43	10:42:28.37	10:42:23.57	272.5	1629	411	1.1	11.8	8.4	0.3	2.4	
	scan 44	10:42:29.87	10:42:25.07	258.3	1366	325	0.8	12.8	7.8	2.5	2.7	
	scan 45	10:42:31.37	10:42:26.57	260.1	1059	212	0.4	11.9	6.7	3.4	2.7	
	scan 46	10:42:32.88	10:42:28.08	238.5	1037	212	0.8	13.1	7.1	0.5	2.3	
	scan 47	10:42:34.38	10:42:29.58	223.9	1094	236	0.9	14.4	6.7	1.1	1.2	
	scan 48	10:42:35.89	10:42:31.09	226.7	1117	248	1.2	13.1	7.0	-0.4	2.7	

405  
406 As stated, a second method was also used to analyze/corroborate the different stages of the burn,  
407 whereby mixing ratios of CO<sub>2</sub>, CO and C<sub>2</sub>H<sub>4</sub> were analyzed to find their burn maxima (Viatte et  
408 al., 2015). The CO<sub>2</sub> elevated mixing ratios (esp. relative to CO) are associated with the hottest,  
409 flaming stage of biomass burns (Yokelson et al., 1996). To temporally isolate the flaming stage,  
410 the MCE criteria was employed and values of 88-95% indicative of flaming were found for the

411 region. Having identified the flaming stage, the pyrolysis stage was estimated by subtracting 6-8  
412 seconds from that spectrum with maximal CO/CO<sub>2</sub> emissions, corresponding to ~4 FTIR scans (at  
413 1 cm<sup>-1</sup> resolution). The agreement between the two methods was quite good and helped to  
414 demarcate the stages as seen in Table 4.

415  
416 Figure 4 displays the infrared spectral progression of Burn 87, longleaf pine needles with inkberry,  
417 bed at 1.0 cm<sup>-1</sup> resolution looking at two different spectral regions. The CO (and CO<sub>2</sub>) profiles are  
418 seen in Fig. 4a. Noted on the z-axis is scan 22; scans 20-22 are the time frames where maximal  
419 CO<sub>2</sub> and CO emissions were observed; the region is also denoted by red spectral traces. Once the  
420 flaming stage had been identified, the pyrolysis phase was then demarcated; in the pyrolysis phase  
421 CO was evident (partially from upwind mixing) and was beginning to significantly increase; the  
422 stage is indicated by orange traces (scans 16-19) in Figure 4. Other stages assigned were noted as  
423 the pre-flame stage where ΔCO and ΔCO<sub>2</sub> were near zero in the FTIR data and are seen as scans  
424 0-8 with purple traces. Blue traces correspond to the smoldering phase of combustion, where CO<sub>2</sub>  
425 mixing ratios decreased, the flame front had passed the extractive probe and MCE values were on  
426 the order of 85-75%. The spectral profile and mixing ratios of ethene (C<sub>2</sub>H<sub>4</sub>) were also used to  
427 evaluate the time-resolved FTIR data (Johnson et al. 1993). This lightweight hydrocarbon is a  
428 product of primary pyrolysis and, if detected, can be used to determine certain stages of the burn  
429 (e.g. Yang et al., 2007). Figure 4b displays primarily the ν<sub>7</sub> band of ethene at 949.4 cm<sup>-1</sup>  
430 (Shimanouchi, 1972). Ethene reached its maxima mixing ratio at scan 22 (red traces) before it  
431 quickly disappeared, being a pyrolysis gas that was oxidized by the flame. It was first seen to  
432 appear as early as scan 13 (green traces) but became clearly evident in scan 16 (orange traces,  
433 pyrolysis phase) and continued to grow. The rapid disappearance of C<sub>2</sub>H<sub>4</sub> upon combustion is  
434 similar to that of formaldehyde and acetaldehyde (Table 4) whose concentrations also dropped  
435 after scan 23, but the disappearance is juxtaposed with acetic acid whose values remained  
436 approximately constant throughout the flaming and smoldering phases. As seen in the IR data, the  
437 C<sub>2</sub>H<sub>4</sub> gas signal corroborated that ethene is a key product of the primary pyrolysis phase. Other  
438 compounds showing significant signals in this time domain and described as pyrolysis gases  
439 include acrolein and allene. (Scharko 2019a; Akagi et al., 2013; Frenklach et al. 1983, 1988; Stein  
440 et al., 1983; Koss et al., 2018; Brilli et al., 2014).



441  
 442 **Figure 4.** Burn 87, inkberry with longleaf pine needles: a) CO and CO<sub>2</sub> spectral profile from 2250-2000 cm<sup>-1</sup>.  
 443 Purple traces indicate the ambient stage, green and orange traces indicate the pre-combustion/pyrolysis stage,  
 444 red spectral traces indicate the flaming stage and blue traces indicate smoldering. b) Largely C<sub>2</sub>H<sub>4</sub> spectral  
 445 waterfall plot from 1000-900 cm<sup>-1</sup> with accompanying C<sub>2</sub>H<sub>4</sub> reference spectrum as black trace.

446 The two methods to determine the pyrolysis, flaming, combustion and smoldering phases yielded  
 447 congruent results: The isolated burn stages determined from method one, in which FTIR gas-phase  
 448 data were synchronized to the FLIR thermal imaging, and from method two, using the FTIR time-  
 449 resolved data only, were found to be virtually identical. This is evidenced by linking the scans  
 450 determined to be in the pyrolysis phase (scans 16-19) using method two as seen in Fig. 4, with the  
 451 temperature data recorded by the FLIR using method one and seen in Table 4. For these scans,

452 the temperature profile ranges from 175-220 °C, corresponding to temperatures associated with  
453 first stages of pyrolysis.

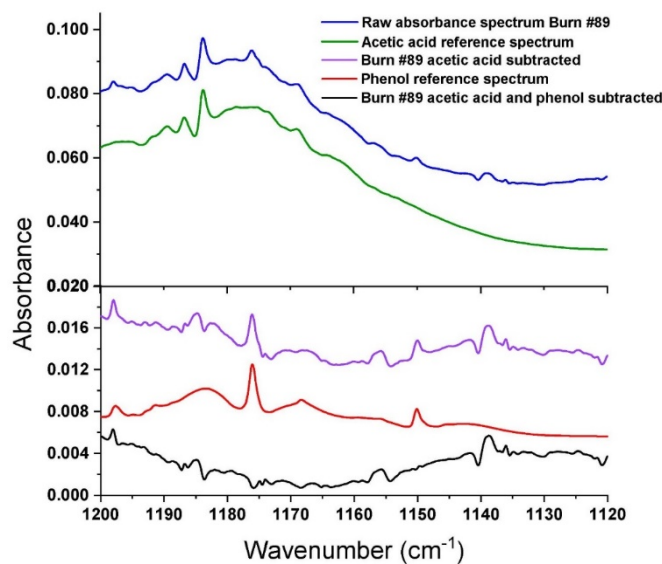
454 For most analytes' biomass burning gas mixing ratios, the concentration values observed at the  
455 peak of the dynamic measurements values were significantly greater than for the concentrations  
456 recorded in the static measurements. The dynamic experiments were of course carried out for the  
457 duration of the burn, whereas the static burns (in an effort to characterize pre-combustion phases)  
458 attempted to isolate a specific time when the pyrolyzate concentrations were maximized.  
459 Analyzing the data using the dynamic technique allowed for confirmation of certain compounds  
460 such as naphthalene, allene, acetaldehyde, and acrolein as compounds that appeared during the  
461 pyrolysis phase. These compounds, which have been previously detected as pyrolysis gases using  
462 FTIR for field plot burns, (Scharcko et al. 2019) were again observed during these laboratory scale  
463 tests and in almost all cases appeared before the flame front encroached on the sampling probe.

