1 New method to determine black carbon mass size distribution

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Abstract. Black carbon (BC) is an important atmospheric component with strong light absorption. Many attempts have been made 10 to measure BC mass size distribution (BCMSD) for its significant impact on climate and public health. Larger-coverage BCMSD, 11 ranging from upper submicron to larger than 1 µm, contributes to substantial proportion of total BC mass and absorption. However, 12 current time resolution of larger-coverage BCMSD measurement was limited to 1 day, which was insufficient to characterize 13 14 variation of larger-coverage BCMSD. In this study, a new method to determine eBCMSD was proposed from size-resolved 15 absorption coefficient measured by an aerodynamic aerosol classifier in tandem with an aethalometer. The proposed method could measure larger-coverage eBCMSD with time resolution as high as 1 hour and was validated by comparing the measurement results 16 with rBCMSD measured by a differential mobility analyzer in tandem with a single particle soot photometer (DMA - SP2) for 17 particle size larger than 300 nm during a field measurement in Yangtze River Delta. Bulk refractory BC mass concentration ($m_{rBC,bulk}$) 18 by DMA – SP2 was underestimated by 33 % compared to bulk equivalent BC mass concentration ($m_{eBC,bulk}$) by this method because 19 of the limited size range of measurement for DMA - SP2. Uncertainty analysis of this method was performed with respect to mass 20 21 absorption cross-section (MAC), transfer function inversion, number fraction of BC-containing particle and instrumental noise. The 22 results indicated that MAC was the main uncertainty source, leading to $m_{eBC,bulk}$ varied from -20 % to 28 %. With the advanatage 23 of wide size coverage up to 1.5 µm, high time resolution, easy operation and low cost, this method is expected to have wide applications in field measurement for better estimating radiative properties and climate effects of BC. 24

25 1 Introduction

Atmospheric black carbon (BC) is emitted through incomplete combustion of carbon-based fuels (Bond, 2001), such as biomass burning and fossil fuel combustion (Bond et al., 2004). BC warms atmosphere and is considered the third important light absorber in the climate system after CO_2 and CH_4 (Bond et al., 2013). BC can induce the respiratory and cardiovascular disease through inhalation (Highwood and Kinnersley, 2006). Plenty of studies have been devoted to BC for its significant impact on the climate and public health.

Bulk BC mass concentration ($m_{BC,bulk}$) is one of the essential parameters for modeling because radiative transfer models calculate

32 BC absorption from $m_{BC,bulk}$ (Bond et al., 2013). A great deal of research has been dedicated to $m_{BC,bulk}$ measurement for both model 33 assimilation and environmental monitoring (Castagna et al., 2019;Helin et al., 2018;Ran et al., 2016). A recent study indicated that 34 the radiative effect of BC was extremely sensitive to its particle size (Matsui et al., 2018). Zhao et al. (2019) further revealed that 35 the variation of BC mass size distribution (BCMSD), namely size-resolved BC mass concentration ($m_{BC-size-resolved}$), led to substantial changes in the radiative effect of BC based on field measurement, highlighting the importance of BCMSD on modeling the radiative 36 37 effect of BC rather than simply $m_{BC,bulk}$. The size of BC affects the deposition rate of BC to the lung (Highwood and Kinnersley, 38 2006), indicating that BCMSD is closely related to health. In the ambient environment, BCMSD is influenced by emission sources. 39 For instance, BCMSD of fossil fuel combustion differs obviously from that of biomass burning (Schwarz et al., 2008), implying that BCMSD is one of the characteristics of emission source. The temporal variation of BCMSD can reflect the atmospheric aging 40 41 of BC, during which BC undergoes remarkable change in size, accompanied by dramatical variation of BC optical properties (Zhang 42 et al., 2008). Therefore, reliable measurement and understanding of BCMSD are highly necessary for estimating the impact of BC on both the earth energy budget and public health (Moosmuller et al., 2009). 43

Quite a few efforts were made to measure BCMSD. The single-particle soot photometer (SP2) was developed using laser-induced 44 45 incandescence (Schwarz et al., 2006), which measured rBCMSD (Petzold et al. (2013)) on a single-particle level. The soot particle 46 aerosol mass spectrometer (SP-AMS) combined laser-induced incandescence and mass spectroscopy, which could determine not only rBCMSD but also the chemical composition of BC-containing particles (Onasch et al., 2012). The SP2 and SP-AMS techniques 47 48 were characterized by high time resolution and high accuracy, but high cost and complicated maintenance, as a result, not widely 49 applied for routine measurement. A more convenient solution was required for wider and better characterization of BCMSD in 50 different regions and emission sources. ECMSD (Petzold et al. (2013)) could be sampled by multi-stage cascade impactor (Viidanoja et al., 2002) combined with off-line analysis, such as thermo/optical organic carbon/elemental carbon method (Chow et al., 2001). 51 52 ECMSD sampled by multi-stage cascade impactor had wide size coverage, but low time resolution, usually from 24 hours (Soto-53 Garcia et al., 2011) to 48 hours (Guo, 2015), which was too low to resolve aging of BC. Differential mobility analyzer (DMA) in 54 tandem with filter-based instrument (Hansen et al., 1984), for instance, micro-aethalometer (MA) (Ning et al., 2013) and particle 55 soot absorption photometer (PSAP) (Tunved et al., 2021), was used to determine eBCMSD (Petzold et al. (2013)) with higher time resolution up to 2 hours (Zhao et al., 2021b). However, the multiple-charge correction and low flow rate of DMA imposed 56 57 restrictions on the accuracy of the measured eBCMSD. The measured size range of DMA was limited to less than about 700 nm, resulting in incomplete measured eBCMSD. Current measurement of larger-coverage BCMSD, ranging from upper submicron to 58 59 larger than 1 µm, was limited in time resolution. Characteristics of larger-coverage BCMSD during atmospheric aging was still unclear. Wang et al. (2022) showed that BC larger than 1 μ m could contribute to as large as 50% of m_{BC,bulk}, highlighting the 60 61 importance of larger-coverage BCMSD. Therefore, it was imperative to measure larger-coverage BCMSD with higher time 62 resolution.

63 In this study, a new method to determine eBCMSD was proposed using size-resolved absorption coefficient ($\sigma_{ab,size-resolved}$)

measured by an aerodynamic aerosol classifier (AAC, Cambustion, UK, Tavakoli and Olfert (2013)) in tandem with an aethalometer (model AE33, Magee, USA, Drinovec et al. (2015)), combined with size-resolved number concentration ($N_{size-resolved}$), simultaneously measured by scanning mobility particle sizer (SMPS, TSI, USA) and an aerodynamic particle sizer (APS, TSI, USA). The proposed method for determining eBCMSD overcame the disadvantages and weighed the advantages of the above-mentioned methods, characterized by high cost-effectiveness, easy and convenient maintenance, high time resolution to 1 hour, and wide size range to up to 1.5 μ m. The proposed method was validated in a field measurement in the Yangtze River Delta and the uncertainty study was carried out based on the measured data.

71 2 Methods

72 2.1 Instrumental setup

73 Figure 1 illustrated the instrumental setup developed in this study, which could be split into two parts, namely the measurement 74 setup and the validation setup. Ambient aerosol particles were drawn through a PM_{10} inlet (16.67 L min⁻¹) and were dried to relative 75 humidity less than 30 % by silica gel diffusion drier beforehand. An advanced flow splitter were used to split aerosol laden flow 76 into different instruments for isokinetic sampling. For the measurement setup, AAC in tandem with AE33 (AAC - AE33) measured 77 $\sigma_{ab,size-resolved}$ at a flow rate of 3 L min⁻¹. Since eBCMSD of larger size coverage was mainly focused in this study, AAC was set to 78 scan 12 particle sizes (D_p) logarithmically equally spaced from 200 nm to 1500 nm (200 nm, 240 nm, 288 nm, 346 nm, 416 nm, 79 499 nm, 600 nm, 720 nm, 865 nm, 1039 nm, 1248 nm, 1500 nm) and smaller size was not scanned. Each size was scanned for 5 minutes and 1 hour was required for a complete cycle. It should be noted that $D_{\rm p}$ was aerodynamic size in this study. Mobility size 80 81 related to DMA was converted to aerodynamic size assuming an effective density of 1.3 g cm⁻³ (Zhao et al., 2019;DeCarlo et al., 2005). AE33 measured absorption coefficient (σ_{ab}) at 7 wavelengths from 370 nm to 950 nm (Drinovec et al., 2015), at which 880 82 nm was adopted in this study because BC dominated particle absorption at 880 nm (Ramachandran and Rajesh, 2007). SMPS and 83 APS measured $N_{\text{size-resolved}}$ for D_{p} less than and greater than about 800 nm at a flow rate of 0.3 L min⁻¹ and 5 L min⁻¹, respectively. 84 85 For validation system, the tandem array of Neutralizer, DMA and SP2 (DMA - SP2) measured rBCMSD (rBCMSD_{DMA-SP2}) at a flow rate of 0.12 L min⁻¹ for purpose of comparing with eBCMSD determined by the proposed method (eBCMSD_{AAC-AE33}). Another 86 AE33 measured bulk absorption coefficient ($\sigma_{ab,bulk}$) simultaneously at a flow rate of 2 L min⁻¹ for closure study with $\sigma_{ab,size-resolved}$. 87

88 2.1.1 Aethalometer model AE33

The principle of obtaining σ_{ab} was well developed for aethalometer (Hansen et al., 1984) and described here in brief. Ambient aerosol particles were drawn into an aethalometer at a flow rate *F* and collected on an area *S* of a filter. A light source illuminated the filter. The transmitted light signal was denoted as $I(I_0)$ for the light which passed through the particle-laden (particle-free) part of the filter. Light attenuation was defined as

93
$$\operatorname{ATN} = -100 \cdot \ln\left(\frac{l}{l_0}\right). \tag{1}$$

94 ATN increased with decreasing I as aerosol particles were loaded on the filter continuously. Therefore, ATN reflected aerosol loading

95 on the filter. If ATN increased Δ ATN during time interval Δt , then attenuation coefficient was defined as

96
$$\sigma_{\text{ATN}} = \frac{S}{100 \cdot F} \cdot \frac{\Delta \text{ATN}}{\Delta t}.$$
 (2)

97 The light attenuation was caused by not only particle absorption, but also scattering by particle and filter matrix. A scattering 98 parameter C_f was introduced to extract σ_{ab} from σ_{ATN} :

99
$$\sigma_{ab} = \frac{\sigma_{ATN}}{c_f},$$
(3)

100 where C_f was set as 2.9 (Zhao et al., 2020b) in this study. Nonlinearity, termed loading effect, became more and more significant 101 with increasing aerosol loading, namely for the same increase in aerosol loading, the corresponding increase in ATN was smaller 102 for heavier aerosol loading. The "dual-spot" technique (Drinovec et al., 2015) was proposed to correct the loading effect for AE33 103 and was used in this study. The σ_{ab} measured by AE33 at given particle size selected by AAC was termed binned σ_{ab} ($\sigma_{ab,binned}$) in 104 this study to differentiate from $\sigma_{ab,bulk}$.

105 2.1.2 Aerodynamic aerosol classifer

The principle of AAC was illustrated detailedly by Tavakoli and Olfert (2013) and was introduced here concisely. The setup of AAC could be simplified to two coaxial cylinders, where the inner radius, outer radius and length was denoted as r_i , r_o and L. Polydisperse particles flowed into the space between the inner cylinder and outer cylinder from one end of the inner cylinder at a flow rate of Q_{sample} . Meawhile, Particle free sheath flow streamed in the space between the inner cylinder and outer cylinder in the direction of the axis of the coaxial cylinders at a flow rate of Q_{sheath} . The sheath flow carried the particles along the coaxial cylinders. At the same time, the two coaxial cylinders rotated with respect to their axis at a rotational speed of ω . Therefore, the particles was migrated outwards across the sheath flow by the centrifugal force acting on them. Relaxation time (τ) was defined as

114 where $C_c(D_p)$ was the Cunningham slip correction factor as a function of D_p (Kim et al., 2005), $\rho_0 = 1$ g cm⁻³ was the reference 115 density and μ was the gas viscosity. It could be seen that τ was directly related to D_p . Dynamic analysis proved that only particles 116 with certain relaxation time τ , and hence certain D_p , could migrate to another end of outer cylinder and emerge as monodisperse 117 flow. τ was related to parameters of AAC by

118
$$\tau = \frac{2Q_{\text{sheath}}}{\pi\omega^2(r_1+r_0)^2L}.$$
(5)

119 Therefore, by changing ω and Q_{sheath} , monodisperse particles of desired D_p could be selected by AAC. Unlike DMA, particles did 120 not need to be charged before entering AAC. Consequently, the transmission efficiency (λ_Ω) of AAC was about 4 times higher than 121 that of DMA and multi-charge correction was not required for AE33 (Johnson et al., 2018).

122 **2.1.3 Differential mobility analyzer in tandem with single-particle soot photometer**

Ambient particles were first drawn into DMA, which could select monodispersed particles with D_p from the polydispersed ambient particles. Then the monodispersed particles flew into SP2, which measured rBC mass of BC-containing particles on single

- particle level. rBCMSD_{DMA-SP2} was measured by changing D_p selected by DMA and measuring rBC mass at different D_p s. SP2 alone could also determine rBCMSD (rBCMSD_{SP2}). Considering that D_p of rBCMSD_{SP2} was calculated from Mie theory with assumed
- 127 inputs (Taylor et al., 2015) and D_p of rBCMSD_{DMA-SP2} was directly measured by DMA, rBCMSD_{DMA-SP2} was used in this study.
- The detail of DMA SP2 was described by Zhao et al. (2021a) and introduced here concisely. The incandescence high gain channel and the incandescence low gain channel of SP2 were calibrated by Aquadag soot particles with effective density of 1.8 g cm⁻³ before measurement. DMA was set to scan D_p from 15 nm to 780 nm with 56 bins and 1 scan cycle required 5 minutes. Therefore, time resolution and size range of the measured rBCMSD_{DMA-SP2} were 5 minutes and from 15 nm to 780 nm. Considering that the time resolution of eBCMSD_{AAC-AE33} was 1 hour, the measured rBCMSD_{DMA-SP2} was time averaged to time resolution of 1 hour for comparison. Given that size range of eBCMSD_{AAC-AE33} was from 200 nm to 1500 nm, the size range of the measured rBCMSD_{DMA-SP2} was interpolated to that of eBCMSD_{AAC-AE33} for comparison. Consequently, the size range of rBCMSD_{DMA-SP2} used
- in this study was from 200 nm to 720 nm, corresponding to the 1st and 8th D_p of eBCMSD_{AAC-AE33}, respectively. According to Zhao et al. (2020a), DMA – SP2 could not detect particle less than about 200 nm, consistent with the minimum D_p in this study.
- The number distribution of BC-containing particle measured by DMA SP2 could be considered as a bivariate function $\frac{\partial^2 N_{BC}}{\partial \log D_p \partial \log D_c}$, where N_{BC} was the number concentration of BC-containing particle, and D_c was the core diameter assuming that BCcontaining particle had a core-shell geometry. The asphericity of BC was not considered in this study. D_c was logarithmically equally distributed from 78 nm to 560 nm with 29 bins. At each D_c bin, the number distribution of BC-containing particle could be considered as a univariate function $\frac{dN_{BC}}{dlogD_p}$. The multiple charging correction of 1-dimensional size distribution was developed by Knutson and Whitby (1975), which was adopted to correct $\frac{dN_{BC}}{dlogD_p}$. By correcting $\frac{dN_{BC}}{dlogD_p}$ at each D_c bin, multiple charging correction of $\frac{\partial^2 N_{BC}}{\partial \log D_p \partial \log D_c}$ was achieved.

144 **2.2 Field measurement**

The proposed method was applied to a field measurement in Changzhou, Jiangsu Province, China (119°36'E, 31°43'N), located in the Yangtze River Delta, from May 17th to June 3rd in 2021. Changzhou was between two megacities, about 187 km to the northwest of Shanghai and about 82 km to the southeast of Nanjing, as shown in Fig. S1a. The area between the Shanghai and Nanjing underwent serious environmental pollution in the past few decades with the development of industrialization and urbanization. As shown in Fig. S1b, the pollution condition of Changzhou was overall representative of the regional background pollution in the Yangtze River Delta.

151 **2.3 Size-resolved calculations**

152 2.3.1 Size-resolved absorption coefficient

153 Tavakoli and Olfert (2013) formulated the ideal theoretical model for AAC transfer function inversion, which was adopted by this

study to derive $\sigma_{ab,size-resolved}$ from $\sigma_{ab,binned}$ and presented here in short. The $\sigma_{ab,size-resolved}$ was given by

155
$$\sigma_{ab,size-resolved} = \frac{d\sigma_{ab}}{dlog D_p} = \frac{ln(10)}{\frac{dlog D_p}{dlog \tau} \cdot \beta^*} \cdot \sigma_{ab,binned},$$
(6)

156 where β^* was related to the ratio of Q_{sheath} to Q_{sample} , $R_t = Q_{\text{sheath}}/Q_{\text{sample}} = 1/\beta$, through

157
$$\beta^* = \left(1 + \frac{1}{\beta}\right) \ln(1 + \beta) - \left(1 - \frac{1}{\beta}\right) \ln(1 - \beta).$$
 (7)

158 Johnson et al. (2018) corrected the ideal inversion formula (6) to take particle loss and spectral broadening into account by replacing

159 β^* with a nonideal $\beta^*_{\rm NI}$:

160
$$\beta_{\rm NI}^* = \lambda_{\Omega} \mu_{\Omega} \left[\ln \left(\frac{1 + \beta / \mu_{\Omega}}{1 - \beta / \mu_{\Omega}} \right) + \frac{\mu_{\Omega}}{\beta} \ln \left(1 - \left(\frac{\beta}{\mu_{\Omega}} \right)^2 \right) \right], \tag{8}$$

where μ_{Ω} was the spectral broadening factor. Both λ_{Ω} and μ_{Ω} were dependent on D_{p} as well as flow, and discussed in detail in Sect. 4.2.

163 2.3.2 eBC mass size distribution measured by aerodynamic aerosol classifer in tandem with aethalometer model AE33

- 164 $\sigma_{ab,size-resolved}$ could be converted to eBCMSD_{AAC-AE33} through mass absorption cross-section (MAC) (Bond and Bergstrom, 2006), 165 which was determined based on method proposed by Zhao et al. (2021b). Briefly, a 2-dimensional lookup table of MAC with respect 166 to D_p and D_c was simulated (MAC(D_p , D_c)) by Mie theory assuming a core-shell geometry. At a given size bin centered at D_p , the 167 corresponding σ_{ab} and number concentration (N) at the size bin, denoted as $\sigma_{ab}(D_p)$ and $N(D_p)$, could be derived by $\sigma_{ab,size-resolved}$ and
- 168 $N_{\text{size-resolved}}$:

169
$$\sigma_{ab}(D_{p}) = \sigma_{ab,size-resolved}(D_{p}) \cdot \Delta \log D_{p}, \qquad (9)$$

170
$$N(D_{\rm p}) = N_{\rm size-resolved}(D_{\rm p}) \cdot \Delta \log D_{\rm p},$$
 (10)

where $\Delta \log D_p$ was the logarithmic width of the size bin. The number concentration of BC-containing particle at the size bin $N_{BC}(D_p)$ was determined by

173
$$N_{\rm BC}(D_{\rm p}) = N(D_{\rm p}) \cdot f_{\rm BC}, \qquad (11)$$

where f_{BC} was the number fraction of BC-containing particle, which was assumed as a fixed value in this study and the uncertainty of the fixed- f_{BC} assumption was discussed in Sect. 4.3. An optimal D_c was found so that calculated σ_{ab} at the size bin, denoted as

176 $\sigma_{ab,calc}(D_p, D_c)$, matched $\sigma_{ab}(D_p)$:

177
$$\sigma_{\rm ab,calc}(D_{\rm p}, D_{\rm c}) = \rho_{\rm BC} \frac{\pi}{6} D_{\rm c}^{3} \cdot MAC(D_{\rm p}, D_{\rm c}) \cdot N_{\rm BC}(D_{\rm p}) = \sigma_{\rm ab}(D_{\rm p}), \qquad (12)$$

where ρ_{BC} was the density of BC, and set as 1.8 g cm⁻³ (Bond and Bergstrom, 2006), consistent with the ρ_{BC} assumption when deriving rBCMSD_{DMA-SP2}. Equivalent BC mass concentration (m_{eBC}) at the size bin, denoted as $m_{eBC}(D_p)$, could be determined by

180
$$m_{\text{eBC}}(D_{\text{p}}) = \frac{\sigma_{\text{ab}}(D_{\text{p}})}{MAC(D_{\text{p}},D_{\text{c}})},$$
(13)

181 then the eBCMSD_{AAC-AE33} at D_p , denoted by eBCMSD_{AAC-AE33}(D_p), could be determined by

182
$$\operatorname{eBCMSD}(D_{\mathrm{p}}) = \frac{m_{\mathrm{eBC}}(D_{\mathrm{p}})}{\Delta \log D_{\mathrm{p}}}.$$
 (14)

183 3 Results and discussion

184 Figure 2 presented the time series of the measurement results. There were 4 pollution episodes during the measurement period:

from about 12:00 May 17th to about 00:00 May 20th, from about 00:00 May 21st to about 12:00 May 22nd, from about 00:00 May 185 24th to about 12:00 May 26th, and from about 12:00 May 28th to about 12:00 May 31st, as shown in Fig. 2a. Both eBCMSDAAC-AE33 186 187 (Fig. 2b) and rBCMSD_{DMA-SP2} (Fig. 2c) recorded the 4 pollution episodes simultaneously with higher BCMSD values than clean 188 episodes. eBCMSD_{AAC-AE33} and rBCMSD_{DMA-SP2} were both integrated from 200 nm to 720 nm, which was the D_p range of measurement for DMA – SP2, and the results were denoted as $m_{eBC,bulk,AAC-AE33,200-720}$ and $m_{rBC,bulk,DMA-SP2,200-720}$, respectively. As 189 190 shown in Fig. 2a, $m_{\text{eBC,bulk,AAC-AE33,200-720}}$ compared well with $m_{\text{rBC,bulk,DMA-SP2,200-720}}$ with determination coefficient (R^2), slope (b_1), and intercept (b_0) of 0.8 (accurate to one decimal place), 1.0 and 0.1 μ g m⁻³ (Fig. S2a). R^2 of 0.8 and b_1 of 1.0 indicated the proposed 191 method was capable of capturing the variation of $m_{\rm BC,bulk}$. The mean $m_{\rm eBC,bulk,AAC-AE33,200-720}$ and $m_{\rm rBC,bulk,DMA-SP2,200-720}$ were 0.7 \pm 192 0.4 μ g m⁻³ and 0.6 \pm 0.3 μ g m⁻³, indicating $m_{eBC,bulk,AAC-AE33,200-720}$ was overall 0.1 μ g m⁻³ higher than $m_{rBC,bulk,DMA-SP2,200-720}$, 193 194 consistent with b_0 of 0.1 µg m⁻³. The reason for overall discrepancy of 0.1 µg m⁻³ in $m_{BC,bulk,200-720}$ might be that DMA – SP2 could not detect rBC with D_c less than about 100 nm (Zhao et al., 2020a), resulting in an underestimated $m_{\rm rBC,bulk,DMA-SP2,200-720}$. 195 eBCMSD_{AAC-AE33} was also integrated from 720 nm to 1500 nm, which was the D_p range that DMA – SP2 did not measure, and the 196 197 result was denoted as m_{eBC,bulk,AAC-AE33,720-1500}. m_{eBC,bulk,AAC-AE33,720-1500} was correlated to m_{rBC,bulk,DMA-SP2,200-720} to study whether 198 BCMSD ranging from 720 nm to 1500 nm was connected to that ranging from 200 nm to 720 nm, as shown in Fig. S2b. R² decreased to 0.2 between $m_{\text{eBC,bulk,AAC-AE33,720-1500}}$ and $m_{\text{rBC,bulk,DMA-SP2,200-720}}$, indicating these two size ranges were not well related and both of 199 200 them needed to be measured independently. Observation by transmission electron microscopy showed that these larger BC-201 containing particles could be coated with massizve materials from secondary processes, or superaggregated BC with fractal BC 202 structure (Wang et al., 2022). The mean $m_{eBC,bulk,AAC-AE33,720-1500}$ was 0.2 \pm 0.2 μ g m⁻³, indicating that DMA – SP2 overall underestimated $m_{\rm BC,bulk}$ for ~ 0.2 µg m⁻³ (~ 33 %) in this field measurement considering that DMA – SP2 did not measure BCMSD 203 larger than 720 nm. Therefore, it is highly necessary to measure BCMSD with wider D_p range for better estimation of $m_{BC,bulk}$ and 204 205 deeper understanding of BC evolution in the atmosphere.

206 Figure 3 exhibited the mean eBCMSD_{AAC-AE33} (eBCMSD_{AAC-AE33}) and mean rBCMSD_{DMA-SP2} (rBCMSD_{DMA-SP2}) during the field measurement. It could be seen that when D_p was less than about 300 nm, $\overline{\text{eBCMSD}}_{AAC-AE33}$ was higher than 207 $\overline{\text{rBCMSD}}_{\text{DMA-SP2}}$. The higher $\overline{\text{eBCMSD}}_{\text{AAC-AE33}}$ may be due to particle diffusion at small D_p which was not well corrected by (7) 208 and underestimated MAC. When D_p was greater than 300 nm and less than 720 nm, $\overline{\text{eBCMSD}}_{AAC-AE33}$ was overall consistent with 209 $\overline{\text{rBCMSD}}_{\text{DMA-SP2}}$. When D_p was larger than 720 nm, where DMA – SP2 did not measure, $\overline{\text{eBCMSD}}_{\text{AAC-AE33}}$ decreased with 210 211 increasing D_p when D_p less than about 870 nm, and increased with increasing D_p when D_p was larger than 870 nm. In the study by Yu et al. (2010), three modes of BCMSD were identified: the mode peaked at about 400 nm, 1000 nm and 5000 nm, which were 212 213 termed as condensation mode, droplet mode and coarse mode, respectively. Following the nomenclature proposed by Yu et al. (2010), the increasing (decreasing) $\overline{\text{eBCMSD}}_{AAC-AE33}$ with increasing D_p for D_p larger (less) than 870 nm signified the lower (upper) end 214 215 of droplet mode (condensation mode). However, TBCMSD_{DMA-SP2} only identified condensation mode, which clearly highlighted 216 the necessity to measure BCMSD of wider D_p range for better characterization of BCMSD. Both the proposed method and DMA –

217 SP2 showed that the temporal variation of BCMSD, expressed as standard deviation (std) of BCMSD in Fig. 3, was as large as

218 BCMSD, reflecting the complex mechanism in the variability of BCMSD.

219 4 Uncertainty analysis

Uncertainty analysis was performed with respect to the MAC lookup table, transfer function inversion from $\sigma_{ab,binned}$ to $\sigma_{ab,size-resolved}$, f_{BC} and instrumental noise, respectively. The resulting uncertainty to eBCMSD_{AAC-AE33} was illustrated in Fig. 4 and to $m_{eBC,bulk,AAC-AE33,200-720}$ was shown in table 1. It could be seen from Fig. 4 that the boundary between condensation mode and droplet mode was distinct no matter how the uncertainty sources disturbed eBCMSD_{AAC-AE33}.

224 4.1 Uncertainty from masss absorption cross-section

225 According to Zhao et al. (2021b), the variation in refractive index (RI) dominated the uncertainty of the MAC lookup table. 226 Therefore, the uncertainty from the MAC lookup table was analyzed with respect to RI. The real part of RI (Re[RI]) was reported to vary from 1.5 to 2.0 in general (Liu et al., 2018) and the imaginary part of RI (Im[RI]) ranged from 0.5 to 1.1 commonly (Bond 227 and Bergstrom, 2006). Hence, Re[RI] (Im[RI]) was changed from 1.5 (0.5) to 2.0 (1.1) with step increase of 0.01, the resulting mean 228 MAC (\overline{MAC}) was the MAC lookup table used in this study (Fig. S3a) and the resulting std divided by the \overline{MAC} was considered as 229 230 the uncertainty of the MAC lookup table (Fig. S3b). As shown in Fig. S3b, when D_c was larger than about 400 nm, the uncertainty 231 was less than 10% and influenced by both D_p and D_c . However, when D_c was less than 400 nm, the uncertainty increased rapidly with decreasing D_c and was dominated by D_c . The uncertainty increased to larger than 23 % when D_c was less than about 100 nm. 232 When D_p was less than about 300 nm, the uncertainty varied from 14% to 24% with a mean value of 22 %, indicating large 233 234 uncertainty in eBCMSD_{AAC-AE33} for D_p less than 300 nm.

The MAC lookup table was replaced with original \overline{MAC} minus its std (-stdMAC) and plus its std (+stdMAC). The resulting MAC lookup tables were taken into the method proposed by Zhao et al. (2021b), and applied to the entire measurement period to study the influence of MAC variation on the eBCMSD_{AAC-AE33}. $\overline{eBCMSD}_{AAC-AE33}$ and mean eBCMSD_{AAC-AE33} after \pm stdMAC were shown in Fig. 4a. The uncertainty increased with decreasing D_p , and reached to maximum when D_p was less than 300 nm, indicating the eBCMSD_{AAC-AE33} for D_p less than 300 nm might potentially have nonnegligible uncertainty. The uncertainty in $m_{eBC,bulk,AAC-AE33,200-720}$ was from – 20 % (+stdMAC) to + 28 % (-stdMAC), which was the largest among the four uncertainty sources, as shown in Table 1.

242 **4.2 Uncertainty from the transfer function inversion**

As stated in Sect. 2.4.1, correction for particle loss and spectral broadening was required when $\sigma_{ab,binned}$ was converted to $\sigma_{ab,size-resolved}$. λ_{Ω} was defined to correct particle loss where $\lambda_{\Omega} = 0$ ($\lambda_{\Omega} = 1$) stood for total (no) particle loss. The relationships between λ_{Ω} and D_p as well as Q_{sample} , as shown in Fig. S4a, were well developed (Karlsson and Martinsson, 2003) and consistent with experimental data of AAC (Johnson et al., 2018). Q_{sample} used in this study was 3.0 L min⁻¹. Q_{sample} was changed from – 30 % (2.1 L min⁻¹) to + 30% (3.9 L min⁻¹), and the resulting λ_{Ω} was used to study the fluctuation of Q_{sample} on λ_{Ω} . As shown in Fig. S4a, the

248 variation of λ_{Ω} was less than 0.5 %, which was negligible.

Spectral broadening was caused by small-size particle diffusion as well as disturbance of sheath flow due to complicated rotation fluid dynamics inside AAC and described by μ_{Ω} where $\mu_{\Omega} < 1$ ($\mu_{\Omega} = 1$) was for (no) broadening. Johnson et al. (2018) found that behavior of μ_{Ω} with respect to D_p ($\mu_{\Omega}(D_p)$) depended on both Q_{sheath} and Q_{sample} , and parameterized $\mu_{\Omega}(D_p)$ (Fig. S4b) based on two case studies of ($Q_{\text{sheath}}, Q_{\text{sample}}$), namely (3 L min⁻¹, 0.3 L min⁻¹) and (15 L min⁻¹, 1.5 L min⁻¹):

253
$$\mu_{\Omega,\text{Johnson}}(D_p) = 0.318 \cdot D_p^{0.0946}.$$
 (15)

- $(Q_{\text{sheath}}, Q_{\text{sample}})$ was (7.5 L min⁻¹, 3 L min⁻¹) in this study, which might lead to uncertainty if $\mu_{\Omega,\text{Johnson}}(D_p)$ was directly used in this 254 study. $\mu_{\Omega,Johnson}(D_p)$ was varied from -23 % to +30 % to study the impact of μ_{Ω} on eBCMSD_{AAC-AE33}. The resulting influence on 255 256 the eBCMSD_{AAC-AE33} was shown in Fig. 4b. The uncertainty of μ_{Ω} did not exhibit a significant size dependence. Lower μ_{Ω} led to 257 lower eBCMSD_{AAC-AE33}, and vice versa. The difference in the $m_{eBC,bulk,AAC-AE33,200-720}$ was from -1% (-23% of $\mu_{\Omega,Johnson}$) to +21%(+ 30% of $\mu_{\Omega,Johnson}$), implying systematic overestimation of $m_{eBC,bulk,AAC-AE33,200-720}$. Therefore, $\mu_{\Omega}(D_p) = 77\% \mu_{\Omega,Johnson}(D_p)$ was used 258 in this study to offset the bias. The reason for the lower limit of μ_{Ω} set as -23 % rather than -30 % was that eBCMSD_{AAC-AE33} was 259 invalid mathematically when μ_{Ω} was less than - 23 % under setting of (O_{sheath} , O_{sample}) used in this study. It should be pointed out 260 that AAC – AE33 could still determine valid eBCMSD under the condition of $\mu_{\Omega}(D_p) < 77\%\mu_{\Omega,\text{Johnson}}(D_p)$ by increasing Q_{sheath} . 261 Q_{sheath} was not changed in this study because desired $\mu_{\Omega}(D_p)$ parameterization was found at $\mu_{\Omega}(D_p) = 77\% \mu_{\Omega,\text{Johnson}}(D_p)$. 262
- 263 $\sigma_{ab,size-resolved}$ (Fig. S5b) was integrated and the result, denoted as $\sigma_{ab,bulk,calc}$ was compared to $\sigma_{ab,bulk}$. As shown in Fig. S5a, 264 $\sigma_{ab,bulk,calc}$ was consistent with $\sigma_{ab,bulk}$. R^2 , b_1 and b_0 between $\sigma_{ab,bulk,calc}$ and $\sigma_{ab,bulk}$ was 0.9, 1.1, and 0.5 Mm⁻¹ (Fig. S6), respectively, 265 which validated conversion from $\sigma_{ab,binned}$ to $\sigma_{ab,size-resolved}$.

266 **4.3 Uncertainty from number fraction of BC-containing particle**

BC-containing aerosol particles had complicated mixing states. Even for internally-mixed particles with same $D_{\rm p}$, the internal BC 267 cores might have different D_c , which could not be resolved by AAC – AE33. Field measurement (Liu et al., 2010) revealed that f_{BC} 268 varied with time, D_c and D_p . This complexity was simplified to a parameterized fixed value of f_{BC} in this study. A fixed f_{BC} implied 269 270 that only one $D_{\rm c}$ value corresponded to a given $D_{\rm p}$ and the size-resolved number concentration of BC-containing particle was determined by Nsