

1 **AMT-2020-383 - Authors Response**

2 **Reviewer 1 Comments – Responsenses and Manuscript Revisions**

3 **General comments.**

4 This paper uses measurements in highly concentrated fire plumes (within 100m of wildland grass fires,
5 and in controlled burns at the Missoula Fire Lab) to assess interferences in UV absorption measurements
6 of ozone at 254 nm.

7 This paper is motivated by the ozone measurements of UV absorption instruments, and the health impacts
8 from that ozone. Large increases in ozone may be observed after precursor NO_x and VOC have had time
9 to react. The time scale to produce that ozone is highly dependent on plume dilution, which itself is highly
10 variable in time, but typically takes place over hours since emission. A fundamental question: for direct
11 emissions, the interfering species will also be diluted, such that the lowest interferences may be expected
12 at the highest levels of plume ozone. Secondary production of UV-active hydrocarbons, e.g., production
13 of nitroaromatics following oxidation in the presence of NO₂, may dominate the ozone interference
14 downwind. What balance of directly emitted vs. secondary species are conjectured to lead to interferences
15 in ambient ozone measurements? Regardless of the source of the interference (primary vs. secondary),
16 given the lack of consistency from fire to fire (or even between different implementations of the UV
17 absorption technique) in the level of interferences measured, can the authors say what level of "fire
18 impact" causes a non-negligible interference? Is 1 ppm of ozone acceptable? 10 ppm of ozone?
19 Regardless, despite the experimental detail in this paper, it is not clear what ozone monitoring locations
20 are expected to suffer from significant interferences as a result of wildfires or prescribed burns. Lacking
21 these considerations, the paper's conclusions are qualitative at best, and by implication condemn a much
22 larger portion of the U.S. ozone monitoring network during the fire season than I suspect is warranted.
23 For a given UV absorption monitor, can they recommend what data to retain, and what data to eliminate
24 because of fire impacts? Some additional clarity in the real-world effects of fire smoke on ozone
25 monitoring is needed for this to make a novel and useful contribution to the literature.

26 The paper is overly long and can be shortened by removing extraneous details, repetitive text, and tables
27 that do not provide any usefully generalizable data as suggested below. Earlier literature is not well cited,
28 and additional references are also suggested below.

29 **Response:** The authors appreciate the time required to provide the review and feel that the suggestions
30 provided by the reviewer will result in an improved manuscript for resubmission.

31 The authors do agree that a more detailed look at data collected at sites being impacted by aged smoke
32 (ex. State and local regulatory monitoring sites being impacted by nearby wildfires and long range
33 transport of photochemically aged smoke plumes) and are currently collecting this data as part of the EPA
34 MASIC study in Boise, ID; Missoula, MT; and Reno, NV. This additional data collection will aid in
35 linking these research chamber and near field prescribed grassland burn measurements back to real world

36 regulatory monitoring situations. We will address these issues in a new “implications” section prior to
37 the manuscript conclusion.

38 The authors will, as suggested by the reviewer, attempt to shorten the length of the manuscript by
39 removing text that is tangential to the scope of the paper. In addition, the authors will review earlier
40 literature and cite as appropriate including those references suggested by the reviewer.

41 **Manuscript Revision:** The authors, as suggested by the reviewer removed text from the manuscript to
42 shorten the length of the document. An additional “Implications” section was added prior to the
43 conclusions section to tie the results of the research detailed in this manuscript to real world monitoring
44 applications. Included in this section will be a review of data from monitoring sites downwind of fires to
45 show the impact of the measurement artifacts described in this manuscript.

46

47 **Specific comments.**

48 **line 14:** "... large increases in ozone are also observed downwind ..." Is this always true?

49 **Response:** The authors did not imply that large increases in ozone are always observed down wind of
50 wildfire events. To clarify this and prevent assumptions that these increases in ozone “always occur”, the
51 text will be rewritten to include a statement like "... large increases in ozone have been observed
52 downwind ..."

53 **Manuscript Revision:** text changed to read "... large increases in ozone have been observed downwind
54 ..."

55 **line 32 (and lines 38 and 182 and elsewhere):** The NO-induced chemiluminescence measurement of
56 ozone is repeatedly described as "interference-free", which is misleading - it has a known dependency on
57 water vapor, which can lead to sensitivity variations of up to 8% if not accounted for. Please rephrase.

58 **Response:** The authors will rephrase these statements to emphasize that sample treatment steps, including
59 the use of a drier, must be taken prior to analysis to remove the effects of water vapor.

60 **Manuscript Revision:** Removed “interference free” from line 32. Removed “interference free” from
61 line 38. Clarifying statement inserted in line 62 “Both the ET-CL and NO-CL methods are subject to slight
62 interferences by water vapor. However, these potential interferences can be eliminated through the use of
63 Nafion based drier or equivalent sample water vapor treatment system.” Removed “interference free”
64 from line 182.

