



Long-term NO_x measurements in the remote marine tropical troposphere

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15 Abstract

16	Atmospheric nitrogen oxides $(NO + NO_2 = NO_x)$ have been measured at the Cape Verde
17	Atmospheric Observatory (CVAO) in the tropical Atlantic (16° 51' N, 24° 52' W) since
18	October 2006. These measurements represent a unique time series of NO_x in the background
19	remote troposphere. Nitrogen dioxide (NO2) is measured via photolytic conversion to nitric
20	oxide (NO) by ultra violet light emitting diode arrays followed by chemiluminescence
21	detection. Since the measurements began, a "blue light converter" (BLC) has been used for
22	NO_2 photolysis, with a maximum spectral output of 395 nm from 2006-2015 and of 385 nm
23	from 2015. The original BLC used was constructed with a Teflon-like material and appeared
24	to cause an overestimation of NO_2 when illuminated. To avoid such interferences, a new
25	additional photolytic converter (PLC) with a quartz photolysis cell (maximum spectral output
26	also 385 nm) was implemented in March 2017. Once corrections are made for the NO_2 artefact
27	from the original BLC, the two NO_2 converters are shown to give comparable NO_2 mixing
28	ratios (PLC = $0.92 \times BLC$, $R^2 = 0.92$), giving confidence in the quantitative measurement of
29	NO_x at very low levels. Data analysis methods for the NO_x measurements made at CVAO have
30	been developed and applied to the entire time series to produce an internally consistent and
31	high quality long-term data set. NO has a clear diurnal pattern with a maximum mixing ratio
32	of 2-10 pptV during the day depending on the season and ~0 pptV during the night. NO ₂ shows
33	a fairly flat diurnal signal, although a small increase in daytime NO_x is evident in some months.
34	Monthly average mixing ratios of NO_2 vary between 5 and 30 pptV depending on the season.
35	Clear seasonal trends in NO and NO_2 levels can be observed with a maximum in autumn/winter
36	and a minimum in spring/summer.

37





38 1 Introduction

Atmospheric nitrogen oxides play a key role in tropospheric chemistry. NO_x helps to control the abundance of the two most important oxidants in the atmosphere, ozone (O₃) and the hydroxyl radical (OH). The presence of NO is usually the key limiting factor in the production of tropospheric O₃, which occurs via oxidation of NO to NO₂ by peroxy radicals (RO₂, HO₂) as described in reactions (1-2), followed by photolysis of NO₂ and rapid conversion of the resulting O(³P) to O₃:

$$45 \qquad \text{RO}_2 + \text{NO} \rightarrow \text{RO} + \text{NO}_2 \tag{1}$$

$$46 \qquad \text{HO}_2 + \text{NO} \rightarrow \text{OH} + \text{NO}_2 \tag{2}$$

47
$$\operatorname{NO}_2 + hv \to \operatorname{NO} + \operatorname{O}({}^{3}\mathrm{P})$$
 $(hv \le 410 \text{ nm})$ (3)

48
$$O(^{3}P) + O_{2} + M \rightarrow O_{3} + M$$
 (4)

49 Reaction (2) also offers a route to the OH radical, above its primary production via O_3 50 photolysis (reactions (5) and (6)). If the NO_x mixing ratio is sufficiently low, then peroxy 51 radicals react with themselves instead of NO, and O_3 depleting reactions (reactions (5) to (8)) 52 dominate over O_3 production (Atkinson, 2000).

53
$$O_3 + hv \to O(^1D) + O_2$$
 ($hv \le 335 \text{ nm}$) (5)

54
$$O(^{1}D) + H_{2}O \rightarrow 2 OH$$
 (6)

55
$$OH + O_3 \rightarrow HO_2 + O_2$$
 (7)

$$56 \qquad \text{HO}_2 + \text{O}_3 \rightarrow \text{OH} + 2 \text{ O}_2 \tag{8}$$

57 NO_x mixing ratios below 10-30 pptV are generally sufficiently low for net tropospheric O_3 58 depletion (Atkinson, 2000; Jaeglé et al., 1998; Logan, 1985). These conditions have previously 59 been reported to apply most of the year in the remote Atlantic Ocean (Lee et al., 2009). The 60 mixing ratio of NO_x in the atmosphere varies from a few pptV in remote areas (Lee et al., 2009; Monks et al., 1998; Reed et al., 2017) to >100 ppbV in polluted areas (Carslaw, 2005; Mazzeo 61 62 et al., 2005; Pandey et al., 2008). It is therefore important to have representative NO_x 63 measurements in different regions of the world to be able to understand the chemistry occurring 64 throughout the troposphere.

Long-term remote atmospheric NO_x measurements are rare due to the difficulty measuring
 very low (pptV) mixing ratios. Many different methods of measuring NO_x are available,





67 however, very few have the limit of detection (LOD) and sensitivity needed to measure NOx in remote regions. The most widely used method is NO chemiluminescence, where NO in the 68 69 presence of excess ozone is oxidized into excited state NO₂, which emits photons that can be 70 detected (Fontijn et al., 1970). NO₂ is generally converted into NO either catalytically by a 71 heated molybdenum converter or photolytically, followed by NO chemiluminescence (Kley 72 and McFarland, 1980). The molybdenum converter has historically been preferred due to its 73 high conversion efficiency of at least 95%, but it also converts other reactive nitrogen species 74 (NO_z) such as peroxyacetyl nitrate (PAN), peroxymethacryloyl nitrate (MPAN), acyl peroxy 75 nitrates (APN), HNO₃, p-HNO₃, and HONO, potentially giving an overestimation of NO₂ 76 (Dunlea et al., 2007; Grosjean and Harrison, 1985; Winer et al., 1974). Two separate studies 77 have shown that photolytic converters (PLC) with a wavelength of 385-395 nm have the 78 smallest spectral overlap with interfering compounds (Pollack et al., 2010; Reed et al., 2016). 79 Reed et al., (2016) showed that in some configurations the PLC can heat up the sampled air 80 making it possible for reactive nitrogen compounds such as peroxyacetyl nitrate (PAN) to decompose thermally and cause an overestimation of NO2. This, however, causes only a 81 82 negligible interference in warm regions such as Cabo Verde where PAN levels are extremely 83 low (Reed et al., 2016).

