



Evaluating the PurpleAir monitor as an aerosol light scattering 1 2 instrument

- James R. Ouimette¹, William C. Malm², Bret A. Schichtel³, Patrick J. Sheridan⁴, Elisabeth 3
- Andrews^{4,5}, John A. Ogren⁶, W. Patrick Arnott⁷ 4
- 5 ¹Sonoma Ecology Center, Sonoma Ecology Center, Eldridge, CA 95431, USA
- 6 ²Cooperative Institute for Research in the Atmosphere, Colorado State University, Fort Collins, 7 CO 80523, USA 8
 - ³National Park Service Air Resource Division, Fort Collins, CO 80523, USA
- 9 ⁴ NOAA Global Monitoring Laboratory, Boulder, CO 80305, USA
- ⁵Cooperative Institute for Research in Environmental Sciences, University of Colorado, Boulder, 10 Colorado 80309, USA 11
- ⁶ NOAA Global Monitoring Laboratory, Boulder, CO 80305, USA (Retired) 12
- ⁷Department of Physics, University of Nevada, Reno, NV 89557, USA 13
- 14 Correspondence to Elisabeth Andrews (betsy.andrews@noaa.gov)

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Abstract

- The Plantower PMS5003 sensors (PA-PMS) used in the PurpleAir (PA) monitor PA-II-SD 17
- 18 configuration are equivalent to cell-reciprocal nephelometers using a 657 nm perpendicularly
- 19 polarized light source that integrates light scattering from 18 to 166 degrees. Yearlong field data
- 20 at the National Oceanic and Atmospheric Administration's (NOAA) Mauna Loa Observatory
- 21 (MLO) and Boulder Table Mountain (BOS) sites show that the 1 h average of the PA-PMS first
- 22 size channel, labeled ">0.3 μm" ("CH1") is highly correlated with submicrometer aerosol
- scattering coefficients at the 550 nm and 700 nm wavelengths measured by the TSI 3563 23
- integrating nephelometer, from 0.4 Mm⁻¹ to 500 Mm⁻¹. This corresponds to an hourly average 24
- submicrometer aerosol mass concentration of approximately 0.2 to 200 µg m⁻³. A physical-25
- 26 optical model of the PA-PMS is developed to estimate light intensity on the photodiode,
- 27 accounting for angular truncation as a function of particle size. Predictions are then compared
- 28 with yearlong fine aerosol size distribution and scattering coefficient field data at the BOS site. It
- 29
- is shown that CH1 is linearly proportional to the model-predicted intensity of the light scattered
- 30 by particles in the PA-PMS laser to its photodiode over 4 orders of magnitude. This is consistent
- 31 with CH1 being a measure of the scattering coefficient and not the particle number concentration
- 32 or particulate matter concentration. Field data at BOS confirm the model prediction that the ratio
- 33 of CH1 to the scattering coefficient would be highest for aerosols with median scattering
- 34 diameters <0.3 µm. The PA-PMS detects aerosols smaller than 0.3 µm diameter in proportion to
- 35 their contribution to the scattering coefficient. The model predicts that the PA-PMS response to
- particles >0.3 µm decreases relative to an ideal nephelometer by about 75% for particle 36
- 37 diameters >1.0 µm. This is a result of using a laser that is polarized, the angular truncation of the
- 38 scattered light, and particle loss in the instrument before reaching the laser. The results of this
- 39 study indicate that the PA-PMS is not an optical particle counter and that its six size fractions are
- 40 not an accurate representation of particle size distribution. The relationship between the PA-PMS
- 1 h average CH1 and b_{sp1}, the scattering coefficient in Mm⁻¹ due to particles below 1 μm 41
- 42 aerodynamic diameter, at wavelength 550 nanometers, is found to be $b_{sp1} = 0.015 \pm 2.07 \times 10^{-5} \times 10^{-5}$
- CH1, for relative humidity below 40%. The coefficient of determination R² is 0.97. This 43





- 44 suggests that the low-cost and widely used PA monitors can be used to measure and predict the
- 45 aerosol light scattering coefficient in the mid-visible nearly as well as integrating nephelometers.
- 46 Keywords: PurpleAir, Plantower PMS5003, nephelometer, low-cost sensor, physical-optical
- 47 model, PM2.5, scattering coefficient, visibility, atmospheric aerosol

1. Introduction

- 49 Currently there are tens of thousands of low-cost aerosol monitors used by atmospheric research
- 50 groups, air quality monitoring and regulatory organizations, and individual citizen scientists
- around the world. The recent explosion in the number of these sensors (see, for example, AAQR,
- 52 20(2), "Special Issue on Low-cost Sensors for Air Quality Monitoring" and papers therein) is a
- 53 result of the increased research, regulatory, and citizen interest over the past few years. For
- example, there are over 9,000 active PurpleAir (PA) aerosol monitors (PurpleAir LLC, Draper,
- 55 UT), with sampling locations on almost every continent. The large geographic coverage of this
- 56 array of low-cost sensors presents enormous potential for obtaining valuable information on
- 57 atmospheric aerosol properties and transport processes.
- 58 The majority of these low-cost aerosol sensors are used to monitor the mass concentration of
- 59 particles with aerodynamic diameters <2.5 μm (PM2.5) (Kelly et al., 2017; Gupta et al., 2018;
- 60 Sayahi et al., 2018; Zheng et al., 2018; Malings et al., 2019; Barkjohn et al., 2020; Holder et al.,
- 61 2020; Jayaratne et al., 2020; Mehadi et al., 2020). However, these sensors do not actually
- 62 measure aerosol mass concentrations but light scattered by the aerosols and thus are dependent
- on the aerosol particle size distribution, morphology, and composition. Recently, Hagan and
- 64 Kroll (2020) developed a framework and computer model to estimate the effects of relative
- 65 humidity and aerosol refractive index on PM2.5 estimated by a number of low-cost sensors.
- 66 Their model assumed that the low-cost sensor lasers were not polarized and could be modeled
- 67 with Mie theory. The PMS5003 (PMS) was included in their classification scheme as an example
- of a sensor that behaved more like a nephelometer than an optical particle counter.
- 69 Three recent laboratory studies showed that the PMS response decreases with particle size. He et
- al. (2020) measured the PMS response to monodisperse ammonium sulfate aerosol particles
- 71 having diameters of 0.1, 0.3, 0.5, and 0.7 µm. The PMS was able to detect 0.1 µm particles. They
- 72 derived a transfer function that showed that the PMS >0.3 μm channel (CH1) response was
- 73 maximum at particle diameter 0.26 μm but decreased significantly below this size. They
- 74 concluded that the PMS behaved more like a nephelometer than an optical particle counter.
- 75 Kuula et al. (2020) generated monodisperse dioctyl sebacate oil droplets from 0.5 to 20 µm and
- measured the PMS CH1 response versus particle diameter using an aerosol particle sizer (APS).
- 77 Their data showed that the PMS relative response decreased for particles >0.5 µm diameter.
- Their data showed that the PMS relative response decreased for particles >0.5 µm diameter
- 78 Tryner et al. (2020) evaluated three low-cost particulate matter sensors, including the PMS, by
- exposing them to five different types of aerosols in the laboratory. They found that the ratios of PMS-reported to filter-derived PM2.5 mass concentrations were inversely proportional to mass
- Tivis-reported to inter-derived Tivis.5 in associate attaches were inversely proportional to ind
- median diameter (MMD). Wood smoke had the smallest MMD, $0.42 \mu m$; its PMS PM2.5
- 82 averaged 2.5 times the filter-derived PM2.5. Conversely, oil mist had the largest MMD, 2.9 μm;
- its PMS PM2.5 averaged only 0.23 times the filter-derived PM2.5.
- 84 Climate modeling requires a robust set of models and atmospheric measurements for predicting
- anthropogenic aerosol radiative forcing. Currently, there are uncertainties in the modeling

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- 86 results, due in part to the sparseness of ground-based data used to evaluate and refine the models
- 87 (e.g., Gliss et al., 2021). Satellite observations provide global coverage that can be used for
- 88 model evaluation, but satellite data require further assessment, particularly when trying to
- 89 provide information about surface aerosol properties. The Surface Particulate Matter Network
- 90 (SPARTAN) (https://www.spartan-network.org/; Snider et al., 2015) was specifically designed
- 91 to assess and improve algorithms to relate satellite retrievals to surface aerosols. SPARTAN
- 92 operates collocated filter-based PM2.5, aerosol scattering coefficient via nephelometer, and
- 93 aerosol optical depth (AOD) measurements at approximately 20 sites around the world. Model
- 94 and satellite uncertainties can be reduced using a distributed set of low-cost sensors that can
- 95 provide aerosol light scattering estimates at a higher spatial and temporal resolution than is
- 96 possible using nephelometers alone. Low-cost sensors are increasingly being used along with
- 97 satellite data to estimate global aerosol impacts (Gupta et al., 2018).
- 98 There is ongoing scientific debate about the accuracy and precision of these low-cost sensors and
- 99 their limitations (Morawska et al., 2018; Jayaratne et al., 2020). Many of the recent papers
- discuss performance evaluations or 'calibrations' of these low-cost sensors by comparing their
- measurements with traditional, research-grade aerosol measurements (Papapostolou et al., 2017;
- Barkjohn et al., 2020). The concerns over data quality, stemming largely from inexpensive
- 103 components, lack of transparency of signal processing, and inadequate quality control and testing
- at the factory, must be weighed against the advantages of low cost and wide spatial coverage.
- The actual measurement in the PA monitor with its two PMS5003 sensors (PA-PMS), and in
- many other low-cost aerosol monitors, is of light scattered by particles (Kelly et al., 2017), which
- traditionally has been done in atmospheric research and aerosol monitoring programs using
- integrating nephelometers. Aerosol light scattering extinction measurements are useful in many
- integrating representation. Across right seattering extinction measurements are discreti in main
- applications, including determination of the radiative forcing effects of aerosols on climate
- change, atmospheric visibility, wildfire/smoke impacts, and validation of model outputs and
- satellite retrievals (e.g., Malm et al., 1994; Sherman et al., 2015; Snider et al., 2015; Gliss et al.,
- 112 2021). Even though most low-cost aerosol sensors use light scattering as the basis of their
- 113 operation, almost none have been evaluated as a low-cost nephelometer to estimate atmospheric
- 114 light scattering. Markowicz and Chilinski (2020) conducted a 3-year evaluation of two low-cost
- sensors versus the Aurora 4000 polar integrating nephelometer at a site in southeast Poland.
- 116 They found that the mass concentration of particles with aerodynamic diameters <10 μm (PM10)
- from the DfRobot SEN0177 and the Alphasense OPC-N2 were highly correlated ($R^2 > 0.89$)
- with the aerosol scattering coefficient measured by the nephelometer. They were able to estimate
- the 1 h average aerosol scattering coefficient from the low-cost sensors with a root mean square
- 120 error (RMSE) of 20 Mm⁻¹, corresponding to 27% of the mean aerosol scattering coefficient.
- 121 Unfortunately, due to cost, availability, and the expertise required to run them, integrating
- 122 nephelometers are not operated in great numbers around the world. A recent analysis by Laj et
- 123 al. (2020) showed 56 long-term monitoring stations reporting their nephelometer data to the
- 124 World Meteorological Organization (WMO) Global Atmosphere Watch (GAW) World Data
- 125 Centre for Aerosols. This count includes nephelometers operated in several monitoring
- 126 networks, including the National Oceanic and Atmospheric Administration's (NOAA) Federated
- 127 Aerosol Network (NFAN, Andrews et al., 2019), the Aerosols, Clouds and Trace Gases Research
- 128 Infrastructure (ACTRIS) network (e.g., Pandolfi et al., 2018), and the Interagency Monitoring of
- Protected Visual Environments (IMPROVE) network (Malm et al., 1994). While there are more





- 130 nephelometers in use around the world for short-term field and laboratory studies, the number
- almost certainly does not exceed a few hundred. This is small compared with the number of low-
- 132 cost aerosol monitors in use globally.
- 133 This paper presents an evaluation of the performance characteristics of the low-cost PA-PMS
- monitor to measure the integrated aerosol light scattering coefficient. It is shown that the PA-
- PMS sensor configuration is similar to a cell-reciprocal nephelometer. A physical-optical model
- 136 based on Mie theory and the PMS geometry is created that predicts scattered light intensity on
- 137 the PMS photodiode and aerosol forward and backward light scattering truncation. Model
- predictions are then compared with yearlong field data at NOAA's Mauna Loa Observatory
- 139 (MLO) in Hawaii and the Boulder Table Mountain (BOS) site in Colorado. These PA-PMS
- measurements are also compared to measured light scattering coefficients, and an empirical
- relationship is developed to estimate the light scattering and uncertainty from the PA-PMS data.
- With a better understanding of what the PA measures, how it works, and its uncertainties, the
- large network of PA-PMSs could be used to estimate the submicrometer aerosol scattering
- 144 coefficient at visible wavelengths throughout the world. These data could then be used to
- improve chemical transport and general circulation models, advance climate change predictions,
- and provide for better air quality forecasts.

