1Assessing sub-grid variability within satellite pixels over urban regions using2airborne mapping spectrometer measurements

3

Wenfu Tang^{1,2}, David P. Edwards², Louisa K. Emmons², Helen M. Worden², Laura M.
Judd³, Lok N. Lamsal^{4,5}, Jassim A. Al-Saadi³, Scott J. Janz⁴, James H. Crawford³, Merritt
N. Deeter², Gabriele Pfister², Rebecca R. Buchholz², Benjamin Gaubert², Caroline R.
Nowlan⁶

¹Advanced Study Program, National Center for Atmospheric Research, Boulder, CO, USA

9 ²Atmospheric Chemistry Observations and Modeling, National Center for Atmospheric Research,

- ³NASA Langley Research Center, Hampton, VA 23681, USA
- 12 ⁴NASA Goddard Space Flight Center, Greenbelt, MD 20771, USA
- 13 ⁵Universities Space Research Association, Columbia, MD 21046, USA
- ⁶Harvard-Smithsonian Center for Astrophysics, Cambridge, MA 02138, USA
- 15

16 Abstract

17 Sub-grid variability (SGV) of atmospheric trace gases within satellite pixels is a key issue 18 in satellite design and interpretation and validation of retrieval products. However, characterizing this variability is challenging due to the lack of independent high-resolution measurements. Here 19 20 we use tropospheric NO₂ vertical column (VC) measurements from the Geostationary Trace gas 21 and Aerosol Sensor Optimization (GeoTASO) airborne instrument with a spatial resolution of 22 about 250 m \times 250 m to quantify the normalized SGV (i.e., the standard deviation of the sub-grid 23 GeoTASO values within the sampled satellite pixel divided by the mean of the sub-grid GeoTASO 24 values within the same satellite pixel) for different hypothetical satellite pixel sizes over urban regions. We use the GeoTASO measurements over the Seoul Metropolitan Area (SMA) and Busan 25 region of South Korea during the 2016 KORUS-AO field campaign, and over the Los Angeles 26 Basin, USA, during the 2017 SARP field campaign. We find that the normalized SGV of NO₂ VC 27 28 increases with increasing satellite pixel sizes (from $\sim 10\%$ for 0.5 km \times 0.5 km pixel size to $\sim 35\%$ 29 for 25 km \times 25 km pixel size), and this relationship holds for the three study regions, which are also within the domains of upcoming geostationary satellite air quality missions. We also quantify 30 the temporal variability of the retrieved NO₂ VC within the same hypothetical satellite pixels 31 (represented by the difference of retrieved values at two or more different times in a day). For a 32 given satellite pixel size, the temporal variability within the same satellite pixels increases with 33 the sampling time difference over SMA. For a given small (e.g., <=4 hours) sampling time 34 35 difference within the same satellite pixels, the temporal variability of the retrieved NO₂ VC 36 increases with the increasing spatial resolution over the SMA, Busan region, and the Los Angeles basin. 37

¹⁰ Boulder, CO, USA

The results of this study have implications for future satellite design and retrieval 38 39 interpretation, and validation when comparing pixel data with local observations. In addition, the 40 analyses presented in this study are equally applicable in model evaluation when comparing model 41 grid values to local observations. Results from the Weather Research and Forecasting model 42 coupled with Chemistry (WRF-Chem) model indicate that the normalized satellite SGV of tropospheric NO₂ VC calculated in this study could serve as an upper bound to the satellite SGV 43 of other species (e.g., CO and SO₂) that share common source(s) with NO₂ but have relatively 44 45 longer lifetime.

46

47 **1. Introduction**

48 Characterizing sub-grid variability (SGV) of atmospheric chemical constituent fields is 49 important in both satellite retrievals and atmospheric chemical-transport modeling. This is 50 especially the case over urban regions where strong variability and heterogeneity exist. The 51 inability to resolve sub-grid details is one of the fundamental limitations of grid-based models 52 (Qian et al., 2010) and has been studied extensively (e.g., Boersma et al., 2016; Ching et al., 2006; Denby et al., 2011; Pillai et al., 2010; Qian et al., 2010). Pillai et al. (2010) found that the SGV of 53 54 column-averaged carbon dioxide (CO₂) can reach up to 1.2 ppm in global models that have a 55 horizontal resolution of 100 km. This is an order of magnitude larger than sampling errors that 56 include both limitations in instrument precision and uncertainty of unresolved atmospheric CO₂ 57 variability within the mixed layer (Gerbig et al., 2003). Denby et al. (2011) suggested that the 58 average European urban background exposure for nitrogen dioxide (NO₂) using a model of 50-km 59 resolution is underestimated by ~44% due to SGV.

60 In contrast, much less attention has been paid to the sub-grid variability within satellite pixels (e.g., Broccardo et al., 2018; Judd et al., 2019; Tack et al., 2020). Indeed, some previous 61 62 studies (e.g., Kim et al., 2016; Song et al., 2018; Zhang et al., 2019; Choi et al., 2020) used satellite 63 retrievals to study SGV in models, and calculated representativeness errors of model results with respect to the satellite measurements (e.g., Pillai et al., 2010). Even though satellite retrievals of 64 65 atmospheric composition often have smaller uncertainties than model results, it has not been until recently that the typical spatial resolution of atmospheric composition satellite products has 66 67 reached scales comparable to regional atmospheric chemistry models ($\leq \sim 10$ km).

68 Quantification of satellite SGV has historically been limited by insufficient spatial coverage of in situ measurements, and is a key issue in designing, understanding, validating and 69 correctly interpreting satellite observations. This is especially important in the satellite instrument 70 71 development process during which the required measurement precision and retrieval resolution 72 need to be defined in order to meet the mission science goals. In addition, when validating and 73 evaluating relatively coarse-scale satellite retrievals by comparing with surface in situ observations, 74 SGV introduces large uncertainties on top of the existing uncertainty introduced by imperfect 75 knowledge of the trace gas vertical profiles. Accurate quantification of satellite SGV can therefore facilitate the estimate of sampling uncertainty for satellite product validation/evaluation. Temporal 76 77 variability within sampled satellite pixels is also an important issue in satellite design, validation, 78 and application. For polar-orbiting satellites, knowledge of temporal variability is necessary to 79 analyze the representativeness of satellite retrievals at specific overpass times. For geostationary Earth orbit (GEO) satellites, developing a measure of the temporal variability of fine-scale spatial 80

81 structure will be important for assessing coincidence during validation of the new hourly 82 observations. This work is partly motivated by validation requirements and considerations for the 83 upcoming GEO satellite constellation for atmospheric composition that includes the Tropospheric 84 Emissions: Monitoring Pollution (TEMPO) mission over North America (Chance et al., 2013; 85 Zoogman et al., 2017), the Geostationary Environment Monitoring Spectrometer (GEMS) over 86 Asia (Kim et al., 2020), and the Sentinel-4 mission over Europe (Courrèges-Lacoste et al., 2017).

87 Airborne mapping spectrometer measurements provide dense observations within the 88 several-kilometer footprint of a typical satellite pixel. This feature of airborne mapping 89 spectrometer measurements provides a unique opportunity to estimate satellite SGV in addition to 90 their role in satellite validation. For example, Broccardo et al. (2018) used aircraft measurements 91 of NO₂ from an imaging differential optical absorption spectrometer (iDOAS) instrument to study intra-pixel variability in satellite tropospheric NO₂ column over South Africa, whilst Judd et al. 92 93 (2019) evaluated the impact of spatial resolution on tropospheric NO₂ column comparisons with 94 in situ observations using the NO₂ measurements of the Geostationary Trace gas and Aerosol Sensor Optimization (GeoTASO). GeoTASO is an airborne remote sensing instrument capable of 95 high spatial resolution retrieval of UV-VIS absorbing species such as NO₂ and formaldehyde 96 (HCHO; Nowlan et al., 2018) and sulfur dioxide (SO₂; Chong et al., 2020), and has measurement 97 characteristics similar to the GEMS and TEMPO GEO satellite instruments. The GeoTASO data 98 used here were taken in gapless, grid-like patterns - or "rasters" - over the regions of interest, 99 100 providing essentially continuous spatial coverage that was repeated during multiple flights up to 101 four times a day in some cases. As such, the GeoTASO data (with a spatial resolution of ~250 m 102 \times 250 m) provide a preview of the type of sampling that is expected from the GEO satellite sensors, 103 making the data particularly suitable for our study. We focus on the GeoTASO measurements 104 made during the Korea United States Air Quality (KORUS-AQ) field experiment in 2016 105 (Crawford et al., 2021). The measurements from KORUS-AQ have been widely used by 106 researchers for various air quality topics, including quantification of emissions and model and satellite evaluation (e.g., Deeter et al., 2019; Huang et al., 2018; Kim et al., 2018; Miyazaki et al., 107 108 2019; Spinei et al., 2018; Tang et al., 2018, 2019; Souri et al., 2020, Gaubert et al., 2020). We 109 further compare our findings from KORUS-AQ with flights conducted during the NASA Student 110 Airborne Research Program (SARP) in 2017 over the Los Angeles (LA) Basin to test the general 111 applicability of our findings over a different urban region. The KORUS-AQ mission took place 112 within the GEMS domain, while the SARP in 2017 is within the domain of TEMPO. Given the 113 similarity between the TEMPO and GEMS instruments in terms of spectral ranges, spectral and 114 spatial resolution, and retrieval algorithms (Al-Saadi et al., 2014), such comparison is reasonable 115 and useful in facilitating the generalization of the results from the study.

116 We use the tropospheric NO_2 vertical column (VC) retrieved by GeoTASO as a tool to 117 assess satellite SGV and temporal variability for different hypothetical satellite pixel sizes over 118 urban regions. Because spatial SGV and temporal variability both vary with satellite pixel size, the 119 two need to be considered together to enhance the accuracy of satellite product analyses. NO₂ is 120 an important air pollutant that is primarily generated from anthropogenic sources such as emissions 121 from the energy, transportation, and industry sectors (Hoesly et al., 2018). It is a reactive gas with 122 a typical lifetime of a few hours in the planetary boundary layer (PBL), although it can also be 123 transported over long distance in the form of peroxyacetyl nitrate (PAN) and nitric acid. NO₂ is a 124 precursor of tropospheric ozone and secondary aerosols and has a negative impact on human health

- 125 and the environment (Finlayson-Pitts et al., 1997). The results from this paper's analysis of NO₂ 126 also have implications for other air pollutants that share common source(s) with NO₂, but that have
- somewhat longer lifetimes, for example, carbon monoxide (CO) and SO₂.