In this study we describe a NO_2 converter, similar to that presented by Pollack et al. (2010), which has been implemented on an instrument to measure NO_x at the CVAO. The data analysis procedure is explained in detail and the first two years of results with the new converter are presented and compared to the data obtained using a different converter.

88

89 2 Experimental

90 2.1 Location

The Cape Verde Atmospheric Observatory (CVAO; 16° 51' N, 24° 52' W) is located on the north eastern coast of São Vicente, Cabo Verde. The air masses arriving at the CVAO predominately come from the northeast (>95% of all wind direction measurements, see Figure 1) and have travelled over the Atlantic Ocean for multiple days since their last exposure to anthropogenic emissions, with the potential exception of ship emissions (Carpenter et al., 2010; Read et al., 2008). The UK Meteorological Office NAME dispersion model (Ryall et al., 2001) has previously been used to investigate the origin of the air masses arriving at the CVAO,





98 which have been shown to be very diverse; North America, the Atlantic, Europe, Arctic, and 99 African regions (Lee et al., 2009). During the spring and summer, the air masses predominantly 100 originate from the Atlantic making it possible to investigate long-term remote marine 101 tropospheric background measurements. During the winter, the CVAO receives air mainly 102 from the Sahara, resulting in very high wintertime dust loadings (Chiapello et al., 1995; Fomba 103 et al., 2014; Rijkenberg et al., 2008). The time zone of Cabo Verde is always UTC-1. A full 104 description of the CVAO site and associated measurements is given in Carpenter et al. (2010).

105

106 2.2 Measurement Technique

107 NO_x has been measured at the CVAO since 2006 using a NO_x chemiluminescence 108 instrument manufactured by Air Quality Design Inc. (AQD), USA. The chemiluminescence 109 technique involves the oxidation of NO by excess O₃ to excited NO₂ (Reaction 9) (Clough and 110 Thrush, 1967; Clyne et al., 1964; Fontijn et al., 1970). The excited NO_2 molecules can be 111 deactivated by emitting photons or by being quenched by other molecules (Reaction 10 and 11) such as N_2 , O_2 , and in particular H₂O. The emitted photons are detected by a 112 113 photomultiplier tube detector (PMT), which gives a signal linearly proportional to the mixing 114 ratio of NO sampled. The measurement of NO_x and NO_2 requires photolytic conversion of NO_2 115 to NO (Reaction 3) followed by NO chemiluminescence detection (Kley and McFarland, 116 1980).

117
$$\operatorname{NO} + \operatorname{O}_3 \to \operatorname{NO}_2^* + \operatorname{O}_2$$
 (9)

118
$$NO_2^* \to NO_2 + hv$$
 (*hv* > 600 nm) (10)

$$119 \qquad \text{NO}_2^* + \text{M} \to \text{NO}_2 \tag{11}$$

120
$$NO_2 + hv \to NO + O(^{3}P)$$
 $(hv \le 410 \text{ nm})$ (3)

Further details of the technique are documented in (Carpenter et al., 2010; Drummond et al.,
1985; Fontijn et al., 1970; Lee et al., 2009; Peterson and Honrath, 1999; Reed et al., 2017; Val
Martin et al., 2006).

124





125 2.3 Instrument Set-up

126 Ambient air is sampled from a downward facing inlet placed into the prevailing wind with 127 a fitted hood 10 m above the ground. A centrifugal pump at a flow rate of ~750 litres per minute 128 pulls the air into a 40 mm glass manifold resulting in a linear sample flow of 10 m s⁻¹, giving 129 a residence time to the inlet of the NO_x instrument of 2.3 s. To reduce the humidity and aerosol 130 concentration in the sampled air, dead-end traps are placed at the lowest point of the manifold 131 inside and outside the laboratory. A Nafion dryer (PD-50T-12-MKR, Permapure) is used to 132 additionally dry the sampled air, using a constant sheath flow of zero air (PAG 003, Eco Physics 133 AG) that has been filtered through a Sofnofil (Molecular Products) and activated charcoal 134 (Sigma Aldrich) trap (dewpoint -15° C). The air is sampled perpendicular to the manifold 135 through a 47mm PTFE (polytetrafluoroethylene) filter with a pore size of 1.2 μm.

136 A schematic diagram of the instrument is shown in Figure 2. Sampled air is passed through two different photolytic NO₂ converters, which are placed in series. The first is a commercial 137 138 unit known as a BLC (Blue Light Converter) supplied by Air Quality Design, as described in 139 (Buhr, 2007). An ultra violet light emitting diode (UV-LED, 3 W, LED Engin, Inc.) array is 140 placed in each end of a reaction chamber made of Teflon-like barium doped material (BLC, λ 141 = 385 nm, volume = 16 cm^3). The entire block surrounding the reaction chamber is irradiated, 142 giving the highest possible conversion efficiency of NO_2 . Each array is cooled by a heat sink to maintain an approximately constant temperature inside of the converter when the diode 143 144 arrays turn on. The second converter consists of two diodes (Hamamatsu Lightningcure 145 L11921-500, $\lambda = 385$ nm) and a photolysis cell made of a quartz tube and two quartz windows glued to each end with a volume of 16 cm^3 (PLC) following the design of Pollack et al. (2010). 146 147 Aluminium foil is wrapped around the quartz tube to increase the reflectivity to give the highest 148 conversion efficiency of NO₂. The diodes are placed at each end of the quartz tube, as shown 149 in Figure S2, without touching the windows to avoid increases in the temperature when the 150 diodes turn on. BLCs have been used at the CVAO since the instrument was installed in 2006, 151 with the most recent converter installed in April 2015 (a BLC2 model), where the wavelength 152 was changed to 385 nm from 395 nm. The PLC was installed in March 2017. The air flow 153 through the instrument is controlled at ~1000 sccm by a mass flow controller (MKS, M100B) 154 giving a residence time of 0.96 s through each of the converters.