147 **2. Instrument description**

- 148 In this section we first describe the physical and optical characteristics of the PA-PMS to place it
- in the context of nephelometry. We then provide a brief overview of integrating nephelometers,
- which are instruments designed specifically to measure light scattering.

151 **2.1 PA-PMS nomenclature**

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- 152 The PMS sensor outputs 14 fields that are processed and reported by the PA. Each of these
- fields will be referred to as a channel. For instance, the PA-reported number concentration of
- particles >0.3 µm is referred to as CH1 in the remainder of this paper, number concentrations
- 155 >0.5 μm as channel two (CH2), and so forth. Furthermore, the two PMS sensors embedded in
- 156 the PA will be referred to as either sensor A or sensor B. Therefore, the number concentration of
- 157 particles >0.3 µm derived from sensor A will be referred to as CH1A and those from sensor B as
- 158 CH1B. The average of CH1A and CH1B will be referred to as CH1avg. The PMS reports the
- 159 CH1 units as "#/dl", which is the number of particles having diameters >0.3 μm per deciliter. In
- this paper the PMS units for CH1 are not used.

2.2 Description of the PA and its PMS 5003 sensors

- The PA monitor integrates two PMS sensors, a Bosch BME280 pressure, temperature, and
- relative humidity sensor and an ESP 8266 chip (https://www2.purpleair.com/pages/technology).
- 164 The available specifications of the PMS are incomplete, and the processing algorithms are
- unknown (He et al., 2020). The following is based on available information and, where needed,
- 166 professional judgment. Each PMS includes a small laser, a photodiode, a small fan to draw air
- across the laser, a microprocessor control unit (MCU), and probably an operational amplifier.
- 168 The MCU processes the signal from the photodiode and outputs the following data fields
- 169 approximately once per second: $>0.3 \mu m$, $>0.5 \mu m$, $>1.0 \mu m$, $>2.5 \mu m$, $>5 \mu m$, $>10 \mu m$, PM1,
- 170 PM2.5, and PM10. The PMS denotes the first six data fields as particle number concentrations





- 171 above the designated cutpoint and the last three data fields as mass concentrations of particles
- below the designated cutpoints; the PM data fields are reported for two different conditions, 172
- 173 "standard particles" and "under atmospheric environment". The PA ESP8266 chip calculates 2
- 174 min averages of the PMS and BME280 signals. It transmits them wirelessly and writes them as a
- 175 CSV file on a microSD card.

2.2.1 Airflow and particle losses

- 177 The recommended orientation of the PA results in aerosol being drawn upward by a small fan
- 178 through four 3-mm diameter entrance holes in each PMS. The aerosol then enters a 9.4 cm³
- 179 chamber (Fig. S1a) and flows upward, parallel to and exposed to the circuit board as shown in
- 180 Fig. S1b. Particles then make a 180 degree turn through three exit holes at the top of the chamber
- 181 to emerge on the other side of the circuit board and flow downhill through a channel that is
- illuminated by the laser. The PMS volumetric flow rate is estimated to be 1.5 cm³ s⁻¹ (~0.090 182
- lpm) based on measurements described in Supplemental Materials Sect. S1. The estimated inlet 183
- 184 velocity through the entrance holes is estimated to be 5.3 cm s⁻¹.
- 185 The PMS inlet orientation 90 degrees to the wind, upward flow, and the low inlet velocity
- 186 through the sampling holes can result in significant aspiration losses of larger particles (Hangal
- and Willeke, 1990). Aspiration losses are greater at higher wind speeds because it is more 187
- 188 difficult for the larger particles to follow the streamlines into the low velocity PMS inlet. This
- 189 can result in a lower concentration of larger particles entering the PMS than are in the ambient
- 190 air. Particle aspiration losses are proportional to the particle Stokes number and the ratio of the
- 191 wind velocity to the inlet face velocity (Hangal and Willeke, 1990). More details are provided in
- 192 Supplemental Materials Sect. S1.
- At typical wind velocities of 1–3 m s⁻¹, the ratio of PMS inlet face velocity to wind speed is only 193
- 194 0.02 to 0.05, much lower than typical sampling ratios of 0.5 to 6.0 (Kulkarni et al., 2011). Pawar
- 195 and Sinha (2020) addressed this problem for the Laser Egg low-cost sensor by putting it in a box
- 196 and adding a 40 lpm fan to increase the inlet-to-wind velocity ratio and to direct the airflow
- 197 upward to the Laser Egg inlet. During calm winds, large particle aspiration losses may occur by
- 198 particle gravitational settling, acting against the PMS upward flow (Grinshpun et al, 1993). The
- 199 actual wind conditions in the ambient air and in the PA near the PMS sample inlet are turbulent.
- 200 Hangal and Willeke (1990) found in their wind tunnel experiments that turbulence intensity had
- a negligible effect on aspiration efficiency. Calculations using Eq. S1 (see Fig. S2) predict that at 201
- 202 a wind speed of 1 m s⁻¹, the PMS aspiration losses for particles \geq 2 μ m may be significant.
- 203 However, it must be cautioned that the literature does not include data for the very low 5.3 cm s⁻¹
- 204 PMS face velocity and actual measurements of the PMS aspiration efficiencies were not made.
- They may be significantly different from these calculated efficiencies. 205
- Inside the PMS 9.4 cm³ chamber, the air has an average velocity of 0.57 cm s⁻¹ and Reynolds 206
- 207 number of 6.1, resulting in an average residence time of 6.3 s. The average air velocity in the
- 208 chamber is equal to the sedimentation velocity of a spherical 10 um diameter particle with a
- density of 2 g cm⁻³ in air at STP (standard temperature and pressure; values used in this analysis 209
- are 273.15 K and 1010.25 hPa, respectively). This suggests that some 2 g cm⁻³ density particles 210
- 211 with diameters >10 µm that enter the PMS would settle out in the chamber and not make it to the
- 212 three exit holes at the top of the chamber. Ultrafine particles can also be lost to the walls of the
- 213 chamber and the printed circuit board due to convective diffusion. Calculations using the





- 214 equation for diffusional losses (Friedlander, 1977) show that less than 1% of the 0.01 µm
- diameter aerosols would be lost in the chamber due to convective diffusion, with even smaller
- 216 diffusional losses for larger particles.
- Loss of particles due to inertial impaction on the wall opposite the three holes (Fig. S1b) was
- estimated by the local air flow Reynolds number near the three holes and the aerosol Stokes
- 219 number. The local Reynolds number is calculated to be 23, and the Stokes number for 10 μm
- particles is 8.2×10^{-4} . At these low numbers, the calculated loss to impaction is less than 1% for
- 221 all particles less than 10 μm diameter (Hering, 1995).
- The average flow velocity through the laser beam is approximately 3.0 cm s⁻¹. By the time the
- 223 air flows through the laser beam, it has lost most of the particles over 10 µm diameter. Further
- 224 particle losses due to gravitational settling over the photodiode would be very small, since the
- gravitational force is parallel to the photodiode.

226 **2.2.2** Laser

- 227 The wavelength and power of three PMS diode lasers were measured using an Ocean Optics Red
- 228 Tide USB650 spectrometer and Melles Griot Universal Optical Power Meter, respectively. The
- 229 wavelength averaged 657 +/- 1 nm, and the power averaged 2.36 +/- 0.04 mW. The laser is
- 230 polarized parallel to the plane of the photodiode detector. This results in the aerosol-scattered
- 231 light being polarized perpendicular to the plane of incidence. Figure S3 shows that
- 232 perpendicular polarization results in significantly greater scattering intensity from 0.3 µm
- particles compared to natural or parallel polarization.
- The PMS laser beam profile is not a simple plane wave, but complex in shape. The laser has a 3
- 235 mm diameter lens that focuses the laser over the photodiode. The beam profile evolves
- significantly as it goes through the focal region (Naqwi and Durst, 1990). The laser beam
- 237 diameter in the laser sensing region over the photodiode was not measured. It was estimated by
- eye to be 0.5 to 1.0 mm, with significant uncertainty. The PMS MCU turns the laser on and off
- every 800 msec or 2.5 s, depending on aerosol concentration. The laser pulses are 600–900
- 240 msec, with the laser power on continuously during this time. We hypothesize that the PMS MCU
- 241 gathers data during laser on, processes it during laser off, and uses the difference of the
- 242 photodiode output during these stages to obtain and subtract any electronic or stray light (other
- than the laser) background signal to the photodiode.

244 **2.2.3 Photodiode detector**

- 245 The actual photodiode model in the PMS is unknown. The photodiode appearance is similar to
- 246 the BPW34 silicon PIN photodiode. In this paper the specifications of the BPW34 are used to
- estimate the likely properties of the detector in the PMS. It has a very large dynamic range when
- 248 operated with reverse bias. The dependence of the photodiode current on the light intensity is
- very linear over 6 or more orders of magnitude, e.g., in a range from a few nanowatts to tens of
- 250 milliwatts. Silicon PIN photodiodes have low dark current, a 20 nanosecond rise time, and good
- 251 wavelength sensitivity between roughly 400 and 1000 nm. (https://www.rp-
- 252 <u>photonics.com/photodiodes.html</u>). At a wavelength of 657 nm, the BPW34 produces
- 253 approximately 0.4 microampere current per microwatt of incident radiant power
- 254 (https://www.fiberoptics4sale.com/blogs/archive-posts/95046662-pin-photodetector-

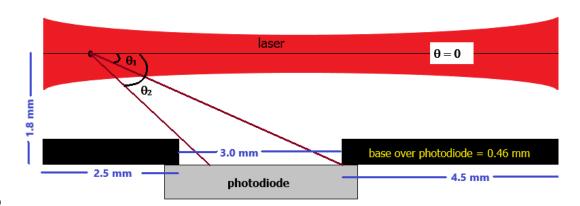




- 255 <u>characteristics-for-optical-fiber-communication</u>). The PMS does not have any optical elements
- 256 to capture and focus the aerosol-scattered light on its photodiode.
- 257 The photodiode does not have a cosine corrector in front and is probably not a true cosine
- detector. However, the relative spectral sensitivity is advertised to be a cosine response by the
- 259 manufacturers
- 260 (https://www.osram.com/ecat/DIL%20BPW%2034%20B/com/en/class pim web catalog 1034
- 261 89/prd pim device 2219537/ and https://www.vishay.com/docs/81521/bpw34.pdf).

2.2.4 Laser and photodiode geometry

- 263 The PMS geometry is very similar to a cell-reciprocal nephelometer. Figure 1 shows the PMS
- laser and photodiode geometry. The distance from the laser exit hole to the light trap is 10 mm;
- the perpendicular distance from the center of the laser to the photodiode is 1.8 mm; the diameter
- of the exposed photodiode area is 3.0 mm; and the thickness of the mask over the photodiode is
- 267 0.46 mm. θ_1 is the lower angular scattering limit, and θ_2 is the upper angular scattering limit for a
- 268 particle in the laser.



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Figure 1. Schematic of the PMS sensor geometry highlighting the dimensions of laser beam (red) and photodiode (gray) and the various relevant distances between the two.

- Due to the PMS geometry, the upper and lower angular scattering limits for θ depend on the
- location, x, of a particle in the laser. This can be seen on Fig. S4. For example, at x = 0 mm, at
- 274 the laser exit, the upper and lower scattering limits for θ are 18–38 degrees. At x = 4.0 mm, over
- 275 the center of the photodiode, the angular integration limits are 50–130 degrees. The PMS
- 276 photodiode is not capable of detecting light scattered from particles at less than 18 degrees.
- Figures S5–S8 provide more detail about the PMS dimensions and geometry.

2.2.5 PMS5003 sensing volume

- The sensing volume is the volume in which the aerosol is irradiated by the laser. The sensing
- volume extends the length of the laser where the aerosol flows through it, approximately 10 mm.
- 281 The sensing volume is shown in Fig. S9. The average residence time of a particle in the laser





- beam is approximately 0.03 s. Some of the scattered light is detected by the photodiode and
- 283 creates a voltage pulse approximately 30 msec wide. It appears that the photodiode is detecting
- 284 either a cloud of particles from the sensing volume or individual pulses, depending on the
- 285 concentration. At low concentrations, the aerosol concentration within the sensing volume is
- unlikely to be uniform, resulting in large relative changes in output per second.

287 2.2.6 Signal processing and electronics

- 288 It is not known how the PMS MCU differentiates and processes the photodiode signals. The
- 289 PMS MCU sends the PA a signal approximately every second in the form of a digital sequence
- of unsigned 16 bit binary data words, and CH1 is thought to be proportional to the photodiode
- 291 current. The photodiode current was not measured in this study. The PA creates 80 s (Firmware
- Version 3) or 120 s (Firmware Version 4 and higher) averages and writes them to its microSD
- 293 card. We measured an average percentage difference of 0.3% between the 2 min averages
- reported by the PA and the 2 min averages calculated from the 1 s values from the PMS. The
- results are shown in Fig. S10. It is apparent that the processing done by the PA to calculate its
- reported 2 min averages does not bias the results.