128 In this study, we apply a satellite pixel random sampling technique and the spatial structure 129 function analysis to GeoTASO data (described in Section 2) to quantify the SGV of satellite pixel 130 NO₂ VC over three urban regions at a variety of spatial resolutions. We analyze the relationship 131 between satellite pixel size and satellite SGV, and the relationship between satellite pixel size and 132 the temporal variability of NO₂ observations (Section 3). We then discuss the implications for 133 satellite design, satellite retrieval interpretation, satellite validation and evaluation, and satellite-134 in situ data comparisons (Section 4). Implications for general local observations and grid data 135 comparisons are also discussed. Section 5 presents our conclusions.

136 **2. Data and methods**

137 In this section, we describe the GeoTASO instrument, campaign flights and the different 138 analysis techniques used to characterize the satellite pixel SGV. We outline two approaches: 139 satellite pixel random sampling to investigate separately both spatial variability and temporal 140 variability, and the construction of spatial structure functions for an alternative measure of spatial 141 variability.

142 2.1 GeoTASO instrument

In this study, we focus on GeoTASO retrievals of tropospheric NO₂ VC. GeoTASO is a hyperspectral instrument (Leitch et al., 2014) that measures nadir backscattered light in the ultraviolet (UV; 290–400 nm) and visible (VIS; 415–695 nm). As one of NASA's airborne UV– VIS mapping instruments, it was designed to support the upcoming GEO satellite missions by acquiring high temporal and spatial resolution measurements with dense sampling for optimizing and experimenting with new retrieval algorithms (Leitch et al., 2014; Nowlan et al., 2016; Lamsal et al., 2017; Judd et al., 2019).

150 NO₂ is retrieved from GeoTASO spectra using the Differential Optical Absorption 151 Spectroscopy (DOAS) technique. The retrieval methods and Level 2 data processing are described 152 in Lamsal et al. (2017) and Souri et al. (2020) for KORUS-AQ and in Judd et al. (2019) for SARP. 153 Although beyond the scope of this work, it is important to recognize that assumptions made in the retrieval process (e.g., assumed vertical distribution of the NO₂ profile) could affect the final 154 variability of the retrieved NO2 fields. GeoTASO has a cross-track field of view of 45° (+/- 22.5° 155 156 from nadir), and the retrieval pixel size is approximately 250 m×250 m from typical flight altitudes 157 of 24,000–28,000 feet (7.3–8.5 km). The dense sampling of airborne remote sensing measurements 158 such as GeoTASO is a unique feature that provides the opportunity to study the expected spatial 159 and temporal variability within satellite retrieved NO₂ pixels at high resolution. We use cloud-free GeoTASO data in this study. GeoTASO NO2 VC retrievals have been validated with aircraft in 160 situ data and ground-based Pandora remote sensing measurements during KORUS-AQ. Validation 161 162 of GeoTASO NO₂ VC retrievals with aircraft in situ data suggested ~25% average difference, 163 while agreement with Pandora is better with a difference of $\sim 10\%$ on average. Mean difference between Pandora and aircraft in situ data is ~20%. These validation results of GeoTASO NO₂ VC 164

retrievals are better than that reported by Nowlan et al. (2016). GeoTASO NO₂ VC retrievals during 2017 SARP have also been validated with Pandora data (Judd et al., 2019).

167 2.2 The 2016 KORUS-AQ field campaign

168 The KORUS-AQ field measurement campaign (Crawford et al., 2021), took place in May-June 2016, to help understand the factors controlling air quality over South Korea. One of the goals 169 170 of KORUS-AO was the testing and improvement of remote sensing algorithms in advance of the 171 launches of the GEMS, TEMPO, and Sentinel-4 satellite missions. It is hoped that the high-quality 172 initial data products from the GEO missions will facilitate their rapid uptake in air quality 173 applications after launch (Al-Saadi et al., 2014; Kim et al., 2020). During KORUS-AQ, GeoTASO flew onboard the NASA LaRC B200 aircraft. We focus on the data taken over the Seoul 174 175 Metropolitan Area (SMA) that is highly urbanized and polluted, and the greater Busan region that is less urbanized and less polluted than SMA (Figure 1). Figure 2 shows the 12 GeoTASO data 176 rasters (i.e., gapless maps) acquired over SMA. It took ~4 hours to sample the large-area rasters 177 (i.e., 0511AM, 0517AM, 0517PM, 0528PM), and ~2 hours to sample small-area rasters (i.e., 178 179 0601PM, 0602AM, 0605AM, 0609AM, and 0609PM). Figure S1 shows the 2 GeoTASO rasters 180 acquired over the Busan region.

181 2.3 The 2017 SARP field campaign

During the NASA Student Airborne Research Program (SARP) flights in June 2017, (https://airbornescience.nasa.gov/content/Student_Airborne_Research_Program), GeoTASO was flown onboard the NASA LaRC UC-12B aircraft over the LA Basin (Figure S2, which also shows the landcover). A detailed description and analysis of these data can be found in Judd et al. (2018; 2019). In this study, we compare our analyses of the KORUS-AQ GeoTASO data with that from SARP over the LA Basin to test the general applicability of our findings.

188 **2.4 Satellite pixel random sampling for spatial variability**

189 The sampling strategy with GeoTASO provides a raster of continuous measurements in a 190 mapped gapless pattern at high spatial resolution (Figures 2, S1, and S2). This dataset allows us to 191 sample and study the SGV of coarser spatial resolution hypothetical satellite pixels sampling the 192 same domain. To mimic satellite observations and quantify the satellite SGV, we randomly sample 193 the GeoTASO data with hypothetical satellite pixels spanning 27 different pixel sizes (0.5 km×0.5 194 km, 0.75 km×0.75 km, 1 km×1 km, 2 km×2 km, up to 25 km×25 km). Because of the transition to 195 better spatial resolution for the future satellite missions, and the coverage limitation in the 196 maximum hypothetical satellite pixel size sampled using the random sampling method, the 197 analysis of SGV only goes up to 25 km × 25 km. This sampling process is conducted for each hour 198 of each selected flight over the regions of interest during the KORUS-AQ and SARP campaigns. 199 For every sampled satellite pixel, the mean (MEAN_{pixel}) and standard deviation (SD_{pixel}) of the 200 GeoTASO tropospheric NO₂ VC data within the pixel are calculated to represent the satellite SGV. Normalized satellite SGV is calculated as the standard deviation of the GeoTASO data within the 201 202 sampled satellite pixel divided by the mean of the GeoTASO data within the same sampled satellite 203 pixel (SD_{pixel}/MEAN_{pixel}).

We use a set of 10,000 hypothetical satellite pixels at each size to include all of the GeoTASO data in the analysis and to cover as many locations as possible. Because the data are located closely in space but may be sampled at slightly different times for the same flight, we separate GeoTASO data into hourly bins for each flight before pixel sampling in order to reduce the impact of temporal variability of the GeoTASO data within a single satellite pixel sample.

As an illustration, we describe the procedure below for the May 17th afternoon flight 209 (Figure 3) that was conducted from 13:00 to 17:00 local time: (1) the GeoTASO data during this 210 211 flight were divided into four hourly groups according to the measurement time, i.e., 13:00-14:00, 212 14:00-15:00, 15:00-16:00, and 16:00-17:00; (2) for each of the 27 hypothetical satellite pixel sizes, 213 we randomly generate 10,000 satellite pixel locations within each hourly group. Therefore, for 214 each hour, we sample 270,000 satellite pixels (27 different satellite pixel sizes and 10,000 samples 215 for each size), and for this example flight, we have a total of up to 1,080,000 possible satellite 216 pixels in each of 4 hourly groups. Note that only $\sim 10\%$ of these samples are used in the analysis because we discarded a sampled satellite pixel if less than 75% of its area is covered by GeoTASO 217 218 data. After applying this 75% area coverage filter, the actual sample size decreases when the pixel 219 size increases. The number of samples is sufficient as our sensitivity tests indicate that the results 220 do not change by halving the sample size. We also tested other choices of the coverage threshold 221 over SMA in addition to 75% (not shown here). The results are similar for small pixels ($< \sim 10$ km²) as they are more likely to be covered by GeoTASO data regardless of the threshold value. 222 223 For larger pixels ($> \sim 15 \text{ km}^2$), the satellite SGV is slightly lower when using 30% or 50% as the 224 area coverage threshold because larger pixels act like smaller pixels when only partially covered. 225 The threshold of 75% was chosen as a trade-off between sample size and representation.

226 **2.5 Satellite pixel random sampling for temporal variability**

227 We also quantify the temporal variability of the retrieved NO₂ VC within the same satellite 228 pixels for different satellite pixel sizes. To calculate temporal variability within a hypothetical 229 satellite pixel, we need GeoTASO data to cover the hypothetical satellite pixel at different times 230 during the day. During the KORUS-AQ and 2017 SARP campaigns, rasters were treated as single 231 units (Judd et al., 2019). Each raster produces a contiguous map of data that we consider as roughly 232 representative of the mid-time of the raster. Unlike the calculation of SGV, which is based on data 233 separated into hourly bins (section 2.4) to reduce the impact of temporal variability in the 234 calculated spatial variability, the satellite pixel random sampling to assess temporal variability is 235 based on rasters, and only conducted for days with multiple rasters. This is to ensure that the 236 sampled hypothetical satellite pixels have multiple values at different times of the day and hence 237 maximize the sample size.

To assess temporal variability within the hypothetical satellite pixels, we randomly select 50,000 pixel locations for each of the 27 hypothetical satellite pixel sizes, and use this same set of pixel locations to sample the GeoTASO data for each raster across all flights for a given day. This process is repeated for all days with multiple rasters, and the 75% of area coverage threshold is also applied. When there are two or more raster values of MEAN_{pixel} for a given pixel location separated by time Dt, the temporal mean difference (TeMD) within the satellite pixel is calculated as:

245
$$TeMD(Dt) = average(|MEAN_{pixel}(t) - MEAN_{pixel}(t + Dt)|)$$
(1)

246 This procedure is repeated for each satellite pixel size.

247 **2.6 Spatial structure function**

Structure functions have been applied to in situ measurements and model-generated 248 tropospheric trace gases to analyze their spatial and temporal variability in previous studies (Harris 249 250 et al., 2001). The Spatial Structure Function (SSF) (Fishman et al., 2011; Follette-Cook et al., 2015) 251 is an alternative measure to the satellite pixel random sampling described above for quantifying 252 spatial variability, and in this work, we apply the SSF to GeoTASO data to assist our analysis of 253 satellite SGV. The main difference between the two measures is that the SSF is based on individual 254 GeoTASO data points, while the results from satellite pixel random sampling are based on sampled 255 satellite pixels. The locations of the GeoTASO pixel centers are used to calculate the distances. 256 The SSF as defined here follows Follette-Cook et al. (2015):

257
$$f(NO_{2,VC}, D) = average(|NO_{2,VC}(x+D) - NO_{2,VC}(x)|)$$
(2)

where $NO_{2,VC}$ is tropospheric NO₂ VC. $f(NO_{2,VC}, D)$ calculates the average of the absolute value of $NO_{2,VC}$ differences across all data pairs (measured in the same hourly bin) that are separated by a distance *D*. To calculate SSF, the first step is the same as the first step of the satellite pixel random sampling: we group GeoTASO data hourly for each flight to reduce the impact of temporal variability of the GeoTASO data, and we only pair each GeoTASO data point with all the other GeoTASO data in the same hourly bin. More details on structure functions can be found in Follette-Cook et al. (2015).