### 464 *3.3 Dynamic Detection of Phenol in the Pre-combustion Through Smoldering Stages*

465 In the present study phenol (C<sub>6</sub>H<sub>6</sub>O) was detected during several burns; its origin ascribed to the  
466 pyrolysis of lignin(s) (Kibet et al. 2012, Hawthorne et al. 1989) and has been mostly observed  
467 using other techniques such as gas chromatography mass spectrometry (GC-MS) (Saiz-Jimenez et  
468 al. 1986). Phenol and phenolic compounds are also known to contribute to the formation of  
469 secondary organic aerosols (Yee et al., 2013). It has been observed in simple pyrolysis  
470 experiments, emanating from both pine and spruce species (e.g. Saiz-Jimenez and De Leeuw,  
471 1986, Ingemarsson et al., 1998). In addition to simply pyrolytic emissions, phenol has also been  
472 identified as a common component of tar as a pyrolysis product. In biomass burning, phenol has  
473 been observed using both FTIR and other methods, (Gilman et al. 2015, Yokelson et al. 2013), e.g.  
474 proton-transfer mass spectrometry (PTR-MS) and GC-MS. In 2013 phenol was detected in a  
475 closed cell, airborne FTIR field experiment but not in an open-path FTIR lab experiment  
476 (Yokelson et al. 2013). The absence of C<sub>6</sub>H<sub>6</sub>O in the lab experiment was attributed to the lack of  
477 consumption of rotten wood as fuel. In those studies, airborne phenol emissions measured in the  
478 field with closed-cell FTIR were also noted as being 2 to 4x greater than the phenol emissions  
479 captured by PTR-MS in the laboratory.

480 Compositional analysis of inkberry, fetterbush, and longleaf pine needles used in the present study  
481 showed that the fuels contained 23 to 30 percent structural lignin in the foliage and 3 to 9 percent

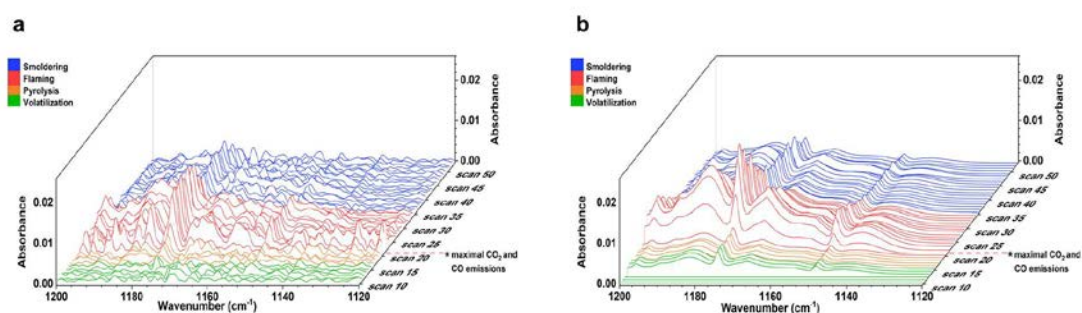
482 phenols (Matt et al 2020). Note that none of the fuels in the wind tunnel experiments or the bench-  
483 scale experiments contained rotten wood. In the present experiments, phenol was detected in 8 of  
484 the 10 static measurements (recall that two of the static measurements only showed ambient gases  
485 due to early closure of the valves). Figure 5 demonstrates the static spectrum from Burn 89,  
486 corresponding to the burning of longleaf pine with inkberry. Seen in Fig. 5 are the experimental  
487 spectrum (blue trace) and also the reference spectrum of acetic acid (green trace). After subtraction  
488 of the  $\text{CH}_3\text{COOH}$  vapor spectrum, the residual contained two small peaks which were readily  
489 identified as phenol vapor via the  $\nu_{15}$  vibrational band near  $1176.2\text{ cm}^{-1}$ , as well as the  $\nu_{16}$  band at  
490  $1150.2\text{ cm}^{-1}$  (Keresztury et al., 1998). The phenol reference spectrum from the PNNL spectral  
491 library (red trace) was then subtracted from that residual (purple trace) with an overall residual  
492 that is mostly noise (black trace). For the dynamic spectra the process is repeated for each of the  
493 individual spectral time slices, represented by scan number using the concentration of phenol  
494 determined by the MALT program. To confirm the spectral analysis, in each case the mixing ratio  
495 calculated by MALT was converted to a spectrum by multiplying by the appropriate concentration  
496 path length factor; the predicted spectrum was visually compared to the actual data.



497  
498 **Figure 5.** Static spectrum obtained from Burn 89 (1 kg longleaf pine needles with inkberry). The blue trace  
499 is the FTIR experimental spectrum, the green trace the reference spectrum of acetic acid, the purple trace  
500 the residual after acetic acid subtraction, the red trace the reference spectrum of phenol and the black trace  
501 residual after phenol subtraction.

502 Phenol was also detected using the dynamic method and Figure 6 displays a series of dynamic  
503 spectra recorded for Burn 87. The spectra in the left frame (a) are individual spectra after the acetic

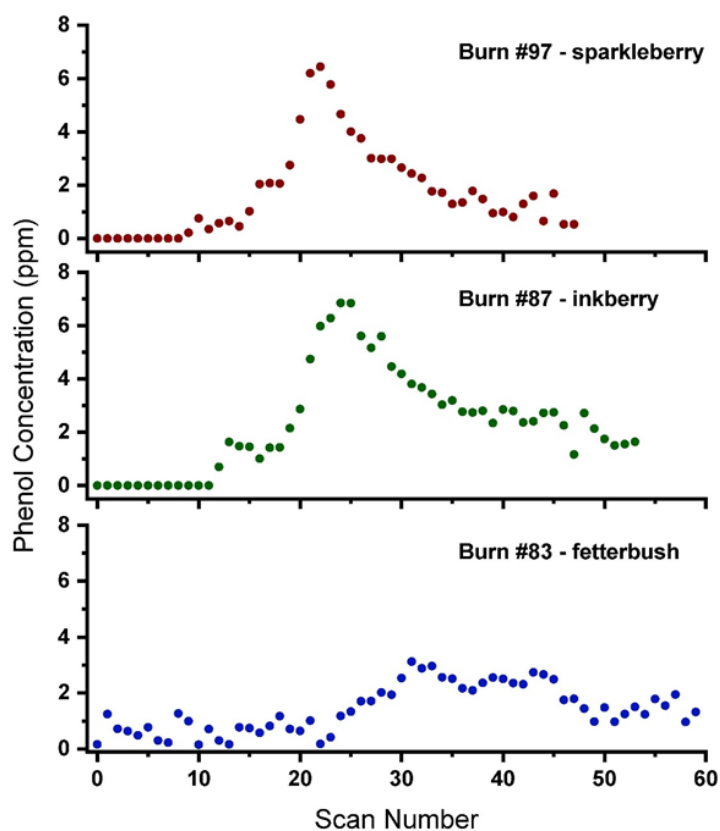
504 acid (CH<sub>3</sub>COOH) spectral component has been subtracted from the spectrum for each time slice,  
 505 all recorded at 1.0 cm<sup>-1</sup> resolution with 54 total measurements recorded at Δt=1.5 seconds. While  
 506 the spectral noise is still significant, the presence of phenol peaks, particularly the ν<sub>16</sub> Q-branch at  
 507 1176.5 cm<sup>-1</sup> and the ν<sub>15</sub> peak at 1150.2 cm<sup>-1</sup>, are evident. Optimization for the phenol mixing ratio  
 508 in each spectrum allowed for its calculation in individual time slices and the derived phenol-only  
 509 spectra are presented as a waterfall plot in the right frame (b). The first clear evidence of phenol is  
 510 seen in scans 14 to 18, before reaching a maximum concentration of 6.9 ppm in scan 24; this is  
 511 observed in the right frame of Figure 6, coinciding approximately with maximal CO<sub>2</sub> concentration  
 512 (scan 22), indicating the greatest ratio of smoke/ambient air in the gas cell.



