# New method to determine equivalent black carbon mass size distribution

- Weilun Zhao<sup>1</sup>, Gang Zhao<sup>2</sup>, Ying Li<sup>3,4</sup>, Song Guo<sup>2</sup>, Nan Ma<sup>5</sup>, Lizi Tang<sup>2</sup>, Zirui Zhang<sup>2</sup>, Chunsheng Zhao<sup>1</sup>
- <sup>1</sup>Department of Atmospheric and Oceanic Sciences, School of Physics, Peking University, Beijing 100871, China
- 4 <sup>2</sup>State Key Joint Laboratory of Environmental Simulation and Pollution Control, College of Environmental Sciences and
- 5 Engineering, Peking University, Beijing 100871, China

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- 6 <sup>3</sup>Department of Ocean Science and Engineering, Southern University of Science and Technology, Shenzhen 518055, China
- <sup>4</sup>Southern Marine Science and Engineering Guangdong Laboratory, Guangzhou 511458, China
- 8 <sup>5</sup>Institute for Environmental and Climate Research, Jinan University, Guangzhou 511443, China
- 9 Correspondence to: Chunsheng Zhao (zcs@pku.edu.cn)

Abstract. Black carbon (BC) is an important atmospheric component with strong light absorption. Many attempts have been made to measure BC mass size distribution (BCMSD) for its significant impact on climate and public health. Larger-coverage BCMSD, ranging from upper submicron to larger than 1 µm, contributes to substantial proportion of total BC mass and absorption. However, current time resolution of larger-coverage BCMSD measurement was limited to 1 day, which was insufficient to characterize variation of larger-coverage BCMSD. In this study, a new method to determine equivalent BCMSD (eBCMSD) was proposed from size-resolved absorption coefficient measured by an aerodynamic aerosol classifier in tandem with an aethalometer. The proposed method could measure larger-coverage eBCMSD with time resolution as high as 1 hour and was validated by comparing the measurement results with refractory BCMSD (rBCMSD) measured by a differential mobility analyzer in tandem with a single particle soot photometer (DMA – SP2) for particle size larger than 300 nm during a field measurement in Yangtze River Delta. Bulk refractory BC mass concentration (m<sub>rBC,bulk</sub>) by DMA – SP2 was underestimated by 33 % compared to bulk equivalent BC mass concentration ( $m_{\rm eBC,bulk}$ ) by this method because of the limited size range of measurement for DMA – SP2. Uncertainty analysis of this method was performed with respect to mass absorption cross-section (MAC), transfer function inversion, number fraction of BC-containing particle and instrumental noise. The results indicated that MAC was the main uncertainty source, leading to mence that make the main uncertainty source, leading to mence the mence the main uncertainty source, and the mence t varied from - 20 % to 28 %. With the advanatage of wide size coverage up to 1.5 µm, high time resolution, easy operation and low cost, this method is expected to have wide applications in field measurement for better estimating radiative properties and climate effects of BC.

## 1 Introduction

Atmospheric black carbon (BC) is emitted through incomplete combustion of carbon-based fuels (Bond, 2001), such as biomass burning and fossil fuel combustion (Bond et al., 2004). BC warms atmosphere and is considered the third important light absorber in the climate system after CO<sub>2</sub> and CH<sub>4</sub> (Bond et al., 2013). BC can induce the respiratory and cardiovascular disease through inhalation (Highwood and Kinnersley, 2006). Plenty of studies have been devoted to BC for its significant impact on the climate and public health.

Bulk BC mass concentration ( $m_{BC,bulk}$ ) is one of the essential parameters for modeling because radiative transfer models calculate BC absorption from  $m_{BC,bulk}$  (Bond et al., 2013). A great deal of research has been dedicated to  $m_{BC,bulk}$  measurement for both model assimilation and environmental monitoring (Castagna et al., 2019;Helin et al., 2018;Ran et al., 2016). A recent study indicated that the radiative effect of BC was extremely sensitive to its particle size (Matsui et al., 2018). Zhao et al. (2019) further revealed that the variation of BC mass size distribution (BCMSD), namely size-resolved BC mass concentration ( $m_{BC,size-resolved}$ ), led to substantial changes in the radiative effect of BC based on field measurement, highlighting the importance of BCMSD on modeling the radiative effect of BC rather than simply  $m_{BC,bulk}$ . The size of BC affects the deposition rate of BC to the lung (Highwood and Kinnersley, 2006), indicating that BCMSD is closely related to health. In the ambient environment, BCMSD is influenced by emission sources. For instance, BCMSD of fossil fuel combustion differs obviously from that of biomass burning (Schwarz et al., 2008), implying that BCMSD is one of the characteristics of emission source. The temporal variation of BCMSD can reflect the atmospheric aging of BC, during which BC undergoes remarkable change in size, accompanied by dramatical variation of BC optical properties (Zhang et al., 2008). Therefore, reliable measurement and understanding of BCMSD are highly necessary for estimating the impact of BC on both the earth energy budget and public health (Moosmuller et al., 2009).

Quite a few efforts were made to measure BCMSD. The single-particle soot photometer (SP2) was developed using laser-induced incandescence (Schwarz et al., 2006), which measured refractory BCMSD (rBCMSD, Petzold et al. (2013)) on a single-particle level. The soot particle aerosol mass spectrometer (SP-AMS) combined laser-induced incandescence and mass spectroscopy, which could determine not only rBCMSD but also the chemical composition of BC-containing particles (Onasch et al., 2012). The SP2 and SP-AMS techniques were characterized by high time resolution and high accuracy, but high cost and complicated maintenance, as a result, not widely applied for routine measurement. A more convenient solution was required for wider and better characterization of BCMSD in different regions and emission sources. Elemental carbon mass size distribution (ECMSD, Petzold et al. (2013)) could be sampled by multi-stage cascade impactor (Viidanoja et al., 2002) combined with off-line analysis, such as thermo/optical organic carbon/elemental carbon method (Chow et al., 2001). ECMSD sampled by multi-stage cascade impactor had wide size coverage, but low time resolution, usually from 24 hours (Soto-Garcia et al., 2011) to 48 hours (Guo, 2015), which was too low to resolve aging of BC. Differential mobility analyzer (DMA) in tandem with filter-based instrument (Hansen et al., 1984), for instance, micro-aethalometer (MA) (Ning et al., 2013) and particle soot absorption photometer (PSAP) (Tunved et al., 2021), was used to determine equivalent BCMSD (eBCMSD, Petzold et al. (2013)) with higher time resolution up to 2 hours (Zhao et al., 2021b). However, the multiple-charge correction and low flow rate of DMA imposed restrictions on the accuracy of the measured eBCMSD. The measured size range of DMA was limited to less than about 700 nm, resulting in incomplete measured eBCMSD. Current measurement of larger-coverage BCMSD, ranging from upper submicron to larger than 1 µm, was limited in time resolution. Characteristics of larger-coverage BCMSD during atmospheric aging was still unclear. Wang et al. (2022) showed that BC larger than 1  $\mu$ m could contribute to as large as 50% of  $m_{BC.bulk}$ , highlighting the importance of larger-coverage BCMSD. Therefore, it was imperative to measure larger-coverage BCMSD with higher time resolution.