To measure NO and NO_x (NO + NO₂ converted into NO) the air is introduced to the chemiluminescent detector (CLD), where NO is oxidized by excess O_3 into excited NO₂ in the





157 reaction volume (241 mL, aluminium with gold coating (Ridley and Grahek, 1990)) shown in Figure 2. The reaction volume is kept at low pressure to minimize quenching of excited NO_2 158 159 and thereby maximize the NO chemiluminescence lifetime. The photons emitted from the 160 excited NO₂ molecules when they relax to ground state are detected by the PMT (Hamamatsu 161 R2257P) to give a signal for NO. NO₂ is converted into NO by the BLC for 1 minute, and then 162 the PLC for 1 minute, each period producing a signal due to $NO + NO_2$. The signal detected 163 by the PMT (S_M) is caused by NO reacting with O_3 (S_{NO}) , dark current from the thermionic emissions from the photocathode of the PMT (S_D), and an interference (S_I) which can be due 164 165 to other gas-phase reactions creating chemiluminescence and from illumination of the chamber walls during NO₂ conversion (Drummond et al., 1985; Reed et al., 2016): 166

$$167 \qquad S_{\rm M} = S_{\rm NO} + S_{\rm D} + S_{\rm I} \tag{I}$$

168 The PMT is cooled to -30°C to reduce the dark current, giving the instrument a higher precision. Other molecules in the atmosphere such as alkenes also react with ozone and emit 169 170 photons to reach their ground state, but at a different time-scale to that of NO₂ (Alam et al., 171 2020; Finlayson et al., 1974). This can give an interfering signal causing the NO and NO_x 172 mixing ratios to be overestimated. However, most of these reactions emit photons at 400-600 173 nm and are therefore filtered by a red transmission cut-off filter (Schott RG-610) placed in 174 front of the PMT (Alam et al., 2020). The filter transmits photons with a wavelength higher 175 than 600 nm (Drummond et al., 1985). A background measurement is therefore required to 176 account for the dark current of the PMT and for the remaining interfering reactions occurring 177 at a different time-scale to that of NO_2 . Background measurements are made by allowing 178 ambient air to interact with O₃ in the zero volume (180 mL, PFA, Savillex, LLC) before 179 reaching the reaction volume (Figure 2). Most excited NO₂ molecules will reach their ground 180 state before the sample reaches the PMT, meaning the signal from NO will not be measured. 181 The efficiency of the reaction between NO and O_3 in the zero volume is calculated from the 182 calibration and will be explained in section 2.4.3.

NO, NO₂ and the background signal are all detected on the same channel, and the
instrument cycle is 1 min of background, 2 min of NO (when the NO₂ converters are off), 1
min of BLC NO_x (the BLC converter is on), and 1 min of PLC NO_x (the PLC is on).

186





187 2.4 Calibration

188 Prior to June 2019, calibrations were performed every 73 hours by standard addition in 189 order to account for temperature and humidity changes in the ambient matrix. In June 2019 the 190 calibration frequency was changed to every 61 hours to ensure that during any given month, 191 calibrations are carried out for approximately equal periods during the night and the day. To 192 calibrate the NO sensitivity, 8 sccm of 5 ppmV NO calibration gas in nitrogen is added to the 193 ambient air flow of ~1000 sccm, giving an NO mixing ratio of approximately 40 ppbV. The 194 mixing ratio used for calibrations are approximately 10,000 times that of the ambient 195 measurements, however, due to reduced cylinder stability for lower NO mixing ratios it is 196 difficult to calibrate at much lower mixing ratios and the chemiluminescence is expected to be 197 linear across the range of expected mixing ratios (Drummond et al., 1985). The calibration gas 198 is added between the PTFE filter and the NO₂ converter as shown in Figure 2. The conversion 199 efficiency of the BLC and the PLC is calibrated by gas phase titration (GPT), where oxygen is 200 added to the sampled NO calibration gas before entering the titration cell, which contains a UV 201 lamp that converts oxygen to ozone. Between 60-80% of the NO calibration gas is oxidized 202 into NO₂, giving a known mixing ratio of NO₂. A theoretical calibration sequence is shown in 203 Figure 3. The first cycle is to calibrate the sensitivity and the second is to calibrate the NO_2 204 conversion efficiency. Each actual calibration includes three cycles of sensitivity calibration 205 and two cycles of conversion efficiency calibration. The signal from NO₂ observed in the NO 206 sensitivity calibration is due to traces of NO2 in the calibration gas. Figure S3 shows the 207 observed percentage of NO₂ in the calibration cylinders from January 2014 to August 2019 208 calculated from the measured sensitivity (sec. 2.4.1) and the conversion efficiencies (CE) of 209 the two converters (sec. 2.4.2):

210 NO₂ in cylinder (pptV) =
$$\frac{(NO.c_{(1)} - NO_{(1)})}{Sensitivity \times CE}$$
 (II)

211 Percentage
$$NO_2 = \frac{NO_2 \text{ in cylinder}}{NO_2 \text{ in cylinder+NO cal conc.}}$$
 (III)

The percentage is stable for both converters, however, the PLC shows approximately 3-4% NO₂ in the NO calibration gas compared to 5-10% for the BLC, which is caused by a BLC artefact. The cylinders used were certified to $\leq 2\%$ NO₂.

215





216 2.4.1 Sensitivity

The sensitivity of the instrument is calculated from the increase in counts per second caused by the calibration gas during NO calibration (untitrated, i.e. without O₃) and from the mixing ratio of the calibration gas as shown by equation (IV). The NO counts per second from the previous measurement cycle before the calibration is subtracted to give the increase due to the calibration gas. The previous cycle needs to be stable and low in NO to give an accurate sensitivity, which is the case at the CVAO.

223
$$Sensitivity = \frac{Counts per second during calibration - Counts per second in previous cycle}{Mixing ratio of calibration gas}$$
(IV)

The sensitivity of the instrument depends on the pressure of the reaction chamber, the ozone mixing ratio in the reaction chamber, the flow of the sample through the reaction chamber, and the temperature of the reaction chamber. To maintain a stable sensitivity, all four parameters should be kept stable (Galbally, 2019). From January 2014 to August 2019 the sensitivity has varied between 2.7 and 7.4 counts s⁻¹ pptV⁻¹ with changes of less than 5% between subsequent calibrations (see Figure S4), unless the instrument has been turned off for a long period of time due to instrumental problems.

