

1 Quantifying the impact of aerosol scattering on the 2 retrieval of methane from airborne remote sensing 3 measurements

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17
18 **Abstract.** As a greenhouse gas with strong global warming potential, atmospheric methane (CH₄)
19 emissions have attracted a great deal of attention. Although remote sensing measurements can provide
20 information about CH₄ sources and emissions, accurate retrieval is challenging due to the influence of
21 atmospheric aerosol scattering. In this study, imaging spectroscopic measurements from the Airborne
22 Visible/Infrared Imaging Spectrometer–Next Generation (AVIRIS-NG) in the short-wave infrared are
23 used to compare two retrieval techniques — the traditional Matched Filter (MF) method and the Optimal
24 Estimation (OE) method, which is a popular approach for trace gas retrievals. Using a numerically
25 efficient two-stream-exact-single-scattering radiative transfer model, we also simulate AVIRIS-NG
26 measurements for different scenarios and quantify the impact of aerosol scattering in the two retrieval
27 schemes by including aerosols in the simulations but not in the retrievals. The presence of aerosols causes
28 an underestimation of CH₄ in both the MF and OE retrievals; the biases increase with increasing surface
29 albedo and aerosol optical depth (AOD). Aerosol types with high single scattering albedo and low
30 asymmetry parameter (such as water soluble aerosols) induce large biases in the retrieval. When
31 scattering effects are neglected, the MF method exhibits lower fractional retrieval bias compared to the
32 OE method at high CH₄ concentrations (2–5 times typical background values), and is suitable for
33 detecting strong CH₄ emissions. For an AOD value of 0.3, the fractional biases of the MF retrievals are

34 between 1.3 and 4.5%, while the corresponding values for OE retrievals are in the 2.8–5.6% range. On
35 the other hand, the OE method is an optimal technique for diffuse sources (<1.5 times typical background
36 values), showing up to five times smaller fractional retrieval bias (8.6%) than the MF method (42.6%)
37 for the same AOD scenario. However, when aerosol scattering is significant, the OE method is superior
38 since it provides a means to reduce biases by simultaneously retrieving AOD, surface albedo and CH₄.
39 The results indicate that, while the MF method is good for plume detection, the OE method should be
40 employed to quantify CH₄ concentrations, especially in the presence of aerosol scattering.

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44 **1 Introduction**

45 Atmospheric methane (CH₄) is about 85 times more potent per unit mass at warming the Earth than
46 carbon dioxide (CO₂) on a 20-year timescale (IPCC, 2013), implying that reduction in CH₄ emissions
47 could be very efficient to slow down global warming in the near term. Global mean CH₄ concentrations
48 have increased from ~700 ppb in the preindustrial era to more than 1860 ppb as of 2019 (NOAA, 2019).
49 The most effective sink of atmospheric CH₄ is the hydroxyl radical (OH) in the troposphere. CH₄ reacts
50 with OH to reduce the oxidizing capacity of the atmosphere and generate tropospheric ozone. Increasing
51 emissions of CH₄ reduce the concentration of OH in the atmosphere. With less OH to react with, the
52 lifespan of CH₄ could also increase, resulting in greater CH₄ concentrations (Holmes et al., 2013). Soils
53 also act as a major sink for atmospheric methane through the methanotrophic bacteria that reside within
54 them.

55 Significant natural CH₄ sources include wetlands (Bubier et al., 1994, Macdonald et al., 1998;
56 Gedney et al., 2004), geological seeps (Kvenvolden and Rogers, 2005; Etiope et al., 2009), ruminant
57 animals, and termites. In addition, increased surface and ocean temperatures associated with global
58 warming may increase CH₄ emissions from melting permafrost (Woodwell et al., 1998; Walter et al.,
59 2006; Schaefer et al., 2014, Schuur et al., 2015) and methane hydrate destabilization (Kvenvolden, 1988;
60 Archer, 2007). Human activity also contributes significantly to the total CH₄ emissions. Rice agriculture
61 is one of the most important anthropogenic sources of CH₄ (Herrero et al., 2016; Schaefer et al., 2016).
62 Other sources include landfills (Themelis and Ulloa, 2007), wastewater treatment, biomass burning, and
63 methane slip from gas engines. Global fugitive CH₄ emissions from coal mining (Kort et al., 2014),
64 natural gas and oil systems (Alvarez et al., 2018), hydraulic fracturing (“fracking”) of shale gas wells
65 (Howarth et al., 2011; Howarth, 2015, 2019), and residential and commercial natural gas distribution
66 sectors (He et al., 2019) are also of increasing concern. Although the sources and sinks of methane are
67 reasonably well known, there are large uncertainties in their relative amounts and in the partitioning
68 between natural and anthropogenic contributions (Nisbet et al., 2014, 2016). This uncertainty is
69 exemplified by the CH₄ “hiatus”, which refers to the observed stabilization of atmospheric CH₄
70 concentrations from 1999–2006, and the renewed rise thereafter (Kirschke et al., 2013).

71 Satellite monitoring of CH₄ can be broadly divided into three categories: solar backscatter, thermal
72 emission and lidar (Jacob et al., 2016). The first solar backscattering mission was SCIAMACHY
73 (Frankenberg et al., 2006), which was operational from 2003–2012 and observed the entire planet once
74 every seven days. It was followed by GOSAT in 2009 (Kuze et al., 2016), and subsequently the next
75 generation GOSAT-2 in 2018 (Glumb et al., 2014). In between, the TROPOMI mission was also
76 launched in 2017, which observes the planet once daily with a high spatial resolution of 7×7 km² (Butz

77 et al., 2012; Veeffkind et al., 2012). CarbonSat (Buchwitz et al., 2013) is another proposed mission to
78 measure CH₄ globally from solar backscatter with a very fine spatial resolution (2×2 km²) and high
79 precision (0.4%). GHGSat-D (McKeever et al., 2017; Varon et al., 2019; Jervis et al., 2020) measures
80 between 1630–1675 nm, with an effective pixel resolution of 50×50 m² over targeted 12×12 km² scenes,
81 and is intended to detect CH₄ emissions from individual industrial sites. In contrast, MethaneSAT (Wofsy
82 and Hamburg, 2019) has a pixel size of 1–2 km² and a wide field of view (200 km²) and can quantify
83 diffuse CH₄ emission sources over large areas. Thermal infrared observations of CH₄ are available from
84 the IMG (Clerbaux et al., 2003), AIRS (Xiong et al., 2008), TES (Worden et al., 2012), IASI (Xiong et
85 al., 2013), and CrIS (Gambacorta et al., 2016) instruments. These instruments provide day/night
86 measurements at spatial resolutions ranging from 5×8 km² (TES) to 45×45 km² (AIRS). GEO-CAPE
87 (Fishman et al., 2012), GeoFTS (Xi et al., 2015), G3E (Butz et al., 2015), and GeoCarb (Polonsky et al.,
88 2014) are proposed geostationary instruments (GeoCarb was selected by NASA under the Earth Venture
89 - Mission program), which when operational will have resolutions of 2–5 km over regional scales. The
90 MERLIN lidar instrument (Kiemle et al., 2014) scheduled for launch in 2021 will measure CH₄ by
91 employing a differential absorption lidar.