513  
 514 **Figure 6.** Burn 87 –longleaf pine needles with inkberry fuel bed during dynamic mode. Measured and  
 515 scaled burn spectra showing the progression of phenol during the time resolved study. Acetic acid and  
 516 water spectral features have been removed in Frame a) with the phenol-only derived mixing ratio  
 517 spectra in Frame b.

518 Figure 6 displays the rapid increase of phenol vapor due to the approaching flame front from scan  
 519 14 (t = 22.5 s) to its maximal mixing ratio in scan 24 (t = 36 s) followed by a longer gradual phenol  
 520 decay with time. This can be juxtaposed with the ethene mixing ratios (seen in Figure 4) that fall  
 521 to nearly zero with the onset of combustion; the ethene is consumed by the flame propagation.  
 522 Prior to scan 14 in Figure 6b, minimal phenol is observed relative to the noise level and is thus fit  
 523 as zero concentration. Phenol contributions for scans 16-19 can be associated with the pyrolysis  
 524 phase of the burn and not combustion. Phenol is one of the major products of 1,2-benzenediol  
 525 pyrolysis with maximum yield reported at 800 °C (Ledesma et al. 2002; Thomas et al. 2007). It is  
 526 important to note that both temperature and rate of heating influence the composition and yield of  
 527 pyrolysis products. Evidenced in Figure 3, the thermal imaging associated with FTIR scan 17  
 528 shows a temperature of ~200 °C which is indicated in Kibet et al. (2012) to be within the  
 529 temperature range of pyrolysis of lignin: 200 to 400 °C. Shortly thereafter, phenol mixing ratios

530 rapidly increase and reach a maximal mixing ratio of 6.9 ppm at scan 24. At scan 24 the flame  
531 front has already reached the extractive probe and thus the maximum intake of smoke and ambient  
532 air is achieved; temperature of the fuel bed is ca. 600 °C, consistent with the flaming phase. The  
533 gradual decay in phenol production as the flame front passes could be due to several factors: (i) an  
534 increased temperature required for complete combustion of the C<sub>6</sub>H<sub>6</sub>O (ii) residence time of  
535 phenol, (iii) phenol production in the smoldering phase as a tar/char, (iv) adsorption to walls of the  
536 stainless steel tubing and cell. A cross-section of Figure 6 shows the rapid onset of phenol  
537 production at the temperatures followed by a gradual decay in concentration: this is indicative of  
538 phenol production throughout the burning of inkberry as a species.



539

540 **Figure 7.** Temporal mixing ratios of phenol for different shrub species. Phenol mixing ratios plotted over  
541 time indicated by scan number for Burn 97 sparkleberry (red), Burn 87 inkberry (green), and Burn 20  
542 fetterbush (blue) all on longleaf pine straw bed.

543 The shape of the temporal profile yields information as to the production of phenol throughout the  
544 evolution of burning. Figure 7 shows the progression of phenol concentration following its first  
545 observed presence in the burn for three different species. These graphs are effectively a cross-  
546 section of Figure 6, showing the progression of the height of the phenol peak (directly correlated



547 to phenol concentration) throughout the burn (with time being represented by scan number in both  
548 cases). The level of phenol generation was observed to vary between plant species. Temporal  
549 profiles of phenol concentration were constructed for burns with three different species:  
550 sparkleberry, inkberry, and fetterbush. These plots illustrate a range of behavior with inkberry and  
551 sparkleberry having similar temporal profiles and similar maxima of ca. 6.5 ppm and fetterbush  
552 having a different temporal profile. It is important to note that Burn 97 (sparkleberry) was  
553 measured at  $0.6\text{ cm}^{-1}$ , while burns 87 and 93 were measured at  $1.0\text{ cm}^{-1}$ , although the profile of  
554 burn 97 is consistent with that of burn 87. That is to say, we do not believe the small change in  
555 resolution affects the recovered mixing ratios. Demonstrated in Figure 7, trace amounts of phenol  
556 appear at the onset of combustion and throughout the pyrolysis phase. Phenol reaches its highest  
557 concentrations, however, during the flaming stage as all three temporal profiles reach maximum  
558 during the latter stages of the burn. Moreover, phenol remains throughout the duration of the burn  
559 and is not consumed by secondary reactions, as is e.g. ethene. For these burns, fetterbush was  
560 observed to have the lowest maximum concentration of phenol, only 3.1 ppm, for the three species,  
561 while sparkleberry and inkberry had similar maxima (as well as similar temporal profiles).

562 The observed differences of phenol in both temporal profile and overall peak concentrations could  
563 arise due to differences in leaf structure and shape, or possibly due to differences in leaf/plant  
564 composition. Pyrolytic production of phenol has been previously attributed to multiple  
565 components of plant composition, including phenol content, lignin content and the amount of  
566 cellulose in each plant species. Therefore, varying phenols, lignin, and cellulose in these plant  
567 species could be the source of phenol concentration variability for each burn. The physical  
568 composition of multiple plant species, including inkberry and fetterbush was analyzed by Matt et  
569 al. (2020); it was shown that inkberry has 2.6 times the percentage of phenol by composition  
570 (9.0%) than fetterbush (3.4%) Although sparkleberry was not included in that study, it can be  
571 suggested that the compositions of inkberry and sparkleberry are similar due to observed phenol  
572 in this experiment as well as plant characteristics. Sparkleberry is a member of the *Vaccinium*  
573 genus which contains many species collectively known as blueberries which are known to contain  
574 high levels of phenolic compounds in the fruits (e.g. Prior et al 1998). This study and the results  
575 of Matt et al. (2020) support the present hypothesis that peak concentrations of phenol are highest  
576 for sparkleberry and inkberry due to higher phenolic content in the plants.



577 **4. Summary**

578 The analytical methods used in this study attempt to provide a detailed view of prescribed burning  
579 by enlisting two different FTIR acquisition modes, static and dynamic. By capturing a “snapshot”  
580 of a single burn experiment, used in the static method, one can discern the gases with higher  
581 specificity and in turn decipher complex spectra by use of chemometrics to extract compounds  
582 with high concentrations leaving behind a residual to be analyzed. Lower resolution may hinder  
583 these efforts and allow compounds that are present at lower mixing ratios to be obscured by higher  
584 absorbing compounds, e.g., carbon dioxide, water, and ethene. In this study we were able to detect  
585 additional compounds e.g. phenol, benzene, and allene with greater confidence. However, in  
586 gaining specificity there is a loss of time resolution and this is where the dynamic method becomes  
587 advantageous. The FTIR dynamic acquisition method when synchronized to thermal imaging,  
588 while lower in sensitivity, allows for an overall profile of the burn and can help assign phases to  
589 the dynamic stages of the flame. That is, the dynamic method in conjunction with thermal IR  
590 imaging provides a more detailed description as temperature and chemical composition profiles  
591 can be correlated and assigned to certain phases of the burns. In this study pyrolysis, flaming and  
592 smoldering combustion were identified using these new techniques, which can aide in the  
593 improvement of fire behavior models used by land managers to conduct prescribed fires.

594 *Data availability.* Data are not publicly available as data release has not been authorized by sponsor  
595 of this research.

596 *Competing interests.* No competing interests.

597 *Acknowledgements.* This work was supported by the Department of Defense’s Strategic  
598 Environmental Research and Development Program (SERDP) within project RC-2640, and we  
599 gratefully acknowledge our sponsors for their support. PNNL is operated for the U.S. Department  
600 of Energy by the Battelle Memorial Institute under contract DE-AC06076RLO 1830.