231

232 2.4.2 Conversion Efficiencies

The conversion efficiency of the BLC and the PLC is calculated based on the titrated (with added O_3) and the untitrated (without added O_3) NO calibration gas as described in equation (V). The numerator gives how much of the NO is titrated into NO₂ and the denominator represents the NO₂ measured when taking the NO₂ content in the NO calibration gas into account. In equation (V), "NO" is the measurement of only NO i.e. when the converters are off, NO.c is when one of the converters are on therefore the measurement is NO + NO₂ and (1) and (2) represent untitrated and titrated NO, respectively.

240
$$CE = \frac{\left[(NO.c_{(2)} - NO_{(2)}) - (NO.c_{(1)} - NO_{(1)})\right]}{\left[NO_{(1)} - NO_{(2)}\right]} = 1 - \frac{NO.c_{(1)} - NO.c_{(2)}}{NO_{(1)} - NO_{(2)}}$$
(V)

The conversion efficiency of the BLC has varied from 82% to 91% between its installation in April 2015 and August 2019 ($j \sim 3 \text{ s}^{-1}$). Prior to April 2015, an older generation BLC ($\lambda = 395$ nm) with a conversion efficiency of 30-35% was used ($j \sim 0.5 \text{ s}^{-1}$). The conversion efficiency of the PLC has varied between 50% and 55% from its installation in March 2017 to August 2019 ($j \sim 1 \text{ s}^{-1}$). See Figure S5 for all the calculated conversion efficiencies.





246

247 2.4.3 Efficiency of the Zero Volume

248 Background measurements are made by reacting NO and interference compounds with O_3 in the zero volume (Figure 2). The system is set up so that NO₂ produced from NO will relax 249 250 to the ground state before it is measured in the downstream reaction chamber, whereas it is 251 assumed that any interfering compounds will emit photons when reaching the reaction chamber 252 and be measured as a background signal (Drummond et al., 1985; Galbally, 2019). If the zero 253 volume is too small or the O₃ mixing ratio is too low, some untitrated NO may lead to NO₂ 254 chemiluminescence within the reaction chamber and the background will be overestimated. On 255 the other hand, if the zero volume is too large, some of the interfering compounds may have 256 relaxed to their ground state before the reaction chamber and the background signal will be 257 underestimated. The residence time of zero volume is 10.8s compared to 14.5s for the reaction 258 volume. The efficiency of the zero volume can be calculated from the calibration cycle. The 259 difference in background counts from before a calibration cycle to during the calibration cycle 260 shows how much of the added NO from the calibration cylinder does not react with O₃ in the 261 zero volume. By dividing this difference by the signal due to NO during the NO measurement 262 of the calibration cycle, which is obtained by subtracting the NO measurement from the 263 previous measurement cycle, the inefficiency of the zero volume is obtained. The efficiency is 264 determined for each calibration cycle (eq. VI) and plotted in Figure S6. It is consistently above 265 98%.

266 Efficiency_{ZV} =
$$1 - \frac{\text{cal zero-measurement zero}}{\text{NO cal-previous NO cycle}}$$
 (VI)

267

268 2.4.4 Artefact Measurements

269 As described in section 2.3, NO_x measurements may have artefacts from 270 chemiluminescence caused by interfering gas-phase reactions and/or from compounds 271 produced by illumination of the reaction chamber walls as well as from pressure differences in 272 the instrument (Drummond et al., 1985; Reed et al., 2016). To estimate artefacts, it is necessary 273 to measure the signal from NO_x -free air. The calibration sequence is followed by sampling 274 NO_x-free air generated from a pure air generator (PAG 003, Eco Physics AG) for 30 minutes. 275 According to the manufacturer, the PAG not only scrubs NO, NO₂ and NO_y from the ambient 276 air but also SO₂, VOCs, H₂O and O₃. An overflow of PAG air is introduced between the aerosol





- 277 filter and the NO₂ converters as shown in Figure 2 and the cycle of background, NO, NO_x BLC,
- and NO_x PLC is used to estimate artefact NO and NO₂ measured by the instrument.

279

280 2.4.4.1 NO Artefact

281 The NO artefact can be caused by two things; alkenes reacting with O₃ and giving 282 chemiluminescence above 600 nm at approximately the same rate as NO₂ or a difference in 283 pressure between the zero volume and the reaction volume. An artefact caused by alkenes will 284 be positive and overestimate the NO mixing ratio, where an artefact due to a pressure difference can be either negative or positive. It can be estimated as the offset from 0 pptV when the mixing 285 286 ratio sampled is 0 pptV. The NO mixing ratio is expected to be 0 pptV when sampling NO_{x} free air or between 22.00 and 04.00 UTC at night. NO generated during the day is rapidly 287 288 oxidized into NO₂ through reactions with O₃ and RO₂ after sunset. During the night, NO is not generated from photolysis of NO₂, and there are no significant local sources of NO at Cabo 289 290 Verde when the air masses come from over the ocean (which is >95% of the time). The average NO mixing ratio between 22.00 and 04.00 UTC and the average NO mixing ratio from the 291 292 PAG zero air tend to be very similar, with the PAG artefact (-3.68 \pm 22.91 pptV (2 σ), January 293 2014 – August 2019) ordinarily lower than the night time artefact $(0.39 \pm 11.92 \text{ pptV} (2\sigma))$, 294 January 2014 - August 2019). Time series of both NO artefact measurements can be found in 295 Figure S7 in the supplementary information. The night time NO artefact is used as it is 296 measured more frequently, it contains the same ambient matrix with nothing scrubbed and to 297 eliminate the possibility of residual NO influencing background measurements determined 298 from the PAG. Since the PAG scrubs VOCs it will also not give an estimate of the artefacts 299 caused by fast reacting alkenes.

300

301 2.4.4.2 NO₂ Artefact

302 NO₂ converters have previously been shown to have artefacts caused by thermal or 303 photolytic conversion of reactive nitrogen compounds (NO_z) other than NO₂ as well as 304 illumination of the chamber walls (Drummond et al., 1985; Reed et al., 2016; Ryerson et al., 305 2000). Fast reacting alkenes, which can cause overestimations of the NO mixing ratios, will 306 not cause the NO₂ mixing ratio to be overestimated, since the NO signal is subtracted from the 307 NO₂ signal.