92 By combining a large number of footprints and high spatial resolution, airborne imaging
93 spectrometers are also well suited for mapping local CH₄ plumes. The Airborne Visible/Infrared Imaging
94 Spectrometer–Next Generation (AVIRIS-NG) measures reflected solar radiance across more than 400
95 channels between 380 and 2500 nm (Green et al., 1998; Thompson et al., 2015). Strong CH₄ absorption
96 features present between 2100 and 2500 nm can be observed at a spectral resolution of 5 nm full width
97 at half maximum (FWHM). A number of approaches have been developed to retrieve CH₄ from such
98 hyperspectral data. Roberts et al. (2010) used a spectral residual approach between 2000 and 2500 nm
99 and Bradley et al. (2011) employed a band ratio technique using the 2298 nm CH₄ absorption band and
100 2058 nm CO₂ absorption band. However, these techniques are not suited for terrestrial locations that
101 have lower albedos and have spectral structure in the SWIR. A cluster-tuned matched filter technique
102 was demonstrated to be capable of mapping CH₄ plumes from marine and terrestrial sources (Thorpe et
103 al., 2013) as well as CO₂ from power plants (Dennison et al., 2013); however, this method does not
104 directly quantify gas concentrations. Frankenberg et al. (2005) developed an iterative maximum *a*
105 *posteriori* differential optical absorption spectroscopy (IMAP-DOAS) algorithm that allows for
106 uncertainty estimation. Thorpe et al. (2014) adapted the IMAP-DOAS algorithm for gas detection in
107 AVIRIS imagery. In addition, they developed a hybrid approach using singular value decomposition and
108 IMAP-DOAS as a complementary method of quantifying gas concentrations within complex AVIRIS
109 scenes.

110 Accurate assessment of CH₄ emissions is particularly challenging in the presence of aerosols
 111 because the latter introduce uncertainties in the light path if not accounted for. In fact, CH₄ emissions are
 112 frequently correlated with pollution due to concurrent aerosol emissions. For large aerosols (such as dust),
 113 the low Ångström exponent values result in high aerosol optical depth (AOD) values even in the
 114 wavelength range from 2000 nm to 2500 nm (Seinfeld and Pandis, 2006; Zhang et al., 2015). Therefore,
 115 it is important to obtain a clear understanding of aerosol impacts on CH₄ retrievals. In this study, SWIR
 116 AVIRIS-NG measurements are used to analyze the impact of aerosol scattering on CH₄ retrievals.
 117 Further, using an accurate but numerically efficient radiative transfer (RT) model (Spurr and Natraj,
 118 2011), we simulate AVIRIS-NG measurements with varying aerosol amounts and quantify the impact of
 119 aerosol scattering using two retrieval techniques, the traditional matched filter (MF) method and the
 120 optimal estimation (OE) method that is widely used in trace gas remote sensing. This article is organized
 121 as follows. The MF and OE retrieval methods are described in Section 2. Section 3 focuses on analysis
 122 of a sample CH₄ plume detected by AVIRIS-NG measurements and compares retrievals using the MF
 123 and OE methods. Section 4 presents a detailed evaluation of aerosol impacts on the two retrieval methods
 124 through simulations of AVIRIS-NG spectra for different geophysical parameters. Section 5 provides a
 125 summary of the work and discusses future research.

126

127 **2 Methods**

128 **2.1 MF method**

129 Real-time remote detection using AVIRIS-NG measurements are traditionally based on the MF
 130 method (Frankenberg et al., 2016). In this method, the background spectra are assumed to be distributed
 131 as a multivariate Gaussian \mathcal{N} with covariance matrix Σ and background mean radiance μ . If H_0 is a
 132 scenario without CH₄ enhancement and H_1 is one with CH₄ enhancement, the MF approach is equivalent
 133 to a hypothesis test between the two scenarios:

$$134 \quad H_0: L_m \sim \mathcal{N}(\mu, \Sigma) \quad (1)$$

$$135 \quad H_1: L_m \sim \mathcal{N}(\mu + t\alpha, \Sigma) \quad (2)$$

136 where L_m is the measurement radiance; t is the target signature, which is defined in Equation (4); α is the
 137 enhancement value, denoting a scaling factor for the target signature that perturbs the background μ . If
 138 x is a vector of measurement spectra with one element per wavelength, $\alpha(x)$ can be written, based on
 139 maximum likelihood estimates (Manolakis et al., 2014), as follows:

$$140 \quad \alpha(x) = \frac{(x - \mu)^T \Sigma^{-1} t}{t^T \Sigma^{-1} t} \quad (3)$$

141 We utilize the same definitions as in Frankenberg et al. (2016). Specifically, the enhancement value $\alpha(\mathbf{x})$
 142 denotes the thickness and concentration within a volume of equivalent absorption, and has units of ppm
 143 \times m. The target signature \mathbf{t} refers to the derivative of the change in measured radiance with respect to a
 144 change in absorption path length due to an optically thin absorbing layer of CH₄. Note that this definition
 145 has the disadvantage that the accuracy of the result degrades when the absorption is strong and further
 146 attenuation becomes nonlinear. At a particular wavelength λ , \mathbf{t} can be expressed as:

$$147 \quad \mathbf{t}(\lambda) = -\kappa(\lambda)\boldsymbol{\mu}(\lambda), \quad (4)$$

148 where κ is the absorption coefficient for a near-surface plume with units of ppm⁻¹ m⁻¹. This is
 149 different from the units of m² · mol⁻¹ traditionally used for the absorption coefficient κ_{trad} in trace
 150 gas remote sensing. Using the ideal gas law to express the volume V (in liters) occupied by one mole of
 151 CH₄ at the temperature and pressure corresponding to the plume altitude ($V = 22.4$ at standard
 152 temperature and pressure), and the relations 1 liter = 10⁻³ m³ and 1 ppm = 10⁻⁶, we obtain the
 153 following expression for unit conversion (units in parentheses):

$$154 \quad \kappa_{trad} [\text{m}^2 \cdot \text{mol}^{-1}] = \kappa [\text{ppm}^{-1} \text{m}^{-1}] \times V [\text{liter mol}^{-1}] \times 10^{-3} [\text{m}^3 \text{liter}^{-1}] / 10^{-6} [\text{ppm}^{-1}] (5)$$

155 Figure 1 shows the target signature, which is calculated based on HITRAN absorption cross-sections
 156 (Rothman et al., 2009). The background mean radiance $\boldsymbol{\mu}$ used in Equation 4 is based on the AVIRIS-
 157 NG measurement shown in Figure 2; this is described in more detail in Section 3.

158 2.2 OE method

159 The OE method is widely used for the remote sensing retrieval of satellite measurements, such as
 160 from the Orbiting Carbon Observatory-2 (OCO-2; O'Dell et al., 2018), the Spinning Enhanced Visible
 161 and Infra-Red Imager (SEVIRI; Merchant et al., 2013), and the Greenhouse Gases Observing Satellite
 162 (GOSAT; Yoshida et al., 2013). It combines an explicit (typically nonlinear) forward model of the
 163 atmospheric state, a (typically Gaussian) prior probability distribution for the variabilities and a (typically
 164 Gaussian) distribution for the spectral measurement errors. In addition, the Bayesian framework used by
 165 the OE approach allows new information (from measurements) to be combined with existing information
 166 (e.g., from models). In many applications, the forward model is nonlinear, and obtaining the optimal
 167 solution requires iterative techniques such as the Levenberg–Marquardt method (Rodgers, 2000), which
 168 has been routinely applied to study the impacts of measurement parameters on the retrieval process (see,
 169 e.g., Zhang et al., 2015). The iteration in this algorithm follows the below procedure.

$$170 \quad \mathbf{x}_{i+1} = \mathbf{x}_i + [(1 + \gamma)\mathbf{S}_a^{-1} + \mathbf{K}_i^T \mathbf{S}_\epsilon^{-1} \mathbf{K}_i]^{-1} \{ \mathbf{K}_i^T \mathbf{S}_\epsilon^{-1} [\mathbf{y} - \mathbf{F}(\mathbf{x}_i)] - \mathbf{S}_a^{-1} [\mathbf{x}_i - \mathbf{x}_a] \} \quad (6)$$

