

# Characterization of soot produced by the ~~MISG-Mini Inverted Soot Generator soot generator~~ with an atmospheric simulation chamber

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## ABSTRACT

The performance of a Mini-Inverted Soot Generator (MISG) has been investigated at ChAMBRé (Chamber for Aerosol Modelling and Bio-aerosol Research) by studying the properties of soot particles generated by ethylene and propane combustion. This work deepens and expands the existing characterization of the MISG this soot generator that is also exploiting coupled with an atmospheric simulation chamber (ASC). Differently from previous works, MISG performance has been also tested at different fuel flows and higher global equivalence ratios. MISG exhausts were investigated after their injection inside the atmospheric simulation chamber: this is another novelty of this work. Starting from an extensive classification of combustion conditions and resulting flame shapes, the MISG exhaust was characterized in terms of concentration of emitted particles and gases, particle size distribution and optical properties. Soot particles were also collected on quartz fibre filters and then analysed by optical and thermal-optical techniques, to measure the spectral dependence of the absorption coefficient  $b_{\text{abs}}$ , and their composition in terms of Elemental and Organic Carbon (EC and OC). Significant differences could be observed when the MISG is fuelled with ethylene and propane ~~both~~ in terms of particle size; in particular, the production of super-micrometric aggregates was observed for ethylene combustion. With equal combustion conditions, ethylene produced higher number concentration of particles and smaller mode diameters. Soot particles produced by propane combustion resulted in higher EC:TC ratios and they were more light absorbing than particles generated by ethylene combustion, and optical behaviour (i.e., absorption coefficient). Values of the Mass Absorption ~~Coefficient~~ Cross Section Coefficient (MAC) and of the Angstrom Absorption Exponent (AAE) turned out to be compatible with the literature, even if with some specific differences. The comprehensive characterization of the MISG soot particles is an important piece of information to design and perform experiments in atmospheric simulation chambers. The soot characterization opens to various kinds of experiments in atmospheric simulation chambers. Particles with well-known properties can be used, for example, to investigate the possible interactions between soot and other atmospheric pollutants, the effects of meteorological variables on soot properties and the oxidative and toxicological potential of soot particles.

## 1. Introduction

“Soot” refers to combustion-generated carbonaceous particles that are a by-product of incomplete combustion of fossil fuels and/or biomass burning (Nordmann et al., 2013; Moore et al., 2014). When investigated by optical techniques, soot particles are generally referred as Black Carbon, BC (Petzold et al. 2013) while the result of thermal - optical characterizations is referred as Elemental Carbon, EC, (Bond and Bergstrom, 2006). However, both BC and EC are defined in operative terms that do not identify the same compounds (Massabò and Prati, 2021) and often produce non-negligible differences in concentration values.

Soot particles constitute an important fraction of anthropogenic particulate matter (PM) especially in urban environments (Weijer et al. 2011), and are emitted by traffic, domestic stoves, industrial chimneys and by any incomplete combustion process. Several works state adverse effects of soot both on climate (Ackerman et al., 2000; Menon et al., 2002; Quinn et al., 2008; Ramanathan and Carmichael, 2008; Bond et al., 2013) and health

47 (Pope et al., 2002; Anenberg et al., 2010; Gan et al., 2011; Cassee et al., 2013; Lelieveld et al., 2015). From  
48 the climatic point of view, soot particles absorb the solar radiation, causing a positive radiative forcing: BC is  
49 considered one of the most significant radiative forcing agent, second only to CO<sub>2</sub> (Ramanathan and  
50 Carmichael, 2008; Bond et al., 2013). Another positive effect on radiative forcing is related to the darkening  
51 of glaciers surface due to the deposition of BC (Skiles et al., 2018). Soot contributes to air pollution also via  
52 reactions with several gas species, as NO<sub>2</sub>, SO<sub>2</sub> and O<sub>3</sub> (Finlayson-Pitts and Pitts, 2000; Nienow and Roberts,  
53 2006). Effects on health include cardiopulmonary morbidity and mortality (Janssen et al., 2012). Soot particles  
54 are suspected to be particularly hazardous to human health, because they are sufficiently small to penetrate the  
55 membranes of the respiratory tract and enter the blood circulation or be transported along olfactory nerves into  
56 the brain (Nemmar et al., 2002; Oberdörster et al., 2005). The understanding of properties and behaviour of  
57 soot particles when they are suspended in the atmosphere is thus necessary to fully assess their adverse effects  
58 and the use of proxies with controlled and known properties can be useful. In this context, soot generators are  
59 employed as stable sources of soot particles. So far, soot generators have been employed for studies on optical  
60 properties (Zhang et al. 2008; Cross et al. 2010; Mamakos et al. 2013; Utry et al. 2014 b; Bescond et al. 2016),  
61 instruments calibration (Onasch et al. 2012; Durdina et al. 2016) and several other purposes, such as studies  
62 on atmospheric processing of soot particles, characterization of uncoated/coated and fresh/denuded of soot  
63 particles (Pagels et al. 2009; Henning et al. 2012; Ghazi et al. 2013; Ghazi and Olfert 2013; [Hu et al., 2021](#)).  
64 The Inverted-Flame Burner (Stipe et al. 2005) is often considered as an ideal soot source (Moallemi et al.,  
65 2019 and references therein), due to its capacity to generate almost pure-EC particles and for the stability of  
66 the flame and of its exhaust (Stipe et al. 2005). To such category belongs the Mini-Inverted Soot Generator,  
67 MISG (Argonaut Scientific Corp., Edmonton, AB, Canada, Model MISG-2), used in this work.

68 The MISG can be operated with different fuels: ethylene (Kazemimanesh et al., 2019), propane (Moallemi  
69 et al., 2019, [Bischof et al., 2019](#)), and theoretically also with ethane or fuel blends with methane and nitrogen,  
70 even if, to our knowledge, no literature is available on such configurations. The air to fuel flow ratio can be  
71 adjusted to control concentration and size of the generated particles. The maximum reachable concentration  
72 declared by the manufacturer is about 10<sup>7</sup> particles cm<sup>-3</sup> (<https://www.argonautscientific.com/>), while particle  
73 size ranges from few tens to few hundreds of nm.

74 The behaviour of soot particles can be efficiently studied in/by ~~atmospheric simulation chambers (ASCs)~~:  
75 these are exploratory platforms which allow to study atmospheric processes under controlled conditions, that  
76 can be maintained for periods long enough to reproduce realistic environments and to study interactions among  
77 their constituents (Finlayson - Pitts and Pitts, 2000; Becker, 2006). ASC experiments are the best compromise  
78 between laboratory and field experiments, since they simulate quasi-real situations but without the  
79 uncertainties and variability of typical field measurements. Recent examples of ASC applications concern the  
80 investigation of the optical properties of mineral dust (Caponi et al., 2017) and wood-burning exhausts (Kumar  
81 et al., 2018, [Hu et al., 2021](#)).

82 Coupling the MISG to an ASC makes possible systematic experiments on the properties of soot particles  
83 exposed and maintained in different conditions. In this work, we mainly investigated the differences between  
84 MISG exhausts produced by ethylene and propane burning. Differently from previous works (Bischof et al.,  
85 2019; Kazemimanesh et al., 2019; Moallemi et al., 2019), the MISG has been connected directly to an  
86 atmospheric simulation chamber; performance has been tested also at different fuel flows and higher global  
87 equivalence ratios. The present characterization deepens and expands the existing knowledge on particles and  
88 gases produced by this soot generator. The comprehensive characterization of the MISG soot particles is an  
89 important piece of information to design the subsequent experiments. Well-characterized soot particles could  
90 be used to investigate the effects that atmospheric parameters can have on soot particles, and to study the  
91 interactions between soot particles and other pollutants.

## 92 2. Materials and methods

### 93 2.1 Mini-Inverted Soot Generator

The MISG, introduced by Kazemimanesh (2019), is a combustion-based soot generator working as an inverted-flame burner (Stipe et al., 2005) where air and fuel flow in an opposite way-direction to the buoyancy force of the hot exhaust gases. ~~This results in a co-flow diffusion flame and leads to a better flame stability by reducing flame tip flickering. The resulting co-flow diffusion flame is more stable bthanks to a reduced flickering of flame tip~~ (Kirchstetter & Novakov, 2007; Stipe et al., 2005) and consequently ~~to a more stable~~ the soot particle generation is more stable.

The MISG is fed with air and fuel supplied by specific cylinders: we used both ethylene and propane, two fuels with a well-known capability of producing soot (Kazemimanesh et al., 2019; Moallemi et al., 2019). Air and fuel flow rates are controlled by two mass flow controllers (MFCs, Bronkhorst High-Tech B.V., Ruurlo, Netherlands, Models F-201CV-10K-MGD-22-V and FG-201CV-MGD-22-V-AA-000, respectively) operated via a home-made National Instruments Labview code. The air and fuel flows can be controlled in the range 0-12 lpm (i.e., litres per minute) and 0-200 mlpm (i.e., millilitres per minute), respectively. Differently from other commercial generators, the MISG does not require a third gas (i.e., N<sub>2</sub>) used as a carrier (quenching gas) and the air flow is internally split between combustion and ~~carriage operations~~ dilution of exhaust product. This implies that the ratio of comburent and carrier gas is not controllable, and the user can only adjust the comburent to fuel ratio.

The efficiency of the combustion process can be given in terms of the global equivalence ratio; that is the ratio between actual and stoichiometric fuel-to-air ratio; starting from the air to fuel ratio (AFR):

$$AFR = \frac{m_A}{m_F} = \frac{n_A * M_A}{n_F * M_F}$$

where:

$m_A$ : air mass;

$m_F$ : fuel mass;

$n_A$ : number of air moles;

$n_F$ : number of fuel moles;

$M_A$ : air molecular weight;

$M_F$ : fuel molecular weight.

~~The stoichiometric AFR value is 15.64 m<sup>3</sup>·m<sup>-3</sup> (inverse value = 0.064 m<sup>3</sup>·m<sup>-3</sup>) and 14.75 m<sup>3</sup>·m<sup>-3</sup> (inverse value = 0.068 m<sup>3</sup>·m<sup>-3</sup>), for propane and ethylene, respectively. Finally, the ratio between stoichiometric and actual AFR corresponds to the global equivalence ratio:~~

$$\varphi = \frac{(m_F/m_A)}{(m_F/m_A)_{st}} \quad \text{Eq.1}$$

where:

$(m_F/m_A)$ : ~~inverse of actual~~ fuel-to-air ratio AFR;

$(m_F/m_A)_{st}$ : ~~inverse of stoichiometric~~ fuel-to-air ratio AFR.

The fuel-to-air ratio is calculated as the opposite of the air-to-fuel ratio (AFR) that is the ratio between air and fuel masses. The stoichiometric AFR value is 15.64 (inverse value = 0.064) and 14.75 (inverse value = 0.068), for propane and ethylene, respectively.

The flame is classified as fuel-rich and fuel-lean when  $\phi > 1$  and  $\phi < 1$ , respectively. ~~It is yet demonstrated (Moore et al., 2014) that fuel lean flames produce soot particles with larger mode diameter (about 100-200 nm) while fuel rich flames lead to an additional mode in the nucleation size range (i.e., 10-30 nm). Finally, Mamakos (2013) reported that low fuel-to-air ratios (i.e.,  $\phi < 1$ ) generate particles with a large fraction of EC~~

139 while semi-volatile organics are generated by high fuel-to-air ratios (i.e.,  $\phi > 1$ ). In this work, fuel-lean  
140 conditions were investigated only.

141 Since the combustion process can produce flame shapes having different characteristics, we first explored  
142 the range of combustion flows from 2 to 10 lpm, in 0.5 lpm steps, and from 30 to 100 mlpm, in 5 mlpm steps,  
143 respectively for air and fuel. Flame types can be distinguished (Kazemimanesh et al., 2019; Moallemi et al.,  
144 2019) as:

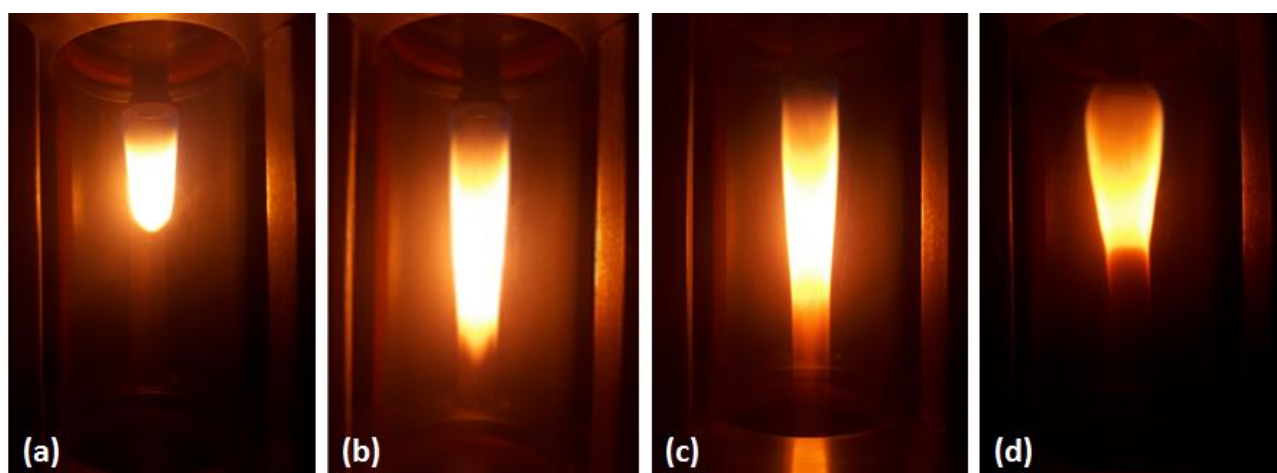
145 - *Closed tip* flame (Fig. 1.a), which generates low concentrations of soot particles (i.e., around  $10^3 \text{ \# cm}^{-3}$ ),  
146 generally forming particle aggregates at the ~~fuel tube nozzle~~ MISG.

147 - *Partially Open tip* flame (Fig. 1.b), the transition between *Open* and *Closed tip*.

148 - *Open tip* flame (Fig. 1.c), which generates high concentrations of soot particles (i.e.,  $> 10^5 \text{ \# cm}^{-3}$ ).

149 - *Asymmetric* flame, which shows a large variability (very short, flickering, etc) and can form particle  
150 aggregates at the ~~MISG~~ fuel tube nozzle.

151 - *Curled Base* flame (Fig. 1.d), a particular shape of the asymmetric flames that can also form particles  
152 aggregates at the ~~MISG~~ fuel tube nozzle.



153

154 *Figure 1: Examples of different flame shapes: (a) Closed tip, (b) Open tip, (c) Partially Open tip, (d) Curled base flame.*

155 By the flames observation (~~Sect. 2.1.2~~), we selected the more interesting combustion conditions (i.e., *Open*  
156 *tip* flames) to perform the characterization experiments. We focused on *Open tip* flames because it is the flame  
157 that generates higher concentrations of soot particles. Operative conditions selected for propane and ethylene  
158 combustion are reported in Tables 1 and 2: we maintained the same air flow and global equivalence ratio with  
159 both the fuels.

160

*Table 1: Combustion parameters and flame shapes selected for propane.*

<b>PROPANE</b>			
<b>AIR flow</b>	<b>FUEL flow</b>	<b>Global Equivalence Ratio</b>	<b>Flame shape</b>
<b>[lpm]</b>	<b>[mlpm]</b>		
7	70	0.244	Partially Open Tip
7	75	0.261	Open Tip
7	80	0.278	Open Tip
7	85	0.296	Open Tip
8	70	0.213	Partially Open Tip
8	75	0.228	Open Tip
8	80	0.244	Open Tip
8	85	0.259	Open Tip

<b>PROPANE</b>			
<b>AIR flow</b>	<b>FUEL flow</b>	<b>Global Equivalence Ratio</b>	<b>Flame shape</b>
<b>[lpm]</b>	<b>[mlpm]</b>		
7	70	0.244	Partially Open Tip
7	75	0.261	Open Tip
7	80	0.278	Open Tip
7	85	0.296	Open Tip
8	70	0.213	Partially Open Tip
8	75	0.228	Open Tip
8	80	0.244	Open Tip
8	85	0.259	Open Tip

Table 2: Combustion parameters and flame shapes selected for ethylene.

<b>ETHYLENE</b>			
<b>AIR flow</b>	<b>FUEL flow</b>	<b>Global Equivalence Ratio</b>	<b>Flame shape</b>
<b>[lpm]</b>	<b>[mlpm]</b>		
7	118	0.244	Partially Open Tip
7	127	0.261	Open Tip
7	135	0.278	Open Tip
7	144	0.296	Open Tip
8	118	0.213	Partially Open Tip
8	127	0.228	Open Tip
8	135	0.244	Open Tip
8	144	0.259	Open Tip

<b>ETHYLENE</b>			
<b>AIR flow</b>	<b>FUEL flow</b>	<b>Global Equivalence Ratio</b>	<b>Flame shape</b>
<b>[lpm]</b>	<b>[mlpm]</b>		
7	118	0.244	Partially Open Tip
7	127	0.261	Open Tip
7	135	0.278	Open Tip
7	144	0.296	Open Tip
8	118	0.213	Partially Open Tip
8	127	0.228	Open Tip
8	135	0.244	Open Tip
8	144	0.259	Open Tip

## 2.2 Chamber setup

Experiments took place at the ChAMBRe (Chamber for Aerosol Modelling and Bio-aerosol Research) facility (Massabò et al., 2018; Danelli et al., 2021) located at the Physics Department of the University of Genoa.

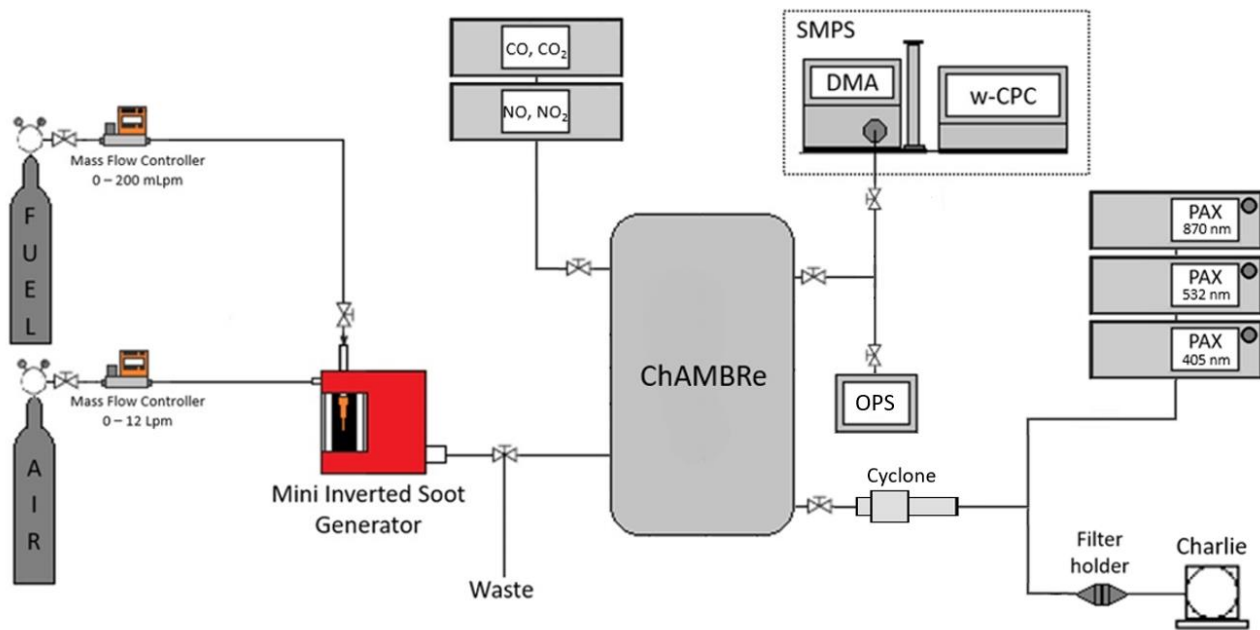
ChAMBRe is a stainless-steel chamber, with a volume of about 2.2 m<sup>3</sup>. Inside the chamber, relative humidity, temperature, and pressure are continuously monitored by a HMT334 Vaisala® Humicap® transmitter and a MKS Instruments 910 DualTrans™ transducer, respectively. Two gas analyzers from Environnement SA, continuously monitored the concentration of NO/NO<sub>2</sub> (model: AC32e), and CO/CO<sub>2</sub> (model: CO12e) inside the chamber or, alternatively, in the laboratory. The mixing of gas and aerosol species is favoured by a fan installed in the bottom of the chamber: mixing time for gaseous species is of about 180 s

178 with a fan rotating speed of 1.6 revolutions per second. A composite pumping system (rotary pump TRIVAC®  
179 D65B, Leybold Vacuum, root pump RUVAC WAU 251, Leybold Vacuum and Leybold Turbovac 1000)  
180 allows to evacuate the internal volume down to  $10^{-5}$  mbar; in this way ChAMBRe is cleaned before each  
181 experiment. Before and during the experiments, ambient air enters the chamber throughout a 5-stage  
182 filtering/purifying inlet (including a HEPA filter, model: PFIHE842, NW25/40 Inlet/Outlet – 25/55 SCFM,  
183 99.97 % efficient at  $0.3 \mu\text{m}$ ). The whole set-up is managed by a custom NI Labview SCADA (Supervisory  
184 Control And Data Acquisition).