265 2.7 WRF-Chem simulation

To briefly demonstrate the application of this technique on model evaluation and other species, we show results of a WRF-Chem simulation (Weather Research and Forecasting model coupled to Chemistry) with a resolution of 3 km × 3 km over SMA in Section 4. The simulation used NCEP GDAS/FNL 0.25 Degree Global Tropospheric Analyses and Forecast Grids as initial and boundary conditions, and the model meteorological fields above the PBL were nudged 6hourly. KORUS version 3 anthropogenic emissions and FINN version 1.5 fire emissions (Wiedinmyer et al., 2011) were used.

273

274 **3. Results**

In this section, we discuss the results for SGV over the different regions considered. Results are presented for the hypothetical satellite pixel random sampling for spatial variability and temporal variability, and for the spatial structure function analysis. We note that the three regions analyzed in this study are urban. Although we expect the results here to be generally applicable over urban regions, we have not tested the approach over cleaner background areas that are characterized by much less heterogeneity.

281 **3.1 Sub-grid variability (SGV) within satellite pixels**

282 SMA, the Busan region, and the LA Basin have different levels of pollution - the average 283 values of the GeoTASO NO₂ VC data over the SMA, the Busan region, and the LA Basin are 2.3×10¹⁶ molecules cm⁻², 1.1×10¹⁶ molecules cm⁻², and 1.3×10¹⁶ molecules cm⁻², respectively. 284 Over the three regions, the mean values (MEAN_{pixel}) and standard deviation (SD_{pixel}) of the 285 286 hypothetical satellite pixels sampled over GeoTASO NO₂ VC data are different (Figure S3). This is consistent with previous studies suggesting SGV can vary regionally (Judd et al., 2019; 287 Broccardo et al., 2018). However, we find that the normalized satellite SGV (calculated as the 288 289 ratio of SD_{pixel} to MEAN_{pixel} for a sampled pixel) is similar over each of the areas, regardless of 290 the absolute level of pollution as represented by MEAN_{pixel} (Figure 4). Over SMA (Figure 4a), the 291 mean normalized satellite SGV of tropospheric NO₂ VC increases smoothly from ~10% for the pixel size of 0.5 km \times 0.5 km, to \sim 35% for the pixel size of 25 km \times 25 km. The interquartile 292 293 variation of the satellite SGV also increases with satellite pixel sizes. The patterns of the sampled 294 satellite pixels over the Busan region (Figure 4b) and LA Basin (Figure 4c) are also found to be 295 similar to those over SMA. Furthermore, Figures S4 and S5 show that even the normalized SGV 296 of individual flights over the three domains generally follow the same pattern, except in the case 297 of the June 9 PM flight.

We also compare normalized satellite SGV for different levels of pollution, regardless of their regions (Figure S6). The normalized satellite SGV for the less polluted pixels (MEAN_{pixel} being lower than the average value of all pixels, i.e., 2×10^{16} molecules cm⁻²) also shows an overall similar pattern as for the more polluted pixels (MEAN_{pixel} being higher than the average value of all pixels). We notice that at small pixel sizes, less polluted pixels have higher normalized satellite SGV, possibly contributed by relatively higher retrieval noise at lower pollution levels.

304 We show the normalized SGV for individual rasters over SMA (Figure 5) to indicate the uncertainty range of the normalized SGV shown in Figure 4. The spread of SGV across different 305 306 individual rasters represents the uncertainties of using the averaged normalized SGV for a specific 307 case. Note that the variation of normalized SGV with pixel size for individual rasters generally 308 follows the same pattern (i.e., increases with satellite pixel size), especially when the pixel size is 309 small (≤ 10 km \times 10 km). The normalized SGV increases from ~10% to ~25%, with the uncertainty 310 range consistently being $\pm 5\%$ when the pixel size is smaller than 10 km \times 10 km. When the pixel size is larger than 10 km \times 10 km, the uncertainty range broadens with pixel sizes from ±5% (10 311 312 km \times 10 km) to \pm 15% (25 km \times 25 km). This means that when the satellite pixel size is large, 313 using the mean normalized SGV in Figure 4 to represent specific cases may lead to larger uncertainties. Below the resolution of 10 km \times 10 km, SGV can be characterized by the mean 314 315 value with relatively smaller uncertainty $(\pm 5\%)$ and hence high confidence, even with large diurnal 316 or day-to-day variations. The spatial resolutions of TEMPO, GEMS, Sentinel-4, and TROPOMI 317 (TROPOspheric Monitoring Instrument, Veefkind et al., 2012; Griffin et al., 2019; van Geffen et al., 2019) are within this ≤ 10 km $\times 10$ km range, while the resolution of OMI (Levelt et al., 2006; 318 319 2018) is not. This means that applying this study (e.g., Figure 4) to OMI for a specific case study 320 (e.g., a specific day) requires extra caution.

The GeoTASO data located closely in space may be sampled at slightly different times for the same flight. To explore the impact of temporal variability on this SGV analysis, we performed two sensitivity tests. The typical time period for a complete flight is ~4 hours. In the first test, we sampled GeoTASO data with hypothetical satellite pixels grouped by each complete flight, rather 325 than grouping the data by each hour (i.e., hourly bins). The resulting patterns and relationships are 326 similar to those derived from grouping data into hourly bins, except that the normalized satellite 327 SGV increases \sim 5% for small pixels due to temporal variability (Figure S7a). In the second test, 328 we sampled GeoTASO data with hypothetical satellite pixels grouped by each raster. The results 329 are still similar to those derived from grouping data into hourly bins (Figure 4), except that the 330 normalized satellite SGV increases ~1% for small pixels due to the inclusion of temporal 331 variability (Figure S7b). This is because sampling by raster includes smaller temporal variability 332 than sampling by flight, but larger temporal variability than sampling by hourly bins.

333 The three regions investigated in this work have different levels of urbanization and air 334 pollution (Figures 1 and S2). PBL conditions are also different in the morning and afternoon 335 (Figure S8). The similarity of the relationships between the satellite pixel size and the normalized 336 satellite SGV over these different regions (Figure 4) suggests that this relationship may be 337 generalizable to NO₂ VC over urban regions with different levels of urbanization and air pollution, 338 and different PBL conditions. Moreover, Figures 4 and 5 point to the possibility of developing a 339 generalized look-up table for the expected normalized satellite SGV for NO₂ VC over urban 340 regions at different satellite pixel sizes, especially for small pixel sizes (e.g., TEMPO, GEMS, and 341 TROPOMI). This would be useful in satellite design, satellite retrieval evaluation and 342 interpretation, and satellite-in situ data comparisons. For example, the satellite pixel size of 343 tropospheric NO₂ VC retrievals from GEMS, TEMPO, TROPOMI, and OMI are highlighted in Figure 4. Following Judd et al. (2019), we choose $3 \text{ km} \times 3 \text{ km}$, $5 \text{ km} \times 5 \text{ km}$, $7 \text{ km} \times 8 \text{ km}$, and 344 345 18 km \times 18 km pixels to represent the expected area of the satellite pixels for TEMPO (2.1 km \times 346 4.4 km), TROPOMI (3.5 km \times 7 km), GEMS (7 km \times 8 km), and OMI (18 km \times 18 km), 347 respectively. The expected normalized satellite SGV for TEMPO, TROPOMI, GEMS, and OMI 348 are 15–20%, ~20%, 20–25%, and ~30%, respectively. Taking the TEMPO example, this implies 349 that the satellite SGV could potentially lead to uncertainties of 15-20% in a validation exercise 350 comparing a satellite retrieval with local measurements of NO₂ VC, from a Pandora spectrometer 351 for example, that may be unrepresentative of the wider pixel area.

352 **3.2 Temporal variability (TeMD) within the same satellite pixels**

353 In addition to satellite spatial SGV, we also analyze the temporal variability (i.e., TeMD) 354 within the same hypothetical satellite pixels. Figure 6 shows TeMD of satellite retrieved 355 tropospheric NO₂ VC over SMA as a function of hypothetical satellite pixel size and the separation time (Dt) between flight rasters as described in Section 2.5. The results for 27 satellite pixel sizes 356 357 analyzed are shown by different colors, while results for selected satellite pixel sizes are 358 highlighted by thicker lines. For all the pixel sizes, TeMD increases monotonically with the time difference Dt between two sampled raster values within the same pixel. The TeMD of tropospheric 359 NO₂ VC is around 0.75×10^{16} molecules cm⁻² for a Dt of 2 hours over SMA for all the sampled 360 satellite pixel sizes and increases to $\sim 2 \times 10^{16}$ molecules cm⁻² for Dt of 8 hours. This indicates that, 361 along with improvements in the satellite retrieval spatial resolution with smaller pixels, improving 362 363 the satellite retrieval temporal resolution with higher frequency measurements is also an effective 364 way to enhance capability in resolving variabilities of NO₂.