65 **line 52:** for clarity, please change to "... generates nitrogen dioxide in an electronically excited state..."
66 The original citation is Clough and Thrush, 1966, Chemical Communications, 728, pp. 783-784.

67 **Response:** The authors agree with this suggestion and will change the text accordingly.

68 **Manuscript Revision:** Text changed to read "... generates nitrogen dioxide in an electronically excited
69 state..."

70 **line 93:** please remove CO₂, as its absorption is negligible at 254 nm.

71 **Response:** The authors agree with this suggestion and will remove CO₂ from the text accordingly.

72 **Manuscript Revision:** CO₂ removed from the text

73 **line 142:** "...a supply of NO gas..." is not always needed - line 213 refers to one implementation of the
74 "scrubberless" UV absorption method uses a supply of N₂O gas and produces NO by photolysis.

75 **Response:** The authors agree with this suggestion and to clarify will rewrite the sentence to read: "The
76 SL-UV method requires a continuous supply of compressed NO or nitrous oxide (N₂O) (which the
77 instrument converts to NO) to serve as the scrubber gas.

78 **Manuscript Revision:** Sentence rewritten to read "Similar to NO-CL, the SL-UV method requires a
79 continuous supply of compressed NO or nitrous oxide (N₂O) (which the instrument converts to NO) to
80 serve as the scrubber gas.

81 **lines 230 - 237:** Details of power, generator, charger, and batteries are tangential to the performance of
82 the analyzers and could be eliminated to shorten the text.

83 **Response:** The authors agree with this suggestion and will review the manuscript and eliminate non-
84 relevant text that will shorten the document.

85 **Manuscript Revision:** Details of power, generator, charger, and batteries and other non-relevant material
86 removed from the manuscript text.

87 **lines 267-8:** "...calibrations for THC were performed using... a methane/propane gas cylinder..." This
88 work eventually concludes that VOCs are "likely to interfere with UV absorption measurements of O₃",
89 no surprise there. What is surprising is the rudimentary approach to quantifying those VOCs in this
90 manuscript. FID response factors vary with carbon number (for example, by up to a factor of 3 between
91 methane and propane!), between aliphatic, aromatic, and cyclic structures, and with heteroatomic
92 functionality. A sentence noting the uncertainty introduced in their measurement of VOCs (here called
93 THC) by using only methane and propane to determine FID sensitivity would be appropriate here.

94 **Response:** The authors agree with this comment and will add a sentence to address the uncertainty
95 associated with our use of the THC method and its calibration procedure to approximate VOC
96 concentrations.

97 **Manuscript Revision:** The THC calibration text was rewritten as follows to emphasize that the THC
98 results are an approximation of THC concentration in smoke "Per the manufacturer provided operators
99 manual, calibrations for THC were performed using the T700U calibrator and a certified EPA
100 methane/propane gas cylinder (Airgas). FID response factors for organic compounds can vary
101 significantly based upon factors such as carbon number and compound class (Tong and Karasek 1984).
102 The carbon numbers for methane and propane vary by a factor of three and the FID response factors for
103 those compounds may also vary by a similar amount. In addition, the complex mixture of hydrocarbons
104 found in smoke will have large variations in carbon number and FID response factors. As such, the results
105 obtained with the THC analyzer are an approximation of THC (and VOC) concentrations in smoke. In

106 addition, for THC calibrations, the T701H zero air generator was replaced with scientific grade zero air
107 compressed gas cylinders (Airgas).”

108 **Figure 2:** This is not a good graphic. There is absolutely no information conveyed by the third dimension
109 of this graph; please turn this into a 2D bar graph and improve the legibility of the different hatches. The
110 high level of interference from the UV-C and UVC- H techniques overwhelms any useful information on
111 the other techniques - suggest plotting only to 50 ppb and annotating the UV-C maxima with text. These
112 data are presented as O3 in ppb - what is the correct, or expectation value? The NO-CL data are lost in
113 this presentation and should be emphasized as the correct value.

114 **Response:** The authors agree the reviewers comment. The figure will be reformatted into 2D and
115 assuming that AMT allows colored figures will include a color scheme to improve clarity and viewability.
116 In addition, the y axis scale will be reduced to 50 ppb and the average values for all methods will be
117 included in the figure as text. The figure caption will be revised to reflect these changes.

118 **Manuscript Revision:** Figure 2 was reformatted into 2D and a color scheme added to improve
119 viewability. The y-axis scale was capped at 50 ppb and the average values for all methods and study
120 periods were included as text in the figure.

121 **Figure 4:** The NO-CL reference trace in the upper figure is the hardest to see; these figures could use
122 some work for legibility. The text refers to positive artifacts for the UV methods during burning periods,
123 ascribed to interferences from VOCs and PM2.5. Another problematic feature is the negative artifact
124 when the chamber is flushed with outside air, where the UV-C method falls below the NO-CL method
125 (bottom panel). Why is that? Did I miss the explanation?