601

602 **REFERENCES**

- 603 Agee, J. K.: Fire and pine ecosystems, in *Ecology and Biogeography of Pinus*, edited by D. M.  
604 Richardson, pp. 193–218, Cambridge University Press, Cambridge, U.K., 2000.
- 605 Aitchison, J.: *The statistical analysis of compositional data*, Chapman and Hall, London ; New  
606 York., 1986.
- 607 Akagi, S. K., Yokelson, R. J., Burling, I. R., Meinardi, S., Simpson, I., Blake, D. R.,  
608 McMeeking, G. R., Sullivan, A., Lee, T., Kreidenweis, S., Urbanski, S., Reardon, J., Griffith, D.  
609 W. T., Johnson, T. J., and Weise, D. R.: Measurements of reactive trace gases and variable O<sub>3</sub>  
610 formation rates in some South Carolina biomass burning plumes, *Atmos. Chem. Phys.*, 13, 1141-  
611 1165, 2013.
- 612 Akagi, S. K., Burling, I. R., Mendoza, A., Johnson, T. J., Cameron, M., Griffith, D. W. T., Paton-  
613 Walsh, C., Weise, D. R., Reardon, J., and Yokelson, R. J.: Field measurements of trace gases  
614 emitted by prescribed fires in southeastern US pine forests using an open-path FTIR system,  
615 *Atmos. Chem. Phys.*, 14, 199-215, 2014.
- 616 Alves, C. A., Gonçalves, C., Pio, C. A., Mirante, F., Caseiro, A., Tarelho, L., Freitas, M. C., and  
617 Viegas, D. X.: Smoke emissions from biomass burning in a Mediterranean shrubland, *Atmos.*  
618 *Environ.*, 44, 3024-3033, 2010.
- 619 Aminfar, A., Cobian-Iñiguez, J., Ghasemian, M., Espitia, N. R., Weise, D. R. and Princevac, M.:  
620 Using Background-Oriented Schlieren to Visualize Convection in a Propagating Wildland Fire,  
621 *Combustion Science and Technology*, 1–21, doi:[10.1080/00102202.2019.1635122](https://doi.org/10.1080/00102202.2019.1635122), 2019.
- 622 Aminfar, A.: *Application of Computer Vision to Transport Phenomena*, Ph.D., University of  
623 California, Riverside., 2019.
- 624 Amini, E., Safdari, M.-S., DeYoung, J. T., Weise, D. R. and Fletcher, T. H.: Characterization of  
625 pyrolysis products from slow pyrolysis of live and dead vegetation native to the southern United  
626 States, *Fuel*, 235, 1475–1491, doi:[10.1016/j.fuel.2018.08.112](https://doi.org/10.1016/j.fuel.2018.08.112), 2019.
- 627 Barbour, M. G. and Billings, W. D., Eds.: *North American terrestrial vegetation*, 2. ed.,  
628 Cambridge Univ. Press, Cambridge., 2000.
- 629 Behm, A., Duryea, M. L., Long, A. J. and Zipperer, W. C.: Flammability of native understory  
630 species in pine flatwood and hardwood hammock ecosystems and implications for the wildland–  
631 urban interface, *International Journal of Wildland Fire*, 13(3), 355–365, doi:[10.1071/WF03075](https://doi.org/10.1071/WF03075),  
632 2004.
- 633 Bist, H. D., Brand, J. C. D. and Williams, D. R.: The Vibrational Spectrum and Torsion of  
634 Phenol, *J. Mole. Spec.*, 24, 402-412, 1967.
- 635 Biswell, H. H.: *Prescribed burning in California wildlands vegetation management*, Berkeley,  
636 CA: University of California Press; p. 255, 1989.

637 Brillì, F., Gioli, B., Ciccioli, P., Zona, D., Loreto, F., Janssens, I. A., and Ceulemans, R.: Proton  
638 Transfer Reaction Time-of-Flight Mass Spectrometric (PTR-TOF-MS) determination of volatile  
639 organic compounds (VOCs) emitted from a biomass fire developed under stable nocturnal  
640 conditions, *Atmos. Environ.*, 97, 54-67, 2014.

641 Burgan, R. E. and Susott, R. A.: Influence of sample processing techniques and seasonal  
642 variation on quantities of volatile compounds of gallberry, saw-palmetto and wax myrtle,  
643 *International Journal of Wildland Fire*, 1(1), 57–62, doi:[10.1071/WF9910057](https://doi.org/10.1071/WF9910057), 1991.

644 Burling, I. R., Yokelson, R. J., Griffith, D. W. T., Johnson, T. J., Veres, P., Roberts, J. M.,  
645 Warneke, C., Urbanski, S. P., Reardon, J., Weise, D. R., Hao, W. M. and de Gouw, J.:  
646 Laboratory measurements of trace gas emissions from biomass burning of fuel types from the  
647 southeastern and southwestern United States, *Atmos. Chem. Phys.*, 10(22), 11115–11130,  
648 doi:[10.5194/acp-10-11115-2010](https://doi.org/10.5194/acp-10-11115-2010), 2010.

649 Burling, I. R., Yokelson, R. J., Akagi, S. K., Urbanski, S. P., Wold, C. E., Griffith, D. W. T.,  
650 Johnson, T. J., Reardon, J., and Weise, D. R.: Airborne and ground-based measurements of the  
651 trace gases and particles emitted by prescribed fires in the United States, *Atmos. Chem. Phys.*,  
652 11, 12197-12216, 2011.

653 Bytnerowicz, A., Arbaugh, M. A., Andersen, C. K. and Riebau, A. R., Eds.: *Wildland fires and  
654 air pollution*, Elsevier, Amsterdam ; Boston., 2009.

655 Carter, M. C. and Foster, C. D.: Prescribed burning and productivity in southern pine forests: a  
656 review, *Forest Ecol. Manage.*, 191, 93–109, 2004.

657 Christensen, N. L.: Vegetation of the Southeastern Coastal Plain, in *North American Terrestrial  
658 Vegetation*, edited by M. G. Barbour and W. D. Billings, pp. 397–448, Cambridge University  
659 Press, New York, NY., 2000.

660 Christian, T. J., Kleiss, B., Yokelson, R. J., Holzinger, R., Crutzen, P. J., Hao, W. M., Shirai, T.,  
661 and Blake, D. R.: Comprehensive laboratory measurements of biomass-burning emissions: 2.  
662 First intercomparison of open-path FTIR, PTR-MS, and GC-MS/FID/ECD, *J. Geophys. Res.*  
663 *Atmos.*, 109, 2004.

664 Chi, C. T., Horn, D. A., Zanders, D. L., Opferkuch, R. E., Nyers, J. M., Pierovich, J. M., Lavdas,  
665 L. G., McMahon, C. K., Nelson, R. M., Jr., Johansen, R. W. and Ryan, P. W.: Source  
666 Assessment: Prescribed Burning, State of the Art, Environmental Protection Technology, United  
667 States Environmental Protection Agency, Research Triangle Park, NC. [online] Available from:  
668 [nepis.epa.gov](http://nepis.epa.gov), 1979.

669 Cohen, S., Hall, J. and Hiers, J. K.: Fire Science Strategy, Strategic Environmental Research and  
670 Development Program, Resource Conservation and Climate Change Program Area, Washington,  
671 D.C. [online] Available from: [https://serdp-  
672 estcp.org/content/download/30210/291748/file/Fire%20Science%20Strategy.pdf](https://serdp-estcp.org/content/download/30210/291748/file/Fire%20Science%20Strategy.pdf), 2014.

673 Collard, F.X. and Blin, J.: A review on pyrolysis of biomass constituents: Mechanisms and  
674 composition of the products obtained from the conversion of cellulose, hemicelluloses and  
675 lignin. *Renewable and Sustainable Energy Reviews*, 38, pp.594-608, 2014.