185 The layout of the experimental configuration adopted for the MISG characterization is shown in Fig. 2.

186 The MISG was warmed for about 45 minutes before injecting soot particles inside the chamber. Injection  
187 of soot particles inside ChAMBRe lasted 2 or 3 minutes, depending on the soot concentration required for each  
188 experiment. We performed some fluid dynamic evaluations with the Particle Loss Calculator software tool  
189 (PLC; von der Weiden et al., 2009). The connection between MISG and ChAMBRe was made by Swagelok  
190 adaptors (size  $\frac{3}{4}$ " ) and ISO-K flanges (16 mm diameter) to avoid any possible leak; the length of the line was  
191 65 cm. The geometry of our experimental setup, combined with particle size and used flow rates, resulted in  
192 particle losses lower than 0.1 % in the dimensional range of 80-2000 nm. All the experiments were performed  
193 at atmospheric pressure,  $19^\circ < T < 21^\circ \text{C}$  and R.H. < 50 %.

194



195

196

Figure 2: Layout of the MISG set-up at ChAMBRe.

197

### 2.3 Size distributions measurements

198 Particle concentration and size distribution inside the chamber were measured by a scanning mobility  
199 particle sizer (SMPS, TSI Inc., Shoreview, MN, USA, Model 3938), composed by a differential mobility  
200 analyzer (DMA, TSI Inc., Shoreview, MN, USA, Model 3081A) and a water condensation particle counter  
201 (w-CPC, TSI Inc., Shoreview, MN, USA, Model 3789). The water-CPC is filled using technical demineralized  
202 water (Conductivity (20°C), max.  $1.5 \mu\text{S}/\text{cm}$ ; VWR Chemicals INTERNATIONAL S.R.L.). The SMPS was  
203 set to measure particles with mobility diameter from 34 nm to 649 nm; aerosol sample and sheath airflow rates  
204 were fixed at 0.17 lpm and 1.60 lpm, respectively, while the scanning period for each cycle was 70 s. The  
205 DMA unit integrates an impactor with an orifice of 0.0508 cm, resulting in cut-off capability at 50 % of 940

206 nm, useful to exclude all the particles larger than this size to enter in the column. Frequent cleaning of this part  
207 was necessary to ensure proper operation and avoid clogging; at the end of each experiment, the whole  
208 impactor system was cleaned using compressed air and isopropyl alcohol.

209 We corrected diffusion losses in the instrument by using the option included in the instrument software;  
210 size distributions were ~~as-well-not~~ corrected for multiple charges effects through the TSI proprietary software  
211 (Aerosol Instrument Manager, Version 11-0-1). An example of comparison between size distribution corrected  
212 and uncorrected by the multiple charge correction algorithm is shown in the Supplementary (see Fig. S.1).

213 Among the other chamber instruments, an Optical Particle Sizer (OPS, TSI Inc., Shoreview, MN, USA,  
214 Model 3330) was used for short times to spot the particle size distribution in the range 0.3-10  $\mu\text{m}$ .

## 215 **2.4 Online optical measurements**

216 Three photoacoustic extinction-meters (PAXs, Droplet Measurement Technologies, Boulder, CO, USA)  
217 were deployed, providing the online determination of the soot particles absorption coefficients at  $\lambda = 870, 532$   
218 and 405 nm. PAXs are constituted by a measurement cell where aerosol optical properties are measured by  
219 two different mechanisms (<https://www.dropletmeasurement.com/> PAX Operator Manual). The sample flow  
220 rate (1 lpm) is split in two different sectors of the cell, both crossed at the same time by the light of a modulated  
221 laser diode. In the absorption sector, soot particles absorb light and release acoustic waves, which are then  
222 detected by an ultra-sensitive microphone. The intensity of the acoustic signal is interpreted to infer the particle  
223 absorption coefficient. In the other sector, a wide-angle reciprocal nephelometer measures the scattering  
224 coefficient instead. It is noteworthy that no correction for the truncation angle is applied by the manufacturer:  
225 this can lead to substantial underestimation of the scattering coefficient, which generally grows as the particle  
226 size increases and the single scattering albedo (SSA) approaches unity. Few papers in literature deal with the  
227 correction for truncation errors in nephelometer measurements (Bond et al., 2009, Modini et al, 2021) for  
228 highly absorbing particles: little is known on the dependency of scattering phase function on the particle  
229 morphology and how this might impact truncation. However, since particles produced by soot generators have  
230 dimensions generally lower than 1  $\mu\text{m}$  and SSA values lower than 0.3 (Moallemi et al., 2019), we disregarded  
231 this issue. At the time of the experiments, the three PAXs had been just calibrated by the manufacturer.

232 In some experiments, soot concentration inside the chamber was too high to be measured directly by PAXs;  
233 and a diluter (eDiluter Pro, Dekati Ltd., Kangasala, Finland) was deployed. Dry air from a cylinder was merged  
234 prior to the PAXs inlet with dilution factor 1:100. Tests performed with and without the diluter demonstrated  
235 a substantial reproducibility of the optical properties measured by the PAXs when the proper dilution factor is  
236 considered.

## 237 **2.5 Offline analysis**

238 Soot particles were also collected on pre-fired 47 mm diameter quartz fibre filters (Pallflex Tissuquartz  
239 2500 QAO-UP) held in a stainless-steel filter holder to allow additional offline analysis. The sampling started  
240 when stable gas and particle concentration values were reached inside the chamber (i.e., about 3 minutes -  
241 corresponding to the chamber mixing time - after the MISG switching off): for each working condition three  
242 filters with different loadings were obtained by a low-volume sampler (TECORA – Charlie HV) working at a  
243 fixed sampling flow (i.e., 10 lpm during experiments without cyclone and 13.67 lpm during experiments with  
244 cyclone).

245 For each sample, the EC and OC mass concentration was determined by thermal-optical transmittance  
246 analysis (TOT) using a Sunlab Sunset EC/OC analyzer and the NIOSH5040 protocol (NIOSH, 1999),  
247 corrected for temperature offsets. We also performed some tests adding a backup filter during the sampling to  
248 determine the volatile fraction of OC.

249 Prior to EC/OC determination, particle-loaded filters were analyzed by the Multi-Wavelength Absorbance  
250 Analyzer (MWAA, Massabò et al., 2013 and 2015), a laboratory instrument for the offline direct quantification



251 of the aerosol absorption coefficients at five different wavelengths ( $\lambda = 850, 635, 532, 405$  and  $375$  nm). Such  
252 peculiar features have been previously exploited in the frame of several field campaigns in urban and rural  
253 sites (Scerri et al., 2018; Massabo et al, 2019; Massabo et al, 2020; Moschos et al., 2021), as well as in peculiar  
254 specific and remote sites (Massabo et al., 2016; Saturno et al., 2017; Baccolo et al., 2021).

## 255 2.6 Cyclone experiments

256 Soot aggregates are also generated by the MISG. Kazemimanesh (2019) retrieved super-aggregates larger  
257 than  $2 \mu\text{m}$  for ethylene combustion while Moallemi (2019) showed aggregate structures larger than  $1 \mu\text{m}$  with  
258 propane. On this basis, confirmed by some short checks by the OPS, we replicated each experiment (see Sect.  
259 2.1) both without and with a cyclone (PM1 Sharp Cut Cyclone - SCC 2.229, MesaLabs, Lakewood, CO, USA)  
260 inserted upstream the PAXs and filters sampler (Fig. 2). The cyclone has a cut-off of  $1 \mu\text{m}$  at a nominal flow  
261 of 16.66 lpm.

## 262 3. Results and Discussion

### 263 3.1 Characterization tests

264 The categories of flame shape observed in the range of air and fuel flows discussed in sect. 2.1.2 are  
265 summarized in Supplementary (see Tables S.13 and S.24), for propane and ethylene respectively. The MISG  
266 characterization with propane has been previously published (Moallemi et al., 2019) and we used it as a  
267 reference. We got some differences especially in the range of transition from Closed tip to Open tip, probably  
268 due to the different setups. Fuel flows higher than 85 mlpm were not investigated due to instrumental  
269 limitation. To our knowledge, no literature information is available for the ethylene in the flow range of Table  
270 4. A similar characterization with ethylene also exists but it only partly covers the flow ranges explored in the  
271 present work. We got some differences especially in the transition range to Open tip flames, probably due to  
272 the different setups. In addition, the subjectivity of the visual determination, that is user dependent, can lead  
273 to differences. It is noteworthy that no correlation could be found between the global equivalence ratio ( $\phi$ ) and  
274 the shape of the corresponding flame. This means that the fundamental parameter of the combustion process  
275 can not be used to predict the flame shape.

276 The repeatability/reproducibility and stability of the MISG emissions were investigated for all the  
277 combustion conditions listed in Table 1 and 2, in terms of number concentration and size distribution of the  
278 generated soot particles. Different combustion conditions were selected, and four experiments were performed  
279 for each combination of air and fuel flows. We chose to keep fixed the air flow to observe the differences  
280 produced by different fuel flows that correspond to different flame shapes (i.e., Partially Open tip or Open  
281 tip). In each test, we recorded the values of total particle number concentration, peak concentration, and mode  
282 diameter. The repeatability/reproducibility was calculated as the percentage ratio between standard deviation  
283 and mean value (i.e., the relative standard deviation) of each series of identical repeated experiments. With  
284 propane, mode reproducibility turned out to be 6 %, while total concentration and peak concentration showed  
285 a 16 % repeatability/reproducibility. With ethylene, the repeatability/reproducibility was 4 % and 10 %,  
286 respectively for mode and total/peak concentration. In addition, we monitored the combustion gases:  $\text{CO}_2$  and  
287 NO concentration varied by about 2 % and 3 %, respectively with propane and ethylene.

288 Table 3: Flame shapes observed for different combustion conditions of propane. Flames are identified as A—symmetric,  
289 CB—Curled Base, CT—Closed tip, POT—Partially Open tip and OT—Open tip; FL indicates if flickering. The dash  
290 indicates that the flame does not ignite.

		FUEL flow [mlpm]											
		30	35	40	45	50	55	60	65	70	75	80	85
AIR flow [lpm]	2	A	A/FL	A	A/FL	A	CB/FL	CB/FL	CB/FL	CB/FL	CB/FL	CB/FL	CB/FL
	2.5	A/FL	A/FL	A	A/FL	A/FL	CB/FL	CB/FL	CB/FL	CB/FL	CB	CB	CB/FL
	3	A/FL	A/FL	A	A/FL	A/FL	CB/FL	CB/FL	CB/FL	CB/FL	CB/FL	CB	CB/FL
	3.5	A/FL	A	A	A	A	CB	CB	CB	CB	CB	CB	CB/FL
	4	A	A	A	A	A	CB	CB	CB/FL	CB	CB	CB	OT
	4.5	A	A	A	A	A	CB	CB	CB	CB	CB	CB	OT
	5	A	A	A	A	A	A/CB	CT	POT	OT	OT	OT	OT
	5.5	A	A	A	A	A	CT	CT	POT	OT	OT	OT	OT
	6	A	A	A	A	CT	CT	CT	CT	POT/OT	OT	OT	OT
	6.5	A	A	A	A	CT	CT	CT	CT	POT	OT	OT	OT
	7	A	A	A	A	A	CT	CT	CT	POT	POT/OT	OT	OT
	7.5	A	A	A	A	A	CT	CT	CT	POT	POT/OT	OT	OT
	8	-	-	A	A	A	CT	CT	CT	POT	POT/OT	OT	OT
	8.5	-	-	A	A	A	CT	CT	CT	POT/OT	POT/OT	OT	OT
	9	-	-	A	A	A	CT	CT	CT	CT	POT	OT	OT
	9.5	-	-	-	A	A	CT	CT	CT	CT	POT	OT	OT
10	-	-	-	A	A	CT	CT	CT	CT	POT	OT	OT	

291  
292  
293

Table 4: Flame shapes observed for different combustion conditions of ethylene. Flames are identified as A – asymmetric; CB – Curled Base, CT – Closed tip, POT – Partially Open tip and OT – Open tip; FL indicates if flickering.

		FUEL flow [mlpm]														
		30	35	40	45	50	55	60	65	70	75	80	85	90	95	100
AIR flow [lpm]	2	A	A	A	A	A/FL	A	A	A	CB	CB	CB	CB	CB	CB	CB
	2.5	A	A	A	A/FL	A/FL	A/FL	A	A	CB	CB	CB	CB	CB	CB	CB
	3	A	A	A	A	A/FL	A/FL	A	A	A/CB	CB	CB	CB	CB	CB	CB
	3.5	A	A	A	A	A/FL	A/FL	A	A	A/CB	CB	CB	CB	CB	CB	CB
	4	A	A	A	A	A	A	A	A	A	A/CB	CB	CB	CB	CB	CB/OT
	4.5	A	A	A	A	A	A	A	A	CB	CB	CB	CB	CB/OT	CB/OT	CB/OT
	5	A	A	A	A	A	A	A	A	CB	CB	CB	CB	CB/OT	CB/OT	CB/OT
	5.5	A	A	A	A	A	A	A	A	CB	CB/OT	CB/OT	CB/OT	CB/OT	CB/OT	CB/OT
	6	A	A	A	A	A	A	CT	CT	CT	CT/POT	CT/POT	POT	POT	OT	OT
	6.5	A	A	A	CT	CT	CT	CT	CT	CT/POT	POT	POT/OT	POT/OT	OT	OT	OT
	7	A	A	A	CT	CT	CT	CT	CT/POT	POT	POT/OT	OT	OT	OT	OT	OT
	7.5	A	A	A	A	CT	CT	CT	CT	POT	POT/OT	OT	OT	OT	OT	OT
	8	A	A	A	CT	CT	CT	CT	CT/POT	POT	POT/OT	OT	OT	OT	OT	OT
	8.5	A	A	A	CT	CT	CT	CT	CT	CT/POT	POT	OT	OT	OT	OT	OT
	9	A	A	CT	CT	CT	CT	CT	CT	POT	OT	OT	OT	OT	OT	OT
	9.5	A	A	CT	CT	CT	CT	CT	CT	CT	POT	OT	OT	OT	OT	OT
10	A	CT	CT	CT	CT	CT	CT	CT/POT	POT	POT/OT	OT	OT	OT	OT	OT	

294

### 3.2 Comparison between propane and ethylene exhausts

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Previous works investigated ~~concentration and mode the~~ exhausts of MISG fuelled by ethylene (Kazemimanesh et al., 2019) and propane (Moallemi et al., 2019) fuelled MISG exhausts. We expand here to a detailed comparison between the two fuels, focusing on ASC experiments. In addition, we reproduced some

of the conditions investigated in the previous works obtaining a good agreement for the mode diameter and SSA figures (see §3 in Supplementary for details-§3).

### 3.2.1 Size distribution

The mean size distributions observed at ChAMBre are given in Fig. 3, for all the selected operative conditions. Data were acquired starting 3 minutes (i.e., after the chamber mixing time) after the MISG switching off, for a specific time interval (i.e., 4 to 10 minutes). All the curves are normalized to the same injection time (i.e., 3 min of injection inside the chamber).