126 **Response:** The authors will work on this time series as well as others to make the figures more legible
127 including looking into using a different scale on the y-axis. The post burn calibration checks on April 23,
128 2018 revealed a +8 % bias in the NO-CL method and a -2 % bias in the UV-C-H method. These biases
129 were evident during the chamber flush periods on that day. Each analyzer was re-zeroed and spanned
130 resulting in the elimination of the bias between the two methods as observed in the results from the
131 subsequent day (April 24, 2018). This will be addressed in the figure caption.

132 **Manuscript Revision:** Figure 4 was reformatted to include a logarithmic scale for O3 concentrations
133 making comparisons between the different methods more clear. The following text was added to the figure
134 caption to address the bias observed during the chamber flush periods “The post burn calibration checks
135 on April 23, 2018 revealed a +8 % bias in the NO-CL method and a -2 % bias in the UV-C-H method.
136 These biases were evident during the chamber flush periods on that day. Each analyzer was re-zeroed and
137 spanned resulting in the elimination of the bias between the two methods as observed in the results from
138 the subsequent day (April 24, 2018).”

139 **Lines 378-388:** I could not follow the confusing thread discussing how and when the MnO2 scrubber
140 failed in these experiments - for clarity I'd recommend deleting this section and removing all data taken
141 with an inoperative scrubber.

142 **Response:** The scope of this paper is a comparison/evaluation of ozone monitoring methods in smoke
143 and the damage to the converter occurred while operating the UV-C analyzer in heavy smoke, the authors
144 feel that this potential measurement issue is very important to those utilizing these instruments and should
145 at a minimum be mentioned in this manuscript. The converter issue is important in that the effect
146 continuous long after the smoke exposure is over and is not obvious when conducting typical QA/QC
147 reviews (e.g., zero/span calibrations and checks). The authors will add/remove text to clarify when the
148 damage occurred and the impact that the damaged converter had on the results obtained with the UV-C
149 method.

150 **Manuscript Revision:** The authors clarified some text in this section but feel the section is well explained
151 as to when the damage occurred and the overall impact. The section now reads “During the 2018 chamber
152 burns the UV-C results were biased high by 15-20 ppb even during non-burn (i.e., overnight) periods as
153 evident in Fig. 4 (top panel) and Fig. S4. The initial hypothesis was that the bias was associated with high
154 chamber backgrounds of interfering species due to years of heavy burning in the chamber. However, it
155 was later discovered during a subsequent summer/fall 2018 ambient air study in North Carolina in the
156 absence of smoke, that sampling heavy smoke plumes during the fall 2017 prescribed grassland burns
157 irreversibly damaged the MnO₂ scrubber in the UV-C instrument. The effect of the bias was observed
158 mainly when sampling ambient air and not readily observed during routine calibration checks (zeroes and
159 spans) except for an increase in the time required to obtain stable zero and span values. During the
160 summer/fall 2018 North Carolina study and prior to the start of the 2019 chamber burns, a new MnO₂
161 scrubber was installed and resulted in a significant and immediate reduction of the observed high bias,
162 shown in Fig. 4 (bottom panel) and Fig. S5.”

163 **Table 3:** Since there appears to be very large fire-to-fire and technique-to-technique variability in the
164 interferences, with no consistent dependence on any of the variables measured, quantifying their precise
165 values in a table seems not very useful. I’m not sure what information this table provides; what
166 quantitative use is it? Recommend deleting.

167 **Response:** The authors disagree with this comment. Regardless of the burning conditions or techniques
168 used, artifacts in the UV photometric methods were observed and are presented in this table. The authors
169 intend to include Table 3 in the manuscript.

170 **Manuscript Revision:** None

171 **line 498:** This section recommends using Nafion dryers to minimize smoke interferences in UV
172 absorption ozone measurements. This begs the question - under what range of conditions does the use of
173 a Nafion dryer allow EPA to actually accept an ozone measurement by the UV absorption measurement?
174 Please discuss.

175 **Response:** This comment goes beyond the scope of this paper which is primarily focused
176 evaluation/comparison of ozone monitoring methods in smoke plumes. However, the authors intend to
177 include an additional implication section that will discuss the potential impact of our findings on real
178 world monitoring application at sites that might be impacted by nearby wildfire smoke plumes.

179 **Manuscript Revision:** An implication section was added immediately preceding the conclusion section
180 that discusses the potential impact of our findings on real world monitoring application at sites that might
181 be impacted by nearby wildfire smoke plumes.

182 **Table 4:** Same comment as for Table 3, above: "Since there appears to be very large fire-to-fire and
183 technique-to-technique variability in the interferences, with no consistent dependence on any of the
184 variables measured, quantifying their precise values in a table seems not very useful. I'm not sure what
185 information this table provides; what quantitative use is it? Recommend deleting."