297 2.2.7 PMS CH1 variability in sampling filtered air

- We found significant variability in PMS response to filtered air. We exposed 21 PAs containing
- 299 42 PMS sensors to filtered air for 2 to 94 hours. The results are summarized in Table S1. Hourly
- 300 average CH1 ranged from 0.10 to 377. Eleven PAs had both PMS CH1A and CH1B averages
- below 2, while seven PAs had at least one CH1 average over 26. We recommend that before
- deployment the PAs sample filtered air for at least 4 hours to identify and eliminate PAs with
- 303 CH1 hourly averages over 2 in filtered air. Removing PAs with high CH1 offsets in filtered air
- reduces uncertainty and improves precision, particularly in cleaner ambient air.

305 2.2.8 PMS CH1 unresponsive to CO₂ and Suva[®]

- Filtered air, CO₂, and Suva[®] (DuPontTM Suva[®] 134a refrigerant) are often used to calibrate
- integrating nephelometers (Anderson et al., 1996). The Rayleigh scattering coefficients of
- 308 filtered air, CO₂, and Suva at 657 nm and at STP (°C and 1013.25 hPa) are 5.5, 13.3, and 46.2
- 309 Mm⁻¹, respectively. We found that the PMS was unresponsive to 100% CO₂ (Fig. S11) and Suva.
- 310 The CH1 for each gas was the same as filtered air. These results indicate that the PMS signal
- 311 processing zeroes out a constant scattering signal and cannot be used to measure the scattering
- 312 coefficient of gases that are commonly used in calibrating nephelometers. Furthermore, the
- method used by the PMS to subtract light scattering by air molecules in the sampling volume is
- 314 unknown.

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2.2.9 PMS CH1 and CH1avg precision

- 316 The PMS CH1 precision was measured by collocating ten PA monitors on the roof of the NOAA
- 317 building in Boulder, Colorado, between 22 January 2021 and 1 February 2021. These monitors
- 318 were not checked with filtered air before deployment. It was found that two of the PMS sensors
- 319 had large offsets and two had moderate offsets at low CH1 values. One PMS sensor was found to
- 320 produce errant data and was removed from the analysis, resulting in valid data from 19 CH1A
- and CH1B sensors in the ten PAs.





The hourly CH1A-CH1B and CH1avg precisions were estimated as the coefficient of variation for the 19 CH1A-CH1B and 9 CH1avg values for each hour, which are plotted against the average CH1 values in Fig. 2. As shown, above CH1 values of 500, the precision is relatively constant with an average of 8% and 4.8% for CH1A-CH1B and CH1avg, respectively. Below CH1 values of 500, the uncertainties increase rapidly with decreasing CH1 values.

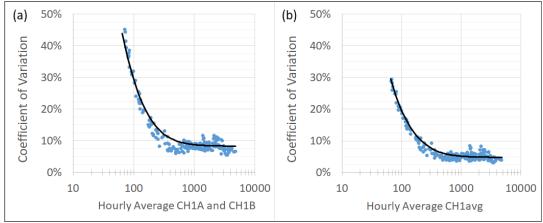


Figure 2. Precision estimated as the coefficient of variation of the hourly CH1A-CH1B (a) and CH1avg values (b) for the 19 collocated sensors and 9 PAs.

The data in Fig. 2 can be modeled by the sum of squares of an additive (Un_{add}) and multiplicative uncertainty (Un_{mult}) (Currie, 1968; JCGM100:GUM, 2008; Hyslop and White, 2008):

Uncertainty =
$$\sqrt{Un_{add}^2 + Un_{mult}^2 * CH1}$$
 (1)

Equation 1 was fitted to the precision data in Fig. 2 where the Un_{mult} was set to the average precision at high CH1 values, and Un_{add} was set to 28 and 19 for the A and B sensors and CH1avg, respectively, to fit the highest variances. The Un_{add}^2 is the variance in CH1 as CH1 approaches zero and is assumed to be equivalent to the uncertainty in values below the instrument minimum detection limit (MDL) or that of blanks (Currie, 1968), which were 0.08 and 0.048 for the A and B sensors and CH1avg, respectively. The coefficient of determination in the model fit for both sets of data was $r^2 = 0.96$. Defining the MDL as the 99% confidence interval of the Un_{add} (Code of Federal Regulations, 40 CFR 136, https://ecfr.io/Title-40/Part-136), MDLs for the individual CH1 sensors and CH1avg were 65 and 44, respectively.

As shown in Sect. S3, the Un_{mult} and Un_{add} are highly dependent on the systematic biases between the individual CH1 sensors and CH1avg and the four CH1 sensors with data offsets as the CH1 approaches zero. Removing these four sensors and normalizing the data for each CH1 sensor by its average reduced the Un_{add} and Un_{mult} to 9% and 3%, respectively, for the CH1 sensors and 6% and 1.9%, respectively, for the CH1avg data. These results correspond to an MDL of 21 and 14 for the normalized CH1 sensor and CH1avg data, respectively. Based on these results, an "off the shelf" PA will have a CH1avg MDL of about 44 and precision of less than 4.3%, but the careful selection of a PA without an offset and that is relatively calibrated will have an MDL of 14 and precision of less than 1.9%.



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2.3 Overview of cell-direct and cell-reciprocal nephelometers

The integrating nephelometer was invented during World War II (Beuttell and Brewer, 1949). It provides a direct measure of aerosol light scattering integrated over a large angular range, the "aerosol light scattering coefficient". This measure requires no assumptions about aerosol composition, size distribution, refractive index, or shape. The most common nephelometer configurations are the "cell-direct" and "cell-reciprocal". Figure 3 presents schematics of the two types of nephelometers. The geometrical relationship between the laser and the photodetector in the PMS resembles a cell-reciprocal nephelometer (Fig. 3b).

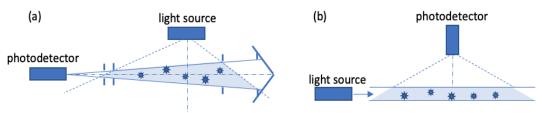


Figure 3. Diagrams of the (a) cell-direct nephelometer and (b) cell-reciprocal nephelometer, simplified from Peñaloza-Murillo (1999).

Middleton (1952) was the first to show that the cell-direct nephelometer with a Lambertian (cosine-adjusted diffuser) light source directly measures the aerosol light scattering coefficient. Anderson et al. (1996), following the derivation in Butcher and Charlson (1972), added geometrical diagrams to make Middleton's derivation much clearer. Mulholland and Bryner (1994) proved that the cell-reciprocal nephelometer with a Lambertian diffuser followed by a photodiode placed at the center of the cell-reciprocal nephelometer also directly measures the aerosol scattering coefficient. This put both the cell-direct and cell-reciprocal nephelometers on equal theoretical footing.

369 There are a number of cell-direct nephelometers in use today. They include the TSI 3563 (St. 370 371 Paul, MN, USA; Anderson et al., 1996), the Ecotech Aurora Models 3000 and 4000 (Knoxfield, 372 Australia; Müller et al., 2011), the Radiance Research M903 (Seattle, WA, USA; Heintzenberg 373 et al., 2006), and the Optec NG-2 (Lowell, MI, USA; Molenar, 1997). In contrast, cell-374 reciprocal nephelometers have more limited commercial availability. The photoacoustic

375 extinctiometer (PAX; Droplet Measurement Technologies, Inc., Longmont, CO, USA) and the

376 three-wavelength photoacoustic soot spectrometer (PASS-3) use a cell-reciprocal nephelometer

377 to measure aerosol light scattering coefficient (Arnott et al., 2006). A cosine corrector followed

378 by a photomultiplier tube is placed at the center of the cell-reciprocal nephelometer (Abu-

379 Rahmah et al., 2006; Nakayama et al., 2015).

380 A "perfect nephelometer" is one in which the nephelometer is able to see the scattered light over 381 the entire angular range from 0 to 180 degrees. In practice, this cannot be achieved for the cell-382 direct and cell-reciprocal nephelometers. Both the forward and backward scattering angles are 383 truncated. For example, the TSI 3563 nephelometer has measured angular truncation below

about 7 degrees in the forward direction and above 170 degrees in the backward direction 384

385 (Anderson et al., 1996; Heintzenberg and Charlson, 1996). For the PASS-3, Nakayama et al.

386 (2015) found that both the large effective truncation angle (21 degrees) as well as the





- 387 perpendicular polarization of the 532 nm laser relative to the scattering plane contribute to the
- 388 large particle size dependence of measured scattering. Light scattering from ammonium sulfate
- particles of 0.71 μm diameter was truncated 50%. Truncation generally results in nephelometers
- 390 underestimating the contribution of particles larger than approximately 1 μm diameter to the
- 391 scattering coefficient, although corrections have been developed to account for angular
- nonidealities (e.g., Anderson and Ogren, 1998; Müller et al., 2011).

3. A physical-optical model of the PMS5003

- To gain insight into how the PMS responds to ambient aerosol properties, a model was
- developed to estimate the intensity of scattered light impinging on the PMS photodiode. The
- primary purpose of the model was to predict how the PMS performance compares to other
- 397 instruments designed to measure the aerosol scattering coefficient, such as integrating
- 398 nephelometers. The model makes simplifying assumptions about the laser that allow the
- 399 application of Mie theory to the light scattered from particles in the laser. Details of the model
- are presented in the Appendix.
- 401 The equation describing the intensity of light scattered from a particle in the laser is (Middleton,
- 402 1952; Anderson et al., 1996)

403
$$I(\theta) = F_{dv} \beta_p(\theta) dv$$
 (2)

- where $I(\theta)$ is the intensity of light at angle θ scattered from a particle in the volume element dv
- (with units of W sr⁻¹); $\beta_p(\theta)$ is the volume scattering function (m⁻¹ sr⁻¹); F_{dv} is the incident laser
- flux density (W m⁻²) impinging on the volume element dy; and dy is the volume element within
- 407 the laser.
- 408 The volume scattering function for a single particle in the laser beam is a function of aerosol
- diameter D_p , complex refractive index m, laser wavelength λ , and scattering angle θ :

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$$\beta_p(\theta) = (\lambda/2\pi)^2 (1/dv) |S_1(m, \lambda, \theta, D_p)|^2$$
 (3)

- where $|S_1(m, \lambda, \theta, D_p)|^2$ is the Mie scattering intensity function for laser light polarized parallel to
- 412 the photodiode surface and perpendicular to the plane of incidence (Bohren and Huffman, 1983).
- The scattered light intensity from a single particle in the laser beam to a narrow strip across the
- 414 middle of the photodiode and from all positions in the scattering volume is integrated to predict
- 415 the total power received by the photodiode as a function of particle diameter D_p and refractive
- 416 index m:

417
$$P(m,D_p) = K \int_{x=0}^{x=10mm} \int_{\theta_1(x)}^{\theta_2(x)} |S_1(m,\theta,D_p)|^2 \sin(\theta) d\theta dx.$$
 (4)

- 418 Due to the PMS geometry, the upper and lower angular scattering limits for θ depend on the
- 419 location, x, of a particle in the laser. Details are provided in the Appendix. This approach can be
- 420 used to estimate the amount of scattered energy detected from mixtures of particles of varying
- diameters and indices of refraction, as shown in Eq. (5)



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$$P = K \int_{Dp} \int_{x=0}^{x=10mm} \int_{\theta_1(x)}^{\theta_2(x)} |S_1(m,\theta,D_p)|^2 \sin(\theta) N(D_p, m) d\theta dx dD_p.$$
 (5)

3.1 Model predictions - Deviation from a perfect cosine response

424 As discussed above, the PMS has a photodetector that is about 1.8 mm below the laser, resulting 425 in forward scattering and backscattering truncation angles of 18 and 166 degrees, respectively. Furthermore, the photodetector is recessed 0.46 mm below the scattering chamber base. 426 427 Equation 4 is used to explore the deviation from a perfect cosine response resulting from the 428 truncated scattering volume and recessed detector. It is shown in Fig. 4. For these calculations, 429 $S_1(m,\theta,D_p)$ is set equal to 1, which corresponds to isotropic scattering or a volume scattering 430 function that is constant over all scattering angles. It is assumed that the detector has a 431 Lambertian response, i.e., the light detected is independent of the direction of the incident energy, which results in a detector cosine response. Figure 4 shows a perfect cosine response in 432 433 yellow, while the red line shows the deviation from a perfect cosine response due to angular 434 truncation. The blue line shows the effect of both angular truncation and an inset detector that is 435 0.46 mm below the chamber base. All curves have been normalized to one at 90 degrees.

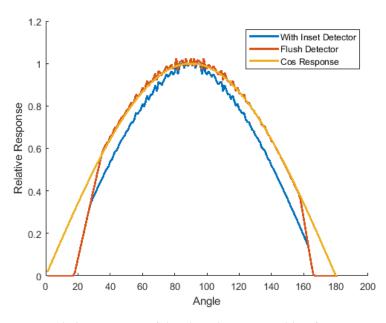


Figure 4. Relative response of the photodetector resulting from truncated scattering angles and a recessed photodetector. See explanation of the different curves in text.

3.2 Model predictions - Intensity versus position on the detector

Figure 5 provides an example of the energy distribution on the photodiode as a function of position in the laser and on the diode resulting from scattering from particles represented by a lognormally distributed aerosol volume size distribution with a volume mean diameter of 0.33 μm and geometric standard deviation of 1.7. Figure 5 shows model predictions of the relative





intensity of scattered light, where the values are proportional to energy flux impinging on the detector.

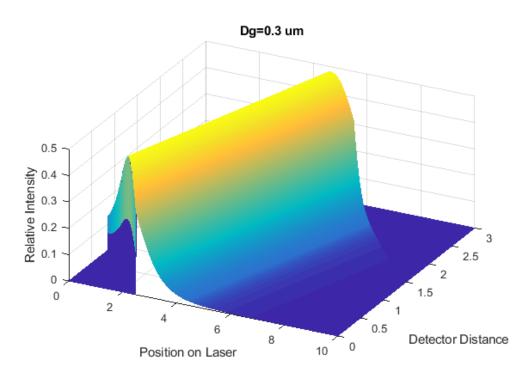


Figure 5. Relative intensity of radiant energy scattered by a lognormally distributed aerosol volume size distribution with a volume mean diameter of 0.33 µm and geometric standard deviation of 1.7 as a function of location of scattering event in the laser and as a function of position on the photodiode. Assumed laser wavelength was 650 nm, and the particle index of refraction was assumed to be 1.53. Positions in the laser and detector are from left to right as in Fig. 1.

The masking resulting from a recessed detector truncates the scattering both in the most forward and backward scattering angles. This masking is shown as the triangular area corresponding to distance down the laser and detector of 0.0-2.5 mm and 0.0-0.78 mm, respectively, for the forward scattering angles and 5.6-10 mm and 1.44-3.0 mm, respectively, for backscattering. Because the laser is parallel to the photodetector, which is assumed to have a $\cos(90-\theta)$ response, the maximum energy scattered to the detector is approximately at $\theta=90$ degrees. However, more energy is scattered to the detector for scattering angles less than 90 degrees, which corresponds to forward scattering, and very little energy is detected by the photodiode for particles in the laser that are greater than about 8 mm down the laser beam, even though the detector is exposed to particles in the laser that are 10 mm away from the laser exit hole. These distances down the laser correspond to backscattering. The total energy detected by the photodiode is the sum or integral across both the detector surface and position in the laser and corresponds to the volume under the curve depicted in Fig. 5.





3.3 Model predictions - Predicted photodiode response as a function of particle diameter

The PMS differs from a perfect nephelometer in at least five important ways:

- 1. The laser is polarized, whereas the nephelometer light source is unpolarized.
- 2. The laser beam profile is not a simple plane wave, but complex in shape. The laser beam profile evolves significantly as it is focused over the photodiode.
- 3. The photodiode likely does not have a perfect cosine response.
- 4. The PMS geometry limits the photodiode to receiving scattered light between approximately 18 and 166 degrees, whereas a perfect nephelometer measures all energy scattered between 0 and 180 degrees.
- 5. The unknown PMS signal processing removes the light scattering signal from CO₂, Suva, and filtered air. These gases are used to calibrate nephelometers, but cannot be used to calibrate the PMS.

The effects of these differences can be seen in Fig. 6, which shows predicted photodiode response as a function of particle diameter. The perfect nephelometer response is in blue, and the PMS response is in yellow. The red line predicts PMS response if the laser were not polarized. Relative intensities have been normalized to an ideal nephelometer measurement of a 0.1 µm diameter particle, which is akin to adjusting the laser power such that the scattered power at a diameter equal to 0.1 µm is the same for all configurations. Scattering as a function of particle diameter is nearly the same for all three configurations from 0.1 µm to about 0.3 µm. At about 0.8 to 1.0 µm, the response of a PMS with an unpolarized laser is about half that of an ideal nephelometer, and the use of a polarized laser reduces its response to about 30% to that of an ideal nephelometer. For particles above 2 µm in diameter, the PMS response compared to an ideal nephelometer is decreased by about 75%. Additionally, the PMS manual (Zhou, 2016) quotes a lower detection limit diameter of 0.3 µm. The model predicts that particles smaller than 0.3 µm in diameter would be detected by the PMS, in direct proportion to their contribution to the scattering coefficient.

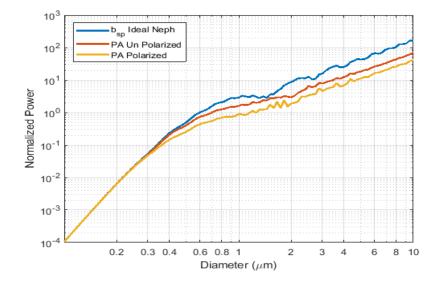






Figure 6. Normalized power detected by an ideal integrating nephelometer, a PMS with an unpolarized light source, and a PMS with a perpendicularly polarized light source plotted as a function of particle diameter. Modeled light source wavelength is 657 nm, and the particle index of refraction is 1.55. See explanation of the different curves in text.

These differences in geometry and optics from an ideal nephelometer are further highlighted in Fig. 7. To highlight the effect of polarization, the blue line shows the ratio of an ideal nephelometer with a laser light source that is perpendicularly polarized to an ideal nephelometer with an unpolarized light source. The red line shows the effects that polarization and PMS geometry have on the measured scattering signal. Again, all hypothetical instrument responses have been normalized to a particle diameter of 0.1 μ m. Relative to scattering for a 0.1 μ m particle, the polarization alone reduces the scattering signal of an ideal nephelometer by 40% for particles with diameters in the 0.8–1.5 μ m size range. The additional effect of PMS scattering geometry reduces the scattering signal at 0.8–1.0 μ m by about another 30% relative to an ideal nephelometer.

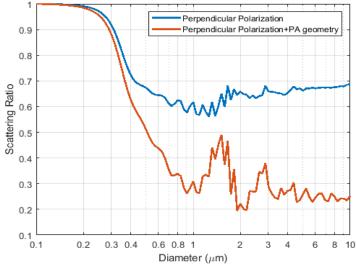


Figure 7. Ratio of scattering of a "perfect" nephelometer to a nephelometer with a light source that is perpendicularly polarized (blue) and to a perpendicularly polarized nephelometer with PMS geometry (red) as a function of particle diameter.

3.4 Model predictions – Differentiating by particle size

The irradiance received by the PMS photodiode from a particle of a given diameter and refractive index depends on the particle's location in the laser beam. The model predicts that particles of different sizes may contribute the same irradiance to the photodiode, depending on their location in the beam, or conversely, light scattered by a particle of a given size can vary by more than an order of magnitude.

As an example, the model predicts that all of the particles in Fig. 8 contribute the same irradiance to the PMS photodiode. The smaller particles contribute the same irradiance by scattering in the





more effective forward scattering regime. The larger particles contribute the same irradiance by scattering in the less effective backscattering regime. The photodiode and its associated electronics would not be able to differentiate between them. As a result, the model predicts that the values reported in the six PA-PMS particle size channels from $>0.3 \mu m$ to $>10 \mu m$ cannot correctly represent the aerosol size distribution.

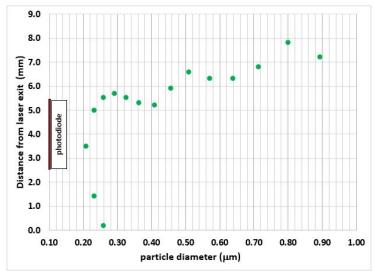


Figure 8. The model predicts that different size particles can generate the same irradiance on the photodiode, depending on their location in the laser beam. In this example, each of the particles would create 1.7×10^{-2} picowatts of scattered irradiance on the photodiode.

4. Experimental – Field studies

Field experiments were conducted at two of the NFAN aerosol monitoring stations: the Mauna Loa Baseline Observatory in Hawaii and the Table Mountain Test Facility in Colorado. Both sites have large suites of aerosol instrumentation and daily access for scientists and technicians to inspect, calibrate, and maintain the instruments. These sites also have integrating nephelometers (TSI 3563, St. Paul, MN, USA) against which to evaluate the PA monitors.

4.1 Description of Mauna Loa site

The Mauna Loa Baseline Observatory (MLO) is located on the north flank of the Mauna Loa volcano, on the Big Island of Hawaii (19.536 °N, 155.576 °W, 3397 m asl). The observatory is a premier atmospheric research facility that has been continuously monitoring and collecting data on global background conditions and atmospheric change since the 1950s (https://www.esrl.noaa.gov/gmd/obop/mlo/). Continuous aerosol measurements at MLO began in the mid-1970s with the installation of condensation particle counters and an integrating nephelometer (Bodhaine and Mendonca, 1974; Bodhaine et al., 1981). MLO lies above the strong marine temperature inversion layer present in the region, which separates the more-polluted lower portions of the island atmosphere in the marine boundary layer from the much cleaner free troposphere. MLO experiences a diurnal wind pattern (Ryan, 1997) that is strongly influenced by the daily heating and nighttime cooling of the dark volcanic lava rock that makes





- 545 up the mountain. This "radiation wind" brings air up from lower elevations during the daytime,
- when atmospheric measurements reflect the local mountain environment. In contrast, during the
- 547 nighttime, downslope winds develop, and the measurements at MLO are typically dominated by
- clean, free-tropospheric conditions (Chambers et al., 2013). At these times, the aerosol
- 549 measurements at MLO often reflect some of the cleanest conditions at any station in the northern
- 550 hemisphere. It has long been known, however, that episodic long-range transport of Asian
- 551 pollution and dust aerosols occurs, most frequently in the springtime (Shaw, 1980; Miller, 1981;
- Harris and Kahl, 1990), and these aerosol events can influence both the daytime and nighttime
- measurements at MLO. Consequently, the aerosol levels at MLO vary over a large range, from
- extremely low to at times mildly elevated. Here we use observations from the MLO integrating
- nephelometer to evaluate the PMS sensor.

4.2 Description of Boulder Table Mountain site

- 557 The Table Mountain Test Facility (BOS) is a large restricted-access federal complex located 14
- 558 km north of Boulder, Colorado (40.125 °N, 105.237 °W, 1689 m asl). NOAA conducts
- atmospheric research at this site, and in addition to its NFAN station, it is one of the Global
- 560 Monitoring Laboratory's seven U.S. Surface Radiation Network (SURFRAD) sites
- 561 (https://www.esrl.noaa.gov/gmd/grad/surfrad/tablemt.html). Many instruments for measuring
- 562 surface and column aerosol properties are maintained at this location and used for long-term
- monitoring of the atmosphere.
- 564 The BOS site lies just east of the Front Range foothills of the Rocky Mountains and is typical of
- a semi-arid, high plains environment. It is a high mesa of predominantly grassland with some
- 566 desert scrub vegetation. The location is well suited for sampling of wildfire smoke plumes during
- 567 fire season in the western United States (summer and autumn), dust events at any time of the
- year, and occasional urban pollution episodes. The NFAN station at BOS
- 569 (https://www.esrl.noaa.gov/gmd/aero/net/bos.html) was completed in September 2019. BOS
- 570 operates an integrating nephelometer and a differential mobility particle spectrometer (DMPS).
- 571 Both provided useful data for evaluating some of the predictions from the physical-optical model
- we developed for the PMS sensor.

4.3 PA monitors

- 574 PA-PMS monitors were installed on the aerosol towers at the MLO and BOS stations, just below
- 575 the main aerosol inlets. MLO had two PA-PMS monitors, one gently heated and one unheated,
- whereas BOS had one gently heated PA-PMS monitor. Prior to deployment, the monitors were
- 577 tested in a filtered air chamber for 4 hours to ensure that the 1-h average CH1 values were less
- 578 than 1 when no particles were present. One of the PMS sensors in the unheated MLO PA had 1-h
- 579 average CH1 values of 27 when no particles were present. The heated monitors were wrapped
- 580 with heating tape and powered by small DC power supplies. All the monitors were covered with
- 581 stainless steel flashing 5 cm below the bottom to prevent rain and snow from entering the inlet
- 582 (Fig. S13).

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- 583 The PA-PMS monitors were warmed in an effort to reduce the sample relative humidity (RH) to
- 584 be closer to that of the nephelometer, which is unavoidably heated by the warmth of the
- 585 laboratory and by the nephelometer's halogen lamp to above ambient temperatures. Because of
- 586 this warming, the RH inside the nephelometers rarely exceeded 40%. Both MLO and BOS are





- low-RH environments under normal conditions, although occasionally moist air masses are
- encountered. The heating of the monitors increased the sample temperatures by 5–8 C, which
- 589 helped to lower the sample RH. While the PA heating was not controlled to achieve an RH
- match with the nephelometer, it brought the sample RH of the two measurements closer together.
- Due to internet protocols at both sites, PA's wireless data transmission feature was not used, and
- 592 the data were stored on the internal micro-SD card. At approximately 1 month intervals, the data
- 593 were downloaded from the micro-SD cards, and the PAs were returned to service. Outputs from
- 594 the two PMS sensors were then compared at these intervals to determine if the PAs were still
- 595 functioning properly. In this study, the 80 s or 2 min averages were used to create 1 h averages to
- 596 compare the PA observations to those of the nephelometer and the DMPS.

4.4 Integrating nephelometer

- 598 The integrating nephelometer (TSI Inc., model 3563) measures the aerosol light scattering
- 599 coefficient at three wavelengths (450, 550, and 700 nm). At both sites, the sample flow path is
- 600 switched every 6 minutes between 1 and 10 µm aerodynamic diameter, multijet, Berner-type
- 601 impactors. Here, the scattering coefficients at 550 nm for both the PM1 and PM10 size fractions
- are used for comparison with the PMS measurements. These are referred to as b_{sp1} and b_{sp10},
- 603 respectively.

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- There are two quality checks of the nephelometer operation made in the field. First, the
- 605 nephelometer automatically samples filtered air once per hour. This provides a record of the
- stability of the instrument background measurement. Second, the nephelometer calibration is
- 607 manually checked on a monthly basis using CO₂ and filtered air (Anderson et al., 1996). The 1 h
- average b_{sp1} in filtered air is 0.01 Mm⁻¹ with a standard deviation of 0.12 Mm⁻¹, based on 125
- 609 hours of sampling filtered air.
- 610 The nephelometer measurements were corrected for angular truncation (Anderson and Ogren,
- 611 1998) and reported at STP. Weekly data review provides quality assurance of the nephelometer
- data. Scattering coefficient data were averaged to 1 min resolution for logging and were further
- averaged to hourly resolution for comparison with the PA data. The 1 h average b_{sp1} uncertainties
- of the nephelometer measurements are ~0.13 Mm⁻¹ for scattering coefficients less than 1.0 Mm⁻¹
- and ~10% for scattering coefficients greater than 1 Mm⁻¹ (Sherman et al., 2015).

4.5 Differential mobility particle spectrometer (DMPS)

- The DMPS was provided by the Institute for Atmospheric and Earth System Research,
- 618 University of Helsinki, Finland. It was checked and calibrated by the World Calibration Centre
- 619 for Aerosol Physics (WCCAP) at Leibniz Institute for Tropospheric Research (IfT), Leipzig,
- 620 Germany, just prior to deployment at NOAA's Table Mountain site. After shipment from IfT to
- NOAA, the DMPS was again checked by aerosolizing polystyrene latex spheres and confirming
- that the peaks occurred in the correct size bins. The DMPS was housed inside the same building
- as the nephelometer at BOS and sampled aerosols through the same inlet, although the DMPS
- flow did not pass through the aerosol impactors.
- The DMPS provides 40 channels of particle concentration versus size, ranging from mobility
- diameters of $0.01 \mu m$ to $0.8 \mu m$. The $0.1 \mu m$ to $0.8 \mu m$ channels of the DMPS were used to





- 627 calculate hourly-average fine aerosol scattering coefficient distributions and the total fine aerosol
- scattering coefficient. The hourly-average, DMPS-calculated fine aerosol scattering coefficients 628
- 629 were compared to the nephelometer-measured fine aerosol scattering coefficients to check
- operational consistency (Fig. S14). No operational changes were made to the DMPS during this 630
- field study. This study did not measure coarse aerosol size distributions. The DMPS hourly-631
- average fine aerosol scattering coefficient distributions were used with the PMS physical-optical 632
- model to predict total 1 h average scattered irradiance on the photodiode. 633

5. Results

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- 635 This section describes our evaluation of the PA-PMS using field data from MLO and BOS. First,
- 636 we provide an overview of the observational data. We then assess how well the model described
- in Sect. 3 is able to represent the observed data and show consistency with results previously 637
- reported in the literature. Next, we present results showing the potential of the PA-PMS to 638
- 639 perform as a nephelometer. Finally, we note how the size information output by the PA is not
- 640 correct due to the PA's primary measurement being a scattering measurement. For the results
- 641 presented below, data from the PA-PMS, nephelometer and DMPS were averaged to hourly
- 642 frequency and merged prior to analysis.

643 5.1 Field data overview

- 644 Heated PA monitors were deployed at the MLO and BOS observatories for 15 and 11 months,
- 645 respectively (Table 1). At both sites weather had no impact on the operation of the PA
- 646 instrument, and downtime only occurred during data downloading.

Table 1. Summary of PA, TSI nephelometer, and DMPS data coverage.

	number of hours				percent coverage			
site	PA-PMS	TSI neph	DMPS	overlap	PA-PMS	TSI neph	DMPS	Time period
MLO	9371	9204	na	9204	97.6	95.9	na	6 May 2019 to 5 June 2020
BOS	7716	7479	7045	6901	97.7	94.7	89.2	13 February 2020 to 6 January 2021

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These two deployments provide an excellent dataset for assessing PA performance in both a clean location (MLO) and in an environment with more elevated particle concentration (BOS).

650 As shown in Table 2, during the field study at MLO, the median CH1 was 26.7. The median b_{sp1} 651

was 0.76 Mm⁻¹ at 550 nm which is approximately 10% of Rayleigh scattering at the MLO 652

altitude. The reported PM2.5 mass concentration from the PA was zero for most of the MLO deployment. The CH1 and b_{sp1} are adjusted to STP in Table 2. The air quality at BOS was less 654

655 pristine than at MLO and is more representative of nonurban continental air quality. The very

high maximum CH1 and $b_{\rm sp1}$ at BOS reported in Table 2 occurred during smoke events in the 656

summer and autumn of 2020. One of the BOS PMS sensors experienced approximately 10% 657

658 degradation in sensitivity after one year in the field (Fig. S15).





Table 2. Summary of PA-PMS and nephelometer hourly observations at MLO and BOS.

		1h median	(average)		1h Range min-max			
	PA-PMS		TSI nephelometer		PA-PMS		TSI nephelometer	
Site	PM2.5 μg m ⁻³	СН1	b _{sp10} Mm ⁻¹	$b_{sp1} \ Mm^{-1}$	PM2.5 μg m ⁻³	СН1	b_{sp10} Mm^{-1}	$b_{sp1} \ Mm^{-1}$
MLO	0.000 (0.12)	26.7 (75.2)	1.19 (2.82)	0.76 (1.50)	0.0 - 21.6	0.26 - 1649	-0.35 - 35.2	-0.29 - 34.2
BOS	3.37 (8.42)	720 (1422)	14.6 (32.4)	9.9 (20.9)	0.0 - 571	7.38 - 63340	-0.11 - 4097	-0.44 - 2596

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5.2 Relationship between model predictions and field data

The PMS sensor is described by the manufacturer as a particle counter that measures particles between 0.3 μm and 10 μm in six size bins. Based on the theoretical characterization of the PMS sensor described in Sect. 3, the sensor is more akin to a polarized, reciprocal integrating nephelometer than a particle counter. Below, the field data and theoretical model are used to demonstrate that the raw PMS sensor signal is an integrated scattering measurement that is sensitive to particles smaller than 0.3 μm but relatively insensitive to particles larger than 1.0 μm .

669 μm

5.2.1 Predicted photodiode irradiance versus CH1 field data at BOS

671 Our model described in Sect. 3 and the Appendix predicts a value proportional to the scattered 672 irradiance impinging on the PMS photodiode as a function of particle diameter and concentration. This was done using the DMPS size distribution data from BOS. The modeled 673 674 PMS photodiode output is plotted against the PMS CH1 output (Fig. 9). The predicted photodiode output is linearly correlated ($R^2 = 0.90$) with CH1 over 4 orders of magnitude. The 675 676 linear relationship between CH1 and modeled photodiode response suggests the likelihood that 677 the CH1 output is directly related to what the photodiode is sensing (i.e., scattering from all particles in the scattering volume). The PA-PMS reported values, such as concentrations of 678 679 particle numbers in various size ranges or PM concentrations, are quantities derived from the 680 scattering signal and the use of an undescribed algorithm.



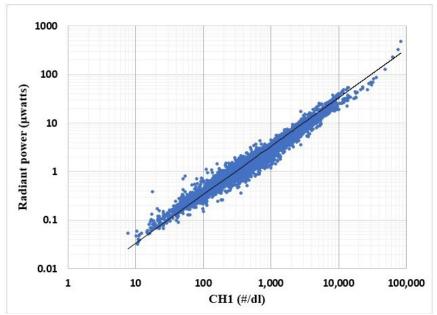


Figure 9. One hour average CH1 reported values are plotted against model-predicted radiant power (or energy) in µwatts on the photodiode. Both the CH1 and DMPS data were adjusted to STP conditions. The ordinary least squares regression line is also shown.

5.2.2 Predicted aerosol size truncation versus field data at BOS

The PMS physical-optical model described in Sect. 3 predicts that if CH1 is proportional to the photodiode power, then its signal will be truncated relative to a perfect nephelometer. Thus, the ratio CH1/b_{sp1} should decrease as median scattering diameter increases. To test this prediction, the DMPS data from BOS were used to calculate hourly-average aerosol scattering coefficient distributions for diameters between 0.1 μm and 0.8 μm . A wavelength of 657 nm and a particle refractive index of 1.53-0.0i were used for the calculations. The median scattering diameter (MSD) was calculated for each hour. The MSD is the aerosol diameter at which approximately half of the light scattering is due to particles smaller than the MSD and the other half to particles larger than the MSD. The MSD was then compared to the ratio of the measured CH1 and $b_{\rm sp1}$ values, i.e., CH1avg/b_{sp1}, for each of these hours. The results shown in Fig. 10 are consistent with the PMS physical-optical model. The highest CH1avg/b_{sp1} ratios tend to occur for aerosols with the lowest MSD and decrease as MSD increases. The results show, as suggested above, that the PMS can efficiently detect particles below 0.3 μm in diameter in proportion to their contribution to the scattering coefficient.



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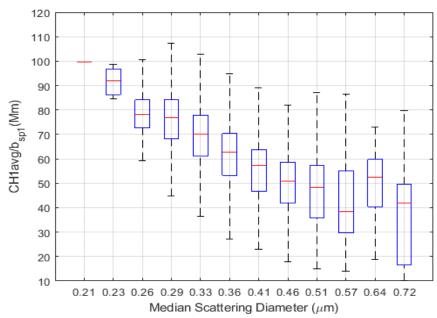


Figure 10. Observed decrease in CH1avg/b_{sp1} ratio as function of median scattering diameter. Red line is the median value, while the upper and lower edges of the blue box represent the 75th and 25th percentile values, respectively. Whiskers extend to the 9th and 91st percentiles. Overall average CH1avg/b_{sp1} ratio for all hours is 65. Data are for 6839 hourly averages at BOS. Approximately 67% of the MSDs observed at BOS were between 0.29 μ m and 0.36 μ m, and 98% of MSDs were between 0.26 μ m and 0.46 μ m.

5.2.3. Estimating the scattering coefficient minimum detection limit of the PA-PMS

The precision analysis in Sect. 2 indicates that the PA monitors used in this study estimated 1 h average CH1 and CH1avg MDLs of approximately 21 and 14, respectively. The estimated 1 h average MDL b_{sp1} of the TSI 3563 nephelometer is approximately 0.20 Mm⁻¹, based on filtered air tests. Further analysis of the relationship between CH1 and b_{sp1} at low levels was performed by plotting the ratio, CH1avg/b_{sp1}, for the combined MLO and BOS dataset, as a function of b_{sp1}. This relationship is shown graphically in Fig. 11. The data values were first averaged over 6 hours because hourly b_{sp1} values near zero included many small negative b_{sp1} values due to the very clean conditions occasionally observed at MLO. The averaging eliminated all but five negative b_{sp1} values, which were removed from the dataset. The CH1avg/b_{sp1} and b_{sp1} values were further averaged over six data points after sorting the data on b_{sp1} levels to more clearly show the relationship between CH1avg and b_{sp1} . At $b_{sp1} > 5$ Mm⁻¹, the CH1avg/ b_{sp1} ratio is relatively constant at 67, the yellow line in Fig. 11. The yellow line is the slope of CH1avg versus b_{sp1} at b_{sp1} values greater than 5 Mm⁻¹. The CH1avg/ b_{sp1} ratio systematically decreases from its highest values to about 35, the slope of CH1avg versus b_{sp1} at $b_{sp1} = 0.4$ Mm⁻¹. For b_{sp1} < 0.4 Mm⁻¹ the CH1avg/b_{sp1} ratio then increases significantly as b_{sp1} decreases, consistent with CH1avg values staying approximately constant below 0.4 Mm⁻¹. Both the CH1avg and b_{sp1} are





- below MDL for $b_{sp1} < 0.2 \text{ Mm}^{-1}$. A CH1avg/ b_{sp1} ratio of approximately 35 at $b_{sp1} = 0.4 \text{ Mm}^{-1}$ and a CH1avg value of about 14±5 is consistent with the estimated CH1avg MDL of 14.
- Based on these results, the 1 h average CH1 sensor MDL for hourly data in units of scattering is approximately 0.4 Mm⁻¹ at MLO. Laboratory tests challenging the PAs with known low-level, spiked aerosol concentrations and defined size distributions are needed to further refine the estimated MDL.