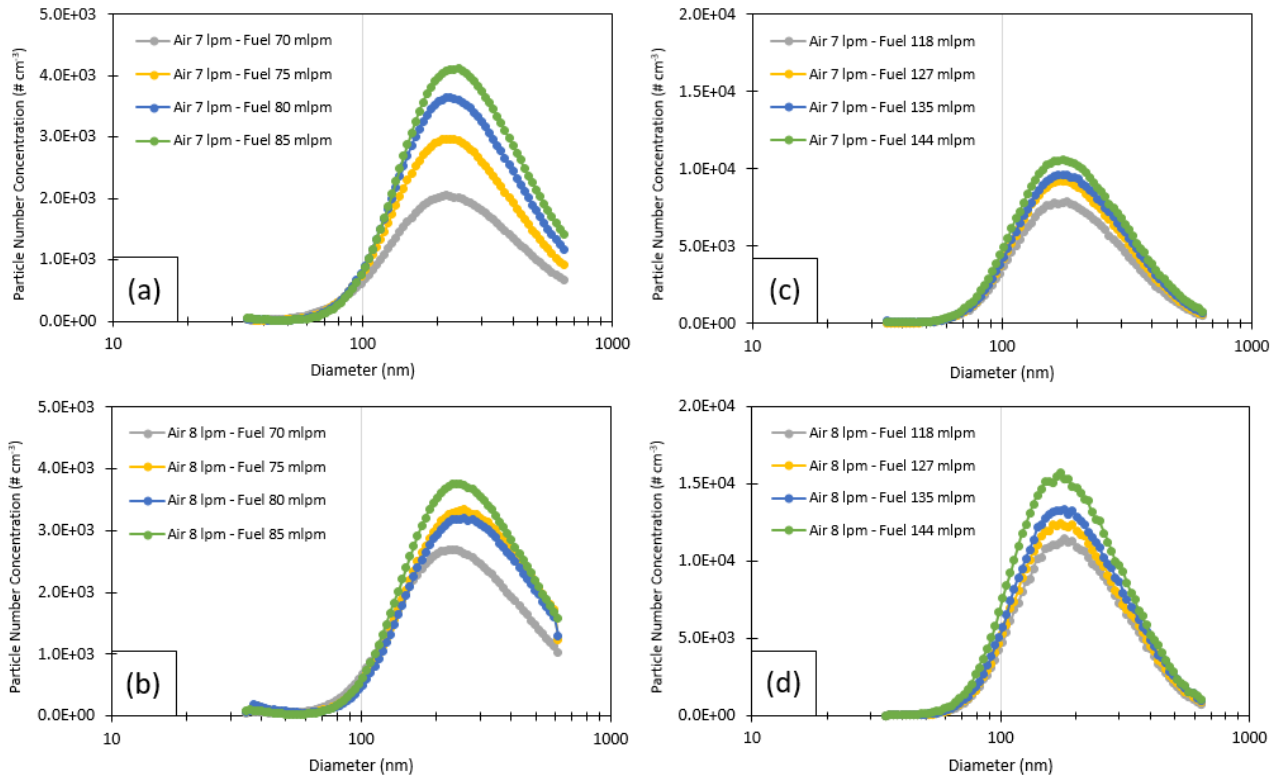
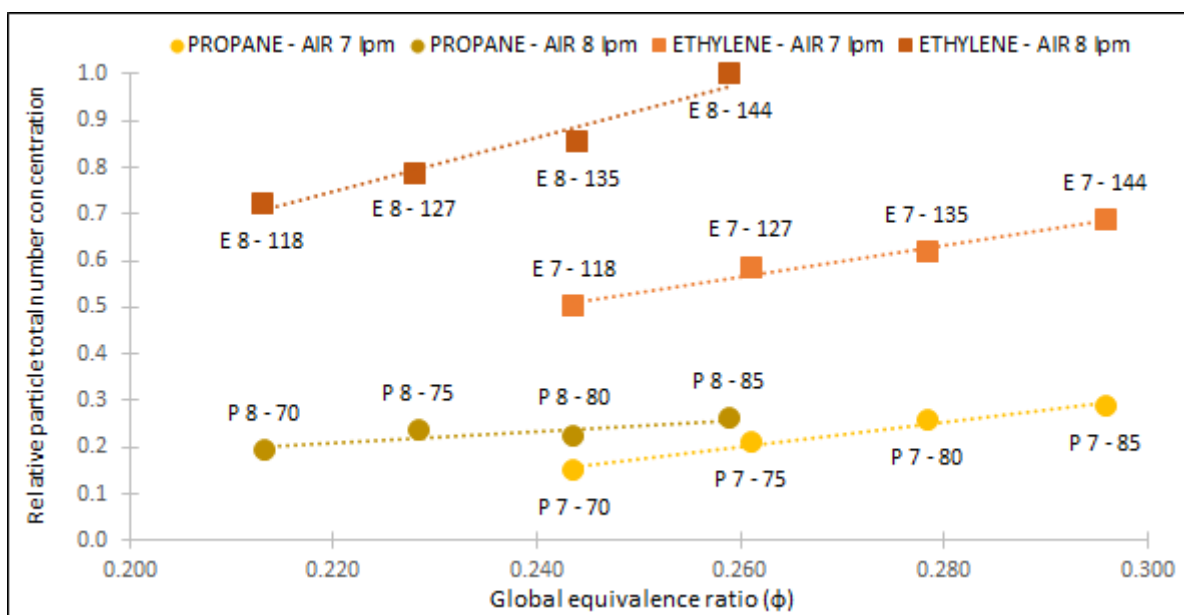
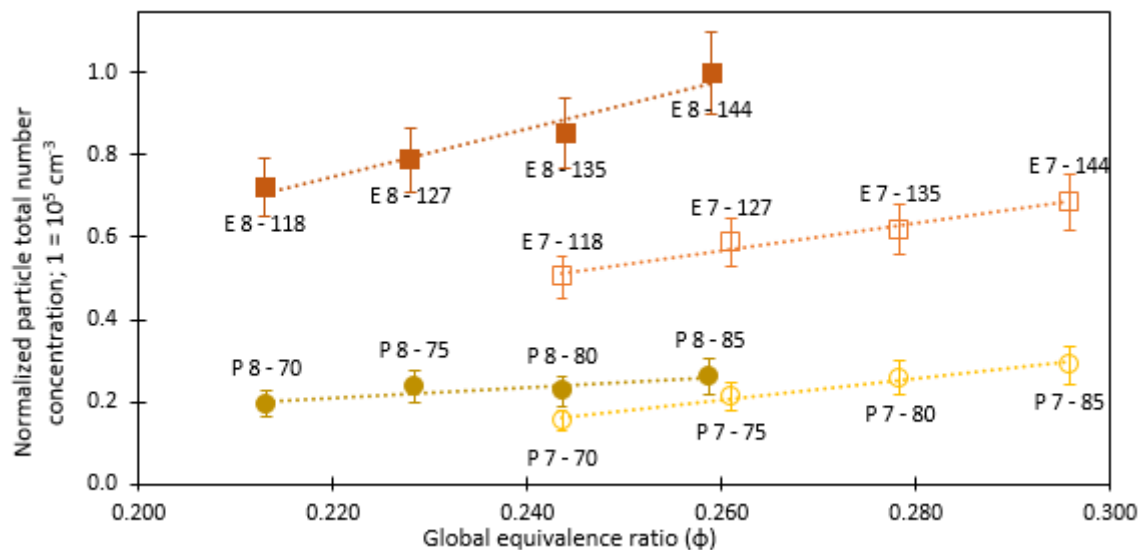


Figure 3: Mean size distributions measured by SMPS. MISG was fuelled with propane (a) and (b) panels and ethylene (c) and (d) with the air and fuel flows indicated in the plots frame.

For a better comparison of between different experiments, particle concentration values were normalized to the maximum recorded in the whole set of tests and therefore varied in the 0-1 range. Fig. 34 shows the result for the total particle number concentration. We can notice that:

- At fixed air flow, the particle number concentration increases with the fuel flow (i.e., with the global equivalence ratio).
- In the same combustion conditions (i.e., same air flow and same global equivalence ratio), ethylene generates more particles than propane.
- With ethylene and at fixed fuel flow, the particle number concentration increases with the air flow. The same holds in some cases with propane but with much smaller variations.

320 ~~At fixed air flow, the particle number concentration increases with the fuel flow (i.e., with the global~~  
 321 ~~equivalence ratio).~~  
 322 ~~In the same combustion conditions (i.e., same air flow and same global equivalence ratio), ethylene generates~~  
 323 ~~more particles than propane.~~  
 324 ~~With ethylene and at fixed fuel flow, the particle number concentration increases with the air flow. The same~~  
 325 ~~holds in some cases with propane but with much smaller variations.~~



327  
 328 *Figure 34: Particle number concentration vs the global equivalence ratio. Values are normalized to the highest of the*  
 329 *whole set. Each point is labelled by E or P (ethylene or propane) and a pair of numbers indicating air and fuel flow*  
 330 *rate, respectively in lpm and mlpm. Dotted lines aim to facilitate the reader eye.*

331  
 332 A similar comparison is shown in Fig. 45 for the particle mode diameter: while the values are basically constant  
 333 for ethylene, the mode diameter with propane slightly increases with air flow (at fixed fuel flow). Furthermore,  
 334 at each phi value, propane generated particles bigger than ethylene.

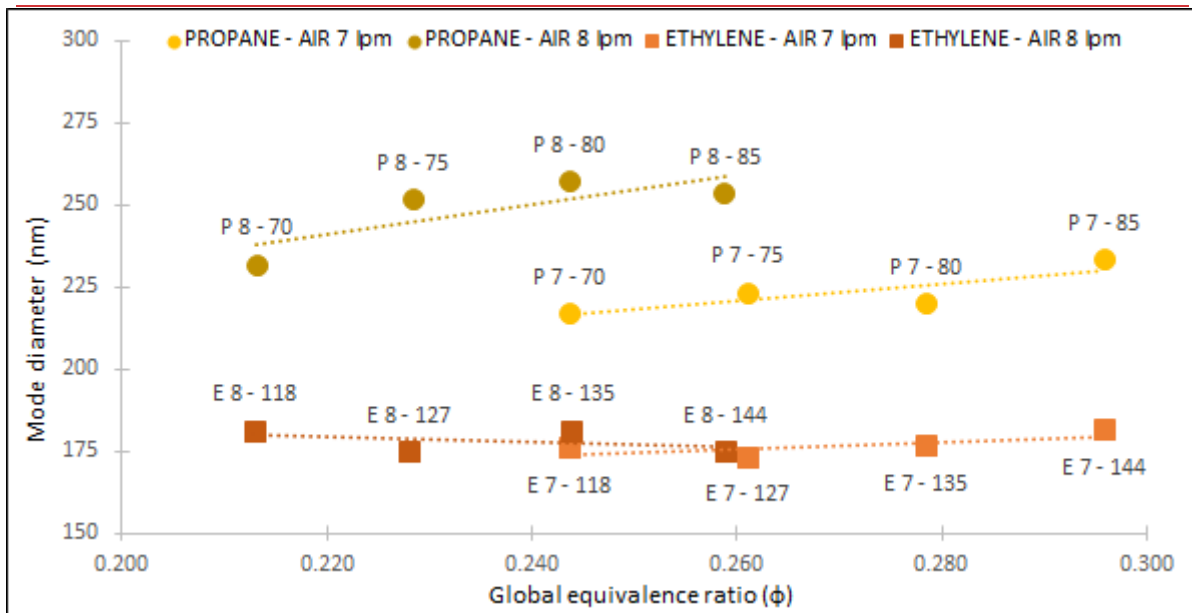
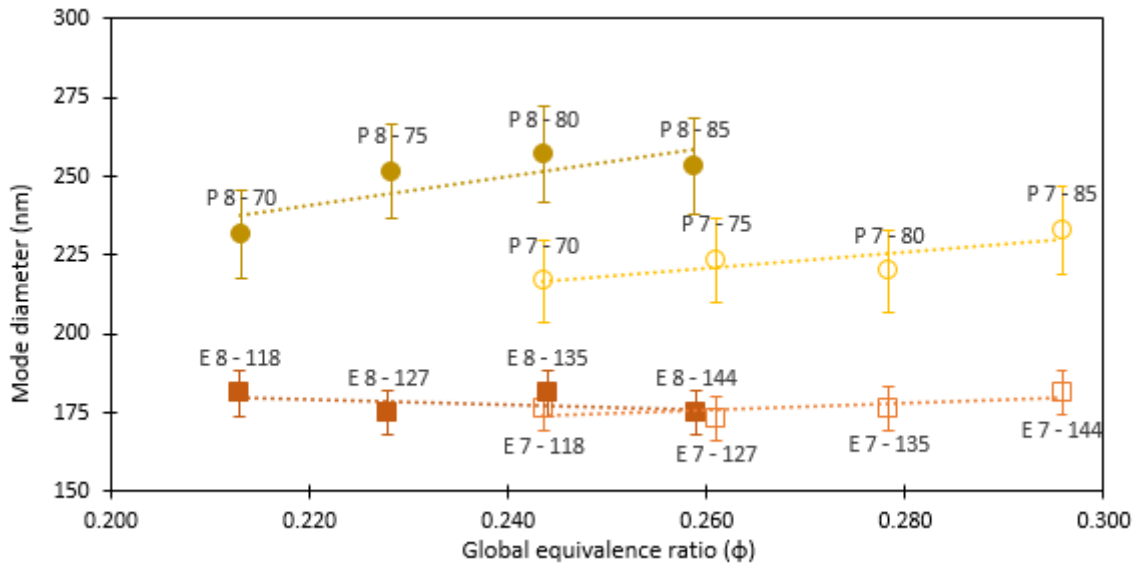


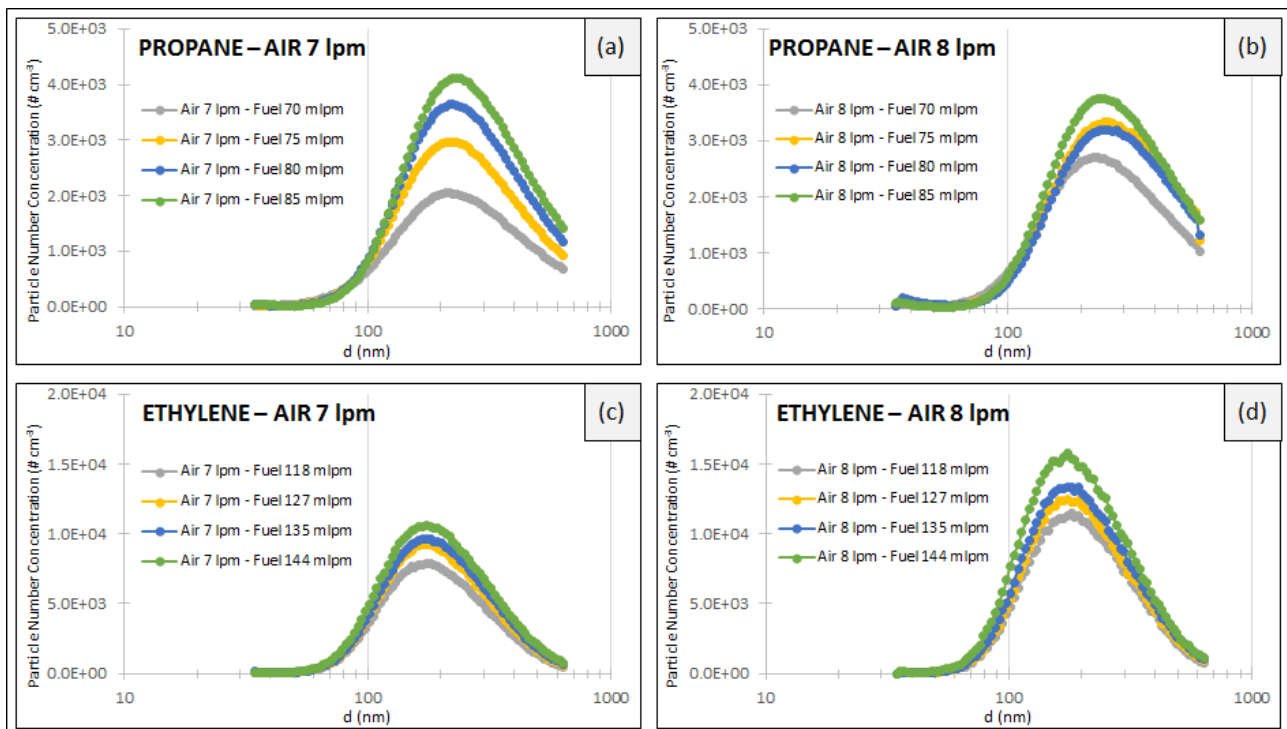
Figure 45: Mode diameter versus the global equivalence ratio. Each point is indicated by E or P (ethylene or propane) and a pair of numbers indicating air and fuel flow rate, respectively, in lpm and mlpm. Dotted lines aim to facilitate the reader eye.

Even if the direct comparison between our findings and results from previous works (Bischof et al., 2019; Kazemimanesh et al., 2019; and Moallemi et al., 2019) are not directly comparable (since feeding flows and global equivalence ratios are different), some similarities can be identified. Previous works observed that by increasing the fuel flow, the particle number concentration increases too, that is what in agreement with what we observed for both the fuels propane. In addition, Bischof (2019) also reported that with propane the particle mode diameter, with propane, did not depend on the global equivalence ratio; as we also observed, but this behaviour for ethylene instead. Kazemimanesh (2019) showed a clear increase in mode diameter, corresponding to an increase of fuel flow rate, that reached a quite constant value (i.e., around 240-270 nm) for ethylene. This trend differs from our observations, since the mode diameter in our case turned out to be quite stable at about 175 nm independently on feeding flows. This difference is probably due to the global equivalence ratios used: while in (Kazemimanesh et al., 2019) global equivalence ratios are lower than 0.206, in our case they are higher than 0.213. In (Moallemi et al., 2019), instead, they observed an opposite behaviour

352 for mode diameters: they retrieved that at fixed fuel flow, a higher air flow produced a slight decrease of the  
 353 mode diameter. Both (Moallemi et al., 2019) and (Bischof et al., 2019) measured mode diameters  $< 200$  nm,  
 354 but they used different combustion conditions (i.e., lower global equivalence ratios resulting from higher air  
 355 flow or lower fuel flow). We can conclude that, as expected, global equivalence ratio is the principal parameter  
 356 affecting size distributions of soot particles.

357 Both in (Kazemimanesh et al., 2019) with ethylene and (Moallemi et al., 2019) and (Bischof et al., 2019)  
 358 with propane, when the fuel flow increased, at a certain air flow, the particle number concentration increased  
 359 too; even if a lower range of global equivalence ratio were considered. In addition, Kazemimanesh (2019) and  
 360 Bischof (2019) also reported that the particle mode diameter, both with ethylene and propane, did not depend  
 361 on the global equivalence ratio, as we also observed. In (Moallemi et al., 2019), instead, they observed an  
 362 opposite behaviour for mode diameters: they retrieved that at fixed fuel flow, a higher air flow produced a  
 363 slight decrease of the mode diameter. Both (Moallemi et al., 2019) and (Bischof et al., 2019) measured mode  
 364 diameters  $< 200$  nm, but this can be due to the specific combustion conditions (i.e., lower global equivalence  
 365 ratios resulting from higher air flow or lower fuel flow).

366 The mean size distributions observed at ChAMBRe are given in Fig. 5, for all the selected operative  
 367 conditions. All the curves are normalized to the same injection time (i.e., 3 min).

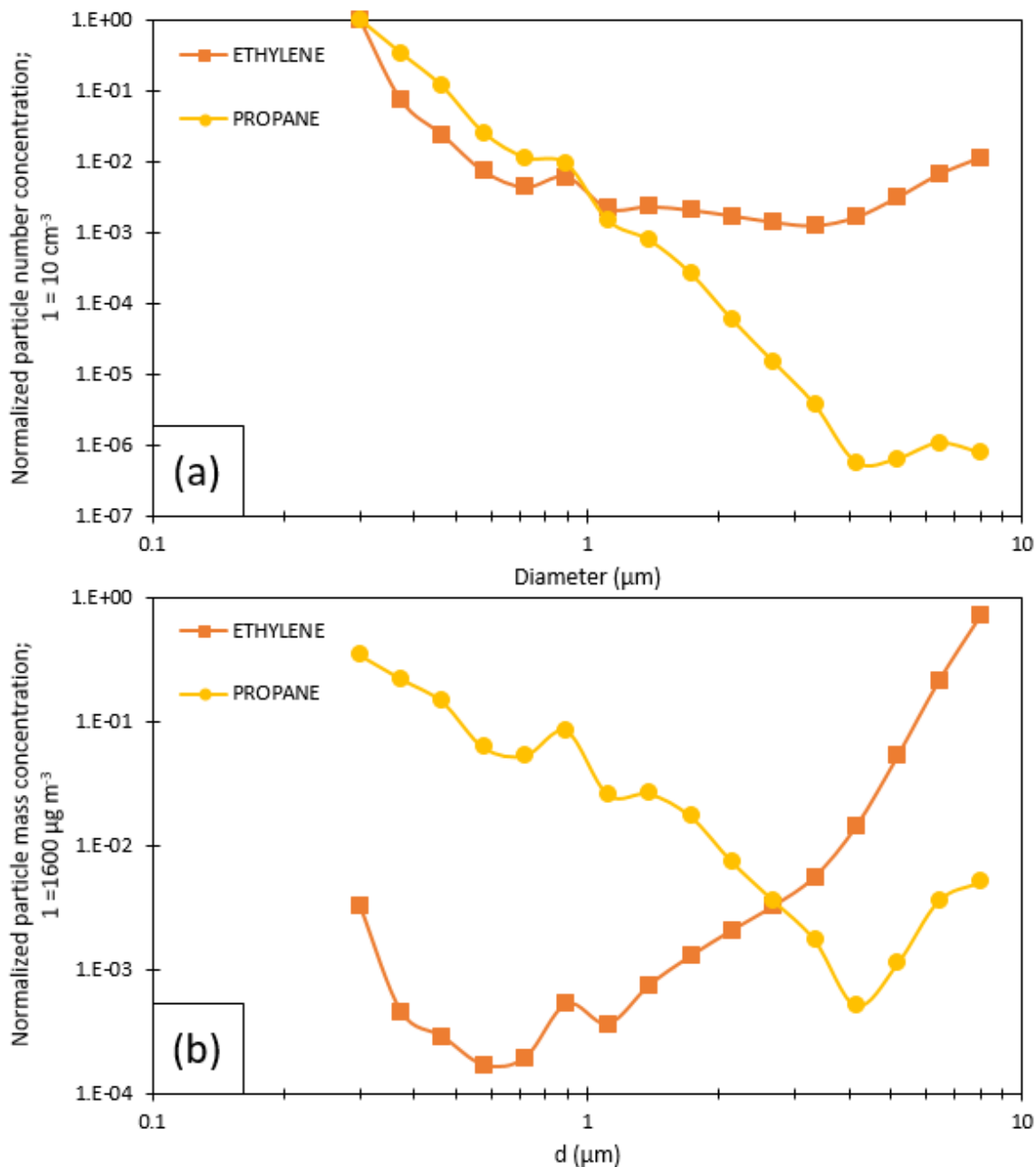


368 *Figure 5: Mean size distributions measured by SMPS. MISG was fuelled with propane (top panels) and ethylene (bottom*  
 369 *panels) with the air and fuel flows indicated in the plots frame.*