308 The spectral output of an NO₂ converter with a wavelength of 385 nm was compared to 309 absorption cross sections of NO₂ and potential interfering species such as BrONO₂, HONO and 310 NO_3 (Reed et al., 2016). The photolytic convertor was shown to have good spectral overlap 311 with the NO₂ cross section with minimal spectral overlap with other NO_z species, except for a 312 small overlap with the absorption cross section of HONO. The interference from BrONO₂, 313 HONO and NO₃ have additionally been evaluated previously for a similar set-up using a Hg 314 lamp (Ryerson et al., 2000). At equal concentrations of NO_2 and NO_2 species, BrONO₂ and NO₃ were estimated to maximum have an interference of 5% and 10%, respectively, using a 315 316 lamp with a wider spectral overlap with the interfering species than what is observed for the LEDs used at the CVAO (Ryerson et al., 2000). At the CVAO, HONO levels have previously 317 been measured to peak at ~3.5 pptV (at noon; (Reed et al., 2017)). For the typical Gaussian 318 319 output of a UV-LED this interference is calculated to be 2.0, 12.6, and 25.7% for UV-LEDs 320 with principle outputs of 395, 385, and 365 nm respectively, resulting in a maximum 321 interference of <0.5 pptV during peak daylight hours. Photolytic conversion of NO_z species is 322 therefore not expected to be an important contributor to the NO₂ artefact at the CVAO due to 323 the narrow spectral output of the LEDs.

Each converter is only on for 1 minute in a 5-minute cycle. For thermal conversion to be a major contributor to the artefact, the converter would have to increase in temperature during that one minute and not the rest of the cycle otherwise an increase in signal should be constant since the air continues to flow through the converters when they are turned off. Thermal decomposition of NO_z species is therefore not expected to have an effect in a climate like the one in Cabo Verde, where the sample temperatures are similar to the ambient temperatures.

330 It has been shown that the walls of a BLC made out of a porous Teflon-like doped block 331 becomes contaminated from the ambient air over time and when the walls are illuminated 332 reactions take place on the surface causing an artefact (Reed et al., 2016; Ryerson et al., 2000). 333 The BLC is similar to the one used by Reed et al. (2016) and it is therefore expected to have 334 an artefact due to reactions taking place on the surface. The PLC is not expected to be 335 contaminated in the same way as it does not have porous chamber walls. Ryerson et al. (2000) observed an increase in artefact over time when sampling ambient air for a similar PLC, 336 337 however, this is not observed for the PLC in the very clean environment at the CVAO (0-10 338 pptV between August 2017 and August 2019, see below) and surface reactions are therefore expected to give a negligible artefact for the PLC. 339





340 The total artefact can be determined by measuring the NO₂ signal when the NO₂ mixing ratio is 0 pptV, however, it is virtually impossible to scrub all NO_x from the ambient air and 341 342 nothing else. To estimate the NO₂ artefact, PAG zero air is measured using both converters. 343 The PLC measures between 0-10 pptV compared to 10-60 pptV using the BLC. Since, as 344 discussed above, the NO₂ artefact of the PLC is believed to be negligible, the signal is believed 345 to represent the remaining NO_2 in the zero air after scrubbing. The signal from the BLC when 346 measuring PAG zero air is expected to be due to the illumination of the chamber walls in addition to the traces of NO₂ left in the zero air. The artefact due to wall reactions in the BLC 347 348 can therefore be estimated by subtracting the signal measured by the PLC.

349

350 3 Data Analysis

351 Time periods with known problems such as maintenance on the manifold, ozone leaks, 352 and periods when the PMT has not reached <-28°C are not included in the dataset. The mean 353 and standard deviation of the zero (background), NO, NO2 BLC and PLC are determined for 354 each 5-minute measurement cycle. To avoid averaging over the time it takes the detector to 355 change and stabilize between the different types of measurements, the last 50 seconds of the 356 measurement cycle are used for the background and the NO counts, and the last 30 seconds for 357 the BLC NO_x and the PLC NO_x counts. Each cycle is filtered based on the percentage standard 358 deviations and differences in counts between subsequent cycles. If the standard deviation or 359 the difference in counts are outside the mean $\pm 2\sigma$ (see Table 1) calculated from a 5-year period 360 between 2014 and 2018, the cycle is not used for further analysis. This removes noisy data as 361 well as sharp spikes but keeps data with sustained increases lasting more than 5 minutes.

To obtain the signals due to NO and NO₂, the interpolated zero and NO measurements are subtracted from the NO and NO_x measurements, respectively. They are converted to a concentration by using the interpolated sensitivity and conversion efficiency as shown in equation VII and VIII:

366 NO mixing ratio =
$$\frac{\text{NO measurement} - \text{background measurement}}{\text{Sensitivity}}$$
 (VII)

367
$$NO_2$$
 Mixing ratio = $\frac{NO_x \text{ measurement} - NO \text{ measurement}}{\text{Sensitivity} \times CE}$ (VIII)

The NO and NO₂ BLC concentrations are corrected by subtracting the interpolated artefacts described in sections 2.4.4.1 and 2.4.4.2. If the difference between two subsequent





NO artefact measurements vary by more than the mean $\pm 2\sigma$ of the differences in NO artefacts determined from January 2014 – August 2019 (0.00 \pm 6.18 pptV), the measurements made between will not be used for further analysis due to a potential step change between the determinations.