To investigate the TeMD shown in Figure 6 we consider the particular factors driving NO₂ variability over SMA. NO₂ has a relatively short lifetime (~ a few hours) and a strong diurnal cycle due to emission activities, chemistry and changing photolysis rate (Fishman et al., 2011; Follette-

Cook et al., 2015). The diurnal cycle of the PBL may also play a large role because horizontal 368 369 dispersion occurs as the PBL thickens during the day. Early in the morning, the PBL is low (~1400 370 m during 9:00-11:00 in SMA during KORUS-AO) and strong source locations are evident such as 371 traffic on major highways, etc. As the day progresses, the PBL height increases (~1800 m during 372 15:00-17:00; Figure S8) due to enhanced convection, which further induces a stronger horizontal 373 divergence at the top of the convective cell that allows for greater horizontal dispersion to take 374 place along with the divergence. By early afternoon, emissions from all the major sources in the 375 central region have mixed together to form a wide area of high pollution over the urban center with 376 strong gradients of decreasing NO₂ out to the surrounding areas. In addition, changing wind 377 conditions (speed and direction; Figure S9) during the day can also lead to a shift in pollution 378 pattern, and result in different pollution conditions for the same pixel at different time of a day. 379 For example, Raster 1 of the 0609AM (9.17 local time) and Raster 2 of 0609PM (17 local time) 380 are used to calculate TeMD for Dt equals 8 hours. The differences in wind conditions (Figure S9) and the pollution patterns (Figure 2) are large. Judd et al. (2018) point out that the topography over 381 382 SMA also plays a role in the ability to mix horizontally as the PBL grows. Therefore, the TeMD 383 can be large between morning and afternoon (i.e., for Dt larger than 6 hours).

384 For a small Dt (2 or 4 hours), TeMD increases at higher spatial resolution (i.e., smaller 385 pixel size). This is especially true for short time periods (e.g., 2 hours and 4 hours), which is more 386 important for the GEO satellite measurements. For example, for Dt of 2 hours, TeMD for satellite pixels of 1 km \times 1 km is about 0.80 \times 10¹⁶ molecules cm⁻², while TeMD for satellite pixels of 25 387 km \times 25 km is about 0.73 \times 10¹⁶ molecules/cm² (~9% lower); when Dt is 4 hours, TeMD for satellite 388 pixels of 1 km \times 1 km is about 1.3 \times 10¹⁶ molecules cm⁻², while TeMD for satellite pixels of 25 km 389 \times 25 km is about 1.1×10¹⁶ molecules/cm² (~15% lower). This indicates that when decreasing pixel 390 391 size, the temporal variability of the retrieved values will increase, even though the normalized 392 satellite spatial SGV decreases. This is expected because averaging over a larger region with high 393 small-scale spatial variability smooths out temporal variability, and therefore produces smaller 394 hourly differences. Our finding here is consistent with that of Fishman et al. (2011).

395 As the time difference Dt increases, the temporal variability TeMD increases for all pixel 396 sizes. However, the TeMD is now greater at large pixel size which is in contrast to the higher 397 TeMD at small pixel size for shorter Dt. This is a result of the pollution pattern that develops over the SMA during the day (June 9th, 2019) as described above. The higher TeMD reflects the fact 398 399 that many of the large pixels now span the strong NO₂ gradient between the urban and surrounding 400 area resulting in a much higher spatial variability than earlier in the day at a spatial scale not 401 captured with the smaller pixels. As a caution, we note that TeMD for 8 hours is determined by only the difference between Raster 1 of the 0609AM and Raster 2 of 0609PM (Figure 2), and that 402 403 the regional coverage for Raster 2 of 0609PM is different from the coverage of the other PM rasters. 404 Therefore, the relationship of TeMD and spatial resolution for a large Dt (e.g., 6 or 8 hours) over 405 SMA requires further study.

GeoTASO data over the Busan region is limited. Given the fewer flights, we are not able to show how TeMD changes with Dt over the Busan region in this study. However, we are able to show the relationship between TeMD and satellite pixel sizes. During KORUS-AQ, there were only two rasters sampled over Busan with a Dt of 2 hours (Figure S10). For this Dt of 2 hours, TeMD increases slightly at higher satellite retrieval spatial resolution (smaller pixel size). More data over the Busan region would help significantly for this analysis. For the LA Basin GeoTASO

data, sampled hypothetical satellite pixels show TeMD increases at higher spatial resolution for 412 413 the available Dt equal to 4 and 8 hours (Figure S11). However, TeMD is fairly constant at these 414 two time differences which is different to what was observed over SMA (Figure 6). We note that 415 with only 2 flight days of flight data, the GeoTASO data over LA is also limited, which may be the main driver of the difference. Besides the limited data, one possible reason is the different wind 416 417 fields over the two regions. As mentioned previously, Raster 1 of the 0609AM and Raster 2 of 418 0609PM are used to calculate TeMD for Dt equals 8 hours over SMA. The differences in wind 419 direction (Figure S9) for the two rasters are large (almost opposite in some cases). However, over 420 LA, the differences in wind direction (Figure S12) for the two rasters (Rasters 1 and 3 for 0627 421 flight) are relatively small, compared to the differences over SMA. Despite the limited sample 422 sizes, TeMD increases when increasing the satellite retrieval spatial resolution over both the Busan 423 region and the LA Basin, which is consistent with the relationships over the SMA for a small Dt.

424 **3.3 Results from Spatial Structure Function (SSF)**

425 In this section, we show the analysis of SSF over SMA (Figure 7) as a complement to our 426 analysis in Section 3.1. As mentioned before, SSF and SGV are different measures of spatial 427 variability and are not directly comparable. This is because SSF is calculated based on differences 428 between a single GeoTASO measurement and all the other GeoTASO measurements on the map, 429 while SGV is derived based on variation among all the GeoTASO measurements within a 430 hypothetical satellite pixel unit. SSF measures the averaged spatial difference at a given distance, while SGV directly quantifies the expected spatial variability within a satellite pixel at a given size. 431 432 As both SSF and SGV are related to spatial variability, we include SSF in this study as an extension 433 to SGV.

Figure 7a shows that the SSF in SMA initially increases with the distance between data 434 435 points, peaks at around 40-60 km during most flights, and then decreases with distance between 436 60 and 140 km. The number of paired GeoTASO data points when the distance is larger than 100 437 km is relatively small (Figure S13) therefore conclusions beyond this distance are not included in 438 this analysis. The increases in SSF for distances in the range of 1-25 km (Figure 7b) are consistent 439 with the relationship between pixel sizes and the normalized satellite SGV shown in Figure 4. For 440 example, over the 1-25 km range, Fig 4a shows the median increases from around 8% to around 441 28%, an increase by a factor of 3.5, and the black line in Figure 7 shows an approximately similar factor (from 0.33×10^{16} molecules/cm² for 1 km to 1.5×10^{16} molecules/cm² for 25 km). This 442 increase of SSF between 1-25 km is also seen over the Busan region and the LA Basin (Figure 443 444 S14). We also notice that SSF shows a relatively strong dependence on the particular GeoTASO 445 flight, while SGV is less sensitive, especially for small pixel sizes.

446 The shapes of the SSF are generally consistent with previous studies for modeled or in situ 447 observations of NO₂ (Fishman et al., 2011; Follette-Cook et al., 2015). Previous studies also suggest that different aircraft campaigns may share the common shape of SSF but different 448 magnitudes, which is strongly related to the fraction of polluted samples versus samples of 449 450 background air in the campaign (Crawford et al., 2009; Fishman et al., 2011). Differences in the 451 shape and size of particular cities also contribute to the differences in the SSF. For example, at a 452 certain distance SSF may compare polluted areas within the same urban region, while over a 453 different smaller city, the comparison at the same distance reveals the gradient between the polluted city and cleaner surrounding background air, so resulting in different peak values. Valin 454

455 et al. (2011) found that the maximum in OH feedback in a NOx-OH steady-state relationship 456 corresponds to a NO₂ e-folding decay length of 54 km in 5m/s winds. This may partially explain 457 the peak between 40~60 km in SSF. As shown in Figures 2 and S7, the overall spatial variability 458 over SMA is higher in the afternoon. Over SMA, the SSF in the morning is generally smaller than 459 in the afternoon, indicating higher spatial variability of tropospheric NO₂ VC in the afternoon (see 460 also Judd et al., 2018). As described in Section 2.6, SSF is calculated based on hourly binned data. However, the overall shapes of SSF (Figure S15) calculated on raster basis are similar to SSF 461 462 calculated on hourly basis (Figure 7).

463 Previous studies (Fishman et al., 2011; Follette-Cook et al., 2015) used SSF values at a 464 particular distance to indicate the satellite precision requirement at a corresponding resolution in order to resolve spatial structure over the pixel scale. For GEMS, the expected spatial differences 465 over the scale of its pixel for the SMA and Busan regions are $\sim 7.5 \times 10^{15}$ molecules cm⁻² and 466 $\sim 3.5 \times 10^{15}$ molecules cm⁻², respectively, taking the SSF values at 5 km to be representative. For 467 TEMPO, the spatial difference is $\sim 2.8 \times 10^{15}$ molecules cm⁻² over LA Basin taking the SSF value 468 at 3 km. Assuming the NO₂ measurement precision requirement to be 1×10^{15} molecules cm⁻² for 469 both TEMPO and GEMS (Chance et al., 2013; Kim et al., 2020), the expected spatial differences 470 471 over the three regions are considerably higher than the precision requirement and should be easily 472 characterized by both the GEMS and TEMPO missions.

473 **4. Discussions and implications**

474 The relationship between satellite pixel sizes and the normalized satellite SGV is fairly 475 robust over the three different urban regions studied here, and Figure 4 points to the possibility of 476 developing a generalized look-up table if more data were available in other urban regions. We note 477 that the GeoTASO data used in this study were sampled during spring and summer. In our future 478 study, we will include more GeoTASO data in the analysis to test the applicability of the look-up 479 table approach under different seasonal conditions and sources. A generalized relationship 480 between satellite pixel sizes and the temporal variability (Figure 6) is not as evident as the 481 relationship between satellite pixel sizes and the normalized satellite SGV due to limited data. 482 However, it is still useful for satellite observations over SMA, which is in the GEMS domain and 483 should be helpful in satellite retrieval interpretation.

Previous studies recognized the challenges in satellite validation/evaluation for NO₂ retrievals due to satellite SGV and representativeness error of in situ measurements (e.g., Nowlan et al., 2016, 2018; Judd et al., 2019; Pinardi et al., 2020; Tack et al., 2020). The gapless airborne mapping datasets of GeoTASO with sufficient spatiotemporal resolution are a promising way to address the issue of satellite SGV and representativeness errors in satellite validation/evaluation (e.g., Nowlan et al., 2016, 2018; Judd et al., 2019).

490 Challenges due to SGV also have implications for other trace gas column measurements. 491 For example, in Tang et al. (2020), satellite SGV and representativeness errors of in situ 492 measurements introduced uncertainties in validation of CO retrievals from the MOPITT 493 (Measurement Of Pollution In The Troposphere) satellite instrument. Normalized SGV of the 494 GeoTASO tropospheric NO₂ VC might serve as an upper bound to the SGV of CO, SO₂ and other 495 species that share common source(s) with NO₂ but with relatively longer lifetimes than NO₂, even 496 if their spatial distributions have different patterns (e.g., Chong et al., 2020). For example, at the 497 resolution of 22 km \times 22 km (resolution of MOPITT CO retrievals), the expected normalized 498 satellite SGV of tropospheric NO₂ VC is ~30%. Therefore, we might expect the normalized 499 satellite SGV for tropospheric CO VC to be lower than this value.