186 **Response:** The authors disagree with this comment. Regardless of the burning conditions or techniques
187 used, artifacts in the UV photometric method were observed and those artifacts are correlated with makers
188 of combustion as illustrated in this table. The authors intend to include Table 4 in the manuscript.

189 **Manuscript Revision:** None

190 **line 581:** I would suggest the authors review and cite the use of perfluorosulfonate membrane tubing to
191 remove UV-active hydrocarbons, e.g., in SO₂ pulsed fluorescence instruments (Luke, W., 1997, JGR,
192 102, 16,255-16,265).

193 **Response:** The authors will review the suggested manuscript and if appropriate cite in the text as a
194 possible solution in mitigating interferences by wildfire generated UV-active hydrocarbons as suggested
195 by the reviewer.

196 **Manuscript Revision:** None. The authors reviewed the suggested manuscript and choose not to cite it
197 in this manuscript. The authors could not find mention of perfluorosulfonate membrane in the manuscript
198 which is similar to the make up of Nafion but did notice several instances of the proprietary "kicker" that
199 may or may not remove interfering hydrocarbons.

200

201 **Reviewer 2 Comments – Responsenses and Manuscript Revisions**

202 **General Comments:** .

203 This study compares O₃ measurement techniques in fresh, concentrate smoke plumes. The authors sample
204 smoke plumes from both prescribed prairie grass burns and controlled chamber burns using a NO
205 chemiluminescence measurement as the interference-free standard with which to compare several
206 iterations of UV absorption-based measurements. This study is motivated by the prevalence of UV-based
207 O₃ analyzers at EPA air quality monitoring stations and the increasing impact of fire emissions on local
208 and regional air quality. Although these comparisons provide insight into the potential for UV-active
209 VOCs in smoke plumes to generate positive artifacts in the UV-based O₃ measurements, a more
210 quantitative assessment is limited by the lack of detailed VOC measurements and the inability to
211 quantitatively disentangle the various CO-O₃ regimes. The authors also suggest the role of Nafion in
212 mitigating potential artifacts, but do not provide enough information on the relative humidity conditions
213 during the various sampling periods or the potential for interactions between water vapor and VOC.
214 Further, the analysis emphasizes the effects of VOC interferences in near-fire smoke plumes but does

215 not provide much discussion on how the potential for interference diminishes with plume age and
216 dispersion. For example, how quickly do VOC react/diffuse to the point where their levels are no longer
217 of concern? How many ozone monitoring sites would be practically affected by these interferences?

218 **Response:** The authors appreciate the time required to provide the review and feel that the suggestions
219 provided by the reviewer will result in an improved manuscript for resubmission.

220 During both the prescribed and chamber burns, data were obtained for RH values and water vapor
221 concentration and is included in the data associated with this paper that will be provided through the EPA
222 Science Hub Web site (<https://catalog.data.gov/dataset/epa-sciencehub>) following the acceptance of this
223 paper. . However the correlations between RH and the magnitude of the ozone artifact were not
224 significant and therefore not included in the manuscript. In general, both the prescribed fire and chamber
225 burns were conducted under dry conditions with $RH \leq 50\%$. Past studies, which are now referenced in the
226 updated manuscript indicate that at those RH values humidity effects are expected to have little to no
227 impact. It is the intention of the authors to add an additional section to this manuscript discussing
228 implications of this research on real world ozone monitoring such as that that occurs at State and local
229 monitoring sites. The authors intend to review data from sites downwind of wildfires that potentially show
230 the artifact in the UV-C O₃ method and how it is correlated with markers of combustion processes. As
231 stated in the text of the manuscript, the authors plan future studies to dig deeper into the hypothesized
232 VOC caused artifact and which will include, as the reviewer suggest looking into interaction between
233 VOCs and water vapor and the capabilities of Nafion in removing certain VOCs.

234 **Manuscript Revision:** An additional “Implications” section was added prior to the conclusions section
235 to tie the results of the research detailed in this manuscript to real world monitoring applications. Included
236 in this section will be a review of data from monitoring sites downwind of fires to show the impact of the
237 measurement artifacts described in this manuscript.

238

239 **Specific Comments:**

240 **L243-244:** Is there any dependence of the artifact magnitude on distance from the active fire line? How
241 quickly do the VOC react/diffuse to the point where their levels are no longer detectable as a positive
242 artifact? All the measurements presented are taken within ~100 m from the fires, but any data collected
243 from aged smoke would be a useful counterpoint.

244 **Response:** The authors did not look at the dependencies of the artifact magnitude on distance from the
245 active fire line. However, the authors do agree that a more detailed look at data collected at sites being
246 impacted by aged smoke (ex. State and local monitoring sites being impacted by nearby wildfires). This
247 would aid in tying these measurements made in or near plume back to real world monitoring situations.
248 Most likely this will be done by adding an implications section prior to the manuscript conclusion.