In this study, a new method to determine eBCMSD was proposed using size-resolved absorption coefficient ( $\sigma_{ab,size-resolved}$ ) measured by an aerodynamic aerosol classifier (AAC, Cambustion, UK, Tavakoli and Olfert (2013)) in tandem with an aethalometer (model AE33, Magee, USA, Drinovec et al. (2015)), combined with size-resolved number concentration ( $N_{size-resolved}$ ), simultaneously measured by scanning mobility particle sizer (SMPS, TSI, USA) and an aerodynamic particle sizer (APS, TSI, USA). The proposed method for determining eBCMSD overcame the disadvantages and weighed the advantages of the above-mentioned methods, characterized by high cost-effectiveness, easy and convenient maintenance, high time resolution to 1 hour, and wide size range to up to 1.5  $\mu$ m. The proposed method was validated in a field measurement in the Yangtze River Delta and the uncertainty study was carried out based on the measured data.

## 2 Methods

## 2.1 Instrumental setup

Figure 1 illustrated the instrumental setup developed in this study, which could be split into two parts, namely the measurement setup and the validation setup. Ambient aerosol particles were drawn througth a PM<sub>10</sub> inlet (16.67 L min<sup>-1</sup>) and were dried to relative humidity less than 30 % by silica gel diffusion drier beforehand. An advanced flow splitter were used to split aerosol laden flow into different instruments for isokinetic sampling. For the measurement setup, AAC in tandem with AE33 (AAC – AE33) measured  $\sigma_{ab,size-resolved}$  at a flow rate of 3 L min<sup>-1</sup>. Since eBCMSD of larger size coverage was mainly focused in this study, AAC was set to scan 12 particle sizes ( $D_p$ ) logarithmically equally spaced from 200 nm to 1500 nm and smaller size was not scanned. Each size was scanned for 5 minutes and 1 hour was required for a complete cycle. It should be noted that  $D_p$  was aerodynamic size in this study. Mobility size related to DMA was converted to aerodynamic size assuming an effective density of 1.3 g cm<sup>-3</sup> (Zhao et al., 2019;DeCarlo et al., 2005). AE33 measured absorption coefficient ( $\sigma_{ab}$ ) at 7 wavelengths from 370 nm to 950 nm (Drinovec et al., 2015), at which 880 nm was adopted in this study because BC dominated particle absorption at 880 nm (Ramachandran and Rajesh, 2007). SMPS and APS measured  $N_{size-resolved}$  for  $D_p$  less than and greater than about 800 nm at a flow rate of 0.3 L min<sup>-1</sup> and 5 L min<sup>-1</sup>, respectively.

For validation system, the tandem array of Neutralizer, DMA and SP2 (DMA – SP2) measured rBCMSD (rBCMSD<sub>DMA-SP2</sub>) at a flow rate of 0.12 L min<sup>-1</sup> for purpose of comparing with eBCMSD determined by the proposed method (eBCMSD<sub>AAC-AE33</sub>). Another AE33 measured bulk absorption coefficient ( $\sigma_{ab,bulk}$ ) simultaneously at a flow rate of 2 L min<sup>-1</sup> for closure study with  $\sigma_{ab,size-resolved}$ .

#### 2.1.1 AE33

The principle of obtaining  $\sigma_{ab}$  was well developed for aethalometer (Hansen et al., 1984) and described here in brief. Ambient aerosol particles were drawn into an aethalometer at a flow rate F and collected on an area S of a filter. A light source illuminated the filter. The transmitted light signal was denoted as  $I(I_0)$  for the light which passed through the particle-laden (particle-free) part of the filter. Light attenuation was defined as

$$ATN = -100 \cdot \ln\left(\frac{I}{I_0}\right). \tag{1}$$

- 95 ATN increased with decreasing I as aerosol particles were loaded on the filter continuously. Therefore, ATN reflected aerosol loading
- on the filter. If ATN increased  $\triangle$ ATN during time interval  $\triangle t$ , then attenuation coefficient was defined as

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$$\sigma_{ATN} = \frac{s}{100 \cdot F} \cdot \frac{\Delta ATN}{\Delta t}$$
 (2)

- 98 The light attenuation was caused by not only particle absorption, but also scattering by particle and filter matrix. A scattering
- 99 parameter  $C_f$  was introduced to extract  $\sigma_{ab}$  from  $\sigma_{ATN}$ :

$$\sigma_{\rm ab} = \frac{\sigma_{\rm ATN}}{c_f},\tag{3}$$

- where  $C_f$  was set as 2.9 (Zhao et al., 2020b) in this study. Nonlinearity, termed loading effect, became more and more significant
- with increasing aerosol loading, namely for the same increase in aerosol loading, the corresponding increase in ATN was smaller
- for heavier aerosol loading. The "dual-spot" technique (Drinovec et al., 2015) was proposed to correct the loading effect for AE33
  - and was used in this study. The  $\sigma_{ab}$  measured by AE33 at given particle size selected by AAC was termed binned  $\sigma_{ab}$  ( $\sigma_{ab,binned}$ ) in
  - this study to differentiate from  $\sigma_{ab,bulk}$ .