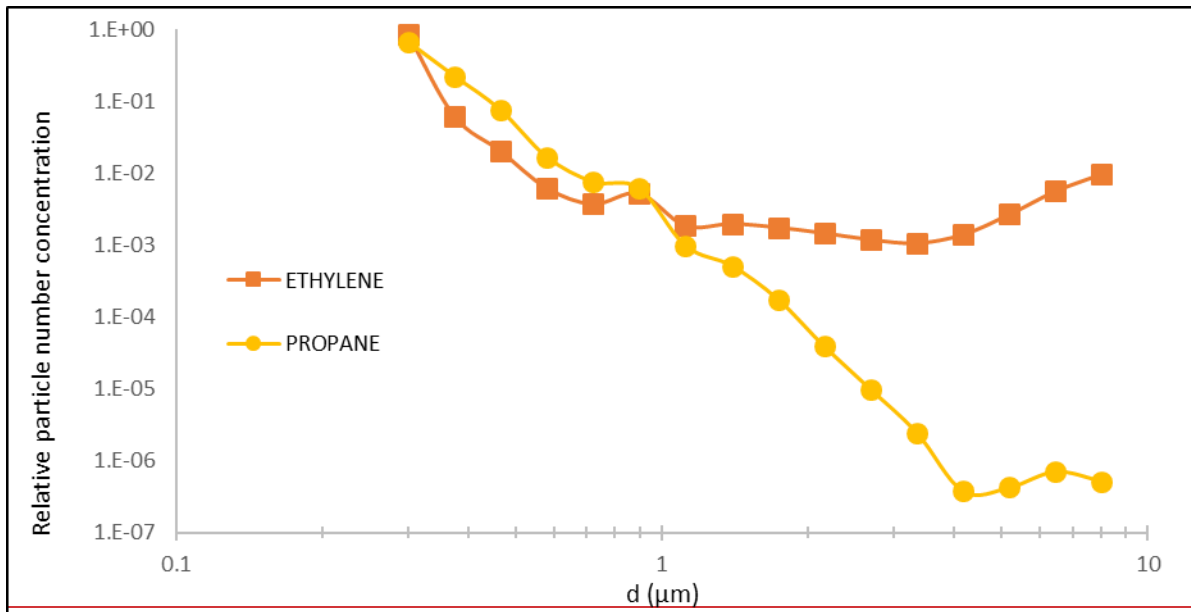
371 Significant differences between the two fuels emerge when considering the ~~particle mass concentration~~  
 372 ~~super-micrometric fraction range~~ (i.e., diameter  $> 1$  extended to  $10 \mu\text{m}$ , including the data collected ~~measured~~  
 373 by the OPS): ethylene combustion produced a ~~limited~~ number of big particles, likely super-aggregates,  
 374 ~~probably formed in the stagnation plane at the bottom part of the combustion cell~~ (Chakrabarty et al., 2012)  
 375 ~~formed directly at the MISG exhaust. This hypothesis was confirmed by dedicated experiments with the setup~~  
 376 ~~specifically modified in respect to the basic one~~ (see Supplementary Fig. S.2). -Kazemimanesh (2019) also  
 377 observed the formation of aggregates, even with smaller dimensions (i.e., about  $2 \mu\text{m}$  ~~of maximum Feret~~  
 378 ~~diameter~~). ~~The particle number concentration, normalized to the total particle number concentration, is shown~~  
 379 ~~in Fig. 6.a~~. We calculated the super-micrometric fraction of the total ~~number concentration~~ measured by the  
 380 OPS with both the fuels (Fig. 6.a): this resulted to be about 3% with ethylene and 0.2% with propane. Particles

381 larger than 4  $\mu\text{m}$  (i.e., optical equivalent diameter) were about 2% with ethylene, with a peak at 8  $\mu\text{m}$ , and  
382 totally negligible with propane. Considering the particle volume-mass distribution (see Fig. 6.b), the latter  
383 difference is obviously enhanced: the super-micrometric fraction is about 99% of the total mass concentration  
384 with ethylene and 9% only with propane. Particles larger than 4  $\mu\text{m}$  contribute to the total volume-mass (and  
385 hence to the soot concentration) for about 98% and 1%, respectively with ethylene and propane.

386 Anyway, super-aggregates formation by ethylene combustion can be partly reduced by using lower air and  
387 fuel flow rates (see Supplementary Fig. S.3 for example).



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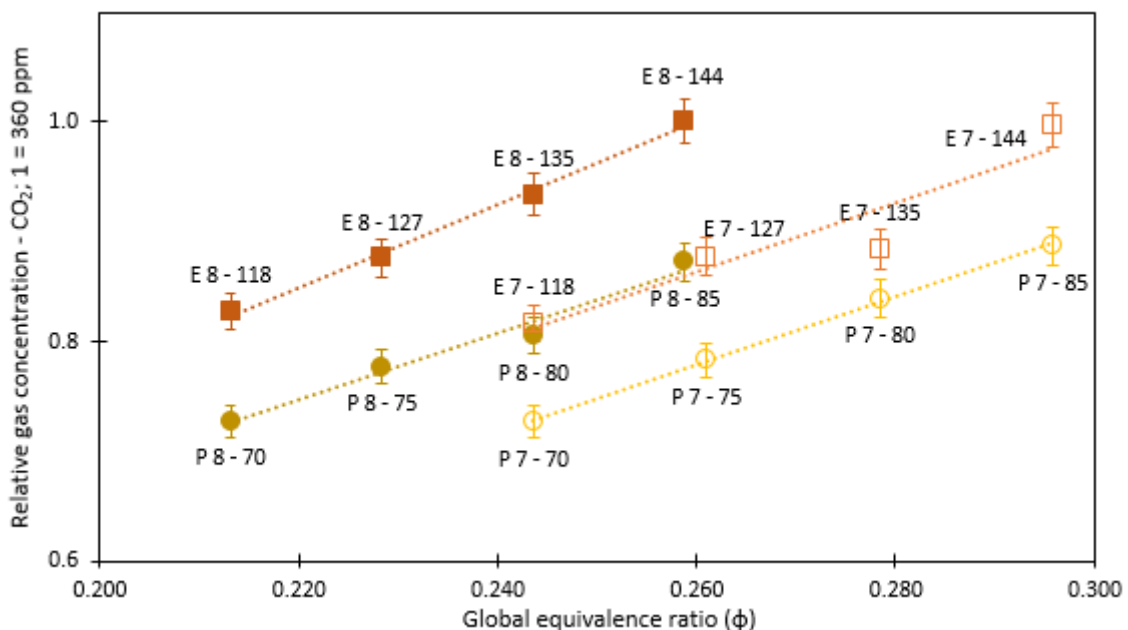
389

390 *Figure 6: Relative particle number concentration normalized to the total number concentration of particles, in function*  
 391 *of particle size. Particle concentration normalized to the total vs. particle diameter, measured by OPS; panel (a) shows*  
 392 *number distribution, panel (b) shows mass distribution. MISG was fuelled with 7 lpm of air and 75 mlpm of fuel during*  
 393 *propane experiment and 127 mlpm of fuel during ethylene experiment. No cyclone used.*

394

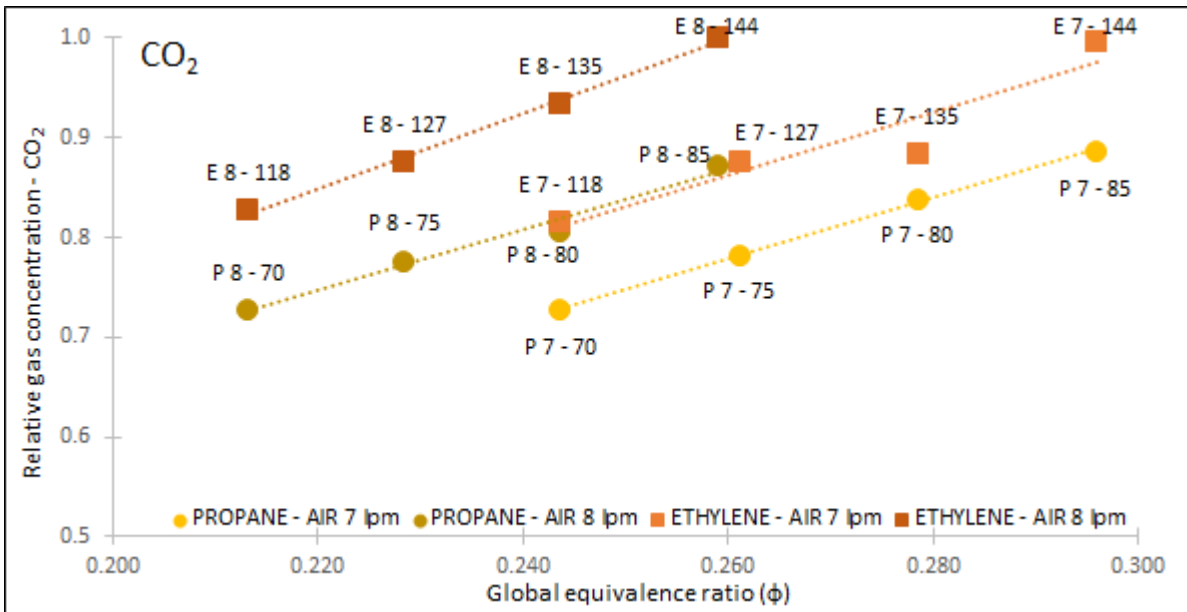
### 3.2.2 Gaseous exhaust

395 Gaseous emissions were characterized too, focusing on the most abundant gases i.e., CO<sub>2</sub> and NO. The  
 396 pattern is similar for both the gases: at fixed air flow rate, gas concentration increased with the fuel flow while  
 397 no significant differences emerged at fixed fuel flow rate and changing the air flow. At equal operative  
 398 conditions (i.e., same combustion conditions, injection time and time from the injection), gaseous emissions  
 399 were higher with ethylene than with propane. With the same normalization introduced in Fig. 3, the CO<sub>2</sub> and  
 400 NO production are compared in Fig. 7 and 8 for each selected MISG configuration. Maximum values were  
 401 360 ppm and 980 ppb, respectively for CO<sub>2</sub> and NO, after 3 minutes of soot injection.



402





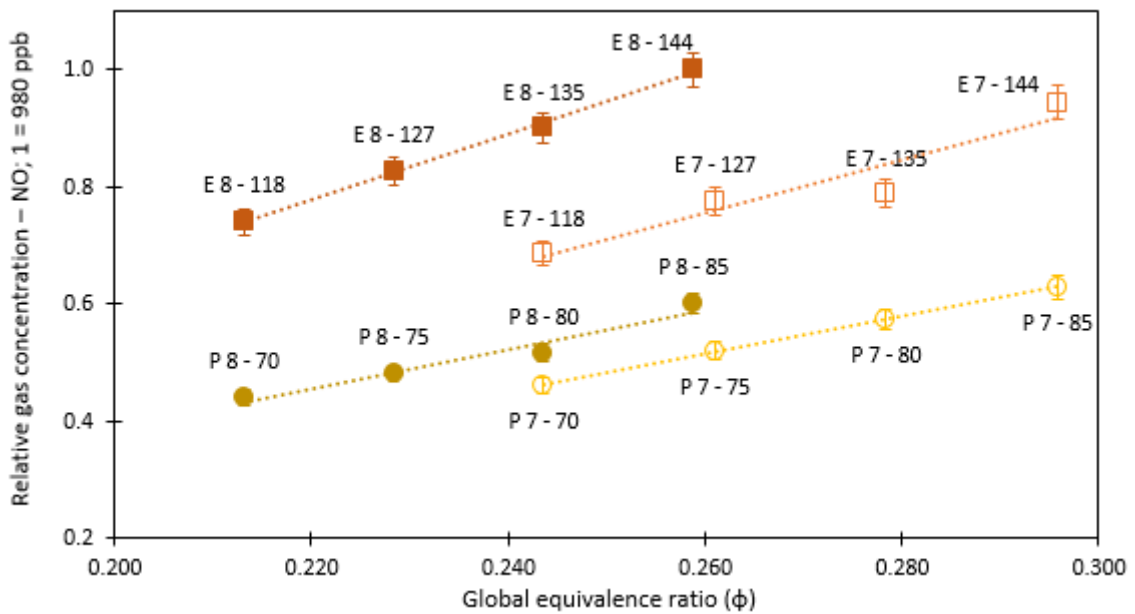
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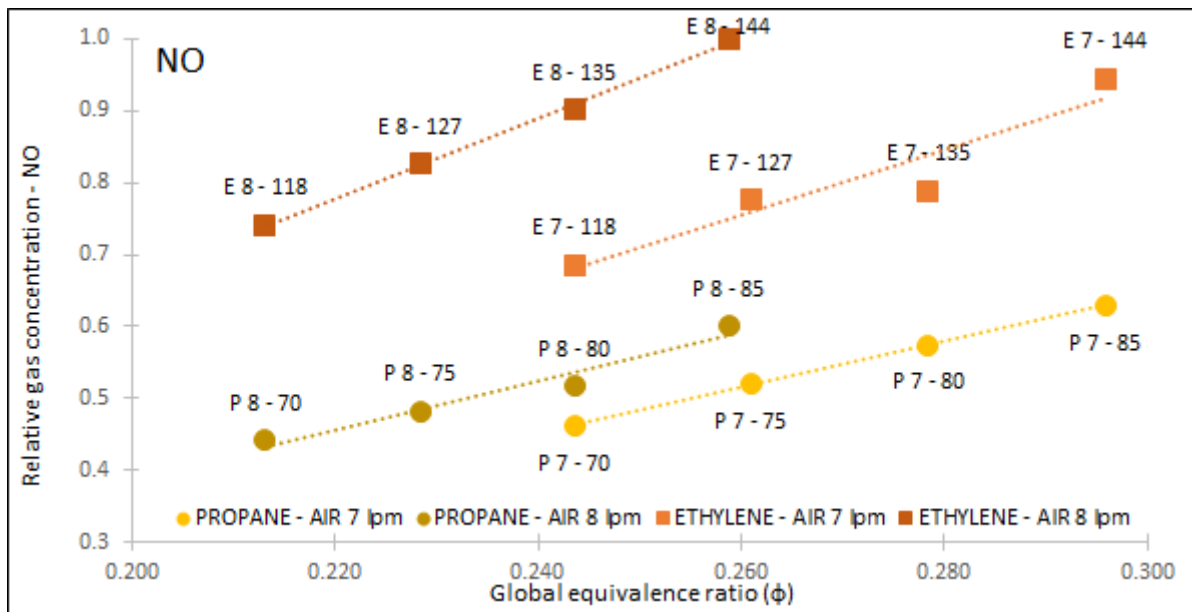
405

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Figure 7: CO<sub>2</sub> concentration versus the global equivalence ratio. Each value was normalized to the highest of the whole set. Data points are labelled by E or P (ethylene or propane) and a pair of numbers indicating air and fuel flow, respectively in lpm and mlpm. Dotted lines aim to facilitate the reader eye.



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Figure 8: NO concentration versus the global equivalence ratio. Each value was normalized to the highest of the whole set. Data points are labelled by E or P (ethylene or propane) and a pair of numbers indicating air and fuel flow, respectively in lpm and mlpm. Dotted lines aim to facilitate the reader eye.

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### 3.2.3 EC/OC quantification

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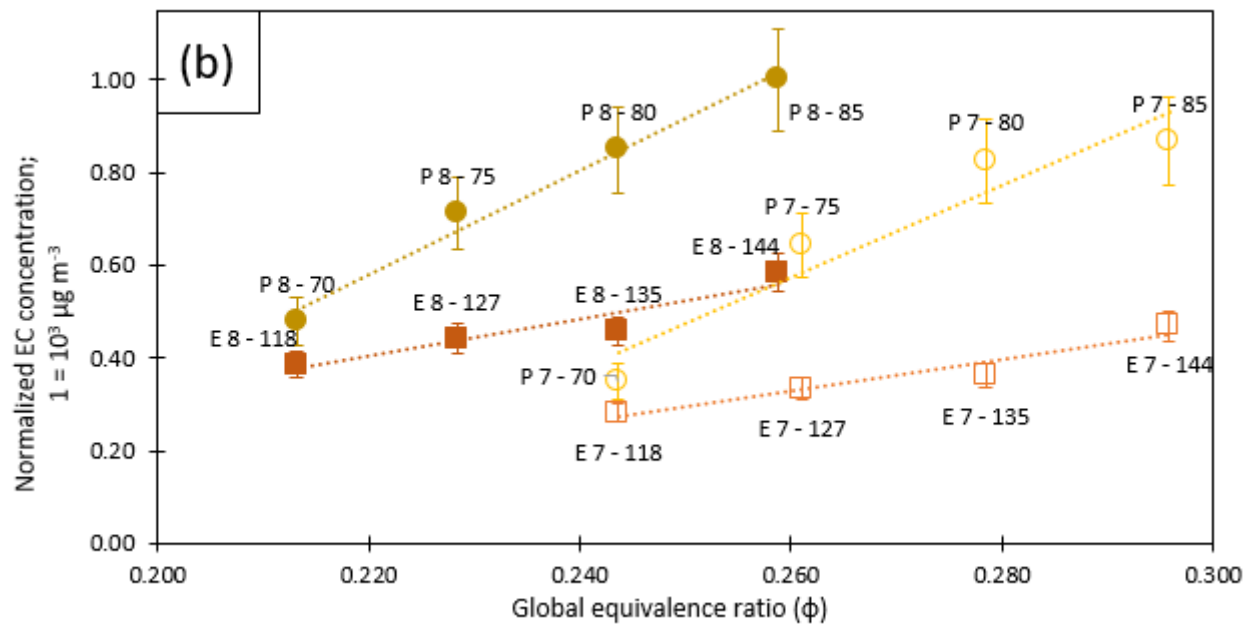
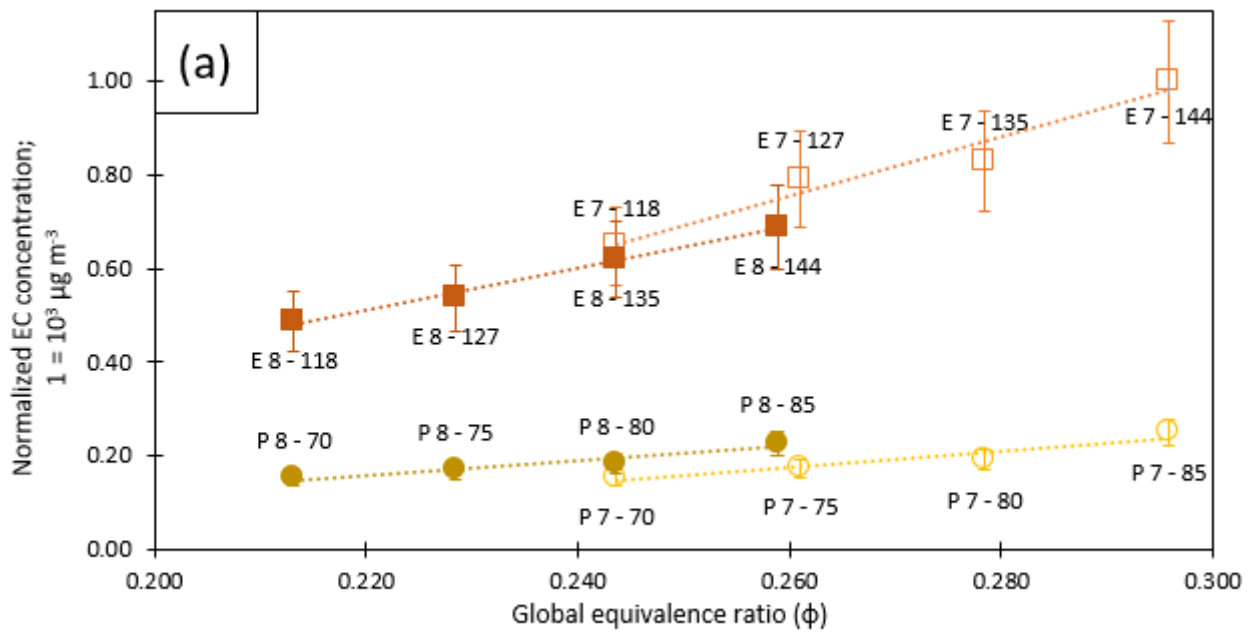
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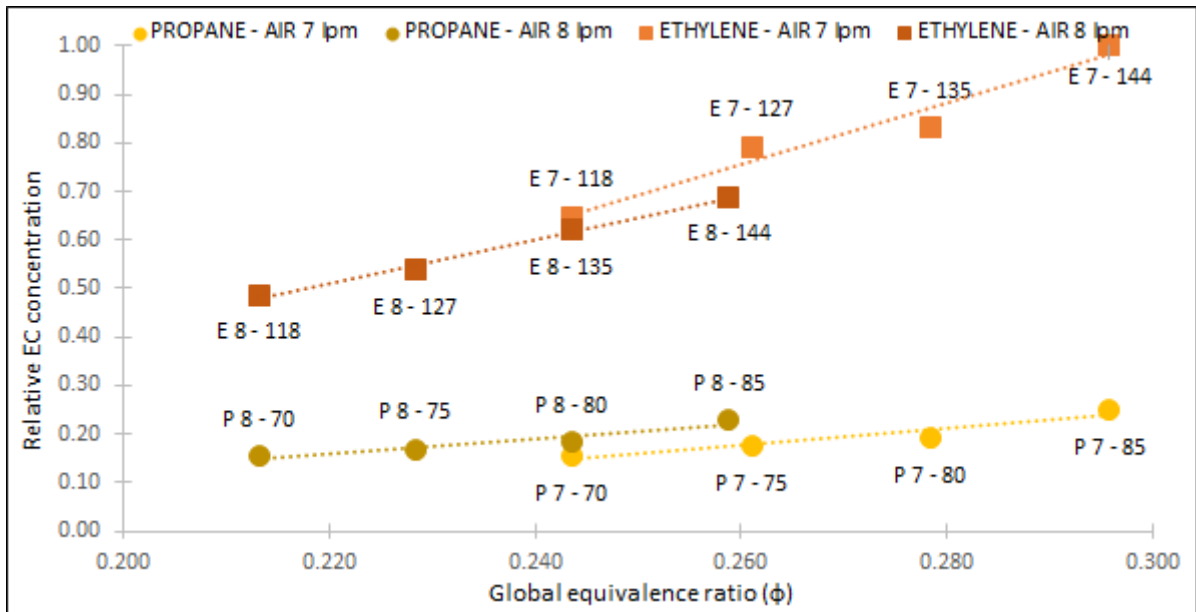
419