171 where \mathbf{x} is a state vector of surface and atmospheric properties, \mathbf{S}_a is the *a priori* covariance matrix, \mathbf{S}_ϵ
 172 is the spectral radiance noise covariance matrix, \mathbf{K} is the Jacobian matrix, \mathbf{x}_a is the *a priori* state vector
 173 and γ is a parameter determining the size of each iteration step. The measured spectral radiance is denoted

174 as \mathbf{y} ; $\mathbf{F}(\mathbf{x})$ is the simulated radiance obtained from the forward model. For the retrieval of CH_4 from
175 AVIRIS-NG measurements, the state vector includes the total column amounts of CH_4 and H_2O , while
176 for the retrievals from synthetic spectra, the H_2O concentration is fixed and the state vector only includes
177 the CH_4 total column. The *a priori* values are within 10% of the true values; *a priori* errors are assumed
178 to be 20% for all state vector elements. The retrieved results are shown as the column averaged mixing
179 ratio (X_{CH_4} , in ppm). Aerosols are not included in the state vector for both the real and synthetic
180 retrievals. They are, however, considered in the forward model for the synthetic simulations. Table 1
181 (WCRP, 1986) lists optical properties for four basic aerosol types (dust, water soluble, oceanic and soot).
182 Table 2 (WCRP, 1986) shows the corresponding properties for three aerosol models that are defined as
183 mixtures of the basic components from Table 1. We employ the Henyey-Greenstein phase function
184 (Henyey and Greenstein, 1941), where aerosol composition is determined by two parameters: single
185 scattering albedo (SSA) and asymmetry parameter (g). The surface albedo is also not retrieved; for both
186 real and synthetic retrievals, it is held fixed and assumed to be independent of wavelength.

187

188 **3 Detection and retrieval of CH_4 from AVIRIS-NG measurements**

189 To illustrate the OE retrieval and its difference from the MF method, we perform retrievals for an
190 AVIRIS-NG measurement made on 4 September 2014 (ang20140904t204546) in Bakersfield, CA, as
191 shown in Figure 2. The location is to the west of the Kern Front Oil field. This detection is a case study
192 from the NASA/ESA CO_2 and MEthane eXperiment (COMEX) campaign in California during June and
193 August/September 2014, which includes airborne *in situ*, airborne non-imaging remote sensing, and
194 ground-based *in situ* instruments to provide a real-time remote detection and measurement for CH_4
195 plumes released from anthropogenic sources. An RGB image of flight data is displayed in Figure 2a; the
196 emission source is a pump jack, as described in Thompson et al. (2015). Figure 2b presents results from
197 the MF method, which shows that the CH_4 plume disperses downwind and has a maximum enhancement
198 value of about $2800 \text{ ppm} \times \text{m}$. Some artifacts caused by surfaces with strong absorption in the 2100–
199 2500 nm wavelength range, such as oil-based paints or roofs with calcite as a component (Thorpe et al.,
200 2013), also produce large α values in the MF method; these can be removed by an optimization method
201 such as the column-wise MF technique (Thompson et al., 2015).

202 Figure 3 displays the measured radiance (a) before normalization and (b) after normalization,
203 corresponding to two detector elements (in plume and out of plume). Every element is a cross-track
204 spatial location. The normalization is done by calculating the ratio of the radiance to the maximum value
205 across the spectral range, such that the values fall between 0 and 1. This is a first order correction for the
206 effects of surface albedo. Comparing the measured spectrum in plume to that out of plume, there is

207 obvious enhancement of CH₄ that is particularly evident in the normalized radiance. CH₄ is the main
208 absorber in the 2100–2500 nm wavelength range, and H₂O is the major interfering gas. Figure 3b
209 indicates the absorption peaks due to H₂O and CH₄.

210 We choose the plume center with 500 elements to illustrate results obtained using the MF and OE
211 methods. The former evaluates the CH₄ α value compared to the background CH₄ concentration, while
212 the latter retrieves XCH₄. In the MF method, the background covariance matrix Σ and mean radiance
213 μ are drawn from a reference region close to the CH₄ emission source. These are shown in Figure 2,
214 where the dashed green box denotes the reference region and the source is located within the solid red
215 box. In the OE method, results are shown as a multiplicative scaling factor compared to a typical XCH₄
216 background of 1.822 ppm. This value is the globally averaged marine surface annual mean for 2014 (Ed
217 Dlugokencky, NOAA/GML, www.esrl.noaa.gov/gmd/ccgg/trends_ch4/), the year corresponding to the
218 AVIRIS-NG measurement being studied. We use an accurate and numerically efficient two-stream-
219 exact-single-scattering (2S-ESS) RT model (Spurr and Natraj, 2011). This forward model is different
220 from a typical two-stream model in that the two-stream approximation is used only to calculate the
221 contribution of multiple scattering to the radiation field. Single scattering is treated in a numerically exact
222 manner using all moments of the phase function. This model has been used for remote sensing of
223 greenhouse gases and aerosols (Xi et al., 2015; Zhang et al., 2015, 2016; Zeng et al., 2017, 2018).
224 Aerosols are neither included in the forward model nor retrieved in this analysis. The surface albedo is
225 set to a wavelength-independent value of 0.5.

226 Results from the two retrieval methods reveal a similar CH₄ plume shape (Figure 4), especially for
227 elements with high CH₄ enhancement. However, larger differences in CH₄ concentrations are evident in
228 the OE retrievals (Figure 4b). Since radiance normalization reduces the impact of surface albedo and
229 aerosols are not included in either retrieval, this might be due to the fact that, in the OE method, H₂O and
230 CH₄ are simultaneously retrieved; the CH₄ retrieval has added uncertainty due to overlapping absorption
231 features between these two gases. The large maximum value of about 3000 in the MF method also
232 contributes to a reduction in relative contrast. While these results provide heuristic information about the
233 relative performance of the two retrieval techniques, it is difficult to compare the CH₄ enhancement
234 directly between the two methods since the background CH₄ concentration used in the MF method cannot
235 be quantified exactly. Further, evaluating retrieval biases due to ignoring aerosol scattering is not trivial
236 when real measurements are used. Therefore, we simulate synthetic spectra (see section 4) using the 2S-
237 ESS RT model to study the impacts of aerosol scattering as a function of different geophysical parameters
238 by varying them in a systematic manner.

239

240 **4 Aerosol impact analysis**

241 **4.1 Synthetic spectra**

242 In a real AVIRIS-NG observation, the exact column concentration of CH₄ cannot be controlled.
243 However, synthetic simulations allow us to manipulate parameters such as CH₄ concentration, surface
244 albedo, AOD, g , and SSA, and thereby test aerosol impacts on CH₄ retrievals. The 2S-ESS RT model is
245 used to simulate the AVIRIS-NG spectral radiance. In this model, a prior atmospheric profile with 70
246 layers from the surface up to 70 km is derived from National Center for Environmental Prediction
247 reanalysis data (Kalnay et al., 1996); absorption coefficients for all relevant gases are obtained from the
248 HITRAN database (Rothman et al., 2009). Monochromatic RT calculations are performed at a spectral
249 resolution of 0.5 cm⁻¹; the radiance spectrum is then convolved using a Gaussian instrument line shape
250 function with a wavelength-dependent full width at half maximum (FWHM) from a calibrated AVIRIS-
251 NG data file. The signal to noise ratio (SNR) is set to be 300, with Gaussian white noise added. This
252 procedure results in a wavelength grid with a resolution of about 5 nm. The spectral wavelength range
253 used to retrieve CH₄ is from 2100 nm to 2500 nm.

254 The additional atmospheric and geometric variables included in the model are listed in Table 3,
255 which are held constant unless otherwise mentioned. The observation geometry parameters are taken
256 from a real AVIRIS-NG measurement. Recent AVIRIS-NG flight campaigns have sensor heights ranging
257 from 0.43 to 3.8 km; we choose a value of 1 km, the same as the highest level where aerosol is present
258 in our simulations. The influence of AOD on CH₄ retrieval as a function of SSA and g is analyzed in
259 Section 4.3; in all other cases, SSA and g are held constant at 0.95 and 0.75, respectively, which is
260 representative of aerosols in the Los Angeles region (Zhang et al., 2015).