#### 2.1.2 AAC

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- The principle of AAC was illustrated detailedly by Tavakoli and Olfert (2013) and was introduced here concisely. The setup of
  - AAC could be simplified to two coaxial cylinders, where the inner radius, outer radius and length was denoted as  $r_i$ ,  $r_o$  and L.
- Polydisperse particles flowed into the space between the inner cylinder and outer cylinder from one end of the inner cylinder at a
  - flow rate of  $Q_{\text{sample}}$ . Meawhile, Particle free sheath flow streamed in the space between the inner cylinder and outer cylinder in the
- direction of the axis of the coaxial cylinders at a flow rate of  $Q_{\text{sheath}}$ . The sheath flow carried the particles along the coaxial cylinders.
- At the same time, the two coaxial cylinders rotated with respect to their axis at a rotational speed of  $\omega$ . Therefore, the particles was
  - migrated outwards across the sheath flow by the centrifugal force acting on them. Relaxation time ( $\tau$ ) was defined as

$$\tau = \frac{c_{\rm c}(D_{\rm p})\rho_0 D_{\rm p}^2}{18\mu},\tag{4}$$

- where  $C_c(D_p)$  was the Cunningham slip correction factor as a function of  $D_p$  (Kim et al., 2005),  $\rho_0 = 1$  g cm<sup>-3</sup> was the reference
- density and  $\mu$  was the gas viscosity. It could be seen that  $\tau$  was directly related to  $D_p$ . Dynamic analysis proved that only particles
- with certain relaxation time  $\tau$ , and hence certain  $D_p$ , could migrate to another end of outer cylinder and emerge as monodisperse
- flow.  $\tau$  was related to parameters of AAC by

$$\tau = \frac{2Q_{\text{sheath}}}{\pi\omega^2(r_1 + r_0)^2 L}.\tag{5}$$

- Therefore, by changing  $\omega$  and  $Q_{\text{sheath}}$ , monodisperse particles of desired  $D_p$  could be selected by AAC. Unlike DMA, particles did
- not need to be charged before entering AAC. Consequently, the transmission efficiency ( $\lambda_{\Omega}$ ) of AAC was about 4 times higher than
- that of DMA and multi-charge correction was not required for AE33 (Johnson et al., 2018).

#### 2.1.3 DMA – SP2

The reason why DMA – SP2 was used rather than SP2 alone was that  $D_p$  could be directly measured by DMA. If SP2 was used

alone,  $D_p$  was calculated from Mie theory with assumed inputs (Taylor et al., 2015). The detail of DMA – SP2 was described by Zhao et al. (2021a) and presented here concisely. The incandescence high gain channel and incandescence low gain channel of SP2 were calibrated by Aquadag soot particles with effective density of 1.8 g cm<sup>-3</sup> before measurement. DMA was set to scan  $D_p$  from 15 nm to 780 nm with 56 bins at time resolution of 5 min.  $D_p$  and time resolution of measured rBCMSD<sub>DMA-SP2</sub> was interpolated and time-averaged to that of eBCMSD<sub>DAAC-AE33</sub> for comparison. It should be noted that 780 nm, the upper  $D_p$  limit scanned by DMA, lied between 720 nm and 865 nm, two of the scanned  $D_p$ s by AAC. Therefore, upper  $D_p$  limit of the interplolated rBCMSD<sub>DMA-SP2</sub> was 200 nm. According to Zhao et al. (2020a), DMA – SP2 could not detect particle less than about 200 nm, consistent with the lower limit of  $D_p$  in this study.

The number distribution of BC-containing particle measured by DMA – SP2 could be presented as a bivariate function  $\frac{\partial^2 N_{\text{BC}}}{\partial \log D_{\text{p}} \partial \log D_{\text{c}}}$ , where  $N_{\text{BC}}$  was the number concentration of BC-containing particle, and  $D_{\text{c}}$  was the core diameter assuming that BC-containing particle had a core-shell geometry. It should be pointed out that asphericity of rBC was not considered in this study.  $D_{\text{c}}$  was logarithmically equally distributed from 78 nm to 560 nm with 29 bins. At a given  $D_{\text{c}}$  bin, the number distribution of BC-containing particle could be considered as a univariate function  $\frac{dN_{\text{BC}}}{d\log D_{\text{p}}}$ . The multiple charging correction of 1-dimensional size distribution was developed by Knutson and Whitby (1975), which was adopted to correct  $\frac{dN_{\text{BC}}}{d\log D_{\text{p}}}$ . By correcting  $\frac{dN_{\text{BC}}}{d\log D_{\text{p}}}$  at each  $D_{\text{c}}$  bin, multiple charging correction of  $\frac{\partial^2 N_{\text{BC}}}{\partial \log D_{\text{p}} \partial \log D_{\text{c}}}$  was achieved.

## 2.2 Field measurement

The proposed method was applied to a field measurement in Changzhou, Jiangsu Province, China (119°36′E, 31°43′N), located in the Yangtze River Delta, from May 17<sup>th</sup> to June 3<sup>rd</sup> in 2021. Changzhou was between two megacities, about 187 km to the northwest of Shanghai and about 82 km to the southeast of Nanjing, as shown in Fig. S1a. The area between the Shanghai and Nanjing underwent serious environmental pollution in the past few decades with the development of industrialization and urbanization. As shown in Fig. S1b, the pollution condition of Changzhou was overall representative of the regional background pollution in the Yangtze River Delta.

## 2.3 Size-resolved calculations

## 2.3.1 $\sigma_{ab, size-resolved}$

Tavakoli and Olfert (2013) formulated the ideal theoretical model for AAC transfer function inversion, which was adopted by this

study to derive  $\sigma_{ab,size-resolved}$  from  $\sigma_{ab,binned}$  and presented here in short. The  $\sigma_{ab,size-resolved}$  was given by

$$\sigma_{\text{ab,size-resolved}} = \frac{d\sigma_{\text{ab}}}{d\log D_{\text{p}}} = \frac{\ln(10)}{\frac{d\log D_{\text{p}}}{d\log \tau}, \beta^{*}} \cdot \sigma_{\text{ab,binned}}, \tag{6}$$

where  $\beta^*$  was related to the ratio of  $Q_{\text{sheath}}$  to  $Q_{\text{sample}}$ ,  $R_t = Q_{\text{sheath}}/Q_{\text{sample}} = 1/\beta$ , through

$$\beta^* = \left(1 + \frac{1}{\beta}\right) \ln(1 + \beta) - \left(1 - \frac{1}{\beta}\right) \ln(1 - \beta). \tag{7}$$

Johnson et al. (2018) corrected the ideal inversion formula (6) to take particle loss and spectral broadening into account by replacing

155  $\beta^*$  with a nonideal  $\beta^*_{NI}$ :

$$\beta_{\text{NI}}^* = \lambda_{\Omega} \mu_{\Omega} \left[ \ln \left( \frac{1 + \beta/\mu_{\Omega}}{1 - \beta/\mu_{\Omega}} \right) + \frac{\mu_{\Omega}}{\beta} \ln \left( 1 - \left( \frac{\beta}{\mu_{\Omega}} \right)^2 \right) \right], \tag{8}$$

where  $\mu_{\Omega}$  was the spectral broadening factor. Both  $\lambda_{\Omega}$  and  $\mu_{\Omega}$  were dependent on  $D_p$  as well as flow, and discussed in detail in Sect.