676 Crutzen, P. J. and Goldammer, J. G., Eds.: *Fire in the environment: the ecological, atmospheric,  
677 and climatic importance of vegetation fires: report of the Dahlem Workshop, held in Berlin, 15-  
678 20 March 1992*, Wiley, Chichester, England ; New York., 1993.

679

680 Depew, C. A., Mann, M. J. and Corlett, R. C.: A Laboratory Simulation of Wood Pyrolysis  
681 Under Field Conditions, *Combustion Science and Technology*, 6(4), 241–246,  
682 doi:[10.1080/00102207208952326](https://doi.org/10.1080/00102207208952326), 1972.

683 Di Blasi, C.: Modeling chemical and physical processes of wood and biomass pyrolysis, *Progress  
684 in Energy and Combustion Science*, 34(1), 47–90, doi:[10.1016/j.pecs.2006.12.001](https://doi.org/10.1016/j.pecs.2006.12.001), 2008.

685

686 Dimitrakopoulos, A. P.: Thermogravimetric analysis of Mediterranean plant species, *Journal of  
687 Analytical and Applied Pyrolysis*, 60(2), 123–130, doi:[10.1016/S0165-2370\(00\)00164-9](https://doi.org/10.1016/S0165-2370(00)00164-9), 2001.

688

689 Fairburn, J. A., Behie, L. A., and Svrcek, W. Y.: Ultraprolysis of n-hexadecane in a novel  
690 micro-reactor, *FUEL*, 69, 1537-1545, 1990.

691

692 Frenklach, M., Taki, S., Durgaprasad, M. B., and Matula, R. A.: Soot formation in shock-tube  
693 pyrolysis of acetylene, allene, and 1, 3-butadiene, *Combust. Flame*, 54, 81–101, 1983.

694

695 Frenklach, M., Yuan, T., and Ramachandra, M. K.: Soot formation in binary hydrocarbon  
696 mixtures, *Energy Fuels*, 2, 462–480, 1988.

697

698 Gilman, J. B., Lerner, B. M., Kuster, W. C., Goldan, P. D., Warneke, C., Veres, P. R., Roberts, J.  
699 M., de Gouw, J. A., Burling, I. R., and Yokelson, R. J.: Biomass burning emissions and potential  
700 air quality impacts of volatile organic compounds and other trace gases from fuels common in  
701 the US, *Atmos. Chem. Phys.*, 15, 13915-13938, 2015.

702

703 Griffith, D. W. T.: MALT5 User guide Version 5.5.9 2016.

704

705 Goode, J. G., Yokelson, R. J., Susott, R. A., and Ward, D. E.: Trace gas emissions from  
706 laboratory biomass fires measured by open-path Fourier transform infrared spectroscopy: Fires  
707 in grass and surface fuels, *J. Geophys. Res. Atmos.*, 104, 21237-21245, 1999.

708

709 Goode, J. G., Yokelson, R. J., Ward, D. E., Susott, R. A., Babbitt, R. E., Davies, M. A., and Hao,  
710 W. M.: Measurements of excess O<sub>3</sub>, CO<sub>2</sub>, CO, CH<sub>4</sub>, C<sub>2</sub>H<sub>4</sub>, C<sub>2</sub>H<sub>2</sub>, HCN, NO, NH<sub>3</sub>, HCOOH,  
711 CH<sub>3</sub>COOH, HCHO, and CH<sub>3</sub>OH in 1997 Alaskan biomass burning plumes by airborne Fourier  
712 transform infrared spectroscopy (AFTIR), *J. Geophys. Res. Atmos.*, 105, 22147-22166, 2000.

713

714 Gordon, I. E., Rothman, L. S., Hill, C., Kochanov, R. V., Tan, Y., Bernath, P. F., Birk, M.,  
715 Boudon, V., Campargue, A., Chance, K. V., Drouin, B. J., Flaud, J.-M., Gamache, R. R.,  
716 Hodges, J. T., Jacquemart, D., Perevalov, V. I., Perrin, A., Shine, K. P., Smith, M.-A. H.,  
717 Tennyson, J., Toon, G. C., Tran, H., Tyuterev, V. G., Barbe, A., Császár, A. G., Devi, V. M.,

718 Furtenbacher, T., Harrison, J. J., Hartmann, J.-M., Jolly, A., Johnson, T. J., Karman, T., Kleiner,  
719 I., Kyuberis, A. A., Loos, J., Lyulin, O. M., Massie, S. T., Mikhailenko, S. N., Moazzen-Ahmadi,  
720 N., Müller, H. S. P., Naumenko, O. V., Nikitin, A. V., Polyansky, O. L., Rey, M., Rotger, M.,  
721 Sharpe, S. W., Sung, K., Starikova, D., S.A. Tashkun, S. A., Van der Auwera, J., Wagner, G.,  
722 Wilzewski, J., Wcisło, P., Yu, S., and Zak, E. J.: The HITRAN2016 molecular spectroscopic  
723 database, *J. Quant. Spectrosc. Radiat. Transfer*, 203, 3-69, 2017.  
724  
725 Guérette, É.-A., Paton-Walsh, C., Desservettaz, M., Smith, T. E. L., Volkova, L., Weston, C. J.  
726 and Meyer, C. P.: Emissions of trace gases from Australian temperate forest fires: emission  
727 factors and dependence on modified combustion efficiency, *Atmos. Chem. Phys.*, 18, 3717-3735,  
728 doi: 10.5194/acp-18-3717-2018, 2018.  
729  
730 Hardy, C. C., Ottmar, R. D., Peterson, J. L., Core, J. E. and Seamon, P.: Smoke management  
731 guide for prescribed and wildland fire; 2001 ed., PMS 420-2 National Wildfire Coordinating  
732 group, Boise, ID. 226 pp., 2001.  
733  
734 Hatch, L. E., Yokelson, R. J., Stockwell, C. E., Veres, P. R., Simpson, I. J., Blake, D. R.,  
735 Orlando, J. J., and Barsanti, K. C.: Multi-instrument comparison and compilation of non-  
736 methane organic gas emissions from biomass burning and implications for smoke-derived  
737 secondary organic aerosol precursors, *Atmos. Chem. Phys.*, 17, 1471-1489, 2017.  
738  
739 Hawthorne, S. B., Krieger, M. S., Miller, D. J., and Mathiason, M. B.: Collection and  
740 quantitation of methoxylated phenol tracers for atmospheric pollution from residential wood  
741 stoves, *Environ. Sci. Technol.*, 23, 470-475, <https://doi.org/10.1021/es00181a013>, 1989.  
742  
743 Hough, W. A.: Caloric value of some forest fuels of the southern United States, Research Note,  
744 USDA Forest Service, Southeastern Forest Experiment Station, Asheville, NC. [online]  
745 Available from: <http://www.treesearch.fs.fed.us/pubs/2778>, 1969.  
746  
747 Hurst, D. F., Griffith, D. W. T., Carras, J. N., Williams, D. J., and Fraser, P. J.: Measurements of  
748 Trace Gases Emitted by Australian Savanna Fires During the 1990 Dry Season, *J. Atmos.*  
749 *Chem.*, 18, 33-56, 1994a.  
750  
751 Hurst, D. F., Griffith, D. W. T., and Cook, G. D.: Trace gas emissions from biomass burning in  
752 tropical Australian savannas, *J. Geophys. Res.*, 99, 16441-16456, 1994b.  
753  
754 Ingemarsson, A., Nilsson, U., Nilsson, M., Pedersen, J. R., and Olsson, J. O.: Slow Pyrolysis of  
755 Spruce and Pine Samples Studied with GC/MS and GC/FTIR/FID, *Chemosphere*, 36-14, 2879-  
756 2889, 1998.  
757  
758 Johnson, T. J., Simon, A., Weil, J. M. and Harris, G. W., “Applications of time-resolved step-  
759 scan and rapid-scan FT-IR spectroscopy: Dynamics from ten seconds to ten nanoseconds”  
760 *Applied Spectroscopy*, 47, 1376, (1993).  
761