500 To demonstrate this idea, we use the WRF-Chem regional model as an intermediary step. 501 At the model resolution, if the SVG of the WRF-Chem model and GeoTASO NO₂ VC agree 502 reasonably well, then the model can be used to predict the SVG of other species that are chemically 503 constrained with NO₂ at the model and coarser resolutions. This is shown in Figure 8 which illustrates how SGV varies with satellite pixel size for NO₂ VC, CO VC, SO₂ VC, and HCHO VC 504 505 calculated from a WRF-Chem simulation. The modeled NO₂, CO, SO₂, and HCHO concentrations are converted to VC, and are filtered to match the rasters of GeoTASO measurements (Figure S16). 506 507 As expected, SGV of modeled NO₂ VC is higher than SGV of modeled CO VC, SO₂ VC, and 508 HCHO VC. We also notice that SGV for modeled NO₂ VC, CO VC, SO₂ VC, and HCHO VC 509 increases with pixel size, which is similar to that for GeoTASO measurements. The SGV for 510 GeoTASO NO₂ shown in this figure (black lines) is calculated based on GeoTASO data that are regridded to the WRF-Chem grid ($3 \text{ km} \times 3 \text{ km}$), making it slightly different from that in Figure 511 512 4. We note that the modelled NO₂ SGV is greater than that calculated from the GeoTASO data 513 indicating that further work is required to reconcile difference due to model descriptions of 514 emissions, chemistry and transport. And ideally, dense GeoTASO-type measurements of CO and 515 other species would allow for a more comprehensive assessment of this approach.

516 This study is also relevant to model comparison and evaluation with in situ observations. 517 Whenever in situ observations are compared to grid data (e.g., comparisons between satellite 518 retrievals and in situ observations, comparisons between grid-based model and in situ observations, and in data assimilation). SGV will introduce uncertainties that need to be quantified to better 519 520 interpret and understand the comparison results. For example, we note that at the resolution of 14 km×14 km (a typical resolution for the forward-looking Multi-Scale Infrastructure for Chemistry 521 522 and Aerosols Version 0; MUSICA-V0, https://www2.acom.ucar.edu/sections/multi-scalechemistry-modeling-musica; Pfister et al. (2020)), Figure 8 shows that the expected normalized 523 524 SGV of tropospheric NO₂ VC is \sim 25-30%. This suggests that when comparing model simulations at coarser resolution with local observations of tropospheric NO₂ VC, a larger normalized SGV 525 526 than this ~25-30% might be expected. If comparing for a specific vertical layer instead of vertical 527 column, an even larger normalized SGV may occur.

528 For data assimilation and inverse modeling application (e.g., top-down emission 529 estimations from satellite observations), it is essential to accurately characterize the observation 530 error covariance matrix **R** (Janjíc et al., 2017). The first component of **R** is the instrument error covariance matrix due to instrument noise and retrieval uncertainty in the case of trace gas satellite 531 532 data. The second component is the representation error covariance matrix, arising from 533 fundamental differences of the atmospheric sampling, typically when assimilating a local point 534 measurement into a grid-based model (Boersma et al, 2016). The observation error covariance due 535 to representativeness error is difficult to define, but can be parameterized when calculating super 536 observations by inflating the observation error variances (Boersma et al., 2016) and quantified by a posteriori diagnostics estimation (Gaubert et al. 2014). Knowledge of the fine-scale model sub-537 538 grid variability is therefore essential to verify those assumptions and inform error statistics for 539 application to chemical data assimilation studies. Our results suggest large potential improvements 540 in emission estimates when assimilating high spatial resolution TROPOMI and GEO satellite data

541 with SGV of $\sim 10\%$ –20% (Figure 4), compared to OMI data with SVG of $\sim 30\%$ (Figure 4), in line 542 with the existing literature for NO₂ (e.g., Valin et al., 2011). We have also shown that significant 543 temporal variability of NO₂ is expected at higher spatial resolutions. This observed signal will 544 open new avenue for space-based monitoring of atmospheric chemistry and will reduce errors of 545 inverse estimates of fluxes.

546 **5. Conclusions**

547 Satellite SGV is a key issue in interpreting satellite retrieval results. Quantifying studies 548 have been lacking due to limited observations at high spatial and temporal resolution. In this study, 549 we have quantified likely GEO satellite SGV by using GeoTASO measurements of tropospheric 550 NO_2 VC over the urbanized and polluted Seoul Metropolitan Area (SMA) and the less-polluted 551 Busan region during KORUS-AQ, and the Los Angeles (LA) Basin during the 2017 SARP 552 campaigns. The main findings of this work are the following:

- 553(1) The normalized satellite SGV increases with pixel size based on random sampling of hourly554GeoTASO data, from ~10% (\pm 5% for specific cases such as an individual day/time of day) for555a pixel size of 0.5 km × 0.5 km to ~35% (\pm 10% for specific cases such as an individual day/time556of day) for the pixel size of 25 km × 25 km. This conclusion holds for all of the three urban557regions in this study despite their different levels of urbanization and pollution, and for time558of day being morning or afternoon.
- (2) Due to its relatively shorter atmospheric lifetime, normalized satellite SGV of tropospheric
 NO₂ VC could serve as an upper bound to satellite SGV of CO, SO₂ and other species that
 share common source(s) with NO₂. This conclusion is supported by high-resolution WRF Chem simulations.
- 563 (3) The temporal variability (TeMD) increases with sampling time differences (Dt) over SMA. 564 TeMD ranges from $\sim 0.75 \times 10^{16}$ molecules cm⁻² at Dt of 2 hours to $\sim 2 \times 10^{16}$ molecules cm⁻² 565 (about three times higher) at Dt of 8 hours. TeMD is caused by temporal variation in emission 566 activities, photolysis, and meteorology throughout the day. Improving the satellite retrieval 567 temporal resolution is an effective way to enhance the capability of satellite products in 568 resolving temporal variability of NO₂.
- (4) Temporal variability (TeMD) increases as pixel size decreases in SMA when time difference is less than 4 hours. Analysis confidence at greater time differences would require more flight datasets with longer time separations during the day. For example, when Dt is 2 hours, TeMD for satellite pixels with the size of 25 km × 25 km is about 20% lower compared to TeMD for satellite pixels with the size of 1 km × 1 km. Thus, ideally, temporal resolution should be increased along with any increase in spatial resolution in order to enhance the accuracy of satellite products.
- (5) The spatial structure function (SSF) at first increases with the distance between points, peaking
 at around 40-60 km during most flight days before decreasing at greater distances. This is
 generally consistent with previous studies.
- 579 (6) SSF analyses suggest that GEMS will encounter NO₂ VC pixel scale spatial differences of 580 $\sim 7.5 \times 10^{15}$ and $\sim 3.5 \times 10^{15}$ molecules cm⁻² over the SMA and Busan regions, respectively. 581 TEMPO will encounter NO₂ VC spatial differences at its pixel scale of $\sim 2.8 \times 10^{15}$ molecules

582 cm^{-2} over the LA Basin. These differences should be easily resolved by the instruments at the 583 stated measurement precision requirement of 1×10^{15} molecules cm⁻².

(7) These findings are relevant to future satellite design and satellite retrieval interpretation,
especially now with the deployment of the high-resolution GEO air quality satellite
constellation, GEMS, TEMPO, and Sentinel-4. This study also has implication for satellite
product validation and evaluation, satellite—in situ data comparisons, and more general pointgrid data comparisons. These share similar issues of sub-grid variability and the need for
quantification of representativeness error.

590 We note that this study has some uncertainties and limitations. (1) The variability at a resolution finer than 250 m × 250 m (i.e., GeoTASO's resolution) may introduce uncertainties to 591 the analysis here, although this is beyond the scope of this study. (2) Even though a large number 592 593 of GeoTASO retrievals have been analyzed in this study, we would still benefit from more 594 GeoTASO flights with a broader spatiotemporal coverage. More GeoTASO-type data over the 595 Busan region and LA Basin will help in testing the consistence in TeMD over different regions. 596 (3) The KORUS-AO campaign was conducted in Spring (May and June), and the 2017 SARP 597 campaign was also conducted in June. More GeoTASO-type measurements over South Korea during different season(s) would be particularly helpful to understand and generalize the findings 598 599 in this study. (4) The three regions analyzed in this study are urban regions, and the results are not 600 tested over cleaner background areas that may be characterized by less heterogeneity.

601 This work demonstrates the value of continued flights of GeoTASO-type instruments for 602 obtaining continuous, high spatial resolution data several times a day for assessing SGV. This will 603 be a particularly useful reference in the comparisons of satellite retrievals and in situ measurements 604 that may have representativeness errors.

605

606 Acknowledgement

607 The authors thank the GeoTASO team for providing the GeoTASO measurements. The authors 608 thank the KORUS-AQ and SARP team for the campaign data. We thank the DIAL-HSRL team the mixing layer height data (available at https://www-air.larc.nasa.gov/cgi-609 for bin/ArcView/korusaq). Tang was supported by a NCAR Advanced Study Program Postdoctoral 610 611 Fellowship. Edwards was partially supported by the TEMPO Science Team under Smithsonian 612 Astrophysical Observatory Subcontract SV3-83021. The National Center for Atmospheric Research (NCAR) is sponsored by the National Science Foundation. The authors thank Ivan 613 614 Ortega and Sara-Eva Martinez-Alonso for helpful comments on the paper.

615

616 Data availability

The KORUS-AQ and SARP data are available at https://www-air.larc.nasa.gov/cgi bin/ArcView/korusaq and https://www-air.larc.nasa.gov/cgi-bin/ArcView/lmos, respectively.

- 620 **Reference**
- 621 Al-Saadi, Jassim, Gregory Carmichael, James Crawford, Louisa Emmons, Saewung Kim, Chang-
- 622 Keun Song, et al.:. KORUS-AQ: An international cooperative air quality field study in Korea, the

623 KORUS-AQ white paper, 2014 (https://espo.nasa.gov/korus-aq/content/ KORUS-624 AQ_White_Paper), 2014.

Boersma, K. F., Vinken, G. C. M., and Eskes, H. J.: Representativeness errors in comparing
chemistry transport and chemistry climate models with satellite UV–Vis tropospheric column
retrievals, Geosci. Model Dev., 9, 875–898, https://doi.org/10.5194/gmd-9-875-2016, 2016.