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- 2.3.2 eBCMSDAAC-AE33
- $\sigma_{ab.size-resolved}$  could be converted to eBCMSD<sub>AAC-AE33</sub> through mass absorption cross-section (MAC) (Bond and Bergstrom, 2006),
- which was determined based on method proposed by Zhao et al. (2021b). Briefly, a 2-dimensional lookup table of MAC with respect
  - to  $D_p$  and  $D_c$  was simulated (MAC( $D_p$ ,  $D_c$ )) by Mie theory assuming a core-shell geometry. At a given size bin centered at  $D_p$ , the
  - corresponding  $\sigma_{ab}$  and number concentration (N) at the size bin, denoted as  $\sigma_{ab}(D_p)$  and  $N(D_p)$ , could be derived by  $\sigma_{ab,size-resolved}$  and
- 164  $N_{\text{size-resolved}}$ :

$$\sigma_{\rm ab}(D_{\rm p}) = \sigma_{\rm ab,size-resolved}(D_{\rm p}) \cdot \Delta \log D_{\rm p},$$
(9)

$$N(D_{p}) = N_{\text{size-resolved}}(D_{p}) \cdot \Delta \log D_{p}, \tag{10}$$

- where  $\Delta \log D_p$  was the logarithmic width of the size bin. The number concentration of BC-containing particle at the size bin  $N_{\rm BC}(D_p)$
- 168 was determined by

$$169 N_{\mathrm{BC}}(D_{\mathrm{p}}) = N(D_{\mathrm{p}}) \cdot f_{\mathrm{BC}}, (11)$$

- where  $f_{BC}$  was the number fraction of BC-containing particle, which was assumed as a fixed value in this study and the uncertainty
- of the fixed- $f_{BC}$  assumption was discussed in Sect. 4.3. An optimal  $D_c$  was found so that calculated  $\sigma_{ab}$  at the size bin, denoted as
- 172  $\sigma_{ab,calc}(D_p, D_c)$ , matched  $\sigma_{ab}(D_p)$ :

$$\sigma_{\mathrm{ab,calc}}(D_{\mathrm{p}}, D_{\mathrm{c}}) = \rho_{\mathrm{BC}} \frac{\pi}{6} D_{\mathrm{c}}^{3} \cdot \mathrm{MAC}(D_{\mathrm{p}}, D_{\mathrm{c}}) \cdot N_{\mathrm{BC}}(D_{\mathrm{p}}) = \sigma_{\mathrm{ab}}(D_{\mathrm{p}}), \tag{12}$$

- where  $\rho_{BC}$  was the density of BC, and set as 1.8 g cm<sup>-3</sup> (Bond and Bergstrom, 2006), consistent with the  $\rho_{BC}$  assumption when
  - deriving rBCMSD<sub>DMA-SP2</sub>. Equivalent BC mass concentration ( $m_{\rm eBC}$ ) at the size bin, denoted as  $m_{\rm eBC}(D_{\rm p})$ , could be determined by

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$$m_{\text{eBC}}(D_{\text{p}}) = \frac{\sigma_{\text{ab}}(D_{\text{p}})}{\text{MAC}(D_{\text{p}},D_{\text{c}})},$$
 (13)

then the eBCMSD<sub>AAC-AE33</sub> at  $D_p$ , denoted by eBCMSD<sub>AAC-AE33</sub>( $D_p$ ), could be determined by

$$eBCMSD(D_p) = \frac{m_{eBC}(D_p)}{\Delta \log D_p}.$$
(14)

- 179 3 Results and discussion
- Figure 2 presented the time series of the measurement results. There were 4 pollution episodes during the measurement period:
- from about 12:00 May 17<sup>th</sup> to about 00:00 May 20<sup>th</sup>, from about 00:00 May 21<sup>st</sup> to about 12:00 May 22<sup>nd</sup>, from about 00:00 May
- 182 24th to about 12:00 May 26th, and from about 12:00 May 28th to about 12:00 May 31st, as shown in Fig. 2a. Both eBCMSD<sub>AAC-AE33</sub>
- 183 (Fig. 2b) and rBCMSD<sub>DMA-SP2</sub> (Fig. 2c) recorded the 4 pollution episodes simultaneously with higher BCMSD values than clean
- episodes. eBCMSD<sub>AAC-AE33</sub> and rBCMSD<sub>DMA-SP2</sub> were both integrated from 200 nm to 720 nm, which was the D<sub>p</sub> range of