762 Johnson, T. J., Masiello, T., and Sharpe, S. W.: The quantitative infrared and NIR spectrum of  
763 CH<sub>2</sub>I<sub>2</sub> vapor: vibrational assignments and potential for atmospheric monitoring, *Atmos. Chem.*  
764 *Phys.*, 6, 2581-2591, 2006.  
765  
766 Johnson, T. J., Profeta, L. T. M., Sams, R. L., Griffith, D.W. T., and Yokelson, R. L.: An  
767 infrared spectral database for detection of gases emitted by biomass burning, *Vib. Spectrosc.*, 53,  
768 97–102, 2010.  
769  
770 Jolly, W. M., Hintz, J., Linn, R. L., Kropp, R. C., Conrad, E. T., Parsons, R. A. and Winterkamp,  
771 J.: Seasonal variations in red pine (*Pinus resinosa*) and jack pine (*Pinus banksiana*) foliar  
772 physio-chemistry and their potential influence on stand-scale wildland fire behavior, *Forest*  
773 *Ecology and Management*, 373, 167–178, doi:[10.1016/j.foreco.2016.04.005](https://doi.org/10.1016/j.foreco.2016.04.005), 2016.  
774  
775 Jolly, W. M., Parsons, R. A., Hadlow, A. M., Cohn, G. M., McAllister, S. S., Popp, J. B.,  
776 Hubbard, R. M. and Negron, J. F.: Relationships between moisture, chemistry, and ignition of  
777 *Pinus contorta* needles during the early stages of mountain pine beetle attack, *Forest Ecology*  
778 *and Management*, 269, 52–59, doi:[10.1016/j.foreco.2011.12.022](https://doi.org/10.1016/j.foreco.2011.12.022), 2012.  
779  
780 Karl, T. G., Christian, T. J., Yokelson, R. J., Artaxo, P., Hao, W. M., and Guenther, A.: The  
781 Tropical Forest and Fire Emissions Experiment: method evaluation of volatile organic compound  
782 emissions measured by PTR-MS, FTIR, and GC from tropical biomass burning, *Atmos. Chem.*  
783 *Phys.*, 7, 5883-5897, 2007.  
784  
785 Keresztury, G., Billes, F., Kubinyi, M. and Sundius, T.: A Density Functional, Infrared Linear  
786 Dichroism, and Normal Coordinate Study of Phenol and its Deuterated Derivatives: Revised  
787 Interpretation of the Vibrational Spectra, *J. Phys. Chem.*, 102, 1371-1380, 1998.  
788  
789 Kibet, J., Khachatryan, L., and Dellinger, B.: Molecular products and radicals from pyrolysis of  
790 lignin, *Environ. Sci. Technol.*, 46, 12994-13001, 2012.  
791  
792 Kochanov, R.V., Gordon, I.E., Rothman, L.S., Shine, K.P., Sharpe, S.W., Johnson, T.J.,  
793 Wallington, T.J., Harrison, J.J., Bernath, P.F., Birk, M. and Wagner, G.: Infrared absorption  
794 cross-sections in HITRAN2016 and beyond: Expansion for climate, environment, and  
795 atmospheric applications. *Journal of Quantitative Spectroscopy and Radiative Transfer*, 230,  
796 pp.172-221. 2019.  
797  
798 Koss, A. R., Sekimoto, K., Gilman, J. B., Selimovic, V., Coggon, M. M., Zarzana, K. J., Yuan,  
799 B., Lerner, B. M., Brown, S. S., Jimenez, J. L., Krechmer, J., Roberts, J. M., Warneke, C.,  
800 Yokelson, R. J., and de Gouw, J.: Non-methane organic gas emissions from biomass burning:  
801 identification, quantification, and emission factors from PTR-ToF during the FIREX 2016  
802 laboratory experiment, *Atmos. Chem. Phys.*, 18, 3299, 2018.  
803  
804 Ledesma, E. B., Marsh, N. D., Sandrowitz, A. K. and Wornat, M. J.: An experimental study on  
805 the thermal decomposition of catechol, *Proceedings of the Combustion Institute*, 29(2), 2299–  
806 2306, doi:10.1016/S1540-7489(02)80280-2, 2002.  
807

808 Liu, X., Huey, L. G., Yokelson, R. J., Selimovic, V., Simpson, I. J., Müller, M., Jimenez, J. L.,  
809 Campuzano-Jost, P., Beyersdorf, A. J., Blake, D. R., Butterfield, Z., Choi, Y., Crouse, J. D.,  
810 Day, D. A., Diskin, G. S., Dubey, M. K., Fortner, E., Hanisco, T. F., Hu, W., King, L. E.,  
811 Kleinman, L., Meinardi, S., Milkoviny, T., Onasch, T. B., Palm, B. B., Peischl, J., Pollack, I. B.,  
812 Ryerson, T. B., Sachse, G. W., Sedlacek, A. J., Shilling, J. E., Springston, S., St. Clair, J. M.,  
813 Tanner, D. J., Teng, A. P., Wennberg, P. O., Wisthaler, A., and Wolfe, G. M.: Airborne  
814 measurements of western US wildfire emissions: Comparison with prescribed burning and air  
815 quality implications, *J. Geophys. Res.-Atmos.*, 122, 6108–6129, doi: 10.1002/2016JD026315,  
816 2017.

817  
818 Melvin, M. A.: 2015 National Prescribed Fire Use Survey Report, Technical Report, Coalition of  
819 Prescribed Fire Councils, Inc. [online] Available from:  
820 [http://stateforesters.org/sites/default/files/publication-](http://stateforesters.org/sites/default/files/publication-documents/2015%20Prescribed%20Fire%20Use%20Survey%20Report.pdf)  
821 [documents/2015%20Prescribed%20Fire%20Use%20Survey%20Report.pdf](http://stateforesters.org/sites/default/files/publication-documents/2015%20Prescribed%20Fire%20Use%20Survey%20Report.pdf), 2015.

822  
823 Matt, F. J., Dietenberger, M. A. and Weise, D. R.: Summative and ultimate analysis of live  
824 leaves from southern U.S. forest plants for use in fire modeling, *Energy Fuels*, (34), 4703–4720,  
825 doi:[10.1021/acs.energyfuels.9b04107](https://doi.org/10.1021/acs.energyfuels.9b04107), 2020.

826  
827 Neuman, J. A., Huey, L. G., Ryerson, T. B., and Fahey, D. W.: Study of Inlet Materials for  
828 Sampling Atmospheric Nitric Acid, *Environ. Sci. Technol.*, 33, 1133-1136, doi:  
829 10.1021/es980767f, 1999.

830  
831 Paton-Walsh, C., Deutscher, N. M., Griffith, D. W. T., Forgan, B. W., Wilson, S. R., Jones, N.  
832 B., and Edwards, D. P.: Trace gas emissions from savanna fires in northern Australia, *J.*  
833 *Geophys. Res.*, 115, doi:10.1029/2009JD013309, 2010.

834  
835 Paton-Walsh, C., Smith, T. E. L., Young, E. L., Griffith, D. W. T. and Guérette, É.-A.: New  
836 emission factors for Australian vegetation fires measured using open-path Fourier transform  
837 infrared spectroscopy- Part 1: Methods and Australian temperate forest fires, *Atmos. Chem.*  
838 *Phys.*, 14, 11313-11333, doi: 10.5194/acp-14-11313-2014, 2014.