261 **4.2 Aerosol impact in the MF method**

262 We simulate synthetic spectra at different AOD, surface albedo and CH₄ concentration values, use
263 the MF method to obtain the CH₄ enhancement, and compare differences in α between scenarios without
264 and with aerosol. The covariance matrix and background mean radiance are calculated from a simulated
265 zero AOD background with surface albedos from 0.1 to 0.5, and XCH₄ set at the typical background
266 value of 1.822 ppm used in Section 3. Figure 5a shows the enhancement value as a function of XCH₄.
267 As the CH₄ concentration increases, the enhancement value obtained by the MF method at first increases
268 approximately linearly. However, the absorption changes in a nonlinear fashion with concentration,
269 whereas the MF method applies a linear formalism to the change. Therefore, the enhancement value
270 (which is correlated with the absorption signature) also shows a deviation from linear behavior at larger
271 XCH₄. Two aerosol scenarios (AOD = 0, 0.3) are compared in Figure 5a, which reveals that the effect of
272 aerosol loading is similar to an underestimation of CH₄ in the retrieval. The underestimation, which is

273 due to the shielding of CH₄ absorption below the aerosol layer and the fact that multiple scattering effects
274 between the aerosol and the surface are ignored, is clearly shown in Figure 5b, where the enhancement
275 value for fixed CH₄ concentration (same concentration as the background) decreases from 0 ppm × m
276 to -1532 ppm × m with increasing AOD. To clarify the impact of AOD at different surface albedo values,
277 zoomed in versions of α as a function of XCH₄ are presented in Figures 5c–f. For the AOD = 0 scenario,
278 the results are independent of surface albedo. This is because there are no multiple scattering effects
279 between the surface and the atmosphere (Rayleigh scattering is negligible in the retrieval wavelength
280 range) when there is no aerosol loading. For the scenarios with aerosol loading, the dispersion in the
281 zero-enhancement XCH₄ value between different surface albedos indicates that results from the MF
282 method are biased more at large AOD and surface albedo values (Figures 5d–f). This is a consequence
283 of increased multiple scattering between the aerosol layer and the surface that is not accounted for by the
284 retrieval algorithm. The maximum bias value is close to -700 ppm × m (equivalent to -0.06×1.822
285 ppm relative to the background concentration of 1.0×1.822 ppm) for an AOD of 0.3 and surface albedo
286 of 0.5 (Figure 5f). The implication of these results is that accurate knowledge of the surface albedo is
287 important for MF retrievals, especially when the aerosol loading is large.

288 A quantitative analysis of underestimation of CH₄ concentration due to aerosol scattering is
289 presented in Figure 6. The color bar shows the α bias — which is defined as the difference between the
290 enhancement value without aerosol (true α value) and that with aerosol — for different CH₄
291 concentrations, surface albedos and AODs. A positive bias means that CH₄ is underestimated. The α bias
292 increases with increasing surface albedo and AOD, reaching a maximum value of about 700 ppm × m
293 for the simulated cases. However, it is interesting that the bias decreases with increasing CH₄
294 concentration, which is different from the results obtained by the OE method (discussed in Section 4.3).
295 This surprising behavior is a direct consequence of the physical basis of the MF method. The rate of
296 increase in enhancement becomes smaller as XCH₄ becomes larger (Figure 5a). Therefore, at higher
297 XCH₄ values, the addition of aerosols (which has a similar effect as a reduction in XCH₄) results in a
298 lower reduction in enhancement compared to that at lower XCH₄ values, resulting in a net decrease in
299 the enhancement bias.

300 4.3 Aerosol impact in the OE method

301 For the simulation of the synthetic spectra, we assume nonzero aerosol loading below 1 km elevation.
302 The OE method is then used to perform retrievals using the same configuration (including, in particular,
303 the same surface albedo) except that AOD is set to zero. This approach is similar to neglecting aerosol
304 scattering in the CH₄ retrieval; the retrieval bias is defined as the difference between the true XCH₄ in
305 the simulation and the retrieved value (positive values refer to underestimation). First, we study the

306 retrieval bias caused by different aerosol types and mixtures. Figure 7a shows CH₄ retrieval biases as a
307 function of SSA and g ; surface albedo and AOD are kept constant at 0.3 and XCH₄ is assumed to be 1.0
308 \times 1.822 ppm. The retrieval bias increases with SSA and decreases with g , with a maximum bias ratio
309 (ratio of retrieval bias to the true value) of about 20%. This behavior can be explained as follows. At
310 higher SSA values, there are more multiple scattering effects (that are ignored in the retrieval). On the
311 other hand, larger values of g imply greater anisotropy of scattering (preference for forward scattering),
312 leading to a reduction in multiple scattering effects. Since the retrieval bias is large for high SSA and
313 low g , the water-soluble aerosol type (Table 1) and the maritime aerosol model (Table 2) can be expected
314 to induce greater biases in the retrieval. In order to compare the impacts of SSA and g in further detail,
315 retrieval results due to a \pm 5% change in SSA and g for the three aerosol models from Table 2 are shown
316 in Figures 7b and 7c. Note that for the maritime aerosol model, the SSA is set to 0.999 for the +5%
317 scenario to ensure physicality. It is clear that (1) the maritime aerosol model induces larger retrieval
318 biases than the other aerosol types, and (2) the retrieval results are more sensitive to changes in g than
319 those in SSA.

320 We then simulate synthetic spectra for different values of CH₄ concentration, surface albedo and
321 AOD. The impacts of aerosol scattering on the retrievals for these scenarios are demonstrated in Figure
322 8. Figure 8a shows a 5 \times 5 panel of boxes. Within each box, XCH₄ is constant, while surface albedo
323 increases from top to bottom and AOD increases from left to right. The variation of XCH₄ across the
324 boxes is shown in Figure 8b. We also show a zoomed in plot of the bottom right box (XCH₄ = 5.8 \times
325 1.822 ppm) in Figure 8c, which illustrates the AOD and surface albedo changes within a box. These
326 changes are identical for all boxes. Figure 8a indicates that OE retrievals produce larger CH₄ biases at
327 higher XCH₄ values, in contrast with MF results. In addition, it is evident that the retrieved CH₄ bias
328 increases with increasing AOD. The CH₄ bias induced by differences in the surface albedo is not as large
329 as that due to AOD variations, but surface albedo effects are noticeable at large AOD. Figure 8d shows
330 the sensitivity of retrieval biases to changes in AOD and surface albedo, again demonstrating the greater
331 impact of AOD than surface albedo in the retrieval.

332 The effects of changing the *a priori*, *a priori* error and RT simulation spectral resolution on the
333 retrieved XCH₄ are shown in Figure 9. For these calculations, the other parameters are set as follows:
334 SSA = 0.95, g = 0.75, AOD = 1.0, surface albedo = 0.5 and true XCH₄ = 5.8 \times 1.822ppm. The parameters
335 were chosen to correspond to the scenario with the largest retrieval bias in Figure 8c (bottom right box
336 in Figure 8c). Figure 9a shows that the retrieved XCH₄ changes by about 9 ppb as the *a priori* changes
337 from half to twice the true XCH₄ value. Similarly, the XCH₄ difference is less than 4 ppb when the *a*
338 *priori* error changes from 0.05 to 0.5 (Figure 9b). Compared to the bias of about 923 ppb induced by

339 neglecting aerosol scattering for this scenario, it is clear that the impacts of the *a priori* and *a priori* error
340 are very small. The effect of spectral resolution is larger, but XCH₄ still changes by only about 100 ppb
341 when the spectral resolution is changed from 0.5 to 0.1 cm⁻¹ (Figure 9c).