measurement for DMA - SP2, and the results were denoted as  $m_{\text{eBC,bulk,AAC-AE33,200-720}}$  and  $m_{\text{rBC,bulk,DMA-SP2,200-720}}$ , respectively. As shown in Fig. 2a,  $m_{\text{eBC,bulk,AAC-AE33,200-720}}$  compared well with  $m_{\text{rBC,bulk,DMA-SP2,200-720}}$  with determination coefficient ( $R^2$ ), slope ( $b_1$ ), and intercept (b<sub>0</sub>) of 0.8 (accurate to one decimal place), 1.0 and 0.1 µg m<sup>-3</sup> (Fig. S2a). R<sup>2</sup> of 0.8 and b<sub>1</sub> of 1.0 indicated the proposed method was capable of capturing the variation of  $m_{\text{BC,bulk}}$ . The mean  $m_{\text{eBC,bulk},AAC-AE33,200-720}$  and  $m_{\text{rBC,bulk},DMA-SP2,200-720}$  were 0.7  $\pm$ 0.4  $\mu$ g m<sup>-3</sup> and 0.6  $\pm$  0.3  $\mu$ g m<sup>-3</sup>, indicating  $m_{\text{eBC,bulk,AAC-AE33,200-720}}$  was overall 0.1  $\mu$ g m<sup>-3</sup> higher than  $m_{\text{rBC,bulk,DMA-SP2,200-720}}$ , consistent with  $b_0$  of 0.1 µg m<sup>-3</sup>. The reason for overall discrepancy of 0.1 µg m<sup>-3</sup> in  $m_{\text{BC,bulk},200-720}$  might be that DMA – SP2 could not detect rBC with D<sub>c</sub> less than about 100 nm (Zhao et al., 2020a), resulting in an underestimated m<sub>rBC,bulk,DMA-SP2,200-720</sub>. eBCMSD<sub>AAC-AE33</sub> was also integrated from 720 nm to 1500 nm, which was the  $D_p$  range that DMA – SP2 did not measure, and the result was denoted as meBC,bulk,AAC-AE33,720-1500. meBC,bulk,AAC-AE33,720-1500 was correlated to mrBC,bulk,DMA-SP2,200-720 to study whether BCMSD ranging from 720 nm to 1500 nm was connected to that ranging from 200 nm to 720 nm, as shown in Fig. S2b. R<sup>2</sup> decreased to 0.2 between  $m_{\rm eBC,bulk,AAC-AE33,720-1500}$  and  $m_{\rm rBC,bulk,DMA-SP2,200-720}$ , indicating these two size ranges were not well related and both of them needed to be measured exclusively. Observation by transmission electron microscopy showed that these larger BC-containing particles could be coated with massizve materials from secondary processes, or superaggregated BC with fractal BC structure (Wang et al., 2022). The mean  $m_{\text{eBC,bulk,AAC-AE33,720-1500}}$  was 0.2  $\pm$  0.2  $\mu$ g m<sup>-3</sup>, indicating that DMA – SP2 overall underestimated  $m_{\text{BC,bulk}}$ for ~ 0.2 μg m<sup>-3</sup> (~ 33 %) in this field measurement considering that DMA – SP2 did not measure BCMSD larger than 720 nm. Therefore, it was highly necessary to measure BCMSD with wider  $D_p$  range for better estimation of  $m_{\rm BC,bulk}$  and deeper understanding of BC evolution in the atmosphere.

Figure 3 exhibited the mean eBCMSD<sub>AAC-AE33</sub> (eBCMSD<sub>AAC-AE33</sub>) and mean rBCMSD<sub>DMA-SP2</sub> (rBCMSD<sub>DMA-SP2</sub>) during the field measurement. It could be seen that when  $D_p$  was less than about 300 nm,  $\overline{\text{eBCMSD}}_{AAC-AE33}$  was higher than  $\overline{\text{rBCMSD}}_{DMA-SP2}$ . The higher  $\overline{\text{eBCMSD}}_{AAC-AE33}$  may be due to particle diffusion at small  $D_p$  which was not well corrected by (7) and underestimated MAC. When  $D_p$  was greater than 300 nm and less than 720 nm,  $\overline{\text{eBCMSD}}_{AAC-AE33}$  was overall consistent with  $\overline{\text{rBCMSD}}_{DMA-SP2}$ . When  $D_p$  was larger than 720 nm, where DMA – SP2 did not measure,  $\overline{\text{eBCMSD}}_{AAC-AE33}$  decreased with increasing  $D_p$  when  $D_p$  less than about 870 nm, and increased with increasing  $D_p$  when  $D_p$  was larger than 870 nm. In the study by Yu et al. (2010), three modes of BCMSD were identified: the mode peaked at about 400 nm, 1000 nm and 5000 nm, which were termed as condensation mode, droplet mode and coarse mode, respectively. Following the nomenclature proposed by Yu et al. (2010), the increasing (decreasing)  $\overline{\text{eBCMSD}}_{AAC-AE33}$  with increasing  $D_p$  for  $D_p$  larger (less) than 870 nm signified the lower (upper) end of droplet mode (condensation mode). However,  $\overline{\text{rBCMSD}}_{DMA-SP2}$  only identified condensation mode, which clearly highlighted the necessity to measure BCMSD of wider  $D_p$  range for better characterization of BCMSD. Both the proposed method and DMA – SP2 showed that the temporal variation of BCMSD, expressed as standard deviation (std) of BCMSD in Fig. 3, was as large as  $\overline{\text{BCMSD}}_p$ , reflecting the complex mechanism in the variability of BCMSD.

#### 4 Uncertainty analysis

Uncertainty analysis was performed with respect to the MAC lookup table, transfer function inversion from  $\sigma_{ab,binned}$  to  $\sigma_{ab,size}$ 

resolved,  $f_{BC}$  and instrumental noise, respectively. The resulting uncertainty to eBCMSD<sub>AAC-AE33</sub> was illustrated in Fig. 4 and to  $m_{eBC,bulk,AAC-AE33,200-720}$  was shown in table 1. It could be seen from Fig. 4 that the boundary between condensation mode and droplet mode was distinct no matter how the uncertainty sources disturbed eBCMSD<sub>AAC-AE33</sub>.

## 4.1 Uncertainty from masss absorption cross-section

According to Zhao et al. (2021b), the variation in refractive index (RI) dominated the uncertainty of the MAC lookup table. Therefore, the uncertainty from the MAC lookup table was analyzed with respect to RI. The real part of RI (Re[RI]) was reported to vary from 1.5 to 2.0 in general (Liu et al., 2018) and the imaginary part of RI (Im[RI]) ranged from 0.5 to 1.1 commonly (Bond and Bergstrom, 2006). Hence, Re[RI] (Im[RI]) was changed from 1.5 (0.5) to 2.0 (1.1) with step increase of 0.01, the resulting mean MAC ( $\overline{\text{MAC}}$ ) was the MAC lookup table used in this study (Fig. S3a) and the resulting std divided by the  $\overline{\text{MAC}}$  was considered as the uncertainty of the MAC lookup table (Fig. S3b). As shown in Fig. S3b, when  $D_c$  was larger than about 400 nm, the uncertainty was less than 10% and influenced by both  $D_p$  and  $D_c$ . However, when  $D_c$  was less than 400 nm, the uncertainty increased rapidly with decreasing  $D_c$  and was dominated by  $D_c$ . The uncertainty increased to larger than 23 % when  $D_c$  was less than about 100 nm. When  $D_p$  was less than about 300 nm, the uncertainty varied from 14% to 24% with a mean value of 22 %, indicating large uncertainty in eBCMSD<sub>AAC-AE33</sub> for  $D_p$  less than 300 nm.