374 Hourly averages of all the measurements are determined. If data coverage during the hour 375 is less than 50%, the hour is flagged and discarded from the data analysis. The hourly NO_x (NO 376 + NO₂ PLC) concentrations between June 2017 and August 2019 are plotted as a function of wind speed and direction in figure 4. It can be observed that the concentrations are enhanced 377 378 at low wind speed and when the air crosses the island (southwest). Measurements made at a 379 wind speed <2 m/s or from a wind direction $>100^{\circ}$ are, therefore, flagged as suspected of local 380 contamination and are not used in the analysis. Extreme mixing ratios outside the mean $\pm 4\sigma$ of the 5-year for NO and 2-year period for NO_2 are flagged as suspicious (see Table 1 for 381 boundaries). Lastly, inconsistence in the measurements such as differences outside the mean \pm 382 383 4σ between the mean and median of a measurement (see Table 1 for boundaries) and 384 differences between the two NO₂ measurements are flagged as suspicious (0.4 ± 32.2 pptV). 385 The data remaining after these removals are 88% of the original NO and NO₂ BLC dataset and 386 83% of the NO₂ PLC dataset remain to analyse.

387

388 3.1 Corrections

As described above, excited NO₂ can be quenched by other sampled molecules, giving a lower observed mixing ratio than the real value. Water molecules are effective quenchers and therefore a correction is usually applied depending on the humidity (Matthews et al., 1977; Ridley et al., 1992). However, since the calibrations at the CVAO are performed by standard addition, and a Nafion dryer is placed in front of the instrument, this is not necessary.

Additionally, NO can react with O_3 in the ambient air in the inlet and manifold giving an overestimation of NO₂ and an underestimation of NO. To correct for this the following equations are used (Gilge et al., 2014):

397
$$[NO]_0 = [NO]_{E1} \times e^{k_{O_3} \times t_{E1}}$$
 (IX)

398
$$[NO_2]_0 = \left(\frac{J_C + k_{O_3}}{J_C}\right) \times \left(\frac{[NO]_{E_2} - [NO]_{E_1} \times e^{-(k_{O_3} \times (t_{C_2} - t_{C_1}) + J_C \times t_{C_2})}}{1 - e^{\left(-(k_{O_3} + J_C) \times t_{C_2}\right)}}\right) - [NO]_0 \qquad (X)$$





399 where [NO]₀ is the corrected NO mixing ratio, [NO]_{E1} is the uncorrected NO mixing ratio, [NO₂]₀ is the corrected NO₂ mixing ratio, [NO]_{E2} is the uncorrected NO mixing ratio when the 400 401 converter is on, k_{03} is the rate of the reaction between NO and O_3 ($k(O_3+NO) \times [O_3] \times 10^{-9} \times$ 402 M), t_{E1} is the sum of the residence time from the inlet to entry of the converter and the time the 403 air is in the converter, t_{C1} and t_{C2} are the time the air is in the converter when the converter is 404 on and off, respectively, and $J_{\rm C}$ is the photolysis rate inside the converter. The residence time 405 from the inlet to the entry of the converter has been 2.3 s since 2015 and the time the air is in each of the converters is 1.0 s (with and without the converter on). The O₃ mixing ratio 406 407 measured at the CVAO has varied between 5 and 60 ppbV (with an uncertainty of 0.07 ppbV) 408 between 2014 and 2019. The ozone correction is calculated for each hour using a rate coefficient of 1.8×10^{-14} cm³ molecule⁻¹ s⁻¹ at 298K (Atkinson et al., 2004). This gives an 409 average O₃ correction $\pm 2\sigma$ of 6.8 $\pm 3.0\%$, 1.7 $\pm 11.0\%$, and 1.3 $\pm 7.1\%$ for NO, NO₂ BLC, and 410 411 NO₂ PLC, respectively, when the mixing ratio measured is above 0.1 pptV (See supplementary 412 information for an example of the calculation). Thus, at the low mixing ratios of O_3 present at 413 Cabo Verde and the short residence time for sampling, the corrections for O_3 are well within 414 the noise of the measurements (see below), but are still included in the final calculated mixing 415 ratios.

416

417 4 Uncertainty Analysis

418 To be able to evaluate the NO_x measurements made at the CVAO an extensive uncertainty 419 analysis is performed. A summary of the analysis can be found in Table 2 and a detailed 420 description in the supplementary information. The hourly precision and uncertainty of the 421 instrument are estimated to characterize the uncertainties at the 95 percent confidence interval 422 (Bell, 1999). The hourly precision is estimated from the zero count variability, which is directly 423 related to the photon-counting precision of the PMT. The uncertainty of the hourly 424 measurements is estimated by combining all the uncertainties associated with the 425 measurements. This includes uncertainties in the calibrations, artefact determinations, and O₃ 426 corrections as well as the precision of the instrument. The precision of the NO and NO₂ 427 measurements are both included in the total uncertainty of the NO₂ measurements as the NO 428 measurements are subtracted from the NO₂ measurements. Each term is converted into pptV 429 to be able to combine them using error propagation.





430 The 25 precision for hourly averaged NO data between January 2014 and August 2019 is 0.96 ± 0.89 pptV. The hourly precisions reported here are in good agreement with our 431 432 previously reported 1 σ precision of the instrument of 0.30 pptV (Reed et al., 2017) and the 2σ 433 precision of 0.6-1.7 pptV (Lee et al., 2009). The NO₂ precisions are determined by taking the 434 conversion efficiency of the respective converters into account. The hourly 2σ precision for 435 hourly averaged NO₂ data between March 2017 and August 2019 becomes 1.45 ± 0.82 pptV 436 and 2.74 ± 2.18 pptV for the BLC and PLC, respectively. The determined NO₂ precisions are 437 within the interval of previously reported precisions for the same instrument (Lee et al., 2009; Reed et al., 2017). 438

439 The total hourly uncertainty for each of the three measurements are determined by 440 combining all the uncertainties described using propagation of uncertainties. The precisions are already calculated as hourly precisions in pptV. The calibration uncertainties are 441 442 interpolated between each calibration and multiplied by the hourly concentrations of NO and 443 NO₂ to calculate hourly uncertainties in pptV. The artefact uncertainties are interpolated 444 between each artefact determination, and the uncertainty due to ozone corrections is determined 445 by multiplying the % uncertainties by the hourly concentrations of NO and NO₂. The hourly uncertainties are determined to be 1.42 ± 1.47 pptV, 8.38 ± 7.46 pptV, and 4.44 ± 5.79 pptV 446 447 for NO, NO₂ BLC, and NO₂ PLC, respectively.