The OC/EC composition was quantified by thermal-optical analysis of samples collected on quartz fibre filters during each experiment. EC:TC concentration ratios resulted to be around 0.7 and 0.9 with propane and ethylene, respectively. In addition, the EC:TC concentration ratios increased with the global equivalence ratio. All the results are given in Fig. 9a and 9b and 10, for experiments without and with cyclone, respectively, adopting the same normalization already introduced in Fig. 3. When removing large particles (see Sect 3.2.1), the EC:TC concentration ratio resulted higher with propane (0.83 against 0.79 measured with ethylene). Actually it is worthy to note that, with ethylene about 40 % of the EC concentration was associated with

420 particles larger than 1  $\mu\text{m}$ . With both fuels, EC:OC ratios increase with the global equivalence ratios whether  
 421 the cyclone is present or not, in agreement with (Kazemimanesh et al., 2019) and (Moallemi et al., 2019).



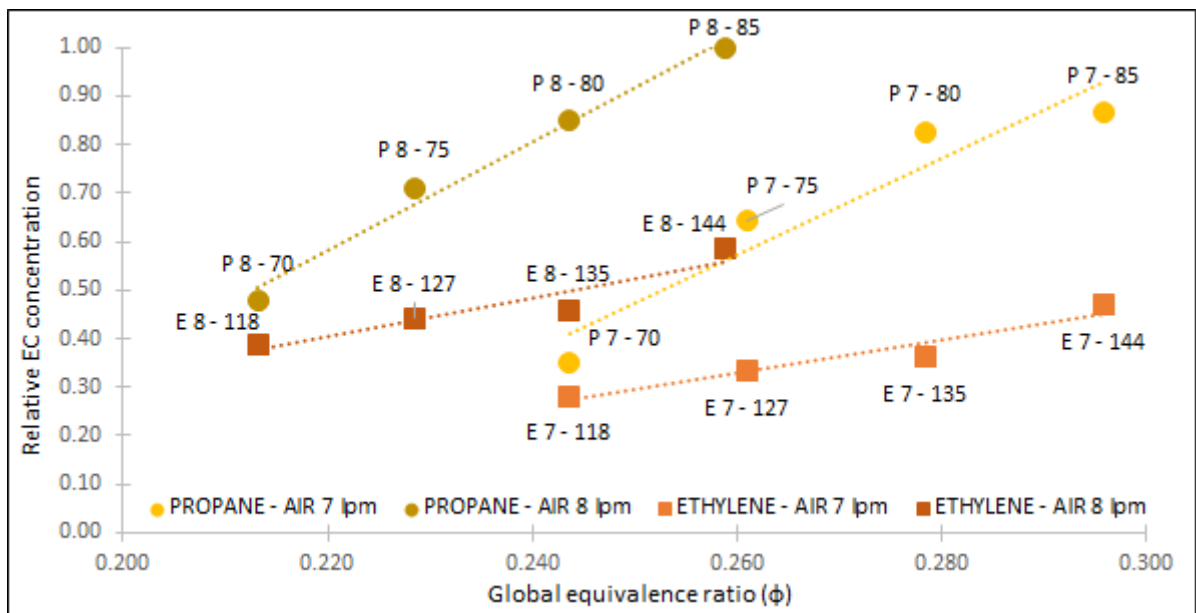
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423

424 Figure 9: EC mass concentration versus the global equivalence ratio, each value was normalized to the highest of the  
 425 whole set. Each point is labelled by E or P (ethylene or propane) and a pair of numbers indicating air and fuel flow rate,  
 426 respectively in lpm and mlpm. (a): ~~No cyclone~~ (a); (b) cyclone impactor upstream the filter (b). Dotted lines aim to  
 427 facilitate the reader eye.

428



429

430 Figure 10: EC mass concentration versus the global equivalence ratio, each value was normalized to the highest of the  
 431 whole set. Each point is labelled by E or P (ethylene or propane) and a pair of numbers indicating air and fuel flow rate,  
 432 respectively in lpm and mlpm. The cyclone impactor upstream the filter removed super-micrometric particles.  
 433

434 The OC:TCEC ratio varies from 0.2734 for propane to 0.119 for ethylene, without cyclone and 0.420 for  
 435 ethylene to 0.16 for propane, when the with and without cyclone was used respectively. In each series of  
 436 experiments (i.e., air flow rate 7 or 8 lpm, ethylene or propane) the OC fraction turned out to be inversely  
 437 proportional to the fuel flow with a minimum at the lowest fuel flow (i.e., 70 lpm with propane and 118 lpm  
 438 with ethylene). This is likely due to the shape of the flame: flames generated by the lowest fuel flow conditions

439 are *Partially Open tip*, with less capability to generate soot particles and hence EC; so that the EC:TCOC ratio  
440 results lower.

441 We also performed some tests ~~adding a backup filter during the sampling to catch to determine~~ the volatile  
442 fraction of OC. The OC concentration values measured on backup filters showed high variability, but they  
443 were compatible with those on not-sampled filters. We analysed 13 blank filters from different bunches and  
444 the average concentration of OC resulted  $\langle \text{OC} \rangle = 0.5 \pm 0.2 \mu\text{g cm}^{-2}$  while OC concentration on backup filters  
445 was  $\langle \text{OC}_{\text{BF}} \rangle = 0.6 \pm 0.2 \mu\text{g cm}^{-2}$  ~~(Since the average OC concentration on the corresponding main filters was~~  
446  ~~$1.4 \pm 0.7 \mu\text{g cm}^{-2}$ ) and the average EC concentration collected on this subset of filter was  $12.3 \pm 0.2 \mu\text{g cm}^{-2}$ ,~~  
447 ~~the volatile fraction phase can be considered negligible.~~ A relationship between OC concentration on the  
448 backup filter and the global equivalence ratio was instead reported in (Kazemimanesh et al., 2019). Actually,  
449 in that study the range of investigated global equivalence ratio values was  $0.129 < \phi < 0.186$  to be compared  
450  $\phi > 0.210$  adopted in this work.

451

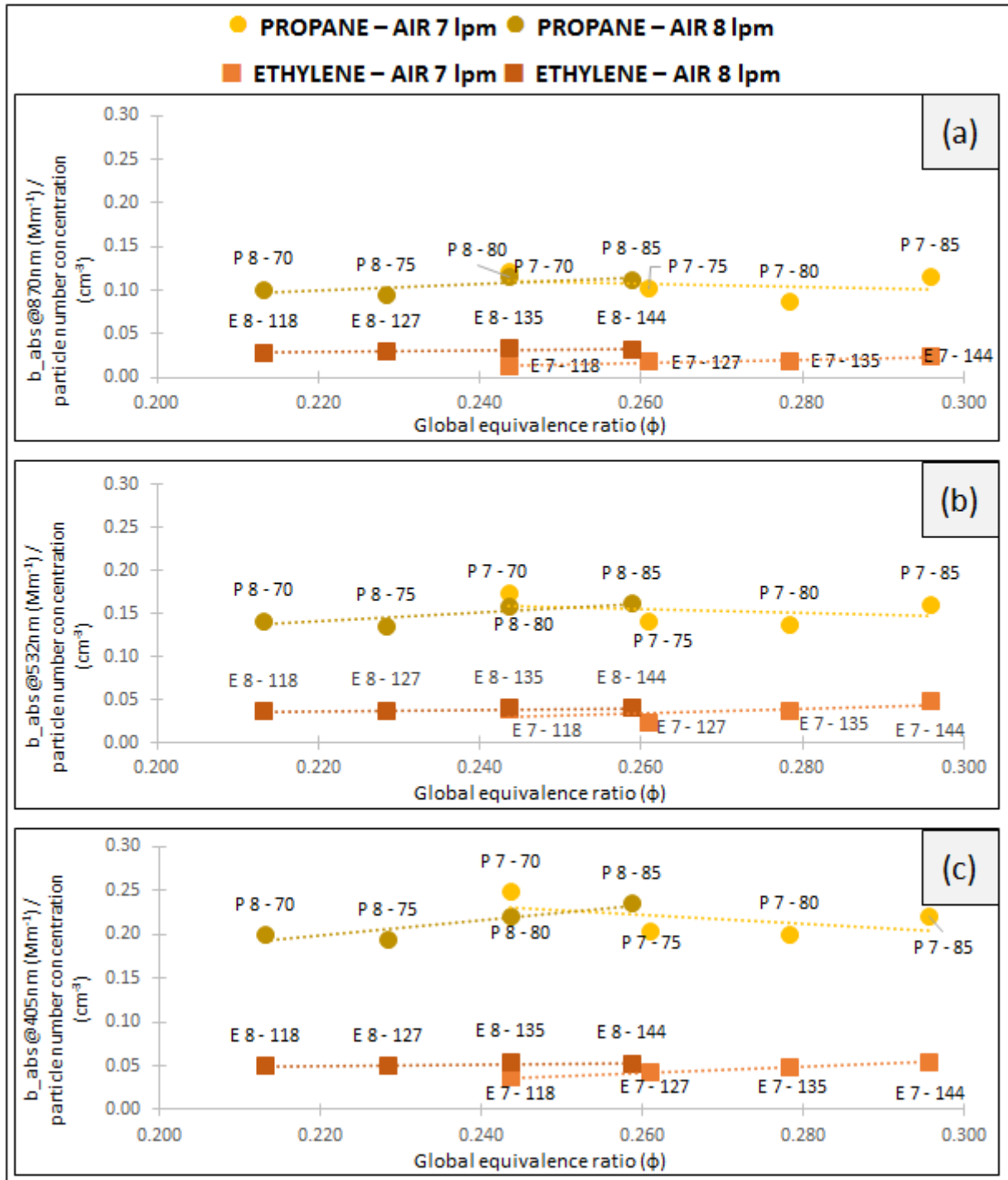
### 452 3.2.4 Optical properties

453 The optical properties of the MISG aerosol were determined in terms of the absorption coefficient ( $b_{\text{abs}}$ ;  
454 i.e. the absorbance per unit length) (Massabò and Prati, 2021). The  $b_{\text{abs}}$  definition applies both to  
455 measurements directly performed on the aerosol dispersed in the atmosphere (by PAXs, in this work) and to  
456 off-line analyses on aerosol sampled on filters (by MWAA, in this work), provided a proper data reduction is  
457 adopted (Massabò and Prati, 2021; and references therein).

458 The online measured  $b_{\text{abs}}$  values were normalized to the total particle number concentration inside  
459 ChAMBRé reached in each single experiment. ~~Absorption coefficients measured at three wavelengths by the~~  
460 ~~PAXs and with the cyclone mounted upstream, are shown in Fig. 11. Similar results were obtained even for~~  
461 ~~experiments without cyclone and for the  $b_{\text{abs}}$  values measured by the MWAA.~~ At each wavelength, the  $b_{\text{abs}}$   
462 values did not show any dependence on the global equivalence ratio, with the propane producing particles  
463 more absorbent than ethylene (see Supplementary Fig. S.4 and S.5, for the experiments without and with  
464 cyclone, respectively). Similar results were obtained even for experiments without cyclone and for the  $b_{\text{abs}}$   
465 values measured by the MWAA. Optical properties such as absorption depend on several parameters, mainly  
466 composition, mixing state, aging, and size. Considering all the experiments reported in this work, no  
467 differences in composition can be expected, since only EC particles were present: this means that differences  
468 in absorption cannot depend on particle composition. Also mixing state and aging cannot explain this  
469 difference: soot inside the chamber was fresh. We can explain the higher light absorbing capability of propane  
470 by considering differences in: size distributions (see Figs. 3-5) and morphology/density of the particles  
471 produced by the burning of the two different fuels.

472 ~~The comparison with previous literature (Moallemi et al., 2019) is hampered by methodological differences~~  
473 ~~(Moallemi and co-workers used the IR-PAX only and reported the Single Scattering Albedo instead of the~~  
474 ~~absorption coefficient). In the literature, only data for the IR-PAX in terms of Single Scattering Albedo (SSA)~~  
475 ~~for propane soot are reported. SSA(IR) values, measured during propane experiments, in our work varied from~~  
476 ~~0.15 to 0.18, in agreement with those obtained by (Moallemi et al., 2019), which ranged from 0.15 to 0.25.~~

477



478

479 *Figure 11: Absorption coefficient @  $\lambda = 870$  (a), 532 (b) and 405 (c) nm, measured by PAXs, versus the global*  
 480 *equivalence ratio.  $b_{abs}$  values are normalized to the total particle number concentration measured by SMPS in the*  
 481 *corresponding experiments. Each point is labelled by E or P (ethylene or propane) and a pair of numbers indicating air*  
 482 *and fuel flow, respectively in lpm and mlpm.*

483

### 3.2.5 Mass Absorption Coefficient

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The  $b_{abs}$  values, together with the EC concentration measured on the filter sampled during each single experiment, can be used to retrieve the Mass Absorption Coefficient (MAC) of the produced aerosol, through the relation:

$$b_{\text{abs}}(\lambda) = \text{MAC} * [\text{EC}] \quad \text{Eq. 2}$$

where:

$b_{\text{abs}}$  [ $\text{Mm}^{-1}$ ]: absorption coefficient

MAC [ $\text{m}^2 \text{g}^{-1}$ ]: Mass Absorption Coefficient

EC [ $\mu\text{g m}^{-3}$ ]: Elemental Carbon concentration

The  $b_{\text{abs}}$  values were calculated directly online by the PAXs and offline by the MWAA analysis, performed at five wavelengths on the sampled filters (see Sect. 2.5). This gave the possibility to extend the characterization of the MISG and to compare two optical analyses on the same carbonaceous aerosol. Since experiments were repeated with two different setups (i.e., with and without the cyclone) and two different fuels (propane and ethylene), four different particle populations can be compared. The comparison was carried out at the three wavelengths (nearly) common to PAXs and MWAA (i.e.,  $\lambda = 870/850, 532$  and  $405 \text{ nm}$ ), as reported in Fig. 11-13. Fig 10 shows the comparison at  $\lambda = 870/850 \text{ nm}$ , while comparison at  $\lambda = 532$  and  $405 \text{ nm}$  are reported in Supplementary (see Fig. S.6 and S.7, respectively). We divided the results by fuel, air flow and with/without cyclone. Each point in the plots sums-up the observations at different global equivalence ratio values. All the measured MAC values, including the other two wavelengths available for the MWAA (i.e.,  $635$  and  $375 \text{ nm}$ ) too, are summarized in Table 5.

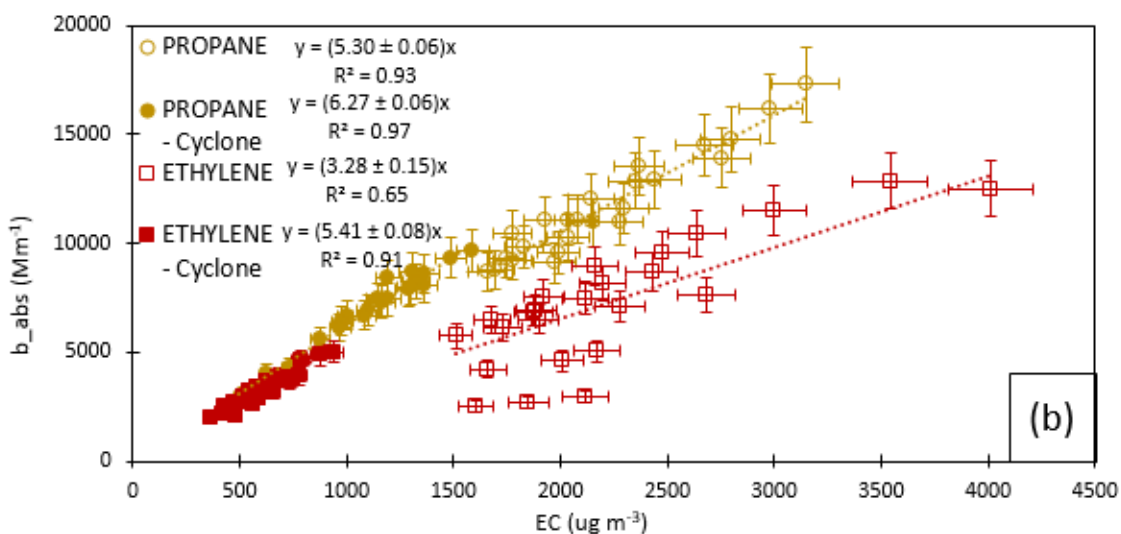
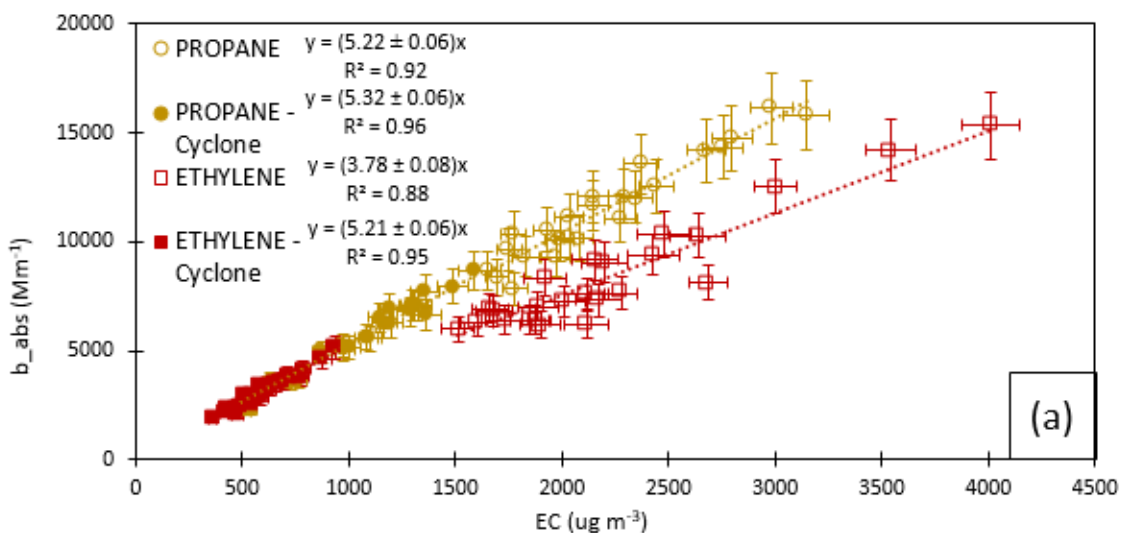
Table 5: Summary of the measured MAC values, in  $\text{m}^2 \text{g}^{-1}$ .