249 **Manuscript Revision:** An implications section was added prior to the conclusion to address some of
250 reviewer 2 comments.

251 **L262:** The authors mention a +/- 10% performance objective between analyzers. Do the calibrations
252 reveal any systematic offset between the CL and UV analyzers? In describing the prescribed and chamber
253 burns, the authors mention varying moisture content in the burn material. Did the authors observe whether
254 the wetter grasses produced more VOC (lower combustion efficiency) in any systematic way?

255 **Response:** The calibrations only revealed a significant offset during one period during this study. The
256 post burn calibration checks on April 23, 2018 revealed a +8 % bias in the NO-CL method and a -2 %
257 bias in the UV-C-H method. These biases were evident during the chamber flush periods on that day.
258 Each analyzer was re-zeroed and spanned resulting in the elimination of the bias between the two methods
259 as observed in the results from the subsequen day (April 24, 2018). All other calibrations did not reveal
260 any systematic offsets or biases between the different analyzers and we will clarify this in the updated
261 version of the manuscript. At present the authors have not investigated the relationship between fuel
262 moisture content and VOC production. In order to simulate a range of natural burning conditions, the
263 chamber burns manipulated the moisture content, fuel type (pine needles, pine needles + fine woody
264 debris), and bulk density of the fuelbeds. These fuelbed properties influence the relative mix of flaming
265 and smoldering combustion and the chamber burns covered a range of combustion efficiencies (modified
266 combustion efficiencies of 0.85 – 0.97). The authors will investigate further and address these findings in
267 a future manuscript.

268 **Manuscript Revision:** The following text was added to the figure caption to address the bias observed
269 during the chamber flush periods “The post burn calibration checks on April 23, 2018 revealed a +8 %
270 bias in the NO-CL method and a -2 % bias in the UV-C-H method. These biases were evident during the
271 chamber flush periods on that day. Each analyzer was re-zeroed and spanned resulting in the elimination
272 of the bias between the two methods as observed in the results from the subsequen day (April 24, 2018).”

273 The following text was also added to section 3.2 “The post burn calibration checks on April 23, 2018
274 revealed a +8 % bias in the NO-CL method and a -2 % bias in the UV-C-H method. These biases were
275 evident during the chamber flush periods on that day. Each analyzer was re-zeroed and spanned resulting
276 in the elimination of the bias between the two methods as observed in the results from the subsequen day
277 (April 24, 2018).” No other calibration corrections werer made during the 2018 and 2019 chamber
278 studies.”

279 **Figure 4:** In general, the scale mismatch on the O3 timeseries makes immediate comparison between
280 methods difficult. The authors should perhaps switch to a log-scale on the y-axis that can effectively
281 compare low and high concentrations and offsets in both smoke plumes and background air. The authors
282 attempt to explain the positive offset of the UV-C method outside of the burning period, but there is also
283 a significant negative offset in the UV-C-H method that is not discussed. Could the authors provide more
284 insight on why the UV-C-H and NO-CL techniques disagree in background air?

285 **Response:** The authors will work on this time series plot as well as others to make the figures more legible
286 including looking into using a different scale on the y-axis. As suggested by the reviewer, the authors will
287 provide more insight into why the UV-C-H and NO-CL techniques disagree in background air.

288 **Manuscript Revision:** Figure 4 was reformatted to include a logarithmic scale for O₃ concentrations
289 making comparisons between the different methods more clear.

290 **L378+:** If the damaged MnO₂ scrubber ineffectively removed O₃, I would expect the UV-C measurement
291 to be biased low in background air rather than high. Please elaborate on the mechanism of MnO₂ damage
292 resulting in a significant positive offset. Also, it's unclear when the scrubber damage became an issue.
293 Did it affect data from the 2017 prescribed burns?

294 **Response:** In order for the scrubber to work correctly, it must remove O₃ and only O₃. Based upon the
295 data, the damage most likely resulted in the scrubber also removing significant amounts of interfering
296 species during the reference measurement which would then be detected as ozone during the sample
297 measurement resulting in the positive artifact. The data collected during the 2017 prescribed burns
298 indicate that the scrubber was functioning properly in that there was excellent agreement between the
299 UV-C and NO-CL methods when sampling out of the smoke plume.