The MAC lookup table was replaced with original  $\overline{\text{MAC}}$  minus its std (-stdMAC) and plus its std (+stdMAC). The resulting MAC lookup tables were taken into the method proposed by Zhao et al. (2021b), and applied to the entire measurement period to study the influence of MAC variation on the eBCMSD<sub>AAC-AE33</sub>.  $\overline{\text{eBCMSD}}_{\text{AAC-AE33}}$  and mean eBCMSD<sub>AAC-AE33</sub> after  $\pm$  stdMAC were shown in Fig. 4a. The uncertainty increased with decreasing  $D_p$ , and reached to maximum when  $D_p$  was less than 300 nm, indicating the eBCMSD<sub>AAC-AE33</sub> for  $D_p$  less than 300 nm might potentially have nonnegligible uncertainty. The uncertainty in  $m_{\text{eBC,bulk,AAC-AE33,200-720}}$  was from -20% (+stdMAC) to +28% (-stdMAC), which was the largest among the four uncertainty sources, as shown in Table 1.

## 4.2 Uncertainty from the transfer function inversion

As stated in Sect. 2.4.1, correction for particle loss and spectral broadening was required when  $\sigma_{ab,binned}$  was converted to  $\sigma_{ab,size-resolved}$ .  $\lambda_{\Omega}$  was defined to correct particle loss where  $\lambda_{\Omega} = 0$  ( $\lambda_{\Omega} = 1$ ) stood for total (no) particle loss. The relationships between  $\lambda_{\Omega}$  and  $D_p$  as well as  $Q_{sample}$ , as shown in Fig. S4a, were well developed (Karlsson and Martinsson, 2003) and consistent with experimental data of AAC (Johnson et al., 2018).  $Q_{sample}$  used in this study was 3.0 L min<sup>-1</sup>.  $Q_{sample}$  was changed from – 30 % (2.1 L min<sup>-1</sup>) to + 30% (3.9 L min<sup>-1</sup>), and the resulting  $\lambda_{\Omega}$  was used to study the fluctuation of  $Q_{sample}$  on  $\lambda_{\Omega}$ . As shown in Fig. S4a, the variation of  $\lambda_{\Omega}$  was less than 0.5 %, which was negligible.

Spectral broadening was caused by small-size particle diffusion as well as disturbance of sheath flow due to complicated rotation fluid dynamics inside AAC and described by  $\mu_{\Omega}$  where  $\mu_{\Omega} < 1$  ( $\mu_{\Omega} = 1$ ) was for (no) broadening. Johnson et al. (2018) found that behavior of  $\mu_{\Omega}$  with respect to  $D_p$  ( $\mu_{\Omega}(D_p)$ ) depended on both  $Q_{\text{sheath}}$  and  $Q_{\text{sample}}$ , and parameterized  $\mu_{\Omega}(D_p)$  (Fig. S4b) based on two case studies of ( $Q_{\text{sheath}}$ ,  $Q_{\text{sample}}$ ), namely (3 L min<sup>-1</sup>, 0.3 L min<sup>-1</sup>) and (15 L min<sup>-1</sup>, 1.5 L min<sup>-1</sup>):

 $\mu_{\Omega,\text{Johnson}}(D_{p}) = 0.318 \cdot D_{p}^{0.0946}.$  (15)

 $(Q_{\text{sheath}}, Q_{\text{sample}})$  was  $(7.5 \text{ L min}^{-1}, 3 \text{ L min}^{-1})$  in this study, which might lead to uncertainty if  $\mu_{\Omega,\text{Johnson}}(D_p)$  was directly used in this study.  $\mu_{\Omega,\text{Johnson}}(D_p)$  was varied from -23% to +30% to study the impact of  $\mu_{\Omega}$  on eBCMSD<sub>AAC-AE33</sub>. The resulting influence on the eBCMSD<sub>AAC-AE33</sub> was shown in Fig. 4b. The uncertainty of  $\mu_{\Omega}$  did not exhibit a significant size dependence. Lower  $\mu_{\Omega}$  led to lower eBCMSD<sub>AAC-AE33</sub>, and vice versa. The difference in the  $m_{\text{eBC,bulk,AAC-AE33,200-720}}$  was from -1% (-23% of  $\mu_{\Omega,\text{Johnson}}$ ) to +21% (+30% of  $\mu_{\Omega,\text{Johnson}}$ ), implying systematic overestimation of  $m_{\text{eBC,bulk,AAC-AE33,200-720}}$ . Therefore,  $\mu_{\Omega}(D_p) = 77\%\mu_{\Omega,\text{Johnson}}(D_p)$  was used in this study to offset the bias. The reason for the lower limit of  $\mu_{\Omega}$  set as -23% rather than -30% was that eBCMSD<sub>AAC-AE33</sub> was invalid mathematically when  $\mu_{\Omega}$  was less than -23% under setting of  $(Q_{\text{sheath}}, Q_{\text{sample}})$  used in this study. It should be pointed out that AAC - AE33 could still determine valid eBCMSD under the condition of  $\mu_{\Omega}(D_p) < 77\%\mu_{\Omega,\text{Johnson}}(D_p)$  by increasing  $Q_{\text{sheath}}$ .  $Q_{\text{sheath}}$  was not changed in this study because desired  $\mu_{\Omega}(D_p)$  parameterization was found at  $\mu_{\Omega}(D_p) = 77\%\mu_{\Omega,\text{Johnson}}(D_p)$ .

 $\sigma_{ab,size-resolved}$  (Fig. S5b) was integrated and the result, denoted as  $\sigma_{ab,bulk,calc}$  was compared to  $\sigma_{ab,bulk}$ . As shown in Fig. S5a,  $\sigma_{ab,bulk,calc}$  was consistent with  $\sigma_{ab,bulk}$ .  $R^2$ ,  $b_1$  and  $b_0$  between  $\sigma_{ab,bulk,calc}$  and  $\sigma_{ab,bulk}$  was 0.9, 1.1, and 0.5 Mm<sup>-1</sup> (Fig. S6), respectively, which validated conversion from  $\sigma_{ab,binned}$  to  $\sigma_{ab,size-resolved}$ .