- Broccardo, S., Heue, K.-P., Walter, D., Meyer, C., Kokhanovsky, A., van der A, R., Piketh, S., Langerman, K., and Platt, U.: Intra-pixel variability in satellite tropospheric NO2 column densities
- 630 derived from simultaneous space-borne and airborne observations over the South African
- 631 Highveld, Atmos. Meas. Tech., 11, 2797–2819, https://doi.org/10.5194/amt-11-2797-2018, 2018.
- 632 Chance, K., Liu, X., Suleiman, R. M., Flittner, D. E., Al-Saadi, J. and Janz, S. J.: Tropospheric 633 emissions: Monitoring of pollution (TEMPO), Proceedings of SPIE, Vol. 8866, Earth Observin
- 634 Systems XVIII, 88660D (September 23, 2013), San Diego, CA USA, 2013.
- 635 Ching, J., Herwehe, J., and Swall, J.: On joint deterministic grid modeling and sub-grid variability 636 conceptual framework for model evaluation, Atmos. Environ., 40, 4935–4945, 2006.
- 637 Choi, S., Lamsal, L. N., Follette-Cook, M., Joiner, J., Krotkov, N. A., Swartz, W. H., Pickering,
- 638 K. E., Loughner, C. P., Appel, W., Pfister, G., Saide, P. E., Cohen, R. C., Weinheimer, A. J., and
- 639 Herman, J. R.: Assessment of NO2 observations during DISCOVER-AQ and KORUS-AQ field
- 640 campaigns, Atmos. Meas. Tech., 13, 2523–2546, https://doi.org/10.5194/amt-13-2523-2020, 2020.
- 641 Chong, H.; Lee, S.; Kim, J.; Jeong, U.; Li, C.; Krotkov, N.; Nowlan, C.; Al-Saadi, J.; Janz, S.;
- Kowalewski, M.; et al. High-resolution mapping of SO2 using airborne observations from the
 GeoTASO instrument during the KORUS-AQ field study: PCA-based vertical column retrievals.
 Remote Sens. Environ., 241, 111725, 2020.
- Courrèges-Lacoste, G. B., Sallusti, M., Bulsa, G., Bagnasco, G., Veihelmann, B., Riedl, S., Smith,
 D. J., and Maurer, R.: The Copernicus Sentinel 4 mission: a geostationary imaging UVN
 spectrometer for air quality monitoring, Proceedings Volume 10423, Sensors, Systems, and NextGeneration Satellites XXI, 1042307, https://doi.org/10.1117/12.2282158, 2017.
- 649 Crawford, J.H.: Assessing scales of variability for constituents relevant to future geostationary
 650 satellite observations and models of air quality. AGU, 90(52), Fall Meet. Suppl., Abstract
 651 A53Ae0237, 2009.
- 652 Crawford, J. H., et al., The Korea-United States Air Quality (KORUS-AQ) Field Study, Elementa,
 653 in press, 2021.
- Denby, B., Cassiani, M., de Smet, P., de Leeuw, F., and Horálek, J.: Sub-grid variability and its
 impact on European wide air quality exposure assessment, Atmos. Environ., 45, 4220–4229, 2011.
- 656 Deeter, M. N., Edwards, D. P., Francis, G. L., Gille, J. C., Mao, D., Martínez-Alonso, S., Worden, 657 H. M., Ziskin, D., and Andreae, M. O.: Radiance-based retrieval bias mitigation for the MOPITT 658 version product. Meas. Tech., 4561-4580, instrument: the 8 Atmos. 12. 659 https://doi.org/10.5194/amt-12-4561-2019, 2019.

660 Finlayson-Pitts, B. J. and Pitts, J. N., J.: Tropospheric air pollution: Ozone, airborne toxics, 661 polycyclic aromatic hydrocarbons, and particles, Science, 276, 1045–1052, 1997.

Fishman, J., Silverman, M. L., Crawford, J. H., and Creilson, J. K.: A study of regional-scale
variability of in situ and model-generated tropospheric trace gases: Insights into observational
requirements for a satellite in geostationary orbit, Atmos. Environ., 45, 4682–4694, 2011.

Follette-Cook, M., Pickering, K., Crawford, J., Duncan, B., Loughner, C., Diskin, G., Fried, A.,
and Weinheimer, A.: Spatial and temporal variability of trace gas columns derived from
WRF/Chem regional model output: Planning for geostationary observations of atmospheric
composition, Atmos. Environ., 118, 28–44, doi:10.1016/j.atmosenv.2015.07.024, 2015.

Friedl, M., Sulla-Menashe, D. (2015). MCD12C1 MODIS/Terra+Aqua Land Cover Type Yearly
L3 Global 0.05Deg CMG V006 [Data set]. NASA EOSDIS Land Processes DAAC. Accessed
2019-08-12 from https://doi.org/10.5067/MODIS/MCD12C1.006.

- 672 Gaubert, B., Coman, A., Foret, G., Meleux, F., Ung, A., Rouil, L., Ionescu, A., Candau, Y., and 673 Beekmann, M.: Regional scale ozone data assimilation using an ensemble Kalman filter and the 674 chemical CHIMERE transport model, Geosci. Model Dev., 7, 283 - 302, https://doi.org/10.5194/gmd-7-283-2014, 2014. 675
- Gaubert, B., Emmons, L. K., Raeder, K., Tilmes, S., Miyazaki, K., Arellano Jr., A. F., Elguindi,
 N., Granier, C., Tang, W., Barré, J., Worden, H. M., Buchholz, R. R., Edwards, D. P., Franke, P.,
 Anderson, J. L., Saunois, M., Schroeder, J., Woo, J.-H., Simpson, I. J., Blake, D. R., Meinardi, S.,
 Wennberg, P. O., Crounse, J., Teng, A., Kim, M., Dickerson, R. R., He, H., Ren, X., Pusede, S. E.,
 and Diskin, G. S.: Correcting model biases of CO in East Asia: impact on oxidant distributions
 during KORUS-AQ, Atmos. Chem. Phys., 20, 14617–14647, https://doi.org/10.5194/acp-2014617-2020, 2020.
- Gerbig, C., Lin, J. C., Wofsy, S. C., Daube, B. C., Andrews, A. E., Stephens, B. B., Bakwin, P. S.,
 and Grainger, C. A.: Toward constraining regional-scale fluxes of CO2 with atmospheric
 observations over a continent: 1. Observed spatial variability from airborne platforms, J. Geophys.
 Res.-Atmos., 108, 4756, doi:4710.1029/2002JD003018, 2003.
- 687 Global Modeling and Assimilation Office (GMAO) (2015), MERRA-2 inst3_3d_asm_Nv: 3d,3-
- 688 Hourly,Instantaneous,Model-Level,Assimilation,Assimilated Meteorological Fields V5.12.4,
- 689 Greenbelt, MD, USA, Goddard Earth Sciences Data and Information Services Center (GES DISC),
- 690 Accessed: Mar 29, 2021, 10.5067/WWQSXQ8IVFW8.
- Griffin, D., Zhao, X., McLinden, C. A., Boersma, F., Bourassa, A., Dammers, E., Degenstein, D.,
 Eskes, H., Fehr, L., Fioletov, V., Hayden, K., Kharol, S. K., Li, S.-M., Makar, P., Martin, R. V.,
 Mihele, C., Mittermeier, R. L., Krotkov, N., Sneep, M., Lamsal, L. N., ter Linden, M., van Geffen,
 J., Veefkind, P., and Wolde, M.: High-Resolution Mapping of Nitrogen Dioxide with TROPOMI:
 First Results and Validation Over the Canadian Oil Sands, Geophys. Res. Lett., 46, 1049–1060,
 https://doi.org/10.1029/2018GL081095, 2019.
- Harris, D., Foufoula-Georgiou, E., Droegemeier, K. K., and Levit, J. J.: Multiscale Statistical
 Properties of a High-Resolution Precipitation Forecast, J. Hydrometeor., 2, 406–418, 2001.

Hoesly, R. M., Smith, S. J., Feng, L., Klimont, Z., Janssens-Maenhout, G., Pitkanen, T., Seibert,
J. J., Vu, L., Andres, R. J., Bolt, R. M., Bond, T. C., Dawidowski, L., Kholod, N., Kurokawa, J.I., Li, M., Liu, L., Lu, Z., Moura, M. C. P., O'Rourke, P. R., and Zhang, Q.: Historical (1750–2014)
anthropogenic emissions of reactive gases and aerosols from the Community Emissions Data
System (CEDS), Geosci. Model Dev., 11, 369–408, https://doi.org/10.5194/gmd-11-369-2018,
2018.

- Huang, M., Crawford, J. H., Diskin, G. S., Santanello, J. A., Kumar, S. V., Pusede, S. E., Parrington,
- M., and Carmichael, G. R.: Modeling Regional Pollution Transport Events During KORUS-AQ:
 Progress and Challenges in Improving Representation of Land-Atmosphere Feedbacks, J. Geophys.
 Res.-Atmos., 123, 10–732, 2018.
- 709 Janjic, T., Bormann, N., Bocquet, M., Carton, J. A., Cohn, 'S. E., Dance, S. L., Losa, S. N.,
- Nichols, N. K., Potthast, R., Waller, J. A., and Weston, P.: On the representation error in data assimilation, Q. J. R. Meteorol. Soc., 144, 1257–1278, https://doi.org/10.1002/qj.3130, 2018.
- Judd, L. M., Al-Saadi, J. A., Janz, S. J., Kowalewski, M. G., Pierce, R. B., Szykman, J. J., Valin,
- 713 L. C., Swap, R., Cede, A., Mueller, M., Tiefengraber, M., Abuhassan, N., and Williams, D.:
- 714 Evaluating the impact of spatial resolution on tropospheric NO2 column comparisons within urban
- 715 areas using high-resolution airborne data, Atmos. Meas. Tech., 12, 6091-6111,
- 716 https://doi.org/10.5194/amt-12-6091-2019, 2019.
- Kim, H. C., Lee, P., Judd, L., Pan, L., and Lefer, B.: OMI NO2 column densities over North
 American urban cities: the effect of satellite footprint resolution, Geosci. Model Dev., 9, 1111–
 1123, https://doi.org/10.5194/gmd-9-1111-2016, 2016.

Kim, H., Zhang, Q., and Heo, J.: Influence of intense secondary aerosol formation and long-range
transport on aerosol chemistry and properties in the Seoul Metropolitan Area during spring time:
results from KORUS-AQ, Atmos. Chem. Phys., 18, 7149–7168, https://doi.org/10.5194/acp-187149-2018, 2018.

- 724 Kim, J., Jeong, U., Ahn, M.-H., Kim, J. H., Park, R. J., Lee, H., Song, C. H., Choi, Y.-S., Lee, K.-725 H., Yoo, J.-M., Jeong, M.-J., Park, S. K., Lee, K.-M., Song, C.-K., Kim, S.-W., Kim, Y. J., Kim, S.-W., Kim, M., Go, S., Liu, X., Chance, K., Miller, C. C., Al-Saadi, J., Veihelmann, B., Bhartia, 726 727 P. K., Torres, O., Abad, G. G., Haffner, D. P., Ko, D. H., Lee, S. H., Woo, J.-H., Chong, H., Park, 728 S. S., Nicks, D., Choi, W. J., Moon, K.-J., Cho, A., Yoon, J., Kim, S.-K., Hong, H., Lee, K., Lee, H., Lee, S., Choi, M., Veefkind, P., Levelt, P. F., Edwards, D. P., Kang, M., Eo, M., Bak, J., Baek, 729 730 K., Kwon, H.-A., Yang, J., Park, J., Han, K. M., Kim, B.-R., Shin, H.-W., Choi, H., Lee, E., Chong, 731 J., Cha, Y., Koo, J.-H., Irie, H., Hayashida, S., Kasai, Y., Kanaya, Y., Liu, C., Lin, J., Crawford, J. H., Carmichael, G. R., Newchurch, M. J., Lefer, B. L., Herman, J. R., Swap, R. J., Lau, A. K. 732 733 H., Kurosu, T. P., Jaross, G., Ahlers, B., Dobber, M., McElroy, C. T., and Choi, Y.: New era of air 734 quality monitoring from space, Geostationary Environment Monitoring Spectrometer (GEMS), B. 735 Am. Meteorol. Soc., 101, E1–E22, https://doi.org/10.1175/BAMS-D-18-0013.1, 2020.
- Lamsal, L. N., Janz, S., Krotkov, N., Pickering, K. E., Spurr, R. J. D., Kowalewski, M., Loughner,
 C. P., Crawford, J., Swartz, W. H., and Herman, J. R.: High-resolution NO2 observations from the

- Airborne Compact Atmospheric Mapper: Retrieval and validation, J. Geophys. Res., 122, 1953–
 1970, https://doi.org/10.1002/2016JD025483, 2017.
- Leitch, J. W., Delker, T., Good, W., Ruppert, L., Murcray, F., Chance, K., Liu, X., Nowlan, C., Janz, S. J., Krotkov, N. A., Pickering, K. E., Kowalewski, M., and Wang, J.: The GeoTASO airborne spectrometer project, Earth Observing Systems XIX, 17–21 August 2014, San Diego, California, United States, Proc. SPIE, 9218, 92181H-9, https://doi.org/10.1117/12.2063763, 2014.
- Levelt, P. F., Oord, G. H. J. van den, Dobber, M. R., Malkki, A., Visser, H., Vries, J. de, Stammes,
 P., Lundell, J. O. V., and Saari, H.: The ozone monitoring instrument, IEEE T. Geosci. Remote,
 44, 1093–1101, https://doi.org/10.1109/TGRS.2006.872333, 2006.
- 747 Levelt, P. F., Joiner, J., Tamminen, J., Veefkind, J. P., Bhartia, P. K., Stein Zweers, D. C., Duncan, 748 B. N., Streets, D. G., Eskes, H., van der A, R., McLinden, C., Fioletov, V., Carn, S., de Laat, J., DeLand, M., Marchenko, S., McPeters, R., Ziemke, J., Fu, D., Liu, X., Pickering, K., Apituley, A., 749 González Abad, G., Arola, A., Boersma, F., Chan Miller, C., Chance, K., de Graaf, M., 750 751 Hakkarainen, J., Hassinen, S., Ialongo, I., Kleipool, Q., Krotkov, N., Li, C., Lamsal, L., Newman, 752 P., Nowlan, C., Suleiman, R., Tilstra, L. G., Torres, O., Wang, H., and Wargan, K.: The Ozone Monitoring Instrument: overview of 14 years in space, Atmos. Chem. Phys., 18, 5699-5745, 753 754 https://doi.org/10.5194/acp-18-5699-2018, 2018.
- Miyazaki, K., Sekiya, T., Fu, D., Bowman, K., Kulawik, S., Sudo, K., Walker, T., Kanaya, Y.,
 Takigawa, M., Ogochi, K., Eskes, H., Boersma, K. F., Thompson, A. M., Gaubert, B., Barre, J.,
 and Emmons, L. K.: Balance of Emission and Dynamical Controls on Ozone During the KoreaUnited States Air Quality Campaign From Multiconstituent Satellite Data Assimilation, J.
 Geophys. Res.-Atmos., 124, 387–413, 2019.
- 760 National Centers for Environmental Prediction/National Weather Service/NOAA/U.S. Department of Commerce. 2015, updated daily. NCEP GDAS/FNL 0.25 Degree Global 761 Tropospheric Analyses and Forecast Grids. Research Data Archive at the National Center for 762 763 Atmospheric Research, Computational and Information Systems Laboratory. https://doi.org/10.5065/D65Q4T4Z. 764
- Nowlan, C. R., Liu, X., Leitch, J. W., Chance, K., González Abad, G., Liu, C., Zoogman, P., Cole,
 J., Delker, T., Good, W., Murcray, F., Ruppert, L., Soo, D., Follette-Cook, M. B., Janz, S. J.,
 Kowalewski, M. G., Loughner, C. P., Pickering, K. E., Herman, J. R., Beaver, M. R., Long, R. W.,
 Szykman, J. J., Judd, L. M., Kelley, P., Luke, W. T., Ren, X., and Al-Saadi, J. A.: Nitrogen dioxide
 observations from the Geostationary Trace gas and Aerosol Sensor Optimization (GeoTASO)
 airborne instrument: Retrieval algorithm and measurements during DISCOVER-AQ Texas 2013,
 Atmos. Meas. Tech., 9, 2647–2668, https://doi.org/10.5194/amt-9-2647-2016, 2016.
- Nowlan, C. R., Liu, X., Janz, S. J., Kowalewski, M. G., Chance, K., Follette-Cook, M. B., Fried,
 A., González Abad, G., Herman, J. R., Judd, L. M., Kwon, H.-A., Loughner, C. P., Pickering, K.
 E., Richter, D., Spinei, E., Walega, J., Weibring, P., and Weinheimer, A. J.: Nitrogen dioxide and
 formaldehyde measurements from the GEOstationary Coastal and Air Pollution Events (GEOCAPE) Airborne Simulator over Houston, Texas, Atmos. Meas. Tech., 11, 5941–5964,
 https://doi.org/10.5194/amt11-5941-2018, 2018.

Pfister, G. G., Eastham, S. D., Arellano, A. F., Aumont, B., Barsanti, K. C., Barth, M. C., Conley,
A., Davis, N. A., Emmons, L. K., Fast, J. D., Fiore, A. M., Gaubert, B., Goldhaber, S., Granier, C.,
Grell, G. A., Guevara, M., Henze, D. K., Hodzic, A., Liu, X., Marsh, D. R., Orlando, J. J., Plane,
J. M. C., Polvani, L. M., Rosenlof, K. H., Steiner, A. L., Jacob, D. J., and Brasseur, G. P.: The
Multi-Scale Infrastructure for Chemistry and Aerosols (MUSICA), B. Am. Meteorol. Soc., 101,
1743–1760, 2020.

Pillai, D., Gerbig, C., Marshall, J., Ahmadov, R., Kretschmer, R., Koch, T., and Karstens, U.: High
resolution modeling of CO2 over Europe: implications for representation errors of satellite
retrievals, Atmos. Chem. Phys., 10, 83–94, https://doi.org/10.5194/acp-10-83-2010, 2010.

787 Pinardi, G., Van Roozendael, M., Hendrick, F., Theys, N., Abuhassan, N., Bais, A., Boersma, F., 788 Cede, A., Chong, J., Donner, S., Drosoglou, T., Dzhola, A., Eskes, H., Frieß, U., Granville, J., Herman, J. R., Holla, R., Hovila, J., Irie, H., Kanaya, Y., Karagkiozidis, D., Kouremeti, N., 789 790 Lambert, J.-C., Ma, J., Peters, E., Piters, A., Postylyakov, O., Richter, A., Remmers, J., Takashima, 791 H., Tiefengraber, M., Valks, P., Vlemmix, T., Wagner, T., and Wittrock, F.: Validation of 792 tropospheric NO2 column measurements of GOME-2A and OMI using MAX-DOAS and direct 793 sun network observations, Atmos. Meas. Tech., 13, 6141-6174, https://doi.org/10.5194/amt-13-794 6141-2020, 2020.

Qian, Y., Gustafson Jr., W. I., and Fast, J. D.: An investigation of the sub-grid variability of trace
gases and aerosols for global climate modeling, Atmos. Chem. Phys., 10, 6917–6946,
https://doi.org/10.5194/acp-10-6917-2010, 2010.

Song, H., Zhang, Z., Ma, P.-L., Ghan, S., and Wang, M.: The importance of considering sub-grid
cloud variability when using satellite observations to evaluate the cloud and precipitation
simulations in climate models, Geosci. Model Dev., 11, 3147–3158, https://doi.org/10.5194/gmd11-3147-2018, 2018.

802

Spinei, E., Whitehill, A., Fried, A., Tiefengraber, M., Knepp, T. N., Herndon, S., Herman, J. R.,
Müller, M., Abuhassan, N., Cede, A., Richter, D., Walega, J., Crawford, J., Szykman, J., Valin, L.,
Williams, D. J., Long, R., Swap, R. J., Lee, Y., Nowak, N., and Poche, B.: The first evaluation of
formaldehyde column observations by improved Pandora spectrometers during the KORUS-AQ
field study, Atmos. Meas. Tech., 11, 4943–4961, https://doi.org/10.5194/amt-11-4943-2018, 2018.

Souri, A. H., Nowlan, C. R., Wolfe, G. M., Lamsal, L. N., Chan Miller, C. E., Abad, G. G., Janz,
S. J., Fried, A., Blake, D. R., Weinheimer, A. J., Diskin, G. S., Liu, X., and Chance, K.: Revisiting
the effectiveness of HCHO/NO2 ratios for inferring ozone sensitivity to its precursors using high
resolution airborne remote sensing observations in a high ozone episode during the KORUS-AQ
campaign, Atmos. Environ., 224, 117341, https://doi.org/10.1016/j.atmosenv.2020.117341, 2020.