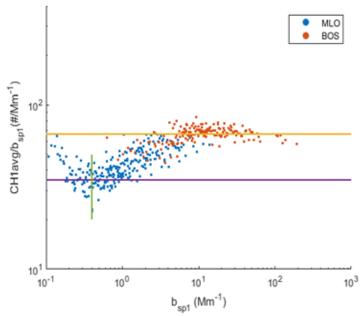


Figure 11. Ratios of CH1avg and measured scattering, b_{sp1} , as a function of measured b_{sp1} for MLO and BOS. Green line corresponds to 0.4 Mm⁻¹ while the purple line, a ratio of 35, corresponds to the additive uncertainty of 14. Yellow line corresponds to a CH1avg/ b_{sp1} ratio of approximately 67, the slope of CH1avg vs. b_{sp1} above about 5 Mm⁻¹.

5.2.4 Evaluating the use of the PA-PMS as a nephelometer

The MLO and BOS hourly-average CH1avg are plotted against b_{sp1} , measured at 550 nm, in Fig. 12. Also shown in Fig. 12 is an ordinary least squares (OLS) regression line with the intercept set equal to zero using the BOS and MLO combined dataset but with values associated with b_{sp1} less than 0.4 Mm⁻¹ and greater than 500 Mm⁻¹ removed. Results of the regression for the combined datasets as well as for the individual BOS and MLO datasets are presented in Table 3. There is good agreement for both datasets (Table 3) with an R^2 of 0.97 and 0.85 for the BOS and MLO datasets, respectively, and 0.97 for the combined datasets. The relationship deviates somewhat from linear with increasing slopes and scatter at lower values of atmospheric scattering coefficient, particularly for the MLO data. The slopes for all data, MLO, and BOS, are $0.015\pm2.07\times10^{-5}$, $0.017\pm5.72\times10^{-5}$, and $0.015\pm2.68\times10^{-5}$, respectively. In the following analysis, a PA-derived atmospheric scattering ($b_{sp1,PA}$) for both MLO and BOS is estimated using $b_{sp1,CH1}=0.015\times CH1$ avg at a wavelength of 550 nm. The "calibration" value of 0.015 corresponds to the





yellow horizontal line in Fig. 11 of 67.0 (1/0.015) and corresponds to a median scattering diameter of about 0.33 μm (Fig. 10).

Figure S16 shows that the submicron aerosol scattering coefficients at 550 nm and 700 nm are highly correlated, with the 700 nm scattering coefficient averaging 52% of the 550 nm scattering coefficient. This results in $b_{\rm sp1,CH1} = 0.0078 \times \text{CH1avg}$ at a wavelength of 700 nm.

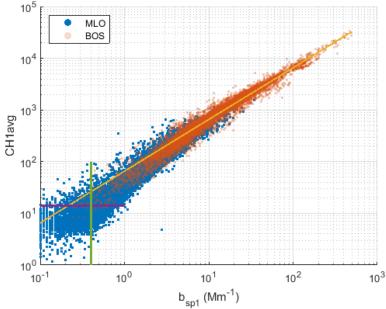


Figure 12. Fine aerosol scattering coefficient from TSI nephelometer vs. CH1avg value from PA. Yellow line represents the fit to all data. Purple line shows the additive uncertainty of 14 while b_{sp1} values less than the green line were removed for the regression analysis.

Table 3. Ordinary least square regression coefficients with a zero intercept and standard error for b_{sp1} and CH1 as the dependent and independent variables, respectively, for the BOS, MLO, and combined datasets. CH1 and b_{sp1} reported at STP. Also shown are the respective coefficients of determination, R^2 .

Site	slope	standard error	1/slope	\mathbb{R}^2
BOS	0.015	2.68×10 ⁻⁵	67.0	0.97
MLO	0.017	5.72×10 ⁻⁵	59.0	0.85
Both BOS&MLO	0.015	2.07×10 ⁻⁵	67.0	0.97

As discussed above, the regression coefficient between b_{sp1} and CH1 for the combined dataset of 0.015 is used to estimate the $b_{sp1,PA}$ derived from the CH1 channel. The data for each dataset and the combined dataset were binned into ten bins based on measured b_{sp1} levels that ranged from 0.4 Mm⁻¹ to 500 Mm⁻¹. Values of b_{sp1} above 500 Mm⁻¹ were removed from the dataset. For each





bin the normalized root mean square error (NRMSE) between b_{sp1,PA} and measured b_{sp1} was calculated. The NRMSE values as a function of the b_{sp1} bins are plotted in Fig. 13 for the combined dataset represented as the gray bars and BOS and MLO represented by blue and orange bars, respectively.

For b_{sp1} levels less than 0.8 Mm⁻¹, the NRMSE is 45–55%, and for b_{sp1} levels greater than 10 Mm⁻¹, the NRMSE is about 25% or less. For b_{sp1} levels greater than 60 Mm⁻¹, the NRMSE approaches 15%.

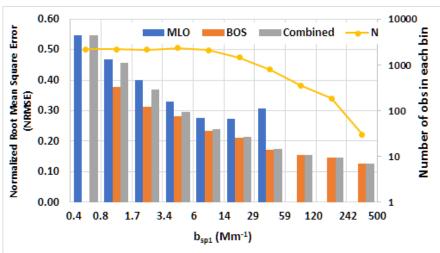


Figure 13. Normalized root mean square error between measured and estimated scattering from CH1 values plotted as a function of binned b_{sp1} for the BOS, MLO, and combined datasets. Yellow line is referenced to the right axis to provide the number of observations in each bin. Numbers on the x-axis represent the lower and upper levels of each scattering bin.

As discussed in Sect. 2.2.9, the uncertainty for high CH1avg values is small (1.9% to 4.8%). The precision of the TSI 3563 nephelometer is also similarly high, and together they account for about 10% NRMSE at high b_{sp1} values.

The overall normalized error is likely due to a variety of sources, primarily the variability in the CH1 values due to using a polarized light source and truncation errors due to the geometry of the PA-PMS sensors. Also, the variability in aerosol characteristics such as size distribution, refractive index, and shape may be important. At extremely low levels, uncertainty may also be due to a nonuniform distribution of particles in the PMS laser beam.

5.2.5 PA-PMS size distributions and PM2.5

The aerosol number concentrations from the six PMS size channels are unrealistic. The BOS field data showed that the concentration of particles larger than 0.3 µm diameter calculated from the DMPS averaged 10 times higher than CH1 (Fig. S17). The other PMS size channels are so highly correlated with CH1 that they provide no additional information (Table S3). Furthermore, it appears that the PMS creates an approximately invariant normalized aerosol number distribution across a wide range of sites (Table S4, Fig. S18). Although the overall CH1





- concentration can vary over 6 orders of magnitude (column 3 in Table S4), the shape of the PMS
- size distribution remains fairly constant. This suggests that the values in the channels above CH1
- are software generated and indicates that the most relevant output from the PMS is from the CH1
- 796 channel. The bottom row of Table S4 shows that the PMS bin fractions above 1 μm increased by
- only a factor of 2–5 in a high-PM2.5 windblown dust episode at Keeler, California. This is
- 798 consistent with the PMS model prediction that PMS coarse aerosol response is small relative to a
- 799 perfect nephelometer.
- 800 The PM2.5 was not measured by Federal Reference Method (FRM) or Federal Equivalent
- 801 Method (FEM) instruments at MLO and BOS during this study. As a result, this study cannot
- 802 compare the PA-PMS PM2.5 channel results with these methods. Figure S19 shows that the PA-
- PMS PM2.5 channel is reasonably well correlated with b_{sp1} for values greater than about 10–20
- 804 μg m⁻³, typical of many moderately polluted locations, with a calculated mass scattering
- efficiency of approximately 2.5 m² g⁻¹. However, it is likely that the PA-PMS underestimates
- PM2.5 for very clean areas where b_{sp1} is often less than 10 Mm⁻¹. For example, the PA-PMS
- PM2.5 was zero for 1099 of the hours in this study when b_{sp1} was greater than 1 Mm⁻¹.
- The results above indicate that CH1 is the primary source of aerosol information from the PMS
- 809 sensor. However, consistent with the sensor behaving like a cell-reciprocal nephelometer, it was
- 810 found that CH1 was not the number concentration of particles having diameters greater than 0.3
- 811 µm. CH1 was approximately a factor of 10 lower than the DMPS number concentration for a
- similar size range.

6. Summary, Discussion, and Future Work

- We have demonstrated that the PMS sensor inside the PA monitor (PA-PMS) appears to behave
- as an imperfect, reciprocal integrating nephelometer. As a scattering sensor, the PMS cannot
- directly count nor size particles in the air stream. The PMS uses an unknown algorithm to
- convert the scattering signal to a near-constant normalized number distribution from which PM
- 818 concentrations are derived.
- The scattering coefficient that is measured by an ideal integrating nephelometer does not need
- 820 correction for any aerosol attributes such as shape, chemical composition, refractive index, or
- 821 diameter. It is a valuable measure for visibility and global climate monitoring. Simple low-cost
- sensors such as the PA-PMS can play a role in estimating aerosol scattering coefficients and
- 823 improving global coverage. Yearlong field data at NOAA's Mauna Loa Observatory and
- 824 Boulder Table Mountain sites show that the 1 h average of the PA-PMS CH1 is highly correlated
- with a nephelometer-measured fine aerosol scattering coefficient at 550 nm, b_{sp1}, over a wide
- scattering coefficient range of 0.4 Mm⁻¹ to 500 Mm⁻¹. The relationship between CH1 and b_{sp1} at
- 827 550 nm is found to be $b_{sp1} = 0.015 \times CH1$ when both quantities are adjusted to the same
- 828 temperature and pressure.
- The physical-optical model developed in this paper for the PMS and the agreement with field
- data may motivate users of other low-cost sensors to develop similar models. It is possible that
- some of the other low-cost sensors also use polarized lasers in a cell-reciprocal configuration like
- the PMS. Such models would improve the understanding of sensor operation and help users
- better recognize the opportunities and limitations of other low-cost sensors in applications such
- as monitoring the scattering coefficient.





data support characterizing the PA-PMS as an imperfect truncated cell-reciprocal nephelometer. 836 837 The results demonstrate that it is possible to use the PA-PMS to estimate the 1 h average fine 838 aerosol scattering coefficient across a wide range of aerosol scattering concentrations, provided 839 the aerosol median scattering diameter is between 0.26 µm and 0.46 µm. The CH1 and b_{sp1} 840 relationship is dependent on the size distribution, and it is expected to change for locations and 841 times where the particle size shifts to larger or smaller sizes than those measured at BOS and 842 MLO. 843 We found that the PA-PMS has important limitations compared to integrating nephelometers. It 844 measures the light scattering over a smaller angular range, causing a significant truncation of the scattering signal in the forward and backward directions. Nephelometers calibrate their scattering 845 846 coefficient with CO₂ or Suva, but the PMS is unresponsive to these gases. As a result, there is 847 currently no convenient way to calibrate the PMS to ensure its accuracy. Neither PA nor 848 Plantower provide technical support. Quality assurance and control are not as robust as one 849 encounters for regulatory and scientific monitoring instruments. For this reason, it is useful to 850 test the PMS sensors in filtered air before using them and to limit field use to those sensors that 851 have 1 h average CH1 values less than 2. While sampling, it is necessary to compare 1 h averages from the two PMS sensors in each PA monitor to become aware of any changes and, if 852 853 needed, to replace them in a timely fashion. 854 This study limited its findings to low-RH air, because both the PA monitors and the 855 nephelometers were heated to reduce RH. Since RH plays such an important role in water uptake 856 by hygroscopic aerosols and the concomitant increase in the scattering coefficient, future work is 857 planned to compare unheated PA monitors with an unheated nephelometer that does not reduce 858 RH before sampling. Our model predicts that the PMS may not be as responsive to hygroscopic 859 growth as an unheated nephelometer. This would imply that the PMS might underestimate high-860 RH, low-visibility aerosol scattering coefficients. 861

The strong relationship between b_{sp1} and CH1 and the agreement between the model and field





- 862 Appendix
- The PMS physical-optical model makes some simplifying assumptions. The actual PMS laser
- 864 beam profile is not a simple plane wave but complex in shape. The model assumes the laser is a
- plane wave with a constant laser beam irradiance profile. This allows the use of Mie theory to
- 866 predict the light scattered by particles in the laser. Secondly, the model calculates the light
- 867 scattered to a narrow strip across the middle instead of the entire photodiode. It assumes that the
- irradiance received by the narrow strip is representative of the entire photodiode.
- The intensity of light scattered from a particle in the laser is

870
$$I(\theta) = F_{dv} \beta(\theta) dv$$
 (A1)

- where $I(\theta)$ is the intensity of light at angle θ scattered from a particle in the volume element dv
- (with units of Watt sr⁻¹); $\beta(\theta)$ is the volume scattering function (m⁻¹ sr⁻¹); F_{dv} is the incident laser
- 873 flux density (Watt m⁻²) impinging on the volume element dv; and dv is the volume element
- within the laser.
- The volume scattering function for a monodisperse aerosol having a diameter D_p and number
- 876 concentration $N(D_p)$ in the PMS laser is

877
$$\beta(m, \lambda, \theta, D_p) = (\lambda/2\pi)^2 N(D_p) |S_1(m, \lambda, \theta, D_p)|^2$$
(A2)

- where $|S_1(m, \lambda, \theta, D_p)|^2$ is the perpendicular scattering intensity function; λ is the laser
- wavelength; m is the particle complex refractive index; θ is the scattering angle; and D_p is the
- aerosol diameter. Note that $\theta = 0$ in the direction of the laser, and $\theta = 90$ degrees perpendicular to
- the laser and photodiode.
- For one particle of size Dp in the volume element dv, $N(Dp) dv = (1/dv) \times (dv) = 1$.
- The incremental power dP (Watt) scattered from a particle in the volume element dv across a
- 884 solid angle $d\Omega$ subtended on the surface of a sphere at distance r from the particle, and normal to
- 885 r, is

886
$$dP = I(\theta) d\Omega$$
. (A3)

- 887 $d\Omega = dA_0 / r^2$, where dA_0 is the incremental area on the sphere at distance r from the particle and
- 888 normal to r. dP is then
- 889 $dP = I(\theta) dA_0 / r^2$.
- For the PMS model, dA_0 is a small rectangle with width w and height $rd\theta$, where w is the width
- of the strip on the photodiode, and $d\theta$ is the differential scattering angle.
- 892 $dA_0 = r d\theta \times w$, where w is the width of the strip on the photodiode. From Fig. A1, $r = b/\sin(\theta)$,
- where b is the distance from the laser to the photodiode.