300 **Manuscript Revision:** To clarify the section describing the bias observed during the 2018 chamber
301 studies was re-written as follows: "During the 2018 chamber burns the UV-C results were biased high by
302 15-20 ppb even during non-burn (i.e., overnight) periods as evident in Fig. 4 (top panel) and Fig. S4. The
303 initial hypothesis was that the bias was associated with high chamber backgrounds of interfering species
304 due to years of heavy burning in the chamber. However, it was later discovered during a subsequent
305 summer/fall 2018 ambient air study in North Carolina in the absence of smoke, that sampling heavy
306 smoke plumes during the fall 2017 prescribed grassland burns followed by subsequent storage of the UV-
307 C analyzer, irreversibly damaged the MnO₂ scrubber in the UV-C instrument. It is hypothesized that the
308 damage resulted in the scrubber removing some of the interfering species in addition to ozone, preventing
309 them from being removed in the reference measurement, and subsequent detection as ozone (positive bias)
310 during the measurement cycle. The effect of the bias was observed mainly when sampling
311 ambient/chamber air and not readily observed during routine calibration checks (zeroes and spans) except
312 for an increase in the time required to obtain stable zero and span values. The bias was not observed
313 during any of the 2017 prescribed grassland burns. During the summer/fall 2018 North Carolina study
314 and prior to the start of the 2019 chamber burns, a new MnO₂ scrubber was installed and resulted in a
315 significant and immediate reduction of the observed high bias, shown in Fig. 4 (bottom panel) and Fig.
316 S5."

317 **Figure S9** indicates there is potential artifact even <1-2 ppm CO. Do these plots just use data from the
318 burn periods or include points when chamber is flushed with outside air?

319 **Response:** Figure S9 includes data from the burn periods only. In the figure caption it describes it as "in-
320 plume". The authors will add clarifying text similar to the following, "...and THC for all in-plume (burn
321 period only) measurements...".

322 **Manuscript Revision:** The figure caption was re-written as follows: "Scatter plots between FRM and
323 FEM O₃ differences and CO, NO₂, and THC for all in-plume (burn period only) measurements made
324 during the 2018 and 2019 Missoula Fire Chamber studies. Observation points have been colored by the
325 O₃ instrument. Over all observations there is little correlation between the O₃ instrument differences,

326 but straight line structures within the overall scatters indicate that individual burn events measured in
327 the chamber have good correlations with distinct ratios.”

328 **L459-461:** How does the residence time and sample rate vary for each instrument?

329 **Response:** Sampling rates and hence residence times are going to be similar for all instruments as they
330 all operate with similar flow rates. The authors will address this comment by either adding analyzer flow
331 rate to Table 1 or by inserting text in the Methods section under each corresponding analyzer type.
332 Generally, UV photometric type analyzers require a greater flow rate because the flow is split between
333 the two cells (reference and measurement). The NO-CL method has only a single cell and requires a much
334 smaller flow rate to achieve a similar residence time.

335 **Manuscript Revision:** The flow rates of each method along with manufacturer reported performance
336 specifications were included in Table S1 which was added to the supplementary materials document. In
337 the text describing each method, a sentence similar to the following was added “Manufacturer provided
338 performance specifications for the NO-CL based TAPI T265 are given in Table S1.”

339 **Table 4:** The slope and intercept uncertainties should be included with the fit parameters. How different
340 are the range of fitted slope values statistically? In general, there is lack of uncertainty treatment in the
341 paper. How do the uncertainties compare for each measurement technique? This information should be
342 included in the manuscript.

343 **Response:** The authors agree with this comment and will work to include uncertainties (both in tables
344 and in the text) of measurement methods and in fit parameters associated with regression statistics.

345 **Manuscript Revision:** Data for the Konza March 2017 were re-analyzed and new values included for
346 slope, intercept R2 and n. The previous analysis included a few values that were associated with CO
347 levels that were below 1 ppm (our threshold of sampling in plume). Standard errors for the regression
348 slope and intercept were included in Table 4. In addition, the following text was added to discuss the
349 results of the regression analysis between markers of combustion CO and THC and the magnitude of the
350 ozone artifact: “The slight differences in the magnitude of the artifacts (fitted regression slopes) along
351 with the low uncertainty (standard errors) values indicate that the magnitude of the artifact may be
352 influenced by local conditions that make each burn unique. Such conditions might include meteorological
353 conditions, fuel composition, fuel moisture content, and times spent in combustion phase (flaming vs
354 smoldering).”

355 **L550-552:** See question 1 above. How close to the plume do you have to be for interferences to matter?
356 Is this relevant for air quality monitoring stations not located in the immediate vicinity of the fire line?

357 **Response:** The authors focused on determining if significant ozone measurement artifacts do occur in
358 near-field smoke events and did not look at the dependencies of the artifact magnitude as a function of
359 distance from the active fire line. However the authors do agree that a more detailed look at data collected
360 at sites being impacted by aged smoke (ex. State and local monitoring sites being impacted by nearby
361 wildfires) and are currently collecting this data as part of the EPA MASIC study in Boise, ID; Missoula,
362 MT; and Reno, NV. This additional data collection will aid in linking these research chamber and near

363 field prescribed grassland burn measurements back to real world regulatory monitoring situations. We
364 will address these issues in a new “implications” section prior to the manuscript conclusion.