Tack, F., Merlaud, A., Iordache, M.-D., Pinardi, G., Dimitropoulou, E., Eskes, H., Bomans, B.,
Veefkind, P., and Van Roozendael, M.: Assessment of the TROPOMI tropospheric NO2 product
based on airborne APEX observations, Atmos. Meas. Tech., 14, 615–646,

816 https://doi.org/10.5194/amt-14-615-2021, 2021.

Tang, W., Arellano, A. F., DiGangi, J. P., Choi, Y., Diskin, G. S., Agustí-Panareda, A., Parrington,
M., Massart, S., Gaubert, B., Lee, Y., Kim, D., Jung, J., Hong, J., Hong, J.-W., Kanaya, Y., Lee,
M., Stauffer, R. M., Thompson, A. M., Flynn, J. H., and Woo, J.-H.: Evaluating high-resolution
forecasts of atmospheric CO and CO2 from a global prediction system during KORUS-AQ field
campaign, Atmos. Chem. Phys., 18, 11007–11030, https://doi.org/10.5194/acp-18-11007-2018,
2018.

823 Tang, W., Emmons, L. K., Arellano Jr., A. F., Gaubert, B., Knote, C., Tilmes, S., Buchholz, R. R., 824 Pfister, G. G., Diskin, G. S., Blake, D. R., Blake, N. J., Meinardi, S., DiGangi, J. P., Choi, Y., Woo, J.-H., He, C., Schroeder, J. R., Suh, I., Lee, H.-J., Jo, H.-Y., Kanaya, Y., Jung, J., Lee, Y., and Kim, 825 826 D.: Source contributions to carbon monoxide concentrations during KORUS-AQ based on CAM-827 chem applications. Res.-Atmos., 124, model J. Geophys. 1-27,https://doi.org/10.1029/2018jd029151, 2019. 828

- 829 Tang, W., Worden, H. M., Deeter, M. N., Edwards, D. P., Emmons, L. K., Martínez-Alonso, S., 830 Gaubert, B., Buchholz, R. R., Diskin, G. S., Dickerson, R. R., Ren, X., He, H., and Kondo, Y.: 831 Assessing Measurements of Pollution in the Troposphere (MOPITT) carbon monoxide retrievals 832 urban versus non-urban regions. Atmos. Meas. Tech., 13. over 1337-1356, 833 https://doi.org/10.5194/amt-13-1337-2020, 2020.
- Valin, L. C., Russell, A. R., Hudman, R. C., and Cohen, R. C.: Effects of model resolution on the
 interpretation of satellite NO2 observations, Atmos. Chem. Phys., 11, 11647–11655,
 https://doi.org/10.5194/acp-11-11647-2011, 2011.
- Veefkind, J. P., Aben, I., McMullan, K., Förster, H., de Vries, J., Otter, G., Claas, J., Eskes, H. J.,
 de Haan, J. F., Kleipool, Q., van Weele, M., Hasekamp, O., Hoogeveen, R., Landgraf, J., Snel, R.,
 Tol, P., Ingmann, P., Voors, R., Kruizinga, B., Vink, R., Visser, H., and Levelt, P. F.: TROPOMI
 on the ESA Sentinel-5 Precursor: A GMES mission for global observations of the atmospheric
 composition for climate, air quality and ozone layer applications, Remote Sens. Environ., 120, 70–
 83, doi:10.1016/j.rse.2011.09.027, 2012.
- van Geffen, J., Boersma, K. F., Eskes, H., Sneep, M., ter Linden, M., Zara, M., and Veefkind, J.
 P.: S5P TROPOMI NO2 slant column retrieval: method, stability, uncertainties and comparisons
 with OMI, Atmos. Meas. Tech., 13, 1315–1335, https://doi.org/10.5194/amt-13-1315-2020, 2020.
- Wiedinmyer, C., Akagi, S. K., Yokelson, R. J., Emmons, L. K., Al-Saadi, J. A., Orlando, J. J., and
 Soja, A. J.: The Fire INventory from NCAR (FINN): a high resolution global model to estimate
 the emissions from open burning, Geosci. Model Dev., 4, 625–641, https://doi.org/10.5194/gmd4-625-2011, 2011.
- Zhang, Z., Song, H., Ma, P.-L., Larson, V. E., Wang, M., Dong, X., and Wang, J.: Subgrid
 variations of the cloud water and droplet number concentration over the tropical ocean: satellite
 observations and implications for warm rain simulations in climate models, Atmos. Chem. Phys.,
 19, 1077–1096, https://doi.org/10.5194/acp-19-1077-2019, 2019.
- Zoogman, P., Liu, X., Suleiman, R., Pennington, W., Flittner, D., Al-Saadi, J., Hilton, B., Nicks,
 D., Newchurch, M., Carr, J., Janz, S., Andraschko, M., Arola, A., Baker, B., Canova, B., Miller,
 C. C., Cohen, R., Davis, J., Dussault, M., Edwards, D., Fishman, J., Ghulam, A., Abad, G. G.,

Grutter, M., Herman, J., Houck, J., Jacob, D., Joiner, J., Kerridge, B., Kim, J., Krotkov, N., Lamsal, 857 858 L., Li, C., Lindfors, A., Martin, R., McElroy, C., McLinden, C., Natraj, V., Neil, D., Nowlan, C., O'Sullivan, E., Palmer, P., Pierce, R., Pippin, M., Saiz-Lopez, A., Spurr, R., Szykman, J., Torres, 859 O., Veefkind, J., Veihelmann, B., Wang, H., Wang, J., and Chance, K.: Tropospheric emissions: 860 Ra., Monitoring pollution (TEMPO), J. Quant. Spectrosc. 861 of 186, 17-39, https://doi.org/10.1016/j.jqsrt.2016.05.008, 2017. 862 863



Figure 1. Domain of the study over South Korea and the land cover. Boxes indicate location of the SMA (upper left) and the Busan region (lower right) domains. The bold polygons in the two boxes represents political boundaries (upper left) of Seoul and Busan (lower right). Land cover data are from MODIS Terra and Aqua MCD12C1 L3 product, version V006, annual mean at 0.05° resolution; Friedl et al., 2015.



Figure 2. GeoTASO data of tropospheric NO₂ vertical column (molecules cm⁻²) measured during KORUS-AQ over the Seoul region. Each panel shows a separate raster. Panel titles show month, day, AM/PM, raster number on that date, and mean time of raster acquisition. There were nine flights sampling rasters over Seoul. The May 01 AM, May 17 AM, May 17 PM, May 28 PM, June 01 PM, and June 02 AM flights each sampled one raster. The June 05 AM, June 09 AM, and June 09 PM flights each sampled two rasters. As a result, there were two flights and two rasters on May 17th, one flight and two rasters on June 5th, and two flights and four rasters on June 9th. The bold polygons in each panel represent political boundary of Seoul.







Figure 3. Demonstration of the hypothetical satellite pixel random sampling method. Each subplot 888 is an hour during May 17th PM flight. For each hour, we randomly sample 10000 hypothetical 889 satellite pixels at each different pixel sizes (i.e., 0.5 km×0.5 km, 0.75 km×0.75 km, 1 km×1 km, 2 890 km×2 km, ..., 25 km×25 km) over the GeoTASO data of tropospheric NO2 vertical column 891 (molecules cm⁻²) every hour. The sampled pixel size (from 0.5 km×0.5 km to 25 km×25 km) are 892 893 shown in the lower-left corner of each sub-plot. Only 100 samples for pixel size of 7 km×7 km (thick black box) and 100 samples for $18 \text{ km} \times 18 \text{ km}$ are shown for demonstration purposes. 894 895 Samples that fail to pass the 75% coverage threshold are not shown. Coastlines, Province/Metropolitan City boundaries are shown by gray solid lines. Main roads are shown by 896 blue dashed lines (data are from http://www.diva-gis.org/gdata). 897



Figure 4. Boxplot (with medians represented by red bars, interquartile ranges between 25th and 75th percentiles represented by blue boxes, and the most extreme data points not considered outliers represented by whiskers) for the normalized satellite sub-grid variability (SGV) over the Seoul Metropolitan Area (a), the Busan region (b), and Los Angeles Basin (c). Normalized satellite SGV is calculated as the standard deviation of the GeoTASO data within the sampled satellite pixel divided by the mean of the GeoTASO data within the sampled satellite pixel. The black lines represent the mean of the normalized satellite SGV at a given size. The resolutions of TEMPO, TROPOMI, GEMS, and OMI are highlighted by the yellow shade in the Figure.



Figure 5. Average of the normalized satellite sub-grid variability (SGV) sampled individually from the twelve rasters (represented by the colored lines), and sampled from all the twelve rasters

915 together (represented by the black line) over the Seoul Metropolitan Area during KORUS-AQ.

- 916 Normalized satellite SGV is calculated by the standard deviation of the GeoTASO data within the
- sampled satellite pixel divided by the mean of the GeoTASO data within the sampled satellite
- 918 pixel.
- 919



Figure 6. Temporal mean differences (TeMD) of hypothetical satellite pixels (molecules cm⁻²) over the Seoul Metropolitan Area as a function of time difference (Dt). Results for each pixel size are color-coded, with selected sizes shown with thicker lines for reference. See also text for details.





932 Figure 7. (a) Spatial Structure Function (SSF) for GeoTASO data of tropospheric NO₂ vertical column molecules cm⁻²) over the Seoul Metropolitan Area (SMA) during KORUS-AQ and (b) the zoom-in version of panel (a) for distance range of 1-25 km. The SSF calculates average of absolute value of NO_{2.VC} differences (i.e., mean difference; y-axis) across all data pairs (measured in the same hourly bin) that are separated by different distance (x-axis). The SSF based on GeoTASO data measured during morning flights are in solid colored lines while the SSF based on GeoTASO data measured during afternoon flights are in dashed colored lines. The SSF based on all the data is in the black solid line.



949

Figure 8. Boxplot of hypothetical satellite normalized SGV of NO₂ vertical column (VC), SO₂ VC, CO VC, and formaldehyde (HCHO) VC derived from the WRF-Chem simulation with a resolution of 3 km \times 3 km (colored lines), and GeoTASO NO₂ VC that gridded to the WRF-Chem grid (black lines) over the Seoul Metropolitan Area. Medians are represented by red bars, interquartile ranges between 25th and 75th percentiles by blue boxes, and the most extreme data points not considered outliers by whiskers. The modeled NO₂, CO, SO₂, and HCHO are filtered to match the rasters of GeoTASO measurements.