- 894 $d\mathbf{\Omega} = dA_0 / r^2 = (r d\theta \times w) / r^2 = d\theta \times (w/r) = (w/b) \times \sin(\theta) d\theta.$
- The incremental power across the solid angle $d\Omega$ normal to r is then

896
$$dP = I(\theta) \times dA_0 / r^2 = I(\theta) \times (w/b) \times \sin(\theta) d\theta.$$
 (A4)

897 Substituting for $I(\theta)$,

898
$$dP(g,x,\theta) = [F_0(\lambda/2\pi)^2 |S_1(\theta,D_p)|^2] \times (w/b) \times \sin(\theta) d\theta.$$
 (A5)

- Equation A5 can be further simplified by combining the constants into $K = (\lambda/2\pi)^2 F_0 w/b$, where
- 900 K has units of watts:

901
$$dP(g,x,\theta) = K |S_1(\theta,D_p)|^2 \sin(\theta) d\theta.$$
 (A6)

- The power received by the photodiode from a particle of diameter D_p in the volume element at x
- 903 is obtained by numerically integrating across θ on the photodiode:

904
$$P(m,D_p) = K \int_{\theta_1(x)}^{\theta_2(x)} |S_1(m,\theta,D_p)|^2 \sin(\theta) d\theta.$$
 (A7)

- Due to the PMS geometry, the upper and lower angular scattering integration limits for θ depend
- 906 on the location x. This can be seen in Fig. S4. For example, at x = 0 mm, the upper and lower
- integration limits for θ are 18 to 38 degrees. At x = 4.0 mm, over the center of the photodiode,
- 908 the angular integration limits are 50 to 130 degrees.
- 909 The total power P in Watts received by the photodiode from the light scattered by all the
- 910 particles of diameter D_p in the laser is obtained by carrying out the numerical integration in Eq.
- 911 A7 for all x from 0 to 10 mm:

912
$$P(m,D_p) = K \int_{x=0}^{x=10mm} \int_{\theta_1(x)}^{\theta_2(x)} |S_1(m,\theta,D_p)|^2 \sin(\theta) d\theta dx.$$
 (A8)

- 913 The result for carrying out this calculation for the power per particle of size D_p is in Table S5 for
- 914 wavelength 657 nm and particle refractive index 1.53 0.015i. The total power received at the
- 915 photodiode by a distribution of particles is obtained by summing up the power per particle of size
- D_p times the number of particles $N(D_p, m)$ in the size interval D_p to $D_p + dD_p$.

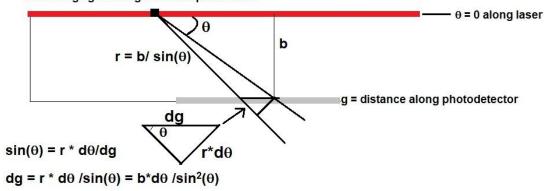
917
$$P = K \int_{Dp} \int_{x=0}^{x=10mm} \int_{\theta_1(x)}^{\theta_2(x)} |S_1(m,\theta,D_p)|^2 \sin(\theta) N(D_p, m) d\theta dx dD_p.$$
 (A9)

- Figure A1 shows the PMS geometry. The distance along the laser is the variable x, which ranges
- 919 from 0 to 10 mm. The distance along the photodiode is the variable g, which ranges from 0 to 3.0
- 920 mm. The distance between the photodiode and the laser is b, approximately 1.8 mm.





differential volume element at distance x along the laser, scattering light at angle $\boldsymbol{\theta}$ to the photodiode



921 **Figure A1.** Sketch of PMS5003 geometry.

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929 **Disclaimer**

- 930 The assumptions, findings, conclusions, judgments, and views presented herein are those of the
- 931 authors and should not be interpreted as necessarily representing NOAA or National Park
- 932 Service policies. Reference to any companies or specific commercial products does not
- 933 constitute endorsement by NOAA or the National Park Service.

934 References

- Abu-Rahmah, A., Arnott, W.P., and H Moosmüller, H. Integrating nephelometer with a low
- 936 truncation angle and an extended calibration scheme. Measurement Science and Technology,
- 937 Volume 17, Number 7. 2006. https://doi.org/10.1088/0957-0233/17/7/010.
- 938 Anderson, T. L. and Ogren, J. A. (1998) Determining Aerosol Radiative Properties Using the
- 939 TSI 3563 Integrating Nephelometer, Aerosol Science and Technology, 29:1, 57-69,
- 940 doi:10.1080/02786829808965551.
- 941 Anderson, T. L., Covert, D. S., Marshall, S. F., Laucks, M. L., Charlson, R. J., Waggoner, A. P.,
- 942 Ogren, J. A., Caldow, R., Holm, R., Quant, F., Sem, G., Wiedensohler, A., Ahlquist, N. A., and
- 943 Bates, T. S. (1996). Performance Characteristics of a High-Sensitivity, Three-Wavelength, Total
- 944 Scatter/Backscatter Nephelometer, J. Atmos. Oceanic Technol. 13: 967-986.





- 945 Andrews, E., Sheridan, P., Ogren, J.A., Hageman, D., Jefferson, A., Wendell, J., Alastuey, A.,
- Alados-Arboledas, L., Bergin, M., Ealo, M., Hallar, A.G., Hoffer, A., Kalapov, I., Keywood, M.,
- 947 Kim, J., Kim, S.-W., Kolonjari, F., Labuschagne, C., Lin, N.-H., Macdonald, A., Mayol-Bracero,
- O.L., McCubbin, I.B., Pandolfi, M., Reisen, F., Sharma, S., Sherman, J. P., Sorribas, M., Sun, J.,
- "Overview of the NOAA/ESRL Federated Aerosol Network" Bull. Amer. Meteor. Soc., 100,
- 950 123-135, doi:10.1175.BAMS-D-17-0175.1, 2019.
- Arnott, W. P.; Walker, J. W.; Moosmuller, H.; Elleman, R. A.; Haflidi, H.; Buzorius, G.; Conant,
- 952 W. C.; Flagan, R. C.; Seinfeld, J. H., (2006). "Photoacoustic Insight for Aerosol Light
- Absorption Aloft from Meteorological Aircraft and Comparison with Particle Soot Absorption
- 954 Photometer Measurements: The DOE Southern Great Plains Climate Research Facility and the
- 955 Coastal Stratocumulus Imposed Perturbation Experiments." Journal of Geophysical Research,
- 956 111, D05S02, doi:10.1029/2005JD005964.
- 957 Barkjohn, K.K., Gantt, B., and Clements, A.L., "Development and application of a United States
- wide correction for PM2.5 data collected with the PurpleAir Sensor," Atmos. Meas. Tech.
- 959 Discuss., [preprint], https://doi.org/10.5194/amt-2020-413, in review, 2020.
- Beuttell, R. G. and Brewer, A.W.: Instruments for the measurement of the visual range, J. Sci.
- 961 Instr. Phys. Ind., 26, 357–359, 1949.
- 962 Bodhaine, B.A. and Mendonca, B.G (1974) Preliminary four wavelength nephelometer
- measurements at Mauna Loa Observatory. Geophys. Res. Lett., 1, #3, 119-122.
- 964 Bodhaine, B.A., Mendonca, B.G., Harris, J.M. and Miller, J.M. (1981). Seasonal variations in
- 965 aerosols and atmospheric transmission at Mauna Loa Observatory. J. Geophys. Res. 86: 7395–
- 966 7398. https://doi.org/10.1029/JC086iC08p07395.
- Bohren, C.F. and Huffman, D.R. Absorption and Scattering of Light by Small Particles. 1983.
- 968 John Wiley & Sons.
- Butcher, S.S., and Charlson, R.J. (1972), An Introduction to Air Chemistry. Academic Press, pp.
- 970 256, https://doi.org/10.1016/B978-0-12-148250-3.X5001-X.
- 971 Chambers, S.D., Zahorowski, W., Williams, A.G., Crawford, J., and Griffiths, A.D. (2013),
- 972 Identifying tropospheric baseline air masses at Mauna Loa Observatory between 2004 and 2010
- 973 using Radon-222 and back trajectories, J. Geophys. Res. Atmos., 118, 992–1004,
- 974 doi:10.1029/2012JD018212.
- 975 Currie, L. "Limits for Qualitative Detection and Quantitation: Application to Radiochemistry",
- 976 Anal. Chem. 40, 586-593 (1968).
- 977 Friedlander, S.K., Smoke, Dust, and Haze: Fundamentals of Aerosol Behavior, 1977, Wiley-
- 978 Interscience. Page 75.
- 979 Gliss, J., Mortier, A., Schulz, M., Andrews, E., Balkanski, Y., Bauer, S.E., Benedictow, A.M.K.,
- 980 Bian, H., Checa-Garcia, R., Chin, M., Ginoux, P., Griesfeller, J.J., Heckel, A., Holben, B.N.,
- 981 Kinne, S., Kipling, Z., Kirkevag, A., Kokkola, H. Laj, P., Le Sager, P., Levy, R., Lund, M.T.,





- 982 Lund Myhre, C., Matsui, H., Myhre, G., Neubauer, D., van Noije, T., North, P., Olivie, D.J.L.,
- 983 Sogacheva, L., Takemura, T., Tsigaridis, K., Tsyro, S.G., "AeroCom phase III multi-model
- 984 evaluation of the aerosol lifecycle and optical properties using ground and space based remote
- sensing as well as surface in situ observations," *Atmos. Chem. Phys.*, 21, 87–128,
- 986 doi:10.5194/acp-21-87-2021, 2021.
- 987 Grinshpun, S., Willeke, K., and Kalatoor, S. A general equation for aerosol aspiration by thin-
- 988 walled sampling probes in calm and moving air. Atmospheric Environment. Part A. General
- 989 Topics. Volume 27, Issue 9, June 1993, Pages 1459-1470. https://doi.org/10.1016/0960-
- 990 1686(93)90132-I.
- Gupta, P., Doraiswamy, P., Levy, R., Pikelnaya, O., Maibach, J., Feenstra, B., et al. (2018).
- 992 Impact of California fires on local and regional air quality: The role of a low-cost sensor network
- and satellite observations. GeoHealth, 2, 172–181.
- Hagan, D. and Kroll, J. 2020. Assessing the accuracy of low-cost optical particle sensors using a
- 995 physics-based approach. Atmos. Meas. Tech., 13, 6343–6355, 2020.
- 996 Hangal, S. and Willeke, K. 1990. Aspiration Efficiency: A Unified Model for all Forward
- 997 Sampling Angles, ES&T, 24, 688-690.
- 998 Harris, J.M. and Kahl, J.D. (1990). A descriptive atmospheric transport climatology for the
- 999 Mauna Loa Observatory, using clustered trajectories. J. Geophys. Res. 95: 13651–13667.
- 1000 https://doi.org/10.1029/JD095iD09p13651.
- He, M., Kuerbanjiang, N. and Dhaniyala, S. Performance characteristics of the low-cost
- 1002 Plantower PMS optical sensor, Aerosol Science and Technology, 54(2), 232–241,
- 1003 doi:10.1080/02786826.2019.1696015, 2020.
- Heintzenberg, J., and Charlson, R. J. (1996). The Integrating Nephelometer: A Review, J.
- 1005 Atmos. Oceanic Technol., 13:987-1000.
- Heintzenberg, J., Wiedensohler, A., Tuch, T. M., Covert, D. S., Sheridan, P., Ogren, J. A., Gras,
- 1007 J., Nessler, R., Kleefeld, C., Kalivitis, N., Aaltonen, V., Wilhelm, R., & Havlicek, M. (2006).
- 1008 Intercomparisons and Aerosol Calibrations of 12 Commercial Integrating Nephelometers of
- 1009 Three Manufacturers, Journal of Atmospheric and Oceanic Technology, 23(7), 902-914.
- Hering, S.V., "Impactors, Cyclones, and Other Inertial and Gravitational Collectors," p 284, in
- 1011 Air Sampling Instruments for Evaluation of Atmospheric Contaminants, 8th ed, B.Cohen, S.V.
- 1012 Hering, Eds., American Conference of Governmental Industrial Hygienists, Cincinnati, OH
- 1013 (1995).
- Holder, A.L.; Mebust, A.K.; Maghran, L.A.; McGown, M.R.; Stewart, K.E.; Vallano, D.M.;
- 1015 Elleman, R.A.; Baker, K.R. Field Evaluation of Low-Cost Particulate Matter Sensors for
- 1016 Measuring Wildfire Smoke. Sensors, 2020, 20, 4796.
- 1017 Hyslop, N. and White, W., An empirical approach to estimating detection limits using collocated
- 1018 data. Environ. Sci. Technology. 2008, 42, 5235-5240.