839  
840 Phillips, M. C., Myers, T. L., Johnson, T. J., and Weise, D. R.: In-situ measurement of pyrolysis  
841 and combustion gases from biomass burning using swept wavelength external cavity quantum  
842 cascade lasers, *Optics Express*, 28(6), 8680-8700, 2020

843  
844 Prior, R. L., Cao, G., Martin, A., Sofic, E., McEwen, J., O'Brien, C., Lischner, N., Ehlenfeldt,  
845 M., Kalt, W., Krewer, G. and Mainland, C. M.: Antioxidant Capacity As Influenced by Total  
846 Phenolic and Anthocyanin Content, Maturity, and Variety of *Vaccinium* Species, *J. Agric. Food*  
847 *Chem.*, 46(7), 2686–2693, doi:10.1021/jf980145d, 1998.

848  
849 Pyne, S. J.: *World fire: the culture of fire on earth*, Pbk. ed., University of Washington Press,  
Seattle., 1997.

850  
851 Rao, P. V. R. and Rao, G. R.: Vibrational analysis of substituted phenols Part I. Vibrational  
spectra, normal coordinate analysis and transferability of force constants of some formyl-,



852 methoxy-, formylmethoxy-, methyl- and halogeno-phenols, *Spectrochimica ACTA Part A*, 58,  
853 3039-3065, 2002.

854 Roscioli, J. R., Zahniser, M. S., Nelson, D. D., Herndon, S. C., and Kolb, C. E.: New Approaches  
855 to Measuring Sticky Molecules: Improvements of Instrumental Response Times Using Active  
856 Passivation, *J. Phys. Chem. A*, 120, 1347-1357, doi: 10.1021/acs.jpca.5b04395, 2015.

857 Safdari, M.-S., Rahmati, M., Amini, E., Howarth, J. E., Berryhill, J. P., Diitenberger, M., Weise,  
858 D. R., and Fletcher, T. H.: Characterization of pyrolysis products from fast pyrolysis of live and  
859 dead vegetation native to the Southern United States, *Fuel*, 229, 151-166, 2018.

860 Safdari, M.-S., Amini, E., Weise, D. R. and Fletcher, T. H.: Comparison of pyrolysis of live  
861 wildland fuels heated by radiation vs. convection, *Fuel*, 268, 117342,  
862 doi:[10.1016/j.fuel.2020.117342](https://doi.org/10.1016/j.fuel.2020.117342), 2020.

863

864 Saiz-Jimenez, C. and De Leeuw, J. W.: Lignin pyrolysis products: Their structures and their  
865 significance as biomarkers, *Organic Geochemistry*, 10(4-6), 869-876, doi:10.1016/S0146-  
866 6380(86)80024-9, 1986.

867

868 Scharko, N. K., Oeck, A. M., Myers, T. L., Tonkyn, R. G., Banach, C. A., Baker, S. P., Lincoln,  
869 E. N., Chong, J., Corcoran, B. M., Burke, G. M., Ottmar, R. D., Restaino, J. C., Weise, D. R. and  
870 Johnson, T. J.: Gas-phase pyrolysis products emitted by prescribed fires in pine forests with a  
871 shrub understory in the southeastern United States, *Atmos. Chem. Phys.*, 19(15), 9681-9698,  
872 doi:[10.5194/acp-19-9681-2019](https://doi.org/10.5194/acp-19-9681-2019), 2019a.

873 Scharko, N. K., Oeck, A. M., Tonkyn, R. G., Baker, S. P., Lincoln, E. N., Chong, J., Corcoran,  
874 B. M., Burke, G. M., Weise, D. R., Myers, T. L., Banach, C. A., Griffith, D. W. T. and Johnson,  
875 T. J.: Identification of gas-phase pyrolysis products in a prescribed fire: first detections using  
876 infrared spectroscopy for naphthalene, methyl nitrite, allene, acrolein and acetaldehyde,  
877 *Atmospheric Measurement Techniques*, 12(1), 763-776, doi:[10.5194/amt-12-763-2019](https://doi.org/10.5194/amt-12-763-2019), 2019b.

878

879 Scott, A. C., Bowman, D. M. J. S., Bond, W. J., Pyne, S. J. and Alexander, M. E.: *Fire on earth:*  
880 *an introduction*, John Wiley & Sons, Inc, Chichester, West Sussex., 2014.

881

882 Sekimoto, K., Koss, A. R., Gilman, J. B., Selimovic, V., Coggon, M. M., Zarzana, K. J., Yuan,  
883 B., Lerner, B. M., Brown, S. S., Warneke, C., Yokelson, R. J., Roberts, J. M., and de Gouw, J.:  
884 High-and low-temperature pyrolysis profiles describe volatile organic compound emissions from  
885 western US wildfire fuels, *Atmospheric Chemistry & Physics*, 18, 2018.

886

887 Selimovic, V., Yokelson, R. J., Warneke, C., Roberts, J. M., Gouw, J. d., Reardon, J., and  
888 Griffith, D. W. T.: Aerosol optical properties and trace gas emissions by PAX and OP-FTIR for  
889 laboratory-simulated western US wildfires during FIREX, *Atmos. Chem. Phys.*, 18, 2929-2948,  
890 2018.

891

892 Sharpe, S.W., Sams, R.L., Johnson, T.J., Chu, P.M., Rhoderick, G.C., and Guenther,  
893 F.R.: Creation of 0.10-cm<sup>-1</sup> resolution quantitative infrared spectral libraries for gas samples,



894 Proc. SPIE 4577, Vibrational Spectroscopy-based Sensor Systems, doi:  
895 <https://doi.org/10.1117/12.455730>, 2002.  
896  
897 Shimanouchi, T.: Tables of Vibrational Frequencies, Consolidated Vol. I. National Bureau of  
898 Standards, 1972.  
899  
900 Smith, T., Paton-Walsh, C. P., Meyer, G. C., Maier, S. W., Russell-Smith, J., Wooster, M., and  
901 Yates, C. P.: New emission factors for Australian vegetation fires measured using open-path  
902 Fourier transform infrared spectroscopy-Part: Australian tropical savanna fire, *Atmos. Chem.*  
903 *Phys.* 14, 14, 11335–11352, 2014.  
904  
905 Smith, T., Clare Paton-Walsh, C. P. Meyer, Garry Cook, Stefan W. Maier, Jeremy Russell-  
906 Smith, Martin Wooster, and Cameron P. Yates. "New emission factors for Australian vegetation  
907 fires measured using open-path Fourier transform infrared spectroscopy-Part 2: Australian  
908 tropical savanna fires." (2014): 11335.  
909  
910 Stein, Y. S., Antal Jr, M. J., and Jones Jr., M.: A study of the gas-phase pyrolysis of glycerol, *J.*  
911 *Anal. Appl. Pyrol.*, 4, 283–296, 1983.  
912  
913 Stockwell, C. E., Yokelson, R., Kreidenweis, S. M., Robinson, A. L., DeMott, P. J., Sullivan, R.  
914 C., Reardon, J., Ryan, K. C., Griffith, D. W. T., and Stevens, L.: Trace gas emissions from  
915 combustion of peat, crop residue, domestic biofuels, grasses, and other fuels: configuration and  
916 Fourier transform infrared (FTIR) component of the fourth Fire Lab at Missoula Experiment  
917 (FLAME-4), *Atmos. Chem. Phys.*, 14, 9727-9754, 2014.  
918  
919 Susott, R. A.: Characterization of the thermal properties of forest fuels by combustible gas  
920 analysis, *For. Sci.*, 28(2), 404–420, 1982.  
921  
922 Thomas, S., Ledesma, E. B. and Wornat, M. J.: The effects of oxygen on the yields of the  
923 thermal decomposition products of catechol under pyrolysis and fuel-rich oxidation conditions,  
924 *Fuel*, 86(16), 2581–2595, doi:10.1016/j.fuel.2007.02.003, 2007.  
925  
926 Tihay, V. and Gillard, P.: Pyrolysis gases released during the thermal decomposition of three  
927 Mediterranean species, *Journal of Analytical and Applied Pyrolysis*, 88(2), 168–174,  
928 doi:[10.1016/j.jaap.2010.04.002](https://doi.org/10.1016/j.jaap.2010.04.002), 2010.  
929  
930 Varhegyi, G., Jakab, E. and Antal, M. J.: Is the Broido-Shafizadeh Model for Cellulose Pyrolysis  
931 True?, *Energy Fuels*, 8(6), 1345–1352, doi:[10.1021/ef00048a025](https://doi.org/10.1021/ef00048a025), 1994.  
932  
933 Viatte, C., Strong, K., Hannigan, J., Nussbaumer, E., Emmons, L. K., Conway, S., Paton-Walsh,  
934 C., Hartley, J., Benmergui, J. and Lin, J.: Identifying fire plumes in the Arctic with tropospheric  
935 FTIR measurements and transport models, *Atmos. Chem. Phys.*, 15, 2227-2246, doi:, 2015.  
936  
937 Waldrop, T. A. and Goodrick, S. L.: Introduction to prescribed fires in southern ecosystems,  
938 *Science Update*, USDA Forest Service, Southern Research Station, Asheville, NC. [online]  
939 Available from: <http://www.treeseearch.fs.fed.us/pubs/41316>, 2012.