FUEL	PAX			MWAA				
	870 nm	532 nm	405 nm	850 nm	635 nm	532 nm	405 nm	375 nm
PROPANE	$5.30 \pm 0.06$	$8.35 \pm 0.08$	$10.55 \pm 0.11$	$5.22 \pm 0.06$	$7.22 \pm 0.09$	$8.81 \pm 0.09$	$10.55 \pm 0.09$	$10.86 \pm 0.12$
PROPANE with cyclone	$6.27 \pm 0.06$	$10.26 \pm 0.06$	$13.48 \pm 0.08$	$5.32 \pm 0.06$	$7.37 \pm 0.07$	$8.95 \pm 0.08$	$10.91 \pm 0.11$	$11.59 \pm 0.14$
ETHYLENE	$3.28 \pm 0.15$	$4.92 \pm 0.19$	$5.89 \pm 0.20$	$3.78 \pm 0.08$	$5.00 \pm 0.09$	$5.91 \pm 0.11$	$6.90 \pm 0.12$	$7.28 \pm 0.14$
ETHYLENE with cyclone	$5.41 \pm 0.08$	$10.42 \pm 0.12$	$15.74 \pm 0.15$	$5.21 \pm 0.06$	$7.62 \pm 0.07$	$9.53 \pm 0.08$	$12.29 \pm 0.10$	$13.03 \pm 0.11$

The MWAA analysis at  $\lambda = 870 \text{ nm}$  (Fig. 12.a) returns compatible MAC values for both the propane series (with/without cyclone) and the ethylene series with cyclone (average MAC =  $5.25 \pm 0.10 \text{ m}^2 \text{g}^{-1}$ ). At  $\lambda = 532$  and  $405 \text{ nm}$  (Fig. 13.a and Fig. 14.a), propane series are still in agreement while the ethylene series with cyclone show a higher MAC value (MAC =  $9.53 \pm 0.08 \text{ m}^2 \text{g}^{-1}$  instead of MAC =  $8.88 \pm 0.13 \text{ m}^2 \text{g}^{-1}$  at  $\lambda = 532 \text{ nm}$  and MAC =  $12.3 \pm 0.1 \text{ m}^2 \text{g}^{-1}$  instead of MAC =  $10.7 \pm 0.2 \text{ m}^2 \text{g}^{-1}$  at  $\lambda = 405 \text{ nm}$ ). The ethylene series without cyclone show consistent variability at all the three wavelengths, with the lowest MAC values of the whole data set (MAC =  $3.78 \pm 0.08 \text{ m}^2 \text{g}^{-1}$ ; MAC =  $5.9 \pm 0.1 \text{ m}^2 \text{g}^{-1}$ ; MAC =  $6.9 \pm 0.1 \text{ m}^2 \text{g}^{-1}$  at  $\lambda = 870, 532$  and  $405 \text{ nm}$  respectively): the differences are probably due to the production of super-micrometric particles (see Sect 3.2.1 and Fig. 6) when the cyclone is not used. With the PAXs analysis (Fig. 12.b, 13.b and 14.b), MAC values turn out higher in the series with cyclone (average values are MAC =  $5.8 \pm 0.4 \text{ m}^2 \text{g}^{-1}$ ;  $10.3 \pm 0.1 \text{ m}^2 \text{g}^{-1}$ ;  $14 \pm 1 \text{ m}^2 \text{g}^{-1}$  at  $\lambda = 870, 532$  and  $405 \text{ nm}$  respectively, with cyclone, MAC =  $4.3 \pm 0.1 \text{ m}^2 \text{g}^{-1}$ ;  $6.6 \pm 1.7 \text{ m}^2 \text{g}^{-1}$ ;  $8 \pm 2 \text{ m}^2 \text{g}^{-1}$  at  $\lambda = 870, 532$  and  $405 \text{ nm}$  respectively, without cyclone), this happens at all the three wavelengths and for both fuels. If series with cyclone are only considered, MAC values do not show any significant differences depending on the fuel. The ethylene series without cyclone shows the lowest MAC values at each wavelength, as observed with the MWAA analysis. PAXs data show a higher variability in MAC values, likely due to a higher sensitivity to particle size than filter based MWAA analysis.

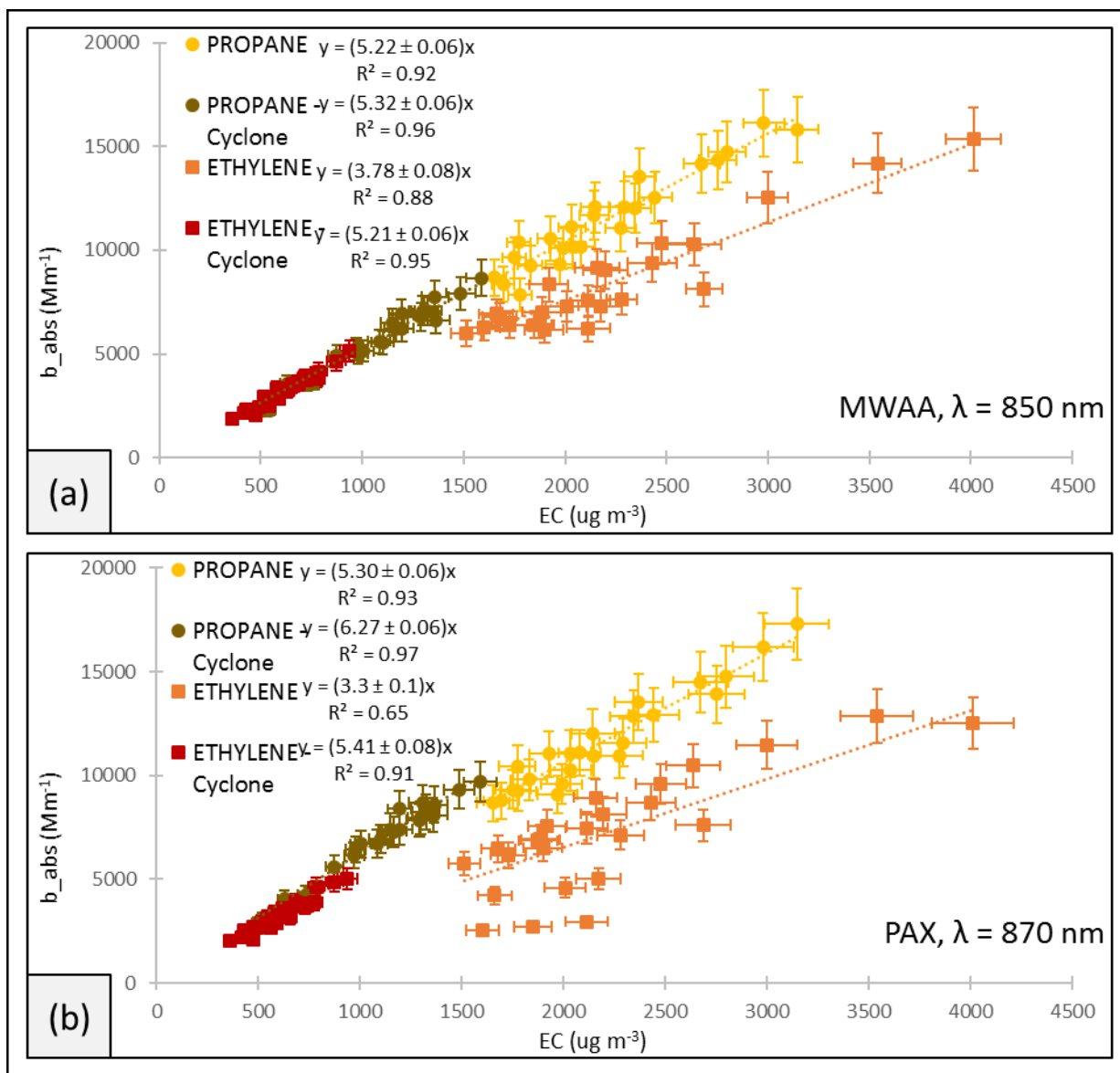
The MWAA analysis at  $\lambda = 870 \text{ nm}$  (Fig. 10.a) returned compatible MAC values for both propane series (with/without cyclone) and ethylene series with cyclone, while a consistently lower MAC value was found for the ethylene series (worse correlation) without the PM1 cutting. The same picture turned out at the other two wavelengths (see Supplementary Fig. S.6 and S.7). By comparing PAX absorption coefficients and EC

525 concentrations at  $\lambda = 870$  nm (Fig. 10.b), obtained MAC values are more variable, with similar values only in  
 526 the case of propane without cyclone and ethylene with cyclone. At  $\lambda = 532$ , in the case of MWAA, similar  
 527 MAC values have been found for both the propane series, while, for ethylene series, MAC values were slightly  
 528 higher when cyclone was used and lower when not. Considering the optical data from PAX, a similar MAC  
 529 for both the fuels was found when the cyclone was present, while it slightly differed in the case of propane  
 530 without cyclone, and it was much lower in the case of ethylene without cyclone. At  $\lambda = 405$  nm, the MWAA  
 531 responses for propane series were still in agreement while the ethylene series showed a higher MAC value  
 532 when using the cyclone, and a lower MAC value without using it. PAX returned a different MAC value for  
 533 each of the four conditions. To summarize, if only series with cyclone are only considered, MAC values show  
 534 only small differences depending on the fuel, larger in the case of PAXs. The ethylene series without cyclone  
 535 showed the lowest MAC values of the whole data set: the most likely reason for this difference is the presence  
 536 of super-micrometric particles (see Sect 3.2.1 and Fig. 6) when the cyclone was not used. With MWAA, the  
 537 MAC values turned out to be the same in all the runs but the case of the ethylene data collected without the  
 538 cyclone. With the PAXs analysis, MAC values turned out higher in the series with cyclone, this happened at  
 539 all the three wavelengths and for both fuels. Since PAXs data showed a higher variability in MAC values,  
 540 photoacoustic measurements are supposed to be more sensitive to particle size than filter based MWAA  
 541 analysis.



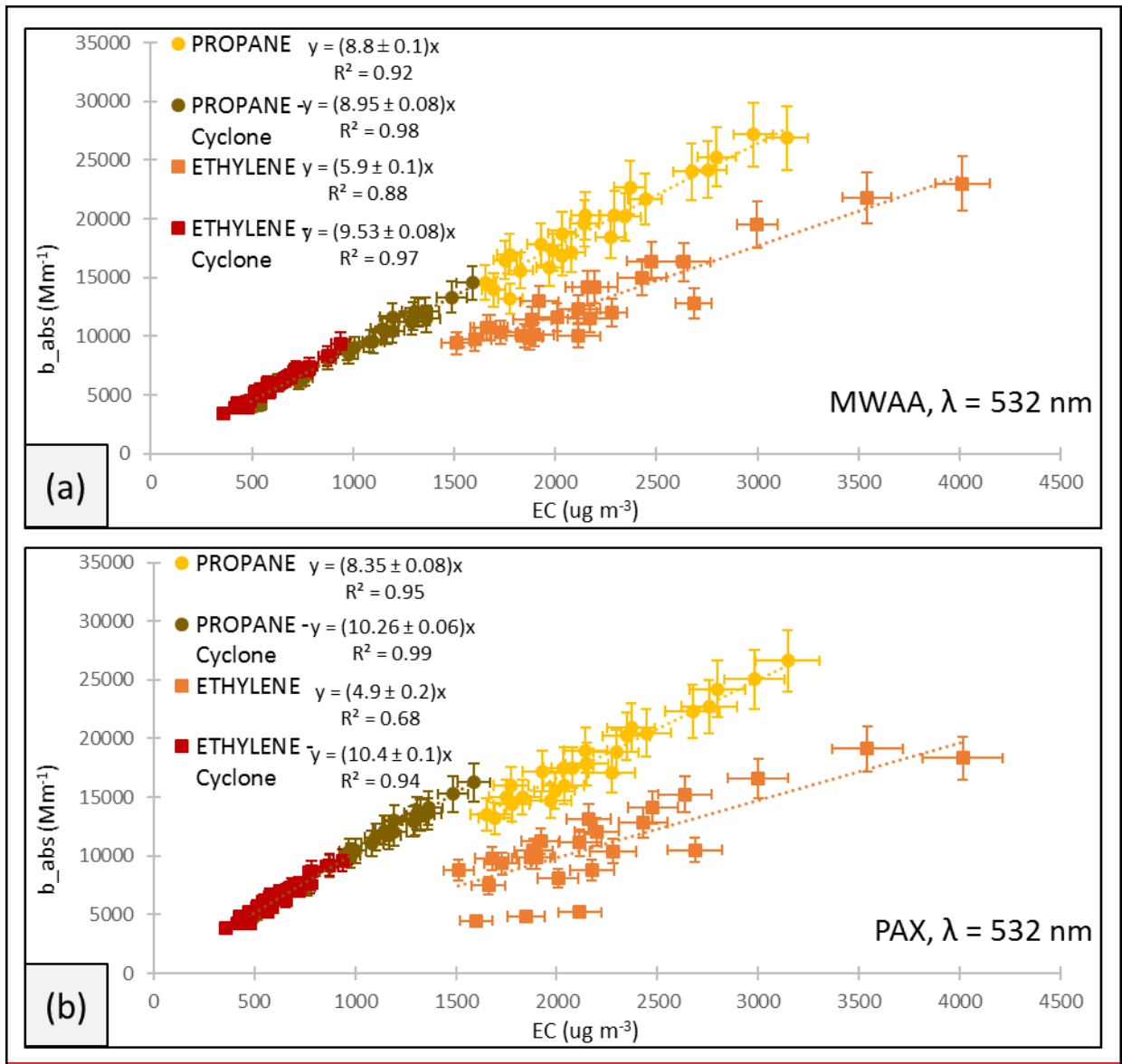
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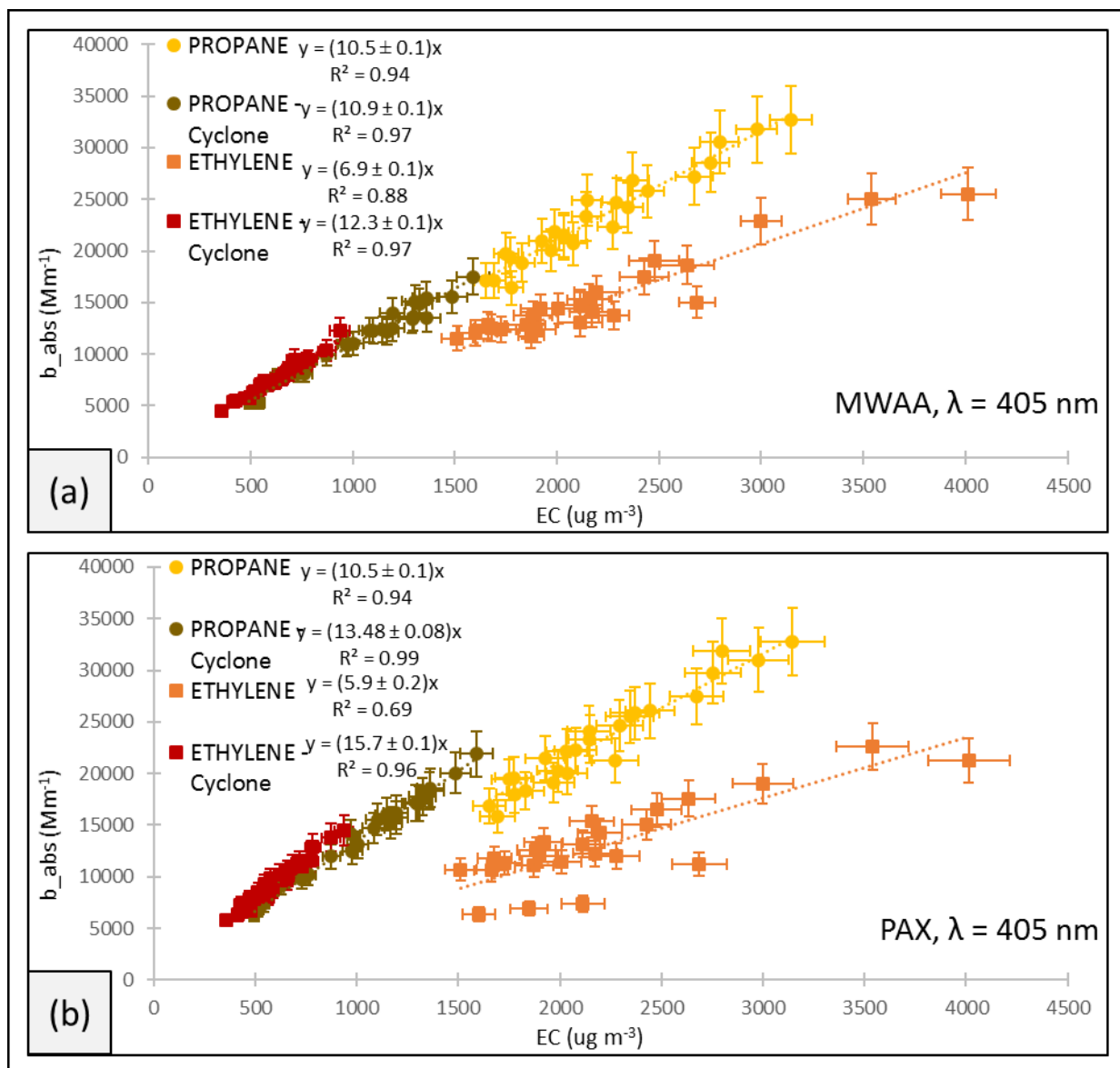
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544 *Figure 102: Absorption coefficient  $\mu_a$  @ 850 nm measured by MWA (a) and @ 870 nm measured by PAX (b) versus EC*  
 545 *concentration. The slope of each fit corresponds to the Mass Absorption Coefficient.*



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547 *Figure 13: Absorption coefficient @ 532 nm, measured by MWA (a) and PAX (b) versus EC concentration. The slope of*  
 548 *each fit corresponds to the Mass Absorption Coefficient.*



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550 *Figure 14: Absorption coefficient at  $\lambda = 405 nm$ , measured by MWA (a) and PAX (b) versus EC concentration. The*  
 551 *slope of each fit corresponds to the Mass Absorption Coefficient.*

552 *A summary of all the measured MAC values, including the other two wavelengths available for the MWA*  
 553 *(i.e., 635 and 375 nm) too, is given in Table 5. Anyway, MAC values are close to theoretical figures for soot*  
 554 *(Bond and Bergstrom, 2006), for both the fuels and at all the wavelengths. IR values are similar to those*  
 555 *obtained by Moallemi (2019) for propane exhaust. With both the fuels MAC values increase when super-*  
 556 *micrometric particles were removed by the cyclone; propane-particles showed higher MAC values than*  
 557 *ethylene ones.*

558 *Table 5: Summary of the measured MAC values, in  $m^2 g^{-1}$ .*

FUEL	PAX			MWAA				
	870 nm	532 nm	405 nm	850 nm	635 nm	532 nm	405 nm	375 nm
PROPANE	5.30 ± 0.06	8.35 ± 0.08	10.5 ± 0.1	5.22 ± 0.06	7.22 ± 0.09	8.8 ± 0.1	10.5 ± 0.1	10.9 ± 0.1
PROPANE with cyclone	6.27 ± 0.06	10.26 ± 0.06	13.48 ± 0.08	5.32 ± 0.06	7.37 ± 0.07	8.95 ± 0.08	10.9 ± 0.1	11.6 ± 0.1
ETHYLENE	3.3 ± 0.1	4.9 ± 0.2	5.9 ± 0.2	3.78 ± 0.08	5.0 ± 0.1	5.9 ± 0.1	6.9 ± 0.1	7.3 ± 0.1
ETHYLENE with cyclone	5.41 ± 0.08	10.4 ± 0.1	15.7 ± 0.1	5.21 ± 0.06	7.62 ± 0.07	9.53 ± 0.08	12.3 ± 0.1	13.0 ± 0.1

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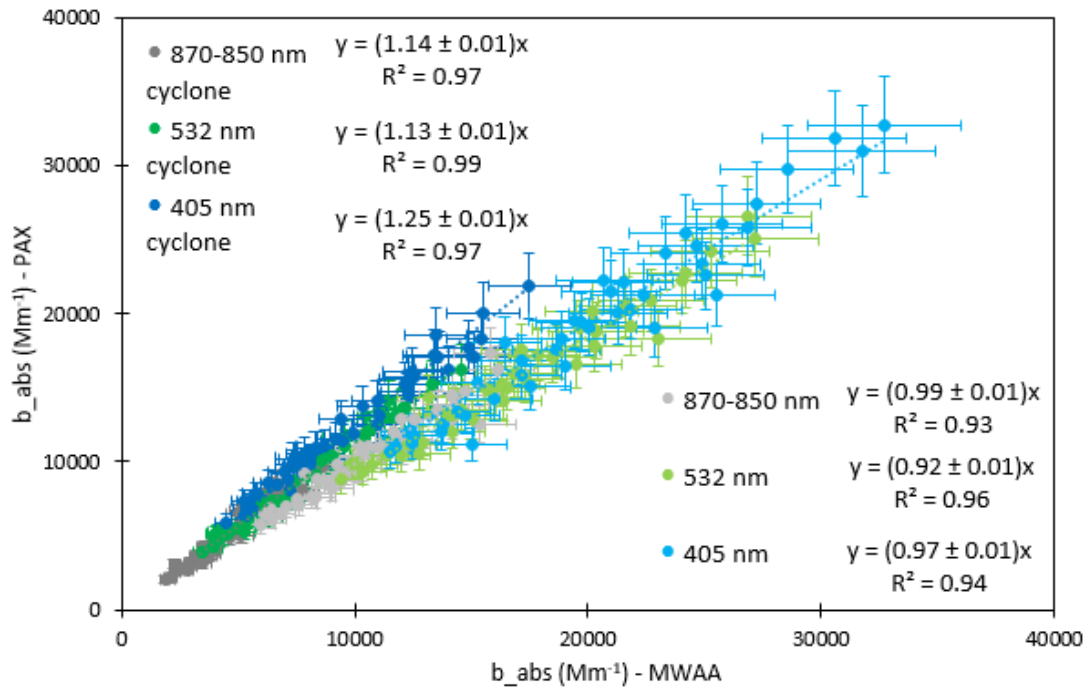
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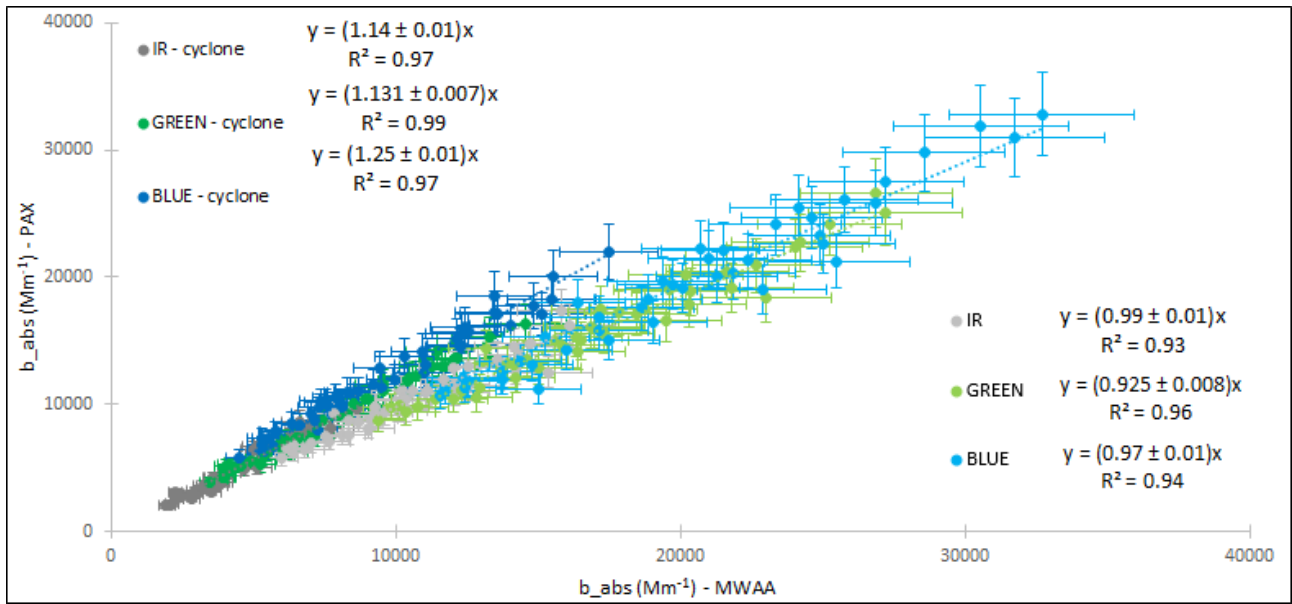
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In (Moallemi et al., 2019), only IR-MAC values are reported for the propane are reported, and they resulted slightly lower than values here quoted reported. This difference could depend on the different techniques used to quantify the EC concentration: we measured used-EC concentration values by from thermal optical analysis while Moallemi (2019) used reported BC concentration measured by LII.

Discrepancies between MAC values obtained by from PAXs and MWAA, for the same experiment, are compatible with the differences of measured  $b_{abs}$  values: the latter. The  $b_{abs}$  values measured by PAXs and MWAA are directly compared in Fig. 115, merging all the data collected by the two setups (i.e., with and without the cyclone) and for the two fuels. The agreement between the two analyses instruments turned out within 25 % and 7 %, respectively without and with the cyclone.



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571 *Figure 115: Correlation study between the absorption coefficient measured by PAX and MWAAs. Colours of dots identify*  
 572 *the wavelength of the analysis: grey refers to 870 nm, green to 532 nm and blue to 405 nm, with cyclone; light grey refers*  
 573 *to 870 nm, light green to 532 nm and light blue to 405 nm; without cyclone; dark and light colours refer to experiments*  
 574 *with and without cyclone, respectively.*

575 In addition, the spectral dependence of the absorption coefficient  $b_{abs}$ , and consequently the Ångström  
 576 Absorption Exponent (AAE, Moosmüller et al., 2011), can be calculated by the power-law:

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$$b_{abs}(\lambda) \approx \lambda^{-AAE}$$

578 where:

- 579  $b_{abs}$  [ $Mm^{-1}$ ]: absorption coefficient
- 580  $\lambda$  [nm]: wavelength used for the analysis
- 581 AAE: Ångström Absorption Exponent.

582 The averages of the resulting AAEs values for the different experimental conditions are reported in Table 6 by  
 583 fitting the data for the 3 and 5 available wavelengths in the case of PAXs and MWAA, respectively.

584 Experimental determinations of the AAE had been reported in the literature as being dependent on aerosol  
 585 chemical composition (Kirchstetter et al., 2004; Utry et al., 2013) and size and morphology (Lewis et al., 2008;  
 586 Lack et al., 2012; Lack and Langridge, 2013; Filep et al., 2013; Utry et al., 2014 a). Particulate generated by  
 587 fossil fuel combustion (i.e., Black Carbon) typically has AAE values close to 1.0 (Harrison et al., 2013, and  
 588 references therein). The AAE values measured in this work for the MISG exhausts are generally close to 1.0  
 589 with higher figures for the cyclone-selected aerosol.

590 *Table 6: AAE values obtained in different experimental conditions through the analysis of PAXs and MWAA raw data.*

EXPERIMENTAL CONDITIONS	AAE - PAX	AAE - MWA
PROPANE 70 to 85 mlpm - AIR 7 lpm	0.88 ± 0.06	0.92 ± 0.04
PROPANE 70 to 85 mlpm - AIR 8 lpm	0.92 ± 0.06	0.91 ± 0.05
PROPANE 70 to 85 mlpm - AIR 7 lpm - cyclone	0.98 ± 0.09	0.99 ± 0.10
PROPANE 70 to 85 mlpm - AIR 8 lpm - cyclone	1.05 ± 0.04	0.97 ± 0.09
ETHYLENE 118 to 144 mlpm - AIR 7 lpm	0.93 ± 0.28	0.84 ± 0.07
ETHYLENE 118 to 144 mlpm - AIR 8 lpm	0.76 ± 0.04	0.81 ± 0.06
ETHYLENE 118 to 144 mlpm - AIR 7 lpm - cyclone	1.40 ± 0.05	1.19 ± 0.09
ETHYLENE 118 to 144 mlpm - AIR 8 lpm - cyclone	1.39 ± 0.04	1.08 ± 0.05

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Since the fit to 3 points could not be reliable, in the Supplementary (Table S.4) we reported the 2-wavelength calculations of the AAE for PAXs.

EXPERIMENTAL CONDITIONS	AAE - PAX	AAE - MWA
PROPANE 70 to 85 mlpm - AIR 7 lpm	0.88 ± 0.06	0.92 ± 0.04
PROPANE 70 to 85 mlpm - AIR 8 lpm	0.92 ± 0.06	0.91 ± 0.05
PROPANE 70 to 85 mlpm - AIR 7 lpm - cyclone	0.98 ± 0.09	1.0 ± 0.1
PROPANE 70 to 85 mlpm - AIR 8 lpm - cyclone	1.05 ± 0.04	0.97 ± 0.09
ETHYLENE 118 to 144 mlpm - AIR 7 lpm	0.9 ± 0.3	0.84 ± 0.07
ETHYLENE 118 to 144 mlpm - AIR 8 lpm	0.76 ± 0.04	0.81 ± 0.06
ETHYLENE 118 to 144 mlpm - AIR 7 lpm - cyclone	1.40 ± 0.05	1.19 ± 0.09
ETHYLENE 118 to 144 mlpm - AIR 8 lpm - cyclone	1.39 ± 0.04	1.08 ± 0.05

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#### 4. Conclusions

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A Mini-Inverted Soot Generator (MISG) was coupled with an atmospheric simulation chamber (ChAMBRe) to compare the emissions ~~of when the burner is fed by~~ two different fuels, ethylene, and propane. Different combustion conditions (i.e., air and fuel flow, global equivalence ratio) were characterized in terms of size distribution, particle and gas composition, optical properties, and EC concentration in the exhausts.

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The MISG turned out to be a stable and reproducible soot particles source, suitable for experiments in atmospheric simulation chambers. In addition, properties of emitted soot particles can be slightly modulated by varying the combustion conditions i.e., tuning the global equivalence ratio and/or varying the fuel used for combustion.

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With equal conditions, ethylene combustion produced s particles with higher number concentration and smaller diameter than propane ~~but is prone to generation of super-aggregates. Anyway, particles generated by both the fuels were larger than the typical exhausts of modern engines, such as aircraft and diesel vehicle engines, which emit ultrafine soot particles. Furthermore, it is noteworthy that ethylene combustion also generates super-micrometric aggregates.~~ These are likely formed in the stagnation plane at the bottom part of the combustion cell ~~directly in the exhaust line where particles density is very high.~~ This information should be kept in mind when planning experiments, since super-aggregates, if not desired, could affect analysis.

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The carbonaceous compounds produced by propane are generally characterized by higher EC to TC ratios than ethylene.

From the optical point of view, particles generated by propane turned out to be more light absorbing than those formed by ethylene, although burning conditions (in terms of global equivalence ratio) were the same. The values of the MAC parameter show a substantial agreement except for those retrieved from the data collected in the ethylene-no cyclone experiments. The latter resulted in lower MAC values, probably due to the presence of super-aggregates in the chamber.

618 This work opens to new and more complex experiments. Well-characterized soot particles could be used to  
619 investigate the effects that atmospheric parameters such as temperature and relative humidity can have on soot  
620 particles, ~~and also~~ to study the interactions between soot particles and gaseous pollutants, solar radiation  
621 or bio-aerosol.

## 622 **Author contribution**

623 VV and DM prepared the experimental setup, performed all the experiments and the data analysis; DM, FP,  
624 [SGD](#) and PP designed and built ChAMBRe; MB designed and implemented the acquisition software; VV and  
625 DM prepared the article with contributions from the other authors.

## 626 **Competing interests**

627 The authors declare that they have no conflict of interest.

## 628 **[Special issue statement:](#)**

629 [-This article is part of the special issue “Simulation chambers as tools in atmospheric research](#)  
630 [\(AMT/ACP/GMD inter-journal SI\)”. It is not associated with a conference.](#)

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## 636 **References**

637 Ackerman, A., Toon, O., Stevens, D., Heymsfield, A., Ramanathan, V., and Welton, E.: Reduction of tropical  
638 cloudiness by soot, *Science*, 288(5468), 1042–1047, doi:10.1126/science.288.5468.1042, 2000.

639 Anenberg, S. C., Horowitz, L. W., Tong, D. Q., and West, J. J.: An estimate of the global burden of  
640 anthropogenic ozone and fine particulate matter on premature human mortality using atmospheric modelling,  
641 *Environ. Health Perspec.*, 118 (9):1189–95, doi:10.1289/ehp.0901220, 2010.

642 Baccolo, G., Nastasi, M., Massabò, D., Clason, C., Di Mauro, B., Di Stefano, E., Łokas, E., Prati, P., Previtali,  
643 E., Takeuchi, N., Delmonte, B., Maggi, V.: Artificial and natural radionuclides in cryoconite as tracers of  
644 supraglacial dynamics: Insights from the Morteratsch glacier (Swiss Alps), *CATENA*, 191:104577,  
645 10.1016/j.catena.2020.104577, 2020.

646 Becker, K. H. (2006): Overview on the Development of Chambers for the Study of Atmospheric Chemical  
647 Processes, in: *Environmental Simulation Chambers: Application to Atmospheric Chemical Processes*, edited  
648 by: Barnes I. and Rudzinski K. J., Springer, Amsterdam, 1–26.

649 Bescond, A., Yon, J., Ouf, F. X., Roze, C., Coppalle, A., Parent, P., Ferry, D., and Laffon, C.: Soot optical  
650 properties determined by analyzing extinction spectra in the visible near-UV: Toward an optical speciation  
651 according to constituents and structure, *J. Aerosol Sci.* 101:118–32, doi:10.1016/j.jaerosci.2016.08.001, 2016.

652 Birch, M. E. and Cary, R. A.: Elemental carbon-based method for occupational monitoring of particulate diesel  
653 exhaust: methodology and exposure issues, *Analyst*, 121, 1183–1190, 1996.

- 654 Bischof, O. F., Weber, P., Bundke, U., Petzold, A., and Kiendler-Scharr, A.: Characterization of the  
655 Miniaturized Inverted Flame Burner as a Combustion Source to Generate a Nanoparticle Calibration Aerosol,  
656 *Emission Contr. Sci. Technol.* 6:37-46, DOI: 10.1007/s40825-019-00147-w, 2019.
- 657 Bond, T. C., and Bergstrom, R. W.: Light Absorption by Carbonaceous Particles: An Investigative Review,  
658 *Aerosol Science and Technology*, 40:1, 27-67, DOI: 10.1080/02786820500421521, 2006.
- 659 [Bond T. C., Covert D. S., and Müller T.: Truncation and Angular-Scattering Corrections for Absorbing](#)  
660 [Aerosol in the TSI 3563 Nephelometer, \*Aerosol Science and Technology\*, 43:9, 866-871, DOI:](#)  
661 [10.1080/02786820902998373, 2009.](#)
- 662 Bond, T. C., Doherty, S. J., Fahey, D. W., Forster, P. M., Berntsen, T., De Angelo, B. J., et al.: Bounding the  
663 role of black carbon in the climate system: A scientific assessment, *Journal of Geophysical Research*  
664 *Atmospheres*, 118(11), 5380–5552. <https://doi.org/10.1002/jgrd.50171>, 2013.
- 665 Caponi, L., Formenti, P., Massabó, D., Di Biagio, C., Cazaunau, M., Pangui, E., Chevaillier, S., Landrot, G.,  
666 Andreae, M. O., Kandler, K., Piketh, S., Saeed, T., Seibert, D., Williams, E., Balkanski, Y., Prati, P., and  
667 Doussin, J.-F.: Spectral- and size-resolved mass absorption efficiency of mineral dust aerosols in the shortwave  
668 spectrum: a simulation chamber study, *Atmos. Chem. Phys.*, 17, 7175–7191, [https://doi.org/10.5194/acp-17-](https://doi.org/10.5194/acp-17-7175-2017)  
669 [7175-2017](https://doi.org/10.5194/acp-17-7175-2017), 2017.
- 670 Cassee, F. R., Héroux, M. E., Gerlofs-Nijland, M. E., and Kelly, F. J.: Particulate matter beyond mass: Recent  
671 health evidence on the role of fractions, chemical constituents and sources of emission, *Inhalation Toxicology*,  
672 25(14), 802–812. <https://doi.org/10.3109/08958378.2013.850127>, 2013.
- 673 [Chakrabarty, R. K., Moosmüller, H., Garro, M. A., and Stipe, C. B.: Observation of Superaggregates from a](#)  
674 [Reversed Gravity Low-Sooting Flame, \*Aerosol Science and Technology\*, 46:1, i-iii, DOI:](#)  
675 [10.1080/02786826.2011.608389, 2012.](#)
- 676 Cross, E. S., ~~T. B.~~ Onasch, T. B., Ahern, A., Wrobel, W., Slowik, J. G., Olfert, J., Lack, D. A., Massoli, P.,  
677 Cappa, C. D., Schwarz, J. P., et al.: Soot particle studies-instrument inter-comparison-project overview,  
678 *Aerosol Sci. Technol.*, 44 (8):592–611. doi:10.1080/02786826.2010.482113, 2010.
- 679 Danelli, S. G., Brunoldi, M., Massabò, D., Parodi, F., Vernocchi, V., and Prati, P.: Comparative  
680 characterization of the performance of bio-aerosol nebulizers in connection with atmospheric simulation  
681 chambers, *Atmos. Meas. Tech.*, 14, 4461–4470, <https://doi.org/10.5194/amt-14-4461-2021>, 2021.
- 682 Durdina, L., Lobo, P., Trueblood, M. B., Black, E. A., Achterberg, S., Hagen, D. E., Brem, B. T., and Wang,  
683 J.: Response of real-time black carbon mass instruments to mini-CAST soot, *Aerosol Sci. Technol.* 50 (9):  
684 906–18, doi: 10.1080/02786826.2016.1204423, 2016.
- 685 Filep, Á., Ajtai, T., Utry, N., Pintér, M. D., Nyilas, T., Takács, S., Máté, Z., Gelencsér, A., Hoffer, A.,  
686 Schnaiter, M., Bozóki, Z., and Szabó, G.: Absorption spectrum of ambient aerosol and its correlation with size  
687 distribution in specific atmospheric conditions after a red mud accident, *Aerosol Air Qual. Res.*, 13, 49–59,  
688 2013.
- 689 Finlayson-Pitts, B. J. and Pitts Jr., J. N.: Chemistry of the upper and lower atmosphere: Theory, experiments  
690 and applications, Academic Press, San Diego, CA, 2000.
- 691 Gan, W. Q., Koehoorn, M., Davies, H. W., Demers, P. A., Tamburic, L., and Brauer, M.: Long-term exposure  
692 to traffic-related air pollution and the risk of coronary heart disease hospitalization and mortality,  
693 *Environmental Health Perspectives*, 119(4), 501–507, <https://doi.org/10.1289/ehp.1002511>, 2011.