365 **Manuscript Revision:** An implications section was added to the manuscript prior to the conclusion to
366 address this and other comments provided by reviewer 2.

367 **L554:** What is estimated CO- Δ O₃ correlation for the chamber studies? It would still be worthwhile to
368 include this information in the supplement.

369 **Response:** Regarding the correlation between Δ O₃ and CO from the chamber based burns, the authors
370 refer the reviewer to the original manuscript text:

371 “As indicated, Δ O₃(UV-C) and CO appear to be correlated in time but when performing linear regression
372 comparisons of Δ O₃(UV-C) and CO during each years chamber burns as a whole, correlations tend to be
373 poor. We suspect the positive O₃ bias is driven by one or more VOCs (likely oxygenated VOCs). In fresh
374 smoke the excess concentrations of individual VOCs (Δ X), and VOC sums (Δ VOC), tend to be highly
375 correlated with Δ CO (Yokelson et al., 1999; Gilman et al. 2015). The emission ratios of individual VOCs
376 to CO (Δ X/ Δ CO) can vary considerably with combustion conditions such as fuel type and condition (e.g.
377 moisture content and decay state), fuel bed properties, such as bulk density, and the relative mix of
378 flaming and smoldering combustion (Gilman et al. 2015; Koss et al., 2017). Additionally, the response
379 of Δ X/ Δ CO to burn conditions varies among VOCs. When each burn is considered individually or in
380 groups with similar conditions, the correlations between Δ O₃, CO, and THC are enhanced. An example
381 of this behavior is shown in Supplementary Fig. S10.”

382 With that being stated, the authors will consider adding the CO- Δ O₃ correlation (both for the entire
383 chamber study period and also a subset of individual burns) either in Table 4 or in the body of the text
384 give evidence to the above statement. Visual representations of the correlations are given in Figures S9
385 and S10.

386 **Manuscript Revision:** The following text was added to section 3.4 to address this comment: “For the
387 chamber burns the magnitude of the ozone artifacts in ppb apparent O₃ per ppm CO, ranges between 6 -
388 210 ppb ppm⁻¹ for the individual burns. R² and standard error values were consistent with those observed
389 during the prescribed burns (see Table 4). “ In addition, the requested information is provided visually in
390 figures S9 and S10.

391 **Figures S9 and S10:** Can you demonstrably separate CO- Δ O₃ regimes based on “burn condition”? The
392 authors allude to this in the text (L563) and show an individual burn in Fig S10, but a more in-depth
393 analysis of the contributing burn condition factors would provide a more quantitative and perhaps
394 predictive assessment of how CO links to O₃ artifacts under the varied burn conditions. The authors also
395 perform separate regressions for NO₂ and THC, but a separate correlation with humidity might be
396 illustrative (if the data exists).

397 **Response:** The authors will consider elaborating further per the reviewers suggestion on CO- Δ O₃
398 regimes based on burn conditions (i.e., individual burns or burns grouped by similar burn conditions).
399 The authors previously attempted to establish a correlation between Δ O₃ and humidity (water vapor

400 concentration) but those correlation were extremely poor. As such the authors chose not to include this
401 analysis.

402 **Manuscript Revision:** The following text was added to section 3.4 to elaborate on the lack of correlation
403 between ΔO_3 and CO when considered as a whole but showing improvements when considering
404 individual burns: “For the chamber burns the magnitude of the ozone artifacts in ppb apparent O_3 per ppm
405 CO, ranges between 6 - 210 ppb ppm⁻¹ for the individual burns. R^2 and standard error values were
406 consistent with those observed during the prescribed burns (see Table 4).”

407 **L571:** Is it possible that interactions between water vapor and VOC somehow compound the VOC effect?
408 In other studies (e.g., Spicer et al. 2010, Turnipseed et al. 2017), Nafion alone seems to play little role in
409 mitigating VOC artifacts but does significantly reduce water vapor artifacts. In drier environments, does
410 adding Nafion affect the positive artifact magnitude? This would be more conclusive evidence that Nafion
411 does in fact remove certain permeable VOC species.

412 **Response:** Both the 2017 prescribed fire and 2018-2019 chamber based burns were conducted under dry
413 conditions ($RH \leq 50\%$) and humidity interferences are expected to be minimal. As stated in the previous
414 comment, the correlation between in plume water vapor concentration and ΔO_3 was not significant. In
415 addition, there is no significant correlation between the magnitude of the artifact and RH. In both the
416 prescribed grassland and chamber burns there was a UV instrument with a Nafion drier and a UV
417 instrument without the drier and they were operated simultaneously. The magnitude of the artifact (both
418 average and maximum) was greatly reduced in the method using the Nafion drier. This is evident in
419 comparing the magnitude of the UV-C artifact with that of the UV-C-H (UV method employing a Nafion
420 based drying system. In all cases, the UV-C artifact was nearly an order of magnitude greater than that of
421 the UV-C-H. This is also became further evident when the Nafion drier was added to the UV-C method
422 on the final day of burning during the 2018 chamber studies, thus reducing the magnitude of the UV-C
423 artifact to a point comparable to that of the UV-C-H method. The effect of Nafion on the magnitude of
424 the artifact is detailed in section 3.3. In section 3.5 of the manuscript, the authors will attempt to clarify
425 that in addition to our hypothesis of certain VOCs being removed by the Nafion, there may also be
426 interactions between water vapor and VOCs that may be confounding the observed artifact.