448

449 5 Results: Examples of Data

The first year of data (August 1st 2017 to July 31st 2018) is chosen as an example of the resulting NO and NO₂ datasets. October 2017, December 2017, and April 2018 are used to highlight the seasonality in the mixing ratios observed during a year of measurements. Panel A in figure 5 and 6 show the full O₃ corrected time series for NO and NO₂, respectively. Panel B, C, and D in the two figures show the time series for the three chosen months and panel E, F, and G show the 3-hour rolling average diurnals for the same months. Monthly diurnals for NO and NO₂ for the entire year can be found in figure S8 and S9, respectively.

457 Clear seasonality can be observed in the diurnal cycles of NO measurements with a 458 maximum of ~10 pptV in Winter and a minimum of ~2 pptV in the spring and summer. This 459 is in good agreement with that reported for previous years (Lee et al., 2009; Reed et al., 2017). 460 The two NO₂ measurements are in general in good agreement when looking at the time series





461 in figure 6. Offsets of up to 10 pptV between the two measurements can be seen over some 462 time periods (E.g. April, Panel D), which are most likely caused by the calculated BLC artefact 463 for those periods either being too high or too low. This is supported by the diurnals having the 464 same shape, but with an offset. Monthly diurnals of the two NO₂ measurements agree within 2 465 standard errors except in August 2017, where the offset between the two measurements is larger than for the remaining months. NO₂ shows a fairly flat diurnal signal, although a small increase 466 467 in daytime NO_2 is evident in some months, which is in agreement with that reported for previous years (Lee et al., 2009; Reed et al., 2017). Spikes in the early morning are noticeable 468 469 in the NO₂ diurnals for July-November, which correspond to the months with an average lower 470 wind speed than the rest of the year (the diurnal for April also shows a spike, however, it is caused by only one morning). These spikes could be caused by local fishing boats passing 471 472 upwind of the observatory in the morning hours, which will give a more prominent spike a low wind speed. Monthly wind speed diurnals can be found in figure S10. The good agreement 473 474 between the two NO₂ measurements observed in figure 6 can also be observed in figure 7, 475 where the two are plotted against each other. The data points are scattered around the 1:1 line 476 shown in black with an overestimation by the BLC. An orthogonal distance regression (ODR) 477 is performed to evaluate the scatter of the data points with uncertainty in both measurements 478 between August 2017 and 2019. The resulting regression line is displayed in red (PLC $NO_2 =$ $0.92 \times BLC \text{ NO}_2, \mathbb{R}^2 = 0.92$). 479

480 The seasonality of the NO measurements can be explained by a combination of the 481 variation of the origin of the air masses arriving at the CVAO, meteorology, photolysis rates, 482 and seasonality of emissions. Back trajectories of the three months used as examples are shown 483 in figure 8. FLEXPART version 10.4 is used in backwards mode, driven by pressure level data 484 from Global Forecast System (GFS) reanalyses at 0.5°×0.5° resolution (Pisso et al., 2019; Stohl 485 et al., 1998). 10-day back-trajectory simulations are initialised every 6 hours, releasing 1000 486 particles from the CVAO site. Further information on FLEXPART can be found in the supplementary information. During the winter maximum (December) the back-trajectories 487 488 indicate that the air reaching CVAO is largely dominated by African air, compared to during 489 the spring minimum (April), which is dominated by Atlantic marine air. Large west African 490 cities such as Dakar and Nouakchott, and/or the shipping lanes to the east/northeast of Cabo 491 Verde, are potential candidates for the source of elevated NO_x. The NO mixing ratios measured 492 in October are higher than those in April and lower than in December. This may be due in part 493 to the influence of polluted African air arriving at Cabo Verde, which is more prominent in





494 October than in April, but less so than in December. The NO_2 and the total NO_x (NO + PLC 495 NO₂, figure 9) similarly show higher levels in December than April, but the mixing ratios 496 observed in October are similar to those in April. It should be noted that some of the days with 497 high percentages of African air have missing data or wind directions from other places than the 498 north east.

From table 3 it can be observed that the NO, NO_2 , and NO_x measurements at the CVAO compare well to the few other measurements in the remote marine boundary layer as well as background sites in Alert, Canada and measurements in the free troposphere. A wintertime seasonal increase in NO, NO_2 , and NO_x can be observed during December-February, which corresponds to the months when surface air masses arrive at Cabo Verde from western Africa (Carpenter et al., 2010; Lee et al., 2009).

505

506 6 Conclusion

507 A photolytic NO₂ converter with external diodes and a quartz photolysis cell (PLC) has 508 been installed at the Cape Verde Atmospheric Observatory and the NO₂ measurements have 509 been compared to those of the historical BLC used at the site, which has internal diodes and a 510 reaction chamber made of Teflon-like barium doped material. The two measurements show good agreement (PLC NO₂ = $0.92 \times BLC$ NO₂, $R^2 = 0.92$) with small differences due to 511 512 uncertainties in the estimations of the BLC NO₂ artefact. Even though the PLC has a lower conversion efficiency (CE= $52 \pm 4\%$) than the BLC (CE= $85 \pm 4\%$), it is preferred due to its 513 514 assumed negligible artefact as a consequence of having non-porous/non-reactive walls. The 515 assumption of a zero artefact causes the hourly uncertainty of the NO₂ measurements to be 516 roughly halved. With 2σ hourly precisions of 0.96 ± 0.89 pptV, 1.45 ± 0.82 pptV, and 2.74 ± 517 2.18 pptV and 2σ hourly uncertainties of 1.42 ± 1.47 pptV, 8.38 ± 7.46 pptV, and 4.44 ± 5.79 pptV for NO, NO₂ BLC, and NO₂ PLC, respectively, the instrument has a high repeatability 518 519 and low uncertainties for all the measurements. The mixing ratios observed at the CVAO (NO: 520 2-10 pptV, NO₂: 5-50 pptV, and NO_x: 7-60 pptV at midday) are in agreement with previous 521 measurements at the CVAO as well as other previous remote measurements around the world.