342 4.4 Comparison of the two retrieval techniques

343 Figure 10 presents the bias ratios for the two retrieval techniques at different AODs (surface albedo
344 = 0.3). In the MF method, the bias ratio is defined as the ratio of the bias to the true value of α . On the
345 other hand, in the OE method, it is the ratio of the bias to the true XCH₄. From Figure 10 it is clear that
346 the bias ratio decreases with increasing CH₄ concentration and has higher values at larger AODs. The
347 bias ratio for the MF method (1.3–4.5%) is up to 53.6% less than that for the OE method (2.8–5.6%) for
348 AOD = 0.3 when the CH₄ concentration is high (2–5 times typical background values). On the other
349 hand, the OE method performs better when enhancements are small and XCH₄ is close to the background
350 value. For example, the bias ratio for the MF method has a high value of about 42.6% at AOD = 0.3 for
351 a 10% enhancement (XCH₄ = 1.1 × 1.822 ppm); the OE value for the same scenario is 8.6%. For scenarios
352 where scattering is ignored, the two retrieval techniques seem to be complementary, with differing
353 utilities for different enhancements. On the other hand, when RT models that account for scattering
354 effects are employed, the MF technique is suboptimal. Further, MF retrievals rely on accurate
355 characterization of the surface albedo, especially when the aerosol loading is large. Finally, the MF
356 method does not retrieve concentrations, which are necessary to infer fluxes. Therefore, the OE technique
357 is in general superior due to its ability to support simultaneous retrieval of aerosols, surface albedo and
358 CH₄ concentration.

359

360 5 Summary and discussion

361 Remote sensing measurements from airborne and satellite instruments are widely used to detect
362 CH₄ emissions. In our study, the traditional MF and the OE methods are used to quantify the effects of
363 aerosol scattering on CH₄ retrievals based on simulations of AVIRIS-NG measurements. The results
364 show that the retrieval biases increase with increasing AOD and surface albedo for both techniques. In
365 the OE method the biases increase with increasing CH₄ concentration and SSA, but decrease with
366 increasing aerosol asymmetry parameter. The CH₄ retrieval bias increases with increasing XCH₄ in the
367 OE method but decreases for the same scenario in the MF method. The surprising MF trend is attributed
368 to the inability of the MF method to treat nonlinear absorption effects at high XCH₄ values. We also
369 present bias ratios for the two techniques. The MF method shows smaller bias ratios at large CH₄
370 concentrations than the OE method; it is, therefore, the optimal method to detect strong CH₄ emission
371 sources when scattering effects can be ignored in the retrieval. For the same retrieval scenario, the OE

372 method seems to be more suitable for detecting diffuse sources. Further, the MF method relies on a
373 comparison with the background CH₄ concentration. It is difficult to get an accurate estimate of the
374 background XCH₄ value in polluted atmospheric environments. In contrast, the OE method provides
375 retrievals based solely on the atmospheric scenario of interest; CH₄, aerosols and surface albedo can be
376 simultaneously inferred. Therefore, when scattering effects need to be considered, the OE method is the
377 appropriate choice. Indeed, the MF method was intended for plume detection. OE enables accurate
378 quantification of XCH₄ in the presence of aerosol scattering.

379 This study focused on a comparison of retrieval techniques. It is also important to accurately
380 represent the physics of atmospheric RT, especially for scenarios with significant aerosol scattering. RT
381 models traditionally used in retrievals of imaging spectroscopic data use simplified radiation schemes
382 and predefined aerosol models, which may introduce inaccuracies in the representation of atmospheric
383 physics. The 2S-ESS model provides the capability to quantify aerosol impacts on CH₄ retrieval for
384 different aerosol types, optical depths and layer heights. In future work, we will compare retrievals using
385 the 2S-ESS model against those from other commonly used models such as MODTRAN. We will also
386 evaluate the impact of varying instrument spectral resolution and signal to noise ratio for simultaneous
387 retrieval of CH₄, surface albedo and AOD. This will be relevant for the design of imaging spectrometers
388 for planned future missions such as the NASA Surface Biology and Geology (SBG) mission.

389

390 **Data availability**

391 The code and data are available from the authors upon request.

392

393 **Author contributions**

394 VN conceived the work, provided the radiative transfer and aerosol models, supervised YH, and
395 assisted with manuscript preparation. YH designed and performed the retrievals, analyzed the results,
396 and prepared the original manuscript. ZZ contributed to retrieval setup and assisted with analysis of the
397 results. PK provided valuable inputs into the science of CH₄ remote sensing. YLY supervised YH and
398 participated in the evaluation of the retrieval results and intercomparison. All listed authors contributed
399 to the review and editing of this manuscript.

400

401 **Competing interests**

402 The authors declare that they have no conflict of interest.

403

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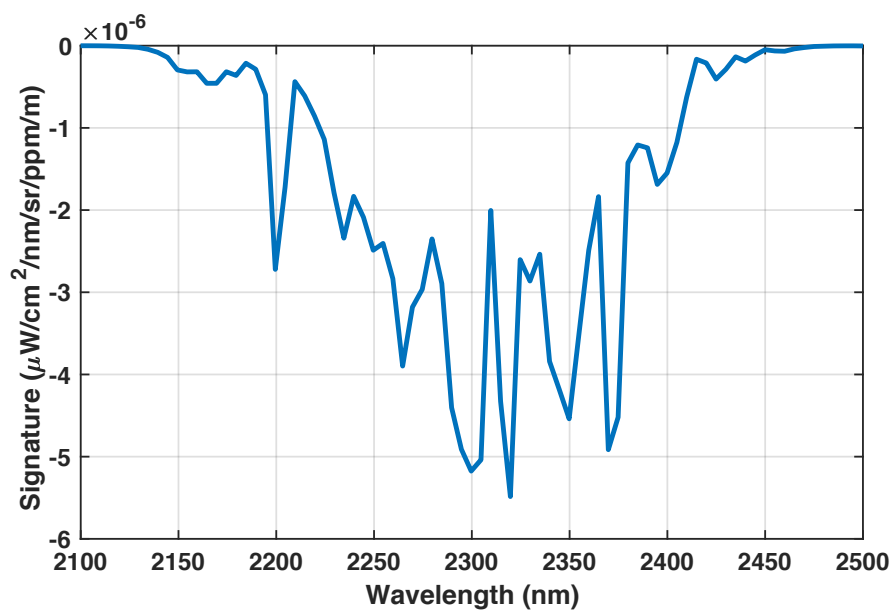
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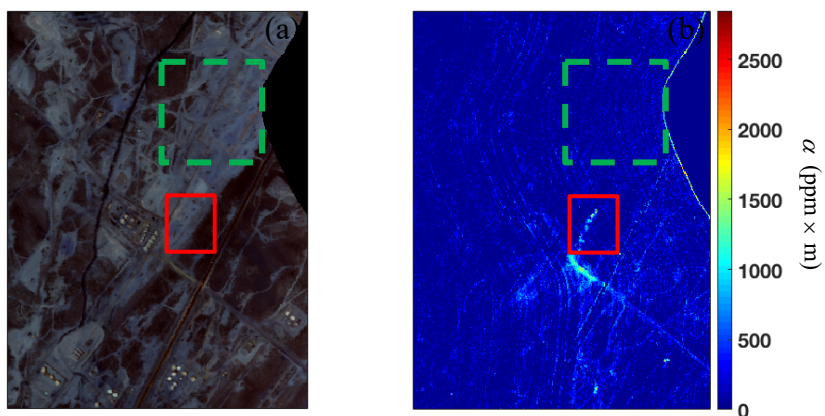
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743 **Figure 1: The target signature used for the Matched Filter method.**