937 Ward, D. E. and Hao, W. M.: Projections of emissions from burning of biomass for use in  
938 studies of global climate and atmospheric chemistry, 19 p., Air and Waste Management  
939 Association, Vancouver, British Columbia, Canada. [online] Available from:  
940 <http://www.treesearch.fs.fed.us/pubs/43258>, 1991.

941 Ward, D. E., and Hardy, C. C.: Smoke emissions from wildland fires, *Environ. Int.*, 17, 117-134,  
942 1991.

943 Ward, D. E. and Radke, L. F.: Emissions measurement from vegetation fires: a comparative  
944 evaluation of methods and results, in *Fire in the environment: the ecological, atmospheric, and*  
945 *climatic importance of vegetation fires: report of the Dahlem Workshop, held in Berlin, 15-20*  
946 *March 1992*, edited by P. J. Crutzen and J. G. Goldammer, pp. 53–76, John Wiley & Sons Ltd.  
947 [online] Available from: [http://www.fs.fed.us/rm/pubs\\_other/rmrs\\_1993\\_ward\\_d001.pdf](http://www.fs.fed.us/rm/pubs_other/rmrs_1993_ward_d001.pdf), 1993.

948 Ward, D. E.: Combustion chemistry and smoke, in *Forest Fires: Behavior and Ecological Effects*,  
949 edited by E. A. Johnson and K. Miyanishi, pp. 55–77, Academic Press, San Diego, CA. [online]  
950 Available from: <http://www.doi.org/10.1016/B978-012386660-8/50006-3>, 2001.

951 Warneke, C., Roberts, J. M., Veres, P., Gilman, J., Kuster, W. C., Burling, I., Yokelson, R., and  
952 de Gouw, J. A.: VOC identification and inter-comparison from laboratory biomass burning using  
953 PTR-MS and PIT-MS, *Int. J. Mass Spectrom.*, 303, 6–14, doi:10.1016/j.ijms.2010.12.002, 2011.

954 Weise, D. R., Fletcher, T. H., Cole, W., Mahalingam, S., Zhou, X., Sun, L., and Li, J.: Fire  
955 behavior in chaparral- Evaluating flame models with laboratory data, *Combust. and Flame*, 191,  
956 500-512, <https://doi.org/10.1016/j.combustflame.2018.02.012>, 2018.

957 Weise, D. R., Johnson, T. J. and Reardon, J.: Particulate and trace gas emissions from prescribed  
958 burns in southeastern U.S. fuel types: Summary of a 5-year project, *Fire Safety Journal*, 74, 71–  
959 81, doi:[10.1016/j.firesaf.2015.02.016](https://doi.org/10.1016/j.firesaf.2015.02.016), 2015.

960 Weise, D. R., Palarea-Albaladejo, J., Johnson, T. J., and Jung, H.: Analyzing Wildland Fire  
961 Smoke Emissions Data Using Compositional Data Techniques, *J. Geophys. Res. Atmos.*, 125(6),  
962 <https://doi.org/10.1029/2019JD032128>, 2020.

963 Williams, S.D., Johnson, T.J., Sharpe, S.W., Yavelak, V., Oates, R.P. and Brauer, C.S.:  
964 Quantitative vapor-phase IR intensities and DFT computations to predict absolute IR spectra  
965 based on molecular structure: I. Alkanes. *Journal of Quantitative Spectroscopy and Radiative*  
966 *Transfer*, 129, 298-307, <https://doi.org/10.1016/j.jqsrt.2013.07.005>, 2013

967 Yang, H., Yan, R., Chen, H., Lee, D. H. and Zheng, C.: Characteristics of hemicellulose,  
968 cellulose and lignin pyrolysis, *Fuel*, 86, 1781–1788, 2007.

969 Yee, L. D., Kautzman, K. E., Loza, C. L., Schilling, K. A., Coggon, M. M., Chhabra, P. S.,  
970 Chan, M. N., Chan, A. W. H., Hersey, S. P., Crouse, J. D., Wennberg, P. O., Flagan, R. C., and  
971 Seinfeld, J. H.: Secondary organic aerosol formation from biomass burning intermediates: phenol  
972 and methoxyphenols, *Atmos. Chem. Phys.*, 13, 8019–8043, [https://doi.org/10.5194/acp-13-8019-](https://doi.org/10.5194/acp-13-8019-2013)  
973 2013, 2013.

974 Yokelson, R. J., Griffith, D. W. T. and Ward, D. E.: Open-path Fourier transform infrared studies  
975 of large-scale laboratory biomass fires, *Journal of Geophysical Research*, 101(D15), 21067,  
976 doi:[10.1029/96JD01800](https://doi.org/10.1029/96JD01800), 1996.

977 Yokelson, R. J., Susott, R., Ward, D. E., Reardon, J., and Griffith, D. W. T.: Emissions from  
978 smoldering combustion of biomass measured by open-path Fourier transform infrared  
979 spectroscopy, *J. Geophys. Res. Atmos.*, 102, 18865-18877, 1997.

980 Yokelson, R. J., Christian, T. J., Bertschi, I. T., and Hao, W. M.: Evaluation of adsorption effects  
981 on measurements of ammonia, acetic acid, and methanol, *J. Geophys. Res.*, 108, 4649,  
982 doi:[10.1029/2003JD003549](https://doi.org/10.1029/2003JD003549), 2003.

983 Yokelson, R. J., Burling, I. R., Gilman, J. B., Warneke, C., Stockwell, C. E., Gouw, J. d., Akagi,  
984 S. K., Urbanski, S. P., Veres, P., Roberts, J. M., Kuster, W. C., Reardon, J., Griffith, D. W. T.,  
985 Johnson, T. J., Hosseini, S., Miller, J. W., Cocker III, D. R., Jung, H., and Weise, D. R.:  
986 Coupling field and laboratory measurements to estimate the emission factors of identified and  
987 unidentified trace gases for prescribed fires, *Atmos. Chem. Phys.*, 13, 89-116, 2013.

988 Zhou, X. and Mahalingam, S.: Evaluation of reduced mechanism for modeling combustion of  
989 pyrolysis gas in wildland fire, *Combustion Science and Technology*, 171(1), 39–70,  
990 doi:[10.1080/00102200108907858](https://doi.org/10.1080/00102200108907858), 2001.

991