522





523 7 Data availability

- 524 The processed data is available through Ebas
- 525 (http://ebas.nilu.no/Pages/DataSetList.aspx?key=45DB99FE2B7F4F97864ECF800E71E5D5
- 526) and through CEDA (Center for Environmental Data Analysis,
- 527 <u>https://catalogue.ceda.ac.uk/uuid/d5422d54d519ed056cc17e97037732b8</u>).

528

529 8 Author contribution

530 LN runs the instrument on a day-to-day basis. MW and STA wrote the script processing the

531 data. MJR ran the back trajectory analysis. BSN developed the photolytic converter setup. All

authors were involved in the analysis, data evaluation and discussion of the results. STA wrote

the paper with contributions from all coauthors. All coauthors proofread and commented on

the paper.

535

536 9 Competing interests

537 The authors declare that they have no conflict of interest.

538

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544





11 Tables

Table 1: The mean $\pm 2\sigma$ of the standard deviation and difference in counts/s between two subsequent measurement cycles.

Measurement	Standard deviation (%) ^a	Difference in counts/s ^b	Extreme values (pptV) ^c	Extreme difference between mean and median (pptV) ^d
Zero	2.4 ± 1.7	-	-	-
NO	2.5 ± 10.6	0 ± 515	1.7 ± 47.9	0.2 ± 4.1
NO ₂ BLC	2.5 ± 7.5	0 ± 1432	16.8 ± 175.2	1.5 ± 33.0
NO ₂ PLC	2.1 ± 2.5	0 ± 738	17.3 ± 176.8	1.7 ± 33.0

^aDetermined as the standard deviation of a cycle divided by the mean. ^bThe difference in counts/s between each cycle. ^cCalculated as the hourly mean ± 4 standard deviations of the hourly mixing ratio. ^dCalculated as the hourly mean ± 4 standard deviations of the differences between the mean and median.

> Table 2: Calculated uncertainties associated with the NO_x measurements. The values in bold are the combined uncertainties for each type of measurement. Each uncertainty is given as the mean uncertainty ± 2 standard deviation of the data between January 2014 and August 2019 for



Source of uncertainty	Probability distribution	Uncertainty (%)	Uncertainty (pptV)
Hourly precision/repeatability NO	Normal		0.96 ± 0.89
Hourly precision/repeatability NO2 BLC	Normal		1.45 ± 0.82
Hourly precision/repeatability NO2 PLC	Normal		2.74 ± 2.18
Total Calibration uncertainty NO ^a		2.78 ± 8.05	0.03 ± 0.27
Total Calibration uncertainty NO ₂ BLC ^a		3.44 ± 9.32	0.33 ± 1.27
Total Calibration uncertainty NO ₂ PLC ^a		3.52 ± 8.67	0.37 ± 1.27
Total NO artefact uncertainty ^b			1.05 ± 3.44
Total NO ₂ artefact uncertainty ^b			7.19 ± 7.24
Hourly O ₃ correction uncertainty NO	Normal	20.00 ± 0.001	0.27 ± 1.14
Hourly O ₃ correction uncertainty NO ₂ BLC	Normal	20.00 ± 0.001	2.47 ± 6.75
Hourly O ₃ correction uncertainty NO ₂ PLC	Normal	20.00 ± 0.001	2.60 ± 6.37
Total hourly uncertainty NO			1.42 ± 1.47
Total hourly uncertainty NO ₂ BLC			8.38 ± 7.46
Total hourly uncertainty NO ₂ PLC			4.44 ± 5.79

determination can be found in table S2. ^a









nade at the Ny-Ålesund Zeppelin mountain station on Svalbard during a spring campaign in 1994. ^kDaytime values.





12 Figures



Figure 1: The frequency of hourly averaged wind speed and direction from January 2014 to August 2019. Each square symbolise 10 degrees of wind direction and 1 m/s wind speed. Each dashed circle show an increase in wind speed of 5 m/s.







Figure 2: Flow diagram of the NOx instrument at the CVAO.







Figure 3: A theoretical calibration cycle. "NO" is the measurement of only NO i.e. when the converters are off, NO.c is when one of the converters are on therefore the measurement is NO + NO₂ and (1) and (2) represent untitrated and titrated NO, respectively.



Figure 4: Total NO_x from June 2017 to August 2019 plotted as a function of wind speed and direction.







Figure 5: Panel A show the time series for filtered O₃ corrected NO from August 1st 2017 to July 31st 2018. Panel B, C and D zoom in on October 2017, December 2017 and April 2018, respectively. Panel E, F and G show the average diurnal of NO for October 2017, December 2017 and April 2018, respectively, with the coloured areas being ± 2 standard errors. If there are less than 15 measurements available for the hour, it is not included in the diurnal.







Figure 6: Panel A show the time series of filtered O_3 corrected NO_2 from August 1st 2017 to July 31st 2018 for the BLC (black) and PLC (red). Panel B, C and D zoom in on October 2017, December 2017 and April 2018, respectively, with the red line being the PLC and the black being the BLC. Panel E, F and G show the average diurnal of NO_2 for October 2017, December 2017 and April 2018, respectively, with the red line being the PLC and the black being the BLC and the coloured areas being ± 2 standard errors. If there are less than 15 measurements available for the hour, it is not included in the diurnal.







Figure 7: The PLC NO_2 concentration is plotted against the BLC NO_2 concentration. The black lines show the 1-to-1 relationship. The red line is the linear regression of the hourly data with uncertainties in both the x and y.







Figure 8: Back trajectories estimated for October 2017, December 2017, and April 2018. FLEXPART version 10.4 is used in backwards mode, driven by pressure level data from Global Forecast System (GFS) reanalyses at $0.5^{\circ} \times 0.5^{\circ}$ resolution (Pisso et al., 2019; Stohl et al., 1998). 10-day back-trajectory simulations are initialised every 6 hours, releasing 1000 particles from the CVAO site.







Figure 9: Panel A show the time series for total NO_x (NO + NO₂ PLC) from August 1st 2017 to July 31st 2018. Panel B, C and D zoom in on October 2017, December 2017 and April 2018, respectively. Panel E, F and G show the average diurnal of NO_x for October 2017, December 2017 and April 2018, respectively, with the coloured areas being ±2 standard errors. If there are less than 15 measurements available for the hour, it is not included in the diurnal.





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