#### 4.3 Uncertainty from number fraction of BC-containing particle

BC-containing aerosol particles had complicated mixing states. Even for internally-mixed particles with same  $D_p$ , the internal BC cores might have different  $D_e$ , which could not be resolved by AAC – AE33. Field measurement (Liu et al., 2010) revealed that  $f_{BC}$  varied with time,  $D_e$  and  $D_p$ . This complexity was simplified to a parameterized fixed value of  $f_{BC}$  in this study. A fixed  $f_{BC}$  implied that only one  $D_e$  value corresponded to a given  $D_p$  and the size-resolved number concentration of BC-containing particle was determined by  $N_{\text{size-resolved}}$  times  $f_{BC}$ . Zhao et al. (2021b) used  $f_{BC}$  value of 0.17 based on SP2 measurement. However, SP2-derived  $f_{BC}$  could not represent the bulk  $f_{BC}$  over the whole particle size spectrum due to the detection limit of SP2. And different regions might have different  $f_{BC}$ . In this study,  $f_{BC}$  was varied and the resulting  $m_{eBC,bulk,AAC-AE33,200-720}$  was compared with  $m_{rBC,bulk,DMA-SP2,200-720}$ ,  $f_{BC}$  was set as 0.35 in this study because  $b_1$  was 1.0 between  $m_{eBC,bulk,AAC-AE33,200-720}$  and  $m_{rBC,bulk,DMA-SP2,200-720}$  when  $f_{BC}$  was varied from 0.25 (– 30 % of 0.35) to 0.46 (+ 30 % of 0.35) to analyze the influence of  $f_{BC}$  on the eBCMSD<sub>AAC-AE33</sub>, as shown in Fig. 4c. eBCMSD<sub>AAC-AE33</sub> was more influenced around 870 nm. The variation of  $f_{BC}$  led to uncertainty of  $\pm 3$  % in  $m_{eBC,bulk,AAC-AE33,200-720}$ , indicating that simplification of  $f_{BC}$  was a minor uncertainty source compared to MAC and transfer function inversion.

## 4.4 Uncertainty from instrumental noise

The instrumental noise stemmed from the fluctuation of the light source and flow of AE33, which was reflected as fluctuation in I,  $I_0$  and F, further leading to the fluctuation in ATN,  $\sigma_{\text{ATN}}$  and  $\sigma_{\text{ab}}$ . The instrumental noise was defined as the std of  $\sigma_{\text{ab,binned}}$  and was shown in Fig. S7b. It could be seen that the instrumental noise did not exhibit significant dependence on  $D_p$ . Comparing Fig. S7a and Fig. S7b, the instrumental noise was irrelevant to the value of  $\sigma_{\text{ab,binned}}$ . Figure S7c illustrated that the instrumental noise was also not correlated to  $\sigma_{\text{ab,bulk}}$  with  $R^2$ ,  $b_1$  and  $b_0$  of 0.0, 0.0 and 0.1 Mm<sup>-1</sup>, respectively, implying that the instrumental noise was not

dependent on the pollution level.

The std of instrumental noise was added to (subtracted from)  $\sigma_{ab,binned}$  and the derived eBCMSD<sub>AAC-AE33</sub> was used to study the influence of instrumental noise on eBCMSD<sub>AAC-AE33</sub>. The mean result was shown in Fig. 4d. eBCMSD<sub>AAC-AE33</sub> larger than 1000 nm was more influenced by the instrumental noise than eBCMSD<sub>AAC-AE33</sub> smaller than 500 nm.  $\sigma_{ab,binned}$  larger than 1000 nm was relatively small (about 0.3 Mm<sup>-1</sup>) compared to  $\sigma_{ab,binned}$  less than 870 nm (about 0.9 Mm<sup>-1</sup>). The mean instrumental noise was 0.1 Mm<sup>-1</sup> and exhibited no significant dependence on  $D_p$ . Therefore,  $\sigma_{ab,binned}$  larger than 1000 nm was more affected by the instrumental noise, resulting in distinct variation in eBCMSD<sub>AAC-AE33</sub>. The uncertainty in  $m_{eBC,bulk,AAC-AE33,200-720}$  was not obvious, which varied from -2% to -1%, also minor compared to MAC and transfer function inversion.

## **5 Summary**

Knowledge of bulk black carbon (BC) characteristics, such as bulk BC mass concentration ( $m_{BC,bulk}$ ), was very limiting for deeper understanding the influence of BC on radiation and health. BC mass size distribution (BCMSD) was one of the BC microphysical characteristics that could indicate emission source, reflect atmospheric aging and effectively reduce uncertainty related to BC radiative effect. However, current BCMSD measurement ranging from upper micron to larger than 1  $\mu$ m was insufficient in time resolution and sophisiticated for routine measurement. In this study, a new method to determine equivalent BCMSD (eBCMSD) was proposed characterized by wide size range of measurement up to 1.5  $\mu$ m, high time resolution up to 1 hour and convenience for extensive measurement. The eBCMSD was retrieved by size-resolved absorption coefficient ( $\sigma_{ab,size-resolved}$ ) measured by an aerodynamic aerosol classifer in tandem with an aethaelometer model AE33 (AAC – AE33), denoted as eBCMSD<sub>AAC-AE33</sub>. Size-resolved number concentration ( $N_{size-resolved}$ ) was measured concurrently by scanning mobility particle sizer (SMPS) and an aerodynamic particle sizer (APS) to model the influence of size on mass absorption cross-section (MAC).

The proposed method was applied to a field measurement in Yangtze River Delta and validated by comparing the refractory BCMSD (rBCMSD) with that measured by an differential mobility analyzer in tandem with a single-particle soot photometer (DMA – SP2), denoted as rBCMSD<sub>DMA-SP2</sub>. The results showed that for particle diameter ( $D_p$ ) less than 300 nm, eBCMSD<sub>AAC-AE33</sub> was higher than rBCMSD<sub>DMA-SP2</sub>, which might be caused by underestimated MAC by the method proposed by Zhao et al. (2021b) or incomplete parameterization of spectral broadening of AAC. eBCMSD<sub>AAC-AE33</sub> was consistent with rBCMSD<sub>DMA-SP2</sub> for  $D_p$  larger than 300 nm.  $m_{BC,bulk}$  integrated over the size range that AAC – AE33 and DMA – SP2 both measured (200 nm – 720 nm), denoted as  $m_{eBC,bulk,AAC-AE33,200-720}$  and  $m_{rBC,bulk,DMA-SP2,200-720}$ , compared well with each other with determination coefficient ( $R^2$ ), slope ( $D_p$ ), and intercept ( $D_p$ ) of 0.8, 1.0 and 0.1 µg m<sup>-3</sup>, respectively. However, DMA – SP2 did not measure  $D_p$  larger than 700 nm, leading to 0.2 µg m<sup>-3</sup> (33 %) underestimation of  $D_p$  larger than 700 nm, leading to