427 **Manuscript Revision:** The authors feel that text and discussion provided in section 3.3 already provide
428 a response to the reviewer 2's comment suggestion. As stated in the response listed above, during this
429 study humidity effects are expected to be at a minimum due to the low RH values that existed during all
430 study periods. As such and to clarify, the following text was inserted in section 2.6: “In general, chamber
431 RH values were below 50% facilitating dry burning condition.” And section 3.1: “In addition, ambient
432 RH values were generally below 50% indicating that the spring and fall 2017 prescribed burns were
433 conducted under dry conditions.”

434 The last sentence of section 3.4 was re-written to read “Considering that the prescribed grassland and
435 chamber burns were conducted under dry conditions, the size of the difference (as large as hundreds of
436 ppb) cannot be explained purely by the previously observed relative humidity effects on measurements
437 (Leston et al., 2005; Wilson et al., 2006), suggesting that the Nafion® dryer is directly impacting the
438 concentrations of other interferences in the sample stream.”

439 **L605:** Could this also be confounded by the faulty MnO₂ scrubber?

440 **Response:** We do know that during the 2018 chamber studies the damaged scrubber did cause an
441 approximate +10-15 ppb bias in the UV-C method which was present even in the absence of smoke. At
442 the end of the 2018 chamber studies, the authors added a Nafion drier to the UV-C method as indicated
443 in Figure 4. The addition of the Nafion to the UV-C method reduced the magnitude of the artifact by a
444 factor of three making it compatible to the artifact observed for the UV C-U method. The addition of the
445 nafion did result in a slight reduction in the bias that we attributed to damaged scrubber but not on the
446 order of 3X. We suspect that the addition of the drier would reduce or remove many of the VOC species
447 prior to also being removed by the faulty scrubber thus resulting in a reduction of the bias but not
448 completely eliminating it. The authors will add clarifying text in the body of the manuscript explaining
449 the damage to the MnO₂ scrubber and its hypothesized effect on the observed bias. The reviewers
450 comment would only apply to the 2018 chamber study as the MnO₂ scrubber in the UV-C method was
451 functioning properly during all other studies.

452 **Manuscript Revision:** Clarifying text was added in section 3.2 to explain the effect that the damaged
453 scrubber had on the UV-C ozone results (positive bias).

454

455 **Technical Corrections:**

456 **Table 1:** Add uncertainty associated with each measurement technique. Sample rate would also be useful.

457 **Response:** The authors will address this comment by either adding analyzer flow rate and uncertainties
458 to table 1 or by inserting text in the Methods section under each corresponding analyzer type.

459 **Manuscript Revision:** An additional table (Table S1) was added to the supplemental materials document
460 containing manufacturer provided performance specifications for each analyzer to address this comment
461 from reviewer 2. In the text describing each method, a sentence similar to the following was added
462 “Manufacturer provided performance specifications for the NO-CL based TAPI T265 are given in Table
463 S1.

464 **Figure S1 and other timeseries in general:** It’s difficult to compare NO-CL and UV measurements of
465 plumes and background air given the large mis-match in scale. Some other way of presenting this material
466 (e.g., semi-log) might help the visual comparison. The lines are also not very easy to distinguish. Using
467 different colors instead of just patterns would help.

468 **Response:** The authors agree with this comment and will take steps to improve the the time series plots,
469 including looking into different scales (e.g. semi-log) and also using colored lines in the figures.

470 **Manuscript Revision:** Figures 4 and S1-5 were reformatted adding logarithmic scales where appropriate
471 and color schemes to improve readability.

472 **Figure 2:** Does not need to be in 3D and could use a color scheme instead of patterns.

473 **Response:** The authors agree with the reviewers comment. The figure will be reformatted into 2D and
474 assuming that AMT allows colored figures will include a color scheme to improve clarity and view ability.
475 In addition, the y axis scale will be reduced to 50 ppb and the average values for all methods will be
476 included in the figure as text. The figure caption will be revised to reflect these changes.

477 **Manuscript Revision:** Figure 2 was reformatted into 2D and a color scheme added to improve
478 viewability. The y-axis scale was capped at 50 ppb and the average values for all methods and study
479 periods were included as text in the figure.

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