- Hyslop, N., and White, W., Estimating precision using duplicate measurements. ISSN:1047-
- 1020 3289 J. Air & Waste Manage. Assoc., 59:1032–1039, DOI:10.3155/1047-3289.59.9.1032, 2009.
- Jayaratne, R., Liu, X., Ahn, K.H., Asumadu-Sakyi, A., Fisher, G., Gao, J., Mabon, A., Mazaheri,
- 1022 M., Mullins, B., Nyaku, M., Ristovski, Z., Scorgie, Y., Thai, P., Dunbabin, M. and Morawska, L.
- 1023 (2020). Low-cost PM2.5 Sensors: An Assessment of their Suitability for Various Applications.
- 1024 Aerosol Air Qual. Res. 20: 520-532. https://doi.org/10.4209/aaqr.2018.10.0390.
- 1025 JCGM100:GUM, "Guide to the Expression of Uncertainty in Measurement," International
- 1026 Standardization Organization: Geneva, Switzerland,
- 1027 https://www.bipm.org/en/publications/guides/gum.html, accessed 22 March 2021, 2008.
- 1028 Kelly, K.E., Whitaker, J., Petty, A., Widmer, C., Dybwad, A., Sleeth, D., Martin, R., and
- 1029 Butterfield, A., "Ambient and laboratory evaluation of a low-cost particulate matter sensor,"
- 1030 Environ. Pollut. 221, 491–500. https://doi.org/10.1016/j.envpol.2016.12.039, 2017.
- 1031 Kulkarni, P., Baron, P., and Willeke, K. Aerosol Measurement: Principles, Techniques, and
- 1032 Applications, 3rd Edition. Section 6.2, "Sample Extraction". John Wiley & Sons, ISBN: 978-0-
- 1033 470-38741-2 July 2011.
- 1034 Kuula, J., Makela, T., Aurela, M., Teinila, K., Varjonen, S., Gonzalez, O., and Timonen, H.:
- Laboratory evaluation of particle-size selectivity of optical low-cost particulate matter sensors,
- 1036 Atmos. Meas. Tech., 13, 2413-2423, https://doi.org/10.5194/amt-13-2413-2020, 2020.
- Laj, P., Bigi, A., Rose, C., Andrews, E., Lund Myhre, C., Collaud Coen, M., Wiedensohler, A.,
- 1038 Schulz, M., and 90 co-authors, "A global analysis of climate-relevant aerosol properties retrieved
- from the network of GAW near-surface observatories," Atmos. Meas. Tech., 13, 4353–4392,
- 1040 https://doi.org/10.5194/amt-13-4353-2020, 2020.
- Malings, Carl, Rebecca Tanzer, Aliaksei Hauryliuk, Provat K. Saha, Allen L. Robinson, Albert
- 1042 A. Presto & R Subramanian (2020) Fine particle mass monitoring with low-cost sensors:
- 1043 Corrections and long-term performance evaluation, Aerosol Science and Technology, 54:2, 160-
- 1044 174, DOI: 10.1080/02786826.2019.1623863.
- Malm, W.C., Sisler, J.F., Huffman, D., Eldred, R.A., Cahill, T.A., 1994. Spatial and seasonal
- trends in particle concentration and optical extinction in the United States. Journal of
- 1047 Geophysical Research 99 (D1), 1347–1370.
- 1048 Markowicz, K.M. and Chilinski, M.T., "Evaluation of Two Low-Cost Optical Particle Counters
- 1049 for the Measurement of Ambient Aerosol Scattering Coefficient and Ångström Exponent,"
- 1050 Sensors, 20, 2617; doi:10.3390/s20092617, www.mdpi.com/journal/sensors, 2020.
- 1051 Mehadi, Ahmed; Hans Moosmüller, David E. Campbell, Walter Ham, Donald Schweizer, Leland
- 1052 Tarnay & Julie Hunter (2020): Laboratory and field evaluation of real-time and near real-time
- 1053 PM2.5 smoke monitors, Journal of the Air & Waste Management Association, DOI:
- 1054 10.1080/10962247.2019.1654036.





- Middleton, W. E. K., (1952), in Vision Through the Atmosphere, pp. 203-206. The University of
- 1056 Toronto Press, Toronto, Canada.
- 1057 Miller, J.M. (1981). A five-year climatology of back trajectories from the Mauna Loa
- 1058 Observatory, Hawaii. Atmos. Environ. 15: 1553–1558. https://doi.org/10.1016/0004-
- 1059 6981(81)90138-4.
- 1060 Molenar, J. V.: Analysis of the real world performance of the Optec NGN-2 ambient
- 1061 nephelometer, Visual Air Quality: Aerosols and Global Radiation Balance, Air and Waste
- Management Association, Pittsburgh, 243–265, 1997.
- 1063 Morawska L, Thai PK, Liu X, Asumadu-Sakyi A, Ayoko G, Bartonova A, Bedini A, Chai F,
- 1064 Christensen B, Dunbabin M, Gao J, Hagler GSW, Jayaratne R, Kumar P, Lau AKH, Louie PKK,
- 1065 Mazaheri M, Ning Z, Motta N, Mullins B, Rahman MM, Ristovski Z, Shafiei M, Tjondronegoro
- 1066 D, Westerdahl D, Williams R. Applications of low-cost sensing technologies for air quality
- monitoring and exposure assessment: How far have they gone? Environ Int., 2018, 116:286-299.
- doi: 10.1016/j.envint.2018.04.018. Epub 2018 Apr 26. PMID: 29704807; PMCID:
- 1069 PMC6145068.
- 1070 Mulholland, G. W., and Bryner, N. P. (1994). Radiometric Model of the Transmission Cell-
- 1071 Reciprocal Nephelometer. Atmos. Environ., 28: 873-887.
- 1072 Müller, T., Laborde, M., Kassell, G., and Wiedensohler, A.: Design and performance of a three-
- 1073 wavelength LED-based total scatter and backscatter integrating nephelometer, Atmospheric
- 1074 Measurement Techniques, 4, 1291-1303, https://doi.org/10.5194/amt-4-1291-2011, 2011.
- 1075 Nakayama, T., Suzuki, H., Kagamitani, S., Ikeda, Y., Uchiyama, A., and Matsumi, Y. (2015).
- 1076 Characterization of a Three Wavelength Photoacoustic Soot Spectrometer (PASS-3) and a
- 1077 Photoacoustic Extinctiometer (PAX), J. Meteorol. Soc. Jpn., 93, 285–308.
- 1078 Naqwi, A. and F. Durst, 1990, Focusing of diode laser beams: a simple mathematical model,
- 1079 Applied Optics, Vol. 29, No. 12, 1780-1785.
- 1080 Pandolfi, M., Alados-Arboledas, L., Alastuey, A., Andrade, M., Angelov, C., Artiñano, B.,
- Backman, J., Baltensperger, U., Bonasoni, P., Bukowiecki, N., Collaud Coen, M., Conil, S., Coz,
- 1082 E., Crenn, V., Dudoitis, V., Ealo, M., Eleftheriadis, K., Favez, O., Fetfatzis, P., Fiebig, M.,
- 1083 Flentje, H., Ginot, P., Gysel, M., Henzing, B., Hoffer, A., Holubova Smejkalova, A., Kalapov, I.,
- 1084 Kalivitis, N., Kouvarakis, G., Kristensson, A., Kulmala, M., Lihavainen, H., Lunder, C., Luoma,
- 1085 K., Lyamani, H., Marinoni, A., Mihalopoulos, N., Moerman, M., Nicolas, J., O'Dowd, C., Petäjä,
- T., Petit, J.-E., Pichon, J. M., Prokopciuk, N., Putaud, J.-P., Rodríguez, S., Sciare, J., Sellegri, K.,
- 1087 Swietlicki, E., Titos, G., Tuch, T., Tunved, P., Ulevicius, V., Vaishya, A., Vana, M., Virkkula,
- 1088 A., Vratolis, S., Weingartner, E., Wiedensohler, A., and Laj, P.: A European aerosol
- phenomenology 6: scattering properties of atmospheric aerosol particles from 28 ACTRIS
- sites, Atmos. Chem. Phys., 18, 7877–7911, https://doi.org/10.5194/acp-18-7877-2018, 2018.
- 1091 Papapostolou, Vasileios, Hang Zhang, Brandon J. Feenstra, Andrea Polidori, Development of an
- 1092 environmental chamber for evaluating the performance of low-cost air quality sensors under





- 1093 controlled conditions, Atmospheric Environment, Volume 171, 2017, Pages 82-90, ISSN 1352-
- 1094 2310, https://doi.org/10.1016/j.atmosenv.2017.10.003.
- 1095 Pawar, H. and Sinha, B (2020) Humidity, density, and inlet aspiration efficiency correction
- improve accuracy of a low cost sensor during field calibration at a suburban site in the North-
- 1097 Western Indo-Gangetic plain (NM-IGP). Aerosol Science and Technology, 54-6 685-703.
- 1098 Peñaloza-Murillo, M. (1999). Deriving the basic cell-reciprocal integrating nephelometer
- 1099 equation and its use for calibration purposes: a comprehensive approach. Measurement Science
- and Technology. 10. R1. doi:10.1088/0957-0233/10/1/003.
- Ryan, S. (1997). The wind field around Mauna Loa derived from surface and balloon
- observations. J. Geophys. Res. 102: 10711–10725. https://doi.org/10.1029/97JD00646
- 1103 Sayahi, T., Butterfield, A. and Kelly, K. E.: Long-term field evaluation of the Plantower PMS
- low-cost particulate matter sensors, Environ. Poll., 245, 932–940,
- 1105 https://doi.org/10.1016/j.envpol.2018.11.065, 2019.
- 1106 Shaw, G.E. (1980). Transport of Asian Desert Aerosol to the Hawaiian Islands. J. Appl.
- 1107 *Meteorol.* 19:1254–1259. https://doi.org/10.1175/1520-
- 1108 0450(1980)019<1254:TOADAT>2.0.CO;2
- 1109 Sherman, J. P., Sheridan, P. J., Ogren, J. A., Andrews, E., Hageman, D., Schmeisser, L.,
- 1110 Jefferson, A., and Sharma, S.: A multi-year study of lower tropospheric aerosol variability and
- 1111 systematic relationships from four North American regions, Atmos. Chem. Phys., 15, 12487–
- 1112 12517, https://doi.org/10.5194/acp-15-12487-2015, 2015.
- 1113 Snider, G., Weagle, C. L., Martin, R. V., Van Donkelaar, A., Conrad, K., Cunningham, D.,
- 1114 Gordon, C., Zwicker, M., Akoshile, C., Artaxo, P., Anh, N.X., Brook, J., Dong, J., Garland,
- 1115 R.M., Greenwald, R., Griffith, D.,He, K., Holben, B.N., Kahn, R., Koren, I., Lagrosas, N.,
- 1116 Lestari, P., Ma, Z., Vanderlei Martins, J., Quel, E.J., Rudich, Y., Salam, A., Tripathi, S.N. Yu,
- 1117 C., Zhang, Q., Zhang, Y., Brauer, M., Cohen, A., Gibson, M.D., and Liu, Y. SPARTAN: A
- 1118 Global Network to Evaluate and Enhance Satellite-Based Estimates of Ground-Level Particulate
- 1119 Matter for Global Health Applications. Atmos. Meas. Tech. 2015, 8, 505–521.
- 1120 Tryner, J., L'Orange, C., Mehaffy, J., Miller-Lionberg, D., Hofstetter, J.C., Wilson, A.,
- Volckens, J., Laboratory evaluation of low-cost PurpleAir PM monitors and in-field correction
- using co-located portable filter samplers. Atmospheric Environment 220 (2020) 117067.
- 1123 Zheng, T., Bergin, M.H., Johnson, K.K., Tripathi, S.N., Shirodkar, S., Landis, M.S., Sutaria, R.,
- 1124 Carlson, D.E., 2018. Field evaluation of low-cost particulate matter sensors in high and low
- 1125 concentration environments. Atmos. Meas. Tech. 11, 4823-4846. https://doi.org/10.5194/amt-11-
- 1126 4823-2018.
- 1127 Zhou, Y., "2016 product data manual of PLANTOWER PMS5003 series data manual,"
- 1128 https://www.aqmd.gov/docs/default-source/aq-spec/resources-page/plantower-pms5003-
- 1129 manual v2-3.pdf, accessed 15 March 2021.