Uncertainty analysis was performed with respect to MAC, transfer function inversion, number fraction of BC-containing particle  $(f_{\rm BC})$  and instrumental noise. MAC was the largest uncertainty source, leading to significant uncertainty for  $D_{\rm p}$  less than 300 nm and about 24% uncertainty in  $m_{\rm eBC,bulk,AAC-AE33,200-720}$ . Transfer function inversion was the second largest uncertainty source, which was induced by incomplete parameterization of spectral broadening. The uncertainty in transfer function inversion led to systematic

- overestimation of  $m_{eBC,bulk,AAC-AE33,200-720}$ , which was corrected in this study. Both  $f_{BC}$  and instrumental noise were minor uncertainty
- 314 sources and one order of magnitude less than MAC and transfer function inversion. f<sub>BC</sub> was simplification of complicated BC mixing
- 315 states, leading to relatively big uncertainty in eBCMSD at 870 nm, around the boundary between condensation mode and droplet
- mode. The eBCMSD for  $D_p$  larger than 1000 nm was more sensitive to instrumental noise.
  - Data availability

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- The measurement data involved in this study are available upon request to the authors.
- 319 Author contributions
- 320 CZ determined the main goal of this study. WZ and GZ designed the methods. WZ carried them out and prepared the paper with
- 321 contributions from all co-authors.
  - Competing interests
  - The authors declare that they have no conflict of interest.
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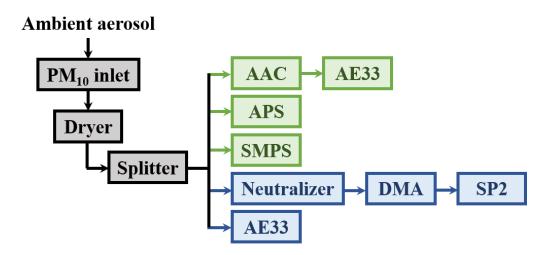


Figure 1: Schematic diagram of the measurement (green) and the validation (blue) setup.



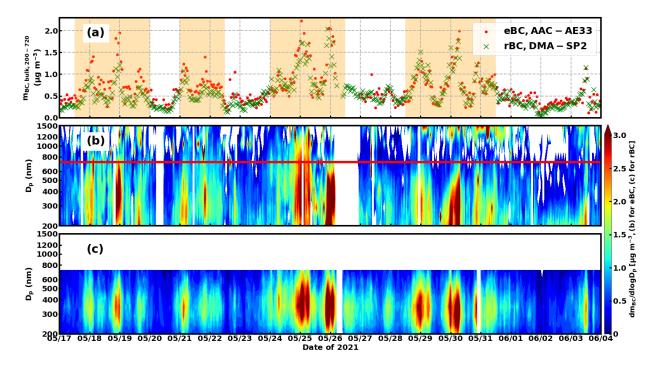


Figure 2: Time series of (a)  $m_{BC,bulk}$  from BCMSD integrated from 200 nm to 720 nm ( $m_{BC,bulk,200-720}$ ) determined by the proposed method (red dot, denoted as "eBC, AAC – AE33") and DMA – SP2 (green cross, denoted as "rBC, DMA – SP2"), BCMSD determined by (b) the proposed method (eBC) and (c) DMA – SP2 (rBC). The red line in (b) marked particle size of 720 nm. The pollution episodes were shaded with orange in (a).

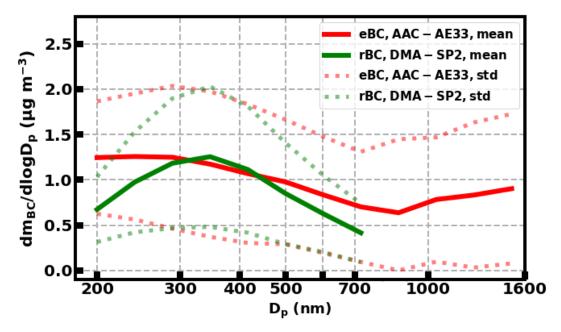


Figure 3: Mean and std of eBCMSD<sub>AAC-AE33</sub> (green) and rBCMSD<sub>DMA-SP2</sub> (red) during the measurement period. Mean BCMSD was denoted by the solid line. Std of BCMSD was denoted by the dotted line.

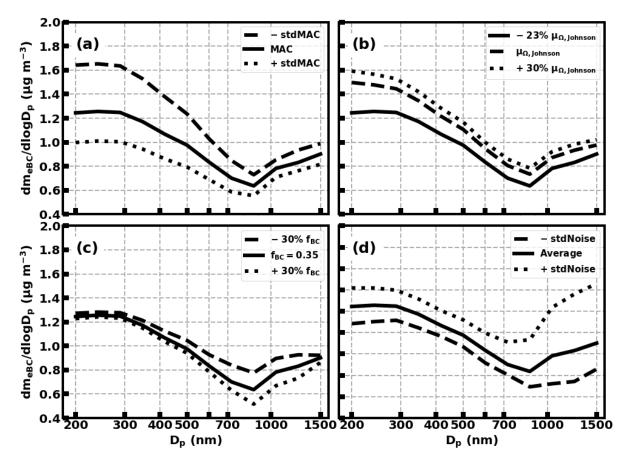


Figure 4: Uncertainty in eBCMSD<sub>AAC-AE33</sub> arising from (a) MAC lookup table, (b) transfer function inversion, (c)  $f_{BC}$  and (d) instrumental noise. The solid lines in each panel were the same and were the mean eBCMSD<sub>AAC-AE33</sub> during the measurement period. The dotted lines and dashed lines in each panel were mean eBCMSD<sub>AAC-AE33</sub>  $\pm$  standard deviation (std) calculated from (a) MAC + std of MAC and MAC – std of MAC, (b)  $\mu_{\Omega,Johnson}$  and 1.3 times  $\mu_{\Omega,Johnson}$ , (c)  $f_{BC}$  of 35% times 0.7 and 1.3, (d) eBCMSD<sub>AAC-AE33</sub> + std of instrumental noise and – std of instrumental noise.

Table 1: The Uncertainty in the  $m_{\rm eBC,bulk,AAC-AE33,200-720}$  resulted from MAC lookup table, transfer function inversion,  $f_{\rm BC}$  and instrumental noise.

Uncertainty source	MAC	Transfer function inversion	$f_{ m BC}$	Instrumental noise
Uncertainty	- 20 % ~ + 28 %	- 1 % ~ + 21 %	-3 % ~ +3 %	-2 % ~ -1 %