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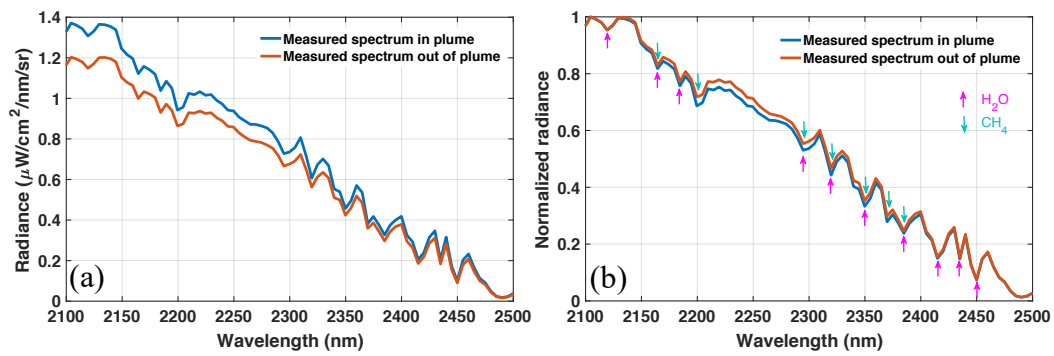


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747 **Figure 2: (a) RGB image of flight data from 4 September 2014 (ang20140904t204546). Adapted from**
748 **Thompson et al. (2015). (b) CH₄ enhancement value α (ppm \times m) obtained by the MF method. An emission**
749 **source is shown in the solid red box and the background region near the target for the MF calculation is**
750 **indicated by the dashed green box.**

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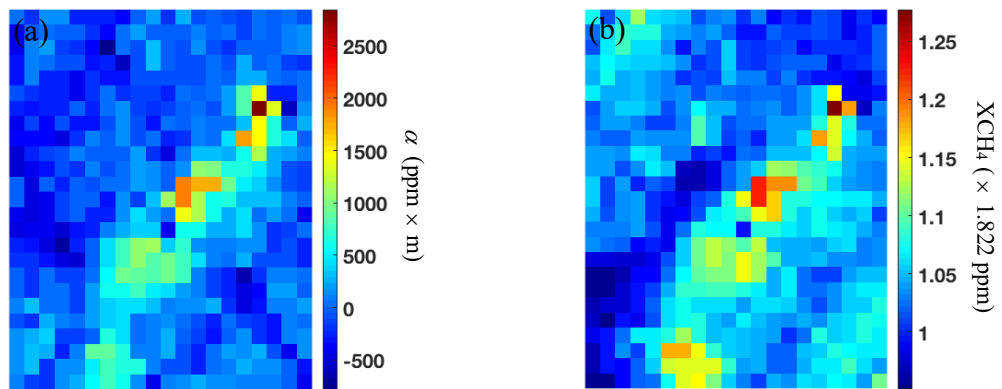
754 **Figure 3: (a) Real radiance and (b) normalized radiance at cross-track detector elements (in and out of plume)**

755 **from the sample AVIRIS-NG measurement. The colored arrows in (b) show the main absorption features due**

756 **to H_2O (purple) and CH_4 (green).**

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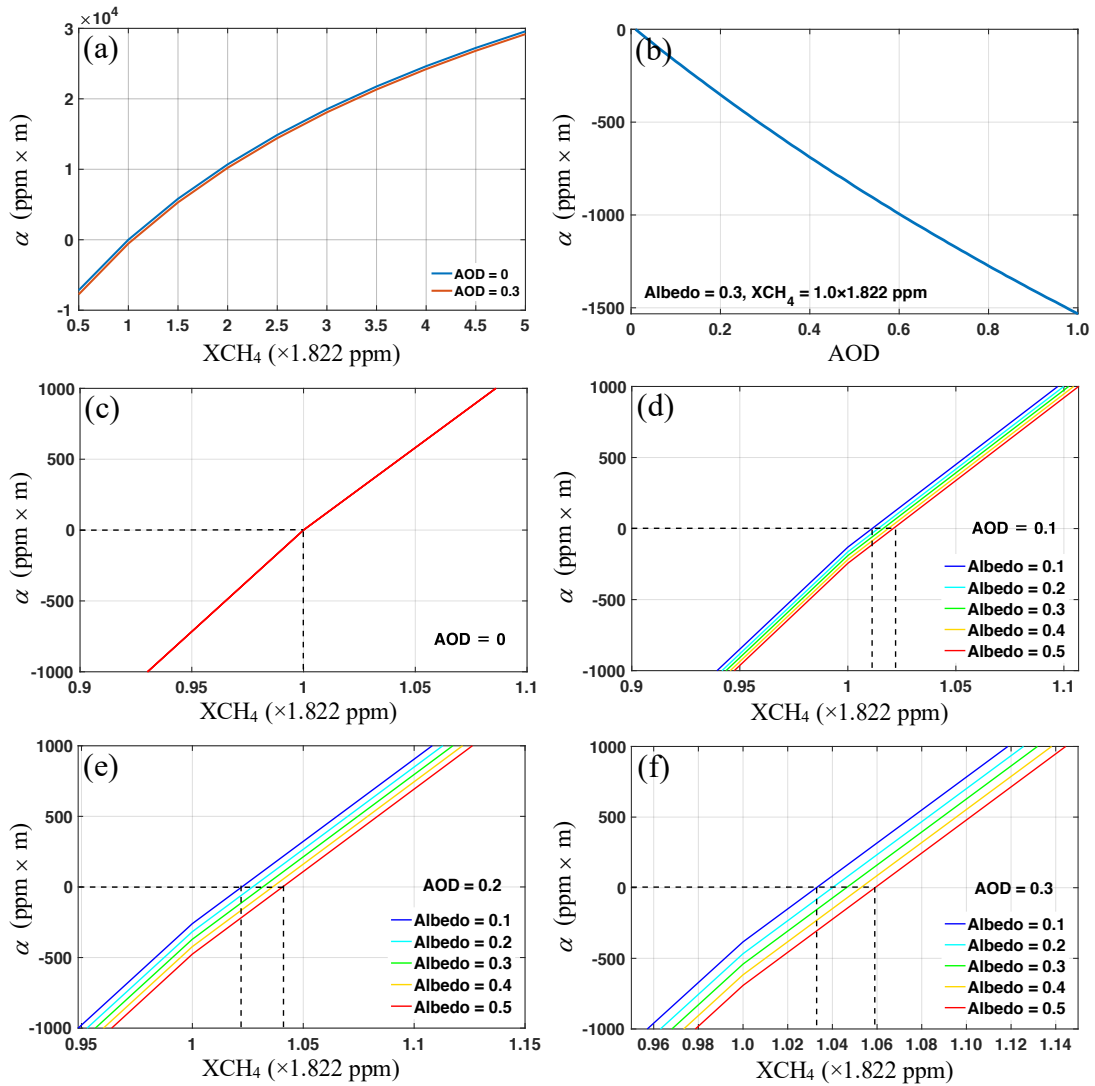


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760 **Figure 4: Retrieval image for the plume center (500 elements) based on the (a) MF method and (b) OE method.**