- 694 Ghazi, R., and Olfert, J. S.: Coating mass dependence of soot aggregate restructuring due to coatings of oleic  
695 acid and dioctyl sebacate, *Aerosol Sci. Technol.* 47 (2):192–200. doi:10.1080/02786826.2012.741273, 2013.
- 696 Ghazi, R., Tjong, H., Soewono, A., Rogak, S. N., and Olfert, J. S.: Mass, mobility, volatility, and morphology  
697 of soot particles generated by a McKenna and inverted burner, *Aerosol Sci. Technol.* 47 (4):395–405.  
698 doi:10.1080/02786826.2012.755259, 2013.
- 699 Harrison, R. M., Beddows, D. C. S., Jones A. M., Calvo A., Alves C., and Pio C.: An evaluation of some issues  
700 regarding the use of aethalometers to measure woodsmoke concentrations, *Atmos. Environ.*, 80, 540–548,  
701 2013.
- 702 Henning, S., Ziese, M., Kiselev, A., Saathoff, H., Möhler, O., Mentel, T. F., Buchholz, A., Spindler, C.,  
703 Michaud, V., Monier, M., et al.: Hygroscopic growth and droplet activation of soot particles: Uncoated,  
704 succinic or sulfuric acid coated, *Atmos. Chem. Phys.* 12 (10):4525–37. doi: 10.5194/acp-12-4525-2012, 2012.
- 705 [Hu, D., Alfarra, M. R., Szpek, K., Langridge, J. M., Cotterell, M. I., Belcher, C., Rule, I., Liu, Z., Yu, C., Shao,  
706 Y., Voliotis, A., Du, M., Smith, B., Smallwood, G., Lobo, P., Liu, D., Haywood, J. M., Coe, H., and Allan, J.  
707 D.: Physical and chemical properties of black carbon and organic matter from different combustion and  
708 photochemical sources using aerodynamic aerosol classification, \*Atmos. Chem. Phys.\*, 21, 16161–16182,  
709 <https://doi.org/10.5194/acp-21-16161-2021>, 2021.](#)
- 710 Janssen, N., Gerlofs-Nijland, M., Lanki, T., Salonen, R., Cassee, F., Hoek, G., Fischer, P., Brunekreef, B., and  
711 Krzyzanowski, M.: Health effects of black carbon, Res. Rep., World Health Organization, Regional Office for  
712 Europe, Copenhagen, Denmark, 2012
- 713 Kazemimanesh, M., Moallemi, A., Thomson, K., Smallwood, G., Lobo, P., and Olfert, J. S.: A novel miniature  
714 inverted-flame burner for the generation of soot nanoparticles, *Aerosol Science and Technology*, 53 (2), 184-  
715 195, [10.1080/02786826.2018.1556774](https://doi.org/10.1080/02786826.2018.1556774), 2019.
- 716 Kirchstetter, T. W., Novakok, T., and Hobbs, P. V.: Evidence that the spectral dependence of light absorption  
717 by aerosols is affected by organic carbon, *J. Geophys. Res.*, 109, D21208,  
718 <https://doi.org/10.1029/2004JD004999>, 2004.
- 719 Kumar, N. K., Corbin, J. C., Bruns, E. A., Massabó, D., Slowik, J. G., Drinovec, L., Močnik, G., Prati, P.,  
720 Vlachou, A., Baltensperger, U., Gysel, M., El-Haddad, I., and Prévôt, A. S. H.: Production of particulate brown  
721 carbon during atmospheric aging of residential wood-burning emissions, *Atmos. Chem. Phys.*, 18, 17843–  
722 17861, <https://doi.org/10.5194/acp-18-17843-2018>, 2018.
- 723 Lack, D. A. and Langridge, J. M.: On the attribution of black and brown carbon light absorption using the  
724 Ångström exponent, *Atmos. Chem. Phys.*, 13, 10535–10543, <https://doi.org/10.5194/acp-13-10535-2013>,  
725 2013.
- 726 Lack, D. A., Langridge, J. M., Bahreini, R., Cappa, C. D., Middlebrook, A. M., and Schwarz, J. P.: Brown  
727 carbon and internal mixing in biomass burning particles, *P. Natl. Acad. Sci. USA*, 109, 14802–14807, 2012.
- 728 Lelieveld, J., Evans, J. S., Fnais, M., Giannadaki, D., and Pozzer, A.: The contribution of outdoor air pollution  
729 sources to premature mortality on a global scale, *Nature* 525 (7569):367–71. doi:10.1038/nature15371, 2015.
- 730 Lewis, K., Arnott, W. P., Moosmüller, H., and Wold, C. E.: Strong spectral variation of biomass smoke light  
731 absorption and single scattering albedo observed with a novel dual-wavelength photoacoustic instrument, *J.  
732 Geophys. Res.*, 113, D16203, <https://doi.org/10.1029/2007JD009699>, 2008.

- 733 Mamakos, A., Khalek, I., Giannelli, R., and Spears, M.: Characterization of Combustion Aerosol Produced by  
734 a Mini-CAST and Treated in a Catalytic Stripper, *Aerosol Science and Technology*, 47:8, 927-936,  
735 DOI:10.1080/02786826.2013.802762, 2013.
- 736 Massabò, D., and Prati P.: An overview of optical and thermal methods for the characterization of  
737 carbonaceous aerosol, *La Rivista del Nuovo Cimento*, vo. 44:3, DOI: 10.1007/s40766-021-00017-8, 2021.
- 738 Massabò, D., Prati, P., Canepa, E., Bastianini, M., Van Eijk, A.M.J., Missamou, T., Piazzola, J.:  
739 Characterization of carbonaceous aerosols over the Northern Adriatic Sea in the JERICO-NEXT project  
740 framework, *Atmospheric Environment*, 228:117449, DOI: 10.1016/j.atmosenv.2020.117449, 2020.
- 741 Massabò, D., Altomari, A., Vernocchi, V., and Prati, P.: Two-wavelength thermal–optical determination of  
742 light-absorbing carbon in atmospheric aerosols, *Atmos. Meas. Tech.*, 12, 3173–3182,  
743 <https://doi.org/10.5194/amt-12-3173-2019>, 2019.
- 744 Massabò, D., Danelli S. G., Brotto P., Comite A., Costa C., Di Cesare A., Doussin J. F., Ferraro F., Formenti  
745 P., Gatta E., Negretti L., Oliva M., Parodi F., Vezzulli L., and Prati P.: ChAMBRé: a new atmospheric  
746 simulation chamber for aerosol modelling and bioaerosol research, *Atmos. Meas. Tech.* 11, 5885-5900, 2018.
- 747 Massabò, D., Caponi, L., Bove, M. C., and Prati, P.: Brown carbon and thermal-optical analysis: a correction  
748 based on optical multiwavelength apportionment of atmospheric aerosols, *Atmos. Environ.*, 125, 119–125,  
749 [2016](https://doi.org/10.1016/j.atmosenv.2015.11.011).<https://doi.org/10.1016/j.atmosenv.2015.11.011>, [2016](https://doi.org/10.1016/j.atmosenv.2015.11.011).
- 750 Massabò, D., Caponi, L., Bernardoni, V., Bove, M. C., Brotto, P., Calzolari, G., Cassola, F., Chiari, M., Fedi,  
751 M. E., Fermo, P., Giannoni, M., Lucarelli, F., Nava, S., Piazzalunga, A., Valli, G., Vecchi, R., and Prati, P.:  
752 Multi-wavelength optical determination of black and brown carbon in atmospheric aerosols, *Atmos. Environ.*,  
753 108, 1–12, 2015.
- 754 Menon, S., Hansen, J., Nazarenko, L., and Luo, Y.: Climate effects of black carbon aerosols in China and  
755 India, *Science* 297 (5590):2250–53. doi:10.1126/science1075159, 2002.
- 756 Moallemi, A., Kazemimanesh, M., Corbin, J. C., Thomson, K., Smallwood, G., Olfert, J. S., and Lobo, P.:  
757 Characterization of black carbon particles generated by a propane-fueled miniature inverted soot generator,  
758 *Journal of Aerosol Science*, 135, 46-57, 10.1016/j.jaerosci.2019.05.004, 2019.
- 759 [Modini, R. L., Corbin, J. C., Brem, B. T., Irwin, M., Bertò, M., Pileci, R. E., Fetfatzis, P., Eleftheriadis, K.,  
760 Henzing, B., Moerman, M. M., Liu, F., Müller, T., and Gysel-Beer, M.: Detailed characterization of the CAPS  
761 single-scattering albedo monitor \(CAPS PM<sub>ss</sub>a\) as a field-deployable instrument for measuring aerosol light  
762 absorption with the extinction-minus-scattering method, \*Atmos. Meas. Tech.\*, 14, 819–851,  
763 <https://doi.org/10.5194/amt-14-819-2021>, 2021.](https://doi.org/10.5194/amt-14-819-2021)
- 764 Moore R. H., Ziemba L. D., Dutcher D., Beyersdorf A. J., Chan K., Crumeyrolle S., Raymond T. M., Thornhill  
765 K. L., Winstead E. L. and Anderson B. E.: Mapping the Operation of the Miniature Combustion Aerosol  
766 Standard (Mini-CAST) Soot Generator, *Aerosol Science and Technology*, 48:5, 467-479, DOI:  
767 10.1080/02786826.2014.890694, 2014.
- 768 Moosmüller, H., Chakrabarty, R.K., Ehlers, K.M., Arnott, W.P.: Absorption Ångström coefficient, brown  
769 carbon, and aerosols: basic concepts, bulk matter, and spherical particles. *Atmos. Chem. Phys.* 11, 1217e1225,  
770 2011.
- 771 Moschos, V., Gysel-Beer, M., Modini, R. L., Corbin, J. C., Massabò, D., Costa, C., Danelli, S. G., Vlachou,  
772 A., Daellenbach, K. R., Szidat, S., Prati, P., Prévôt, A. S. H., Baltensperger, U., and El Haddad, I.: Source-

- 773 specific light absorption by carbonaceous components in the complex aerosol matrix from yearly filter-based  
774 measurements, *Atmos. Chem. Phys.*, 21, 12809–12833, <https://doi.org/10.5194/acp-21-12809-2021>, 2021.
- 775 [Nemmar, A., Hoet, P.H.M., Vanquickenborne, B., Dinsdale, D., Thomeer, M., Hoylaerts, M.F., Vanbilloen,](#)  
776 [H., Mortelmans, L., Nemery, B.: Passage of inhaled particles into the blood circulation in humans. \*Circulation\*](#)  
777 [105, 411–414, <https://doi.org/10.1161/hc0402.104118>, 2002.](#)
- 778 [Nienow, A.M., and Roberts, J.T.: Heterogeneous Chemistry of Carbon Aerosols. \*Annual Review of Physical\*](#)  
779 [\*Chemistry\* 57, 105–128, <https://doi.org/10.1146/annurev.physchem.57.032905.104525>, 2006.](#)
- 780 NIOSH: Method 5040 Issue 3: Elemental Carbon (Diesel Exhaust). In NIOSH Manual of Analytical Methods.  
781 National Institute of Occupational Safety and Health, Cincinnati, OH, 1999.
- 782 Nordmann, S., Birmili, W., Weinhold, K., Müller, K., Spindler, G., and Wiedensohler, A.: Measurements of  
783 the mass absorption cross section of atmospheric soot particles using Raman spectroscopy, *J. Geophys. Res.*  
784 *Atmos.*, 118, 12,075–12,085, doi: 10.1002/2013JD020021, 2013.
- 785 [Oberdörster, G., Oberdörster, E., Oberdörster, J.: Nanotoxicology: An Emerging Discipline Evolving from](#)  
786 [Studies of Ultrafine Particles, \*Environmental Health Perspectives\* 113, 823–839,](#)  
787 [<https://doi.org/10.1289/ehp.7339>, 2005.](#)
- 788 Onasch, T. B., Trimborn, A., Fortner, E. C., Jayne, J. T., Kok, G. L., Williams, L. R., Davidovits, P. and  
789 Worsnop, D. R.: Soot particle aerosol mass spectrometer: Development, validation, and initial application,  
790 *Aerosol Sci. Technol.* 46 (7):804–17, doi:10.1080/02786826.2012.663948, 2012.
- 791 Pagels, J., Khalizov, A. F., McMurry, P. H., and Zhang, R. Y.: Processing of soot by controlled sulphuric acid  
792 and water condensation—Mass and mobility relationship, *Aerosol Sci. Technol.* 43 (7):629–4,  
793 doi:10.1080/02786820902810685, 2009.
- 794 Petzold, A., Ogren, J. A., Fiebig, M., Laj, P., Li, S.-M., Baltensperger, U., Holzer-Popp, T., Kinne, S.,  
795 Pappalardo, G., Sugimoto, N., Wehrli, C., Wiedensohler, A., and Zhang, X.-Y.: Recommendations for  
796 Reporting “Black Carbon”, Measurements. *Atmos. Chem. Phys.*, 13:8365–8379, 2013.
- 797 Petzold, A., Schloesser, H., Sheridan, P.J., Arnott, W.P., Ogren, J.A., Virkkula, A.: Evaluation of multiangle  
798 absorption photometry for measuring aerosol light absorption, *Aerosol Sci. Technol.* 39 (1), 2005.
- 799 Petzold, A., and Schöllner, M.: Multi-angle absorption photometry—a new method for the measurement of  
800 aerosol light absorption and atmospheric black carbon, *Journal of Aerosol Science*, 35, 421–441, 2004.
- 801 Pope, C.A. III, Burnett, R. T., Thun, M. J., Calle, E. E., Krewski, D., Ito, K., et al.: Lung Cancer,  
802 Cardiopulmonary Mortality, and Long-term Exposure to Fine Particulate Air Pollution. *J. Am. Med. Assoc.*,  
803 287(9):1132–41, 2002.
- 804 Quinn, P. K., Bates, T. S., Baum, E., Doubleday, N., Fiore, A. M., Flanner, M., et al.: Short-lived pollutants in  
805 the Arctic: Their climate impact and possible mitigation strategies, *Atmospheric Chemistry and Physics*, 8(6),  
806 1723–1735. <https://doi.org/10.5194/acp-8-1723-2008>, 2008.
- 807 Ramanathan, V., and G. Carmichael, G.: Global and regional climate changes due to black carbon, *Nat. Geosci.*  
808 1 (4):221–27. doi:10.1038/ngeo156, 2008.
- 809 Saturno, J., Pöhlker, C., Massabò, D., Brito, J., Carbone, S., Cheng, Y., Chi, X., Ditas, F., Hrabě de Angelis,  
810 I., Morán-Zuloaga, D., Pöhlker, M. L., Rizzo, L. V., Walter, D., Wang, Q., Artaxo, P., Prati, P., and Andreae,

- 811 M. O.: Comparison of different Aethalometer correction schemes and a reference multi-wavelength absorption  
812 technique for ambient aerosol data, *Atmos. Meas. Tech.*, 10, 2837–2850, [https://doi.org/10.5194/amt-10-2837-](https://doi.org/10.5194/amt-10-2837-2017)  
813 2017, 2017.
- 814 Scerri, M. M., Kandler, K., Weinbruch, S., Yubero, E., Galindo N., Prati, P., Caponi, L., and Massabò, D.:  
815 Estimation of the contributions of the sources driving PM<sub>2.5</sub> levels in a Central Mediterranean coastal town,  
816 *Chemosphere*, 211, 465–481, <https://doi.org/10.1016/j.chemosphere.2018.07.104>, 2018.
- 817 [Skiles, S.M., Flanner, M., Cook, J.M., Dumont, M., Painter, T.H.: Radiative forcing by light-absorbing](https://doi.org/10.1038/s41558-018-0296-5)  
818 [particles in snow. \*Nature Clim Change\* 8, 964–971. <https://doi.org/10.1038/s41558-018-0296-5>, 2018.](https://doi.org/10.1038/s41558-018-0296-5)
- 819 Stipe, C. B., Higgins, B. S., Lucas, D., Koshland, C. P., & Sawyer, R. F. Inverted co-flow diffusion flame for  
820 producing soot, *Review of Scientific Instruments*, 76 (2), <https://doi.org/10.1063/1.1851492>, 2005.
- 821 Utry, N., Ajtai, T., Filep, Á., Dániel P. M., Hoffer, A., Bozoki, Z., and Szabó, G.: Mass specific optical  
822 absorption coefficient of HULIS aerosol measured by a four-wavelength photoacoustic spectrometer at NIR,  
823 VIS and UV wavelengths, *Atmos. Environ.*, 69, 321–324, 2013.
- 824 Utry, N., Ajtai, T., Filep, Á., Pintér, M., Török, Zs., Bozóki, Z., and Szabó, G.: Correlations between absorption  
825 Angström exponent (AAE) of wintertime ambient urban aerosol and its physical and chemical properties,  
826 *Atmos. Environ.*, 91, 52–59, 2014.
- 827 Utry, N., Ajtai, T., Pinter, M., Bozóki, Z., and Szabó, G.: Wavelength-dependent optical absorption properties  
828 of artificial and atmospheric aerosol measured by a multiwavelength photoacoustic spectrometer. *Int. J.*  
829 *Thermophys.*, 35 (12):2246–58. doi:10.1007/s10765-014-1746-6, 2014
- 830 von der Weiden, S.-L., Drewnick, F., and Borrmann, S.: Particle Loss Calculator – a new software tool for the  
831 assessment of the performance of aerosol inlet systems, *Atmos. Meas. Tech.*, 2, 479–494,  
832 <https://doi.org/10.5194/amt-2-479-2009>, 2009.
- 833 Weijer, E. P., Schaap, M., Nguyen, L., Matthijsen, J., Denier van der Gon, H. A. C., et al.: Anthropogenic and  
834 Natural Constituents in Particulate Matter in the Netherlands, *Atmos. Chem. Phys.*, 11:2281–2294, 2011.
- 835 Zhang, R., Khalizov, A. F., Pagels, J., Zhang, D., Xue, H., and McMurry, P. H.: Variability in morphology,  
836 hygroscopicity, and optical properties of soot aerosols during atmospheric processing, *Proc. Natl. Acad. Sci.*  
837 *USA* 105 (30):10291–96. doi:10.1073/pnas.0804860105, 2008.