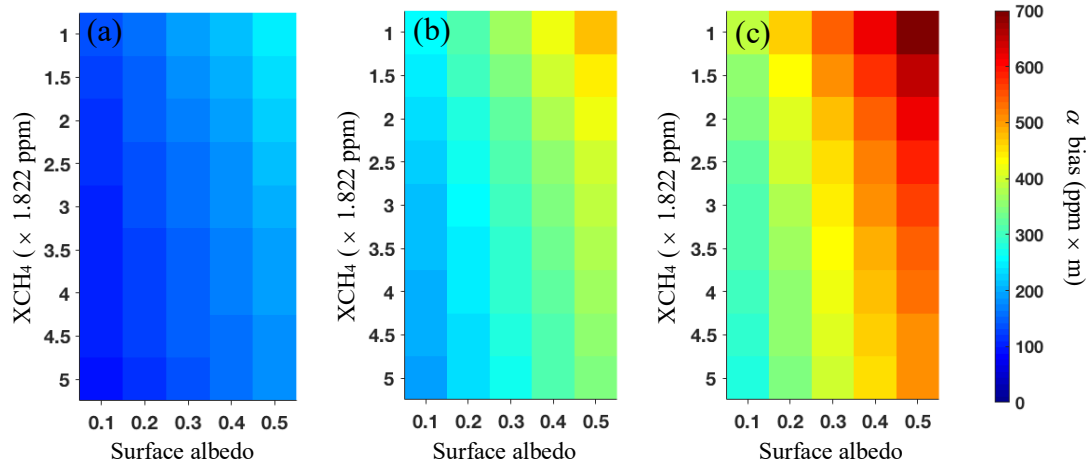
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764 Figure 5: (a) α as a function of XCH_4 for AOD = 0 and AOD = 0.3 (surface albedo = 0.3). (b) α as a function
765 of AOD ($XCH_4 = 1.0 \times 1.822$ ppm, surface albedo = 0.3). Zoomed in versions of α as a function of XCH_4 for
766 different surface albedos (0.1-0.5), where (c) AOD = 0, (d) AOD = 0.1, (e) AOD = 0.2, and (f) AOD = 0.3.
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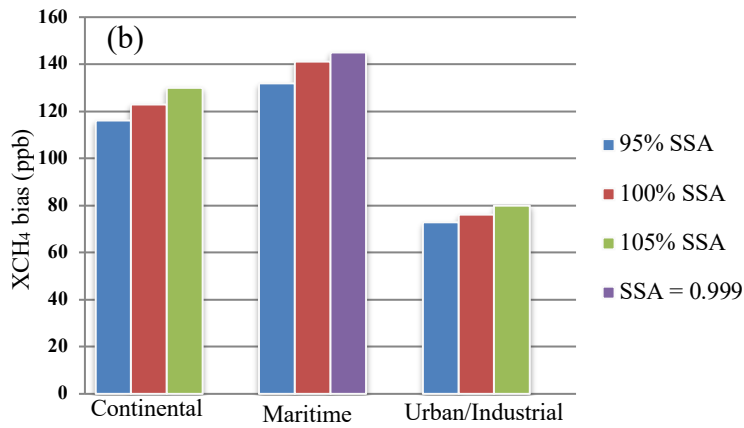
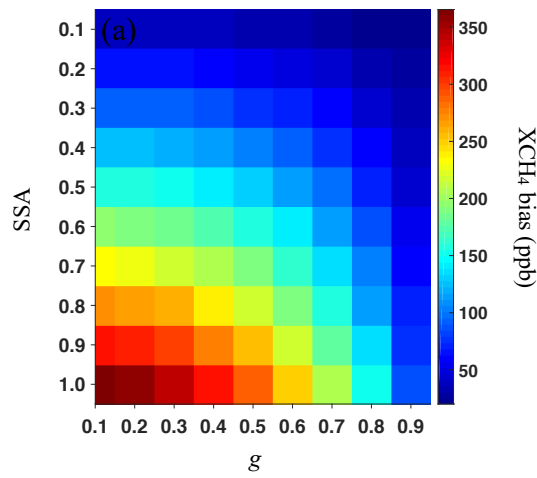
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770 **Figure 6: Bias in α as a function of XCH₄ and surface albedo for (a) AOD = 0.1, (b) AOD = 0.2, and (c) AOD**
771 **= 0.3.**

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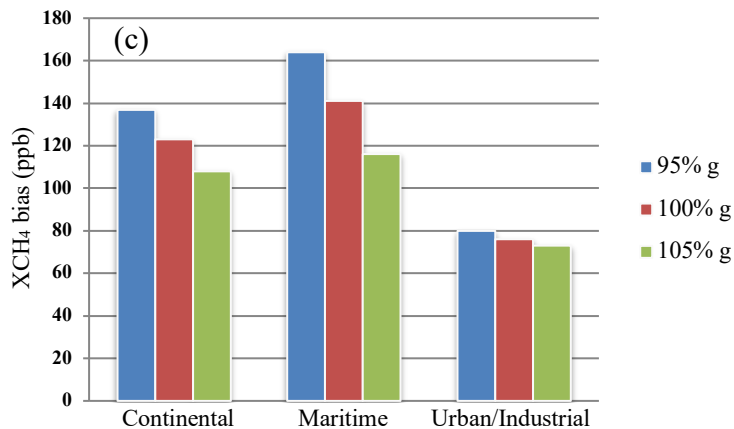
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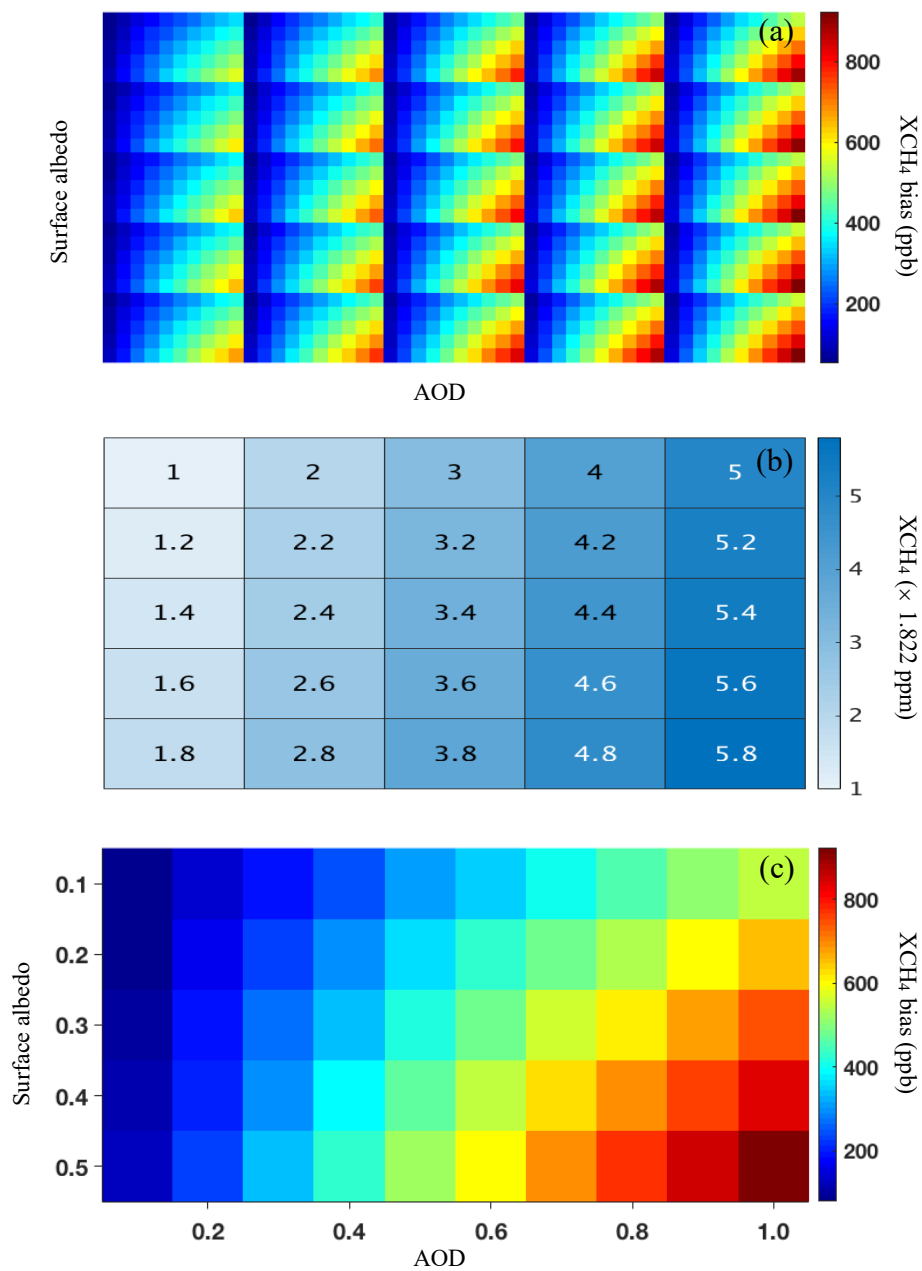
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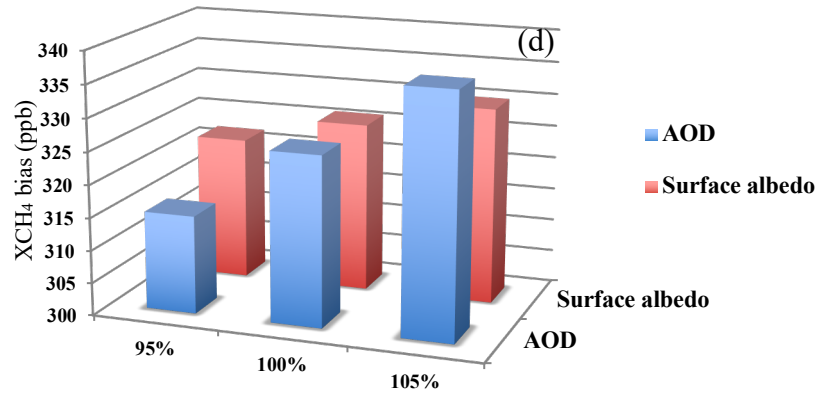
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Figure 7: (a) CH₄ retrieval biases for different values of g and SSA. Surface albedo, AOD = 0.3, XCH₄ = 1.0 × 1.822 ppm. (b) CH₄ retrieval biases for a ± 5% change in SSA for the three aerosol mixture models. (c) Same as (b), but for a ± 5% change in g .





783

784 Figure 8: (a) CH₄ retrieval biases for different values of XCH₄, AOD and surface albedo. $g = 0.75$, $SSA = 0.95$.

785 (b) XCH₄ for each box in (a). (c) Zoomed in plot of bottom right box (XCH₄ = 5.8×1.822 ppm). The x and y

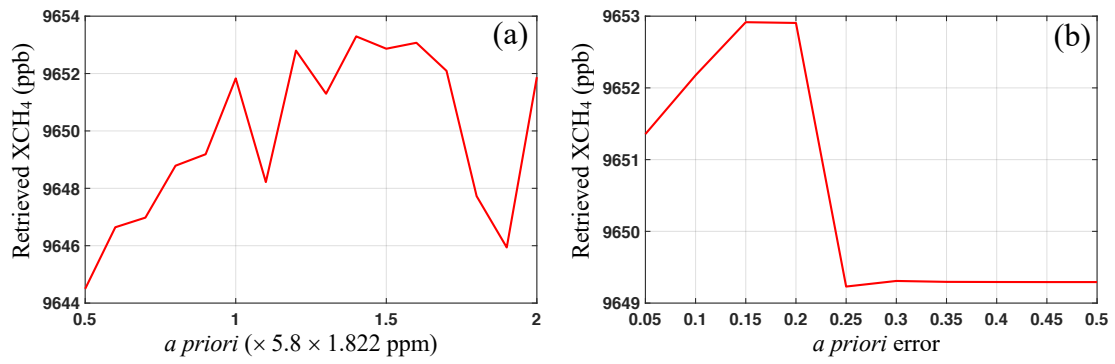
786 axes show the variation of AOD and surface albedo, respectively. These changes are identical for every box

787 in (a). (d) CH₄ retrieval biases for a $\pm 5\%$ change in AOD and surface albedo from a base value of 0.3 ($g =$

788 0.75, $SSA = 0.95$, XCH₄ = 5.8×1.822 ppm).

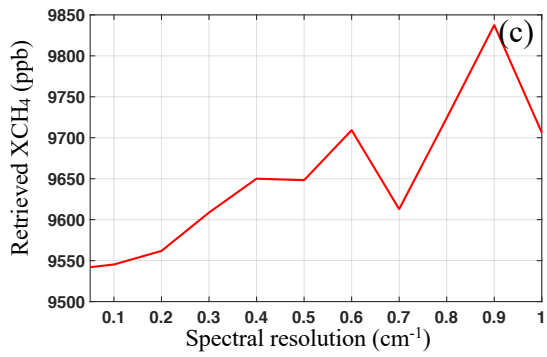
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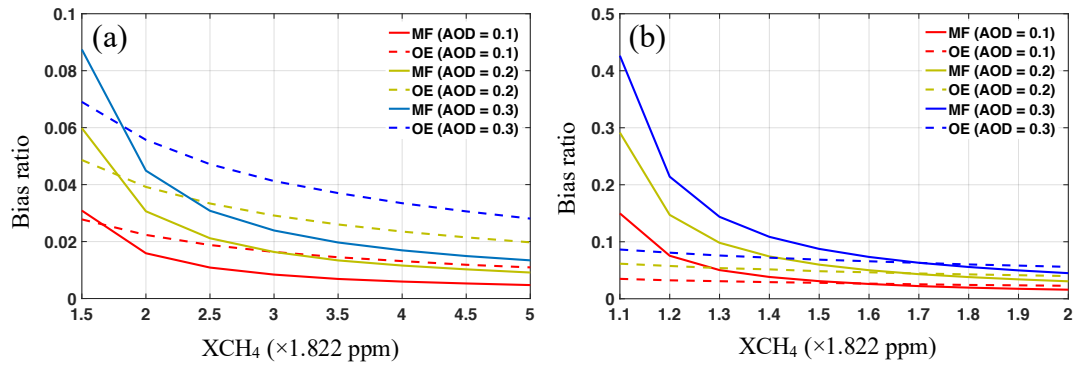
794 **Figure 9: Retrieved XCH₄ for different values of (a) *a priori* (*a priori* error = 0.2), (b) *a priori* error (*a priori* =**

795 **5.5 × 1.822 ppm) and (c) spectral resolution. *g* = 0.75, SSA = 0.95, AOD = 1.0, surface albedo = 0.5, XCH₄ =**

796 **5.8 × 1.822 ppm.**

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798



799 **Figure 10: (a) Bias ratio as a function of CH₄ concentration for the two retrieval techniques, where the XCH₄**
800 **ranges from 1.5 to 5 ($\times 1.822$ ppm). (b) Same as (a), but for XCH₄ ranging from 1.1 to 2 ($\times 1.822$ ppm). Surface**
801 **albedo is set to 0.3 for all cases; results for the MF and OE methods are shown by solid and dashed lines,**
802 **respectively.**
803

804

	Dust-like	Water soluble	Oceanic	Soot
SSA	0.805	0.799	0.970	0.014
<i>g</i>	0.926	0.550	0.816	0.092

805 **Table 1: Optical properties of basic aerosol types (WCRP, 1986).**

806

		Continental	Maritime	Urban/Industrial
Aerosol component	Dust-like	70%		17%
	Water soluble	29%	5%	61%
	Oceanic		95%	
	Soot	1%		22%
SSA		0.746	0.966	0.314
<i>g</i>		0.764	0.810	0.586

807 **Table 2: Optical properties of three aerosol mixture models (WCRP, 1986).**

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Attribute	Values
Sensor height	1 km
View zenith angle	11.91°
Solar zenith angle	30.75°
Relative azimuth angle	22.87°
Aerosol loading region	surface to 1 km
SSA	0.95
<i>g</i>	0.75

810 **Table 3: Inputs for the 2S-ESS model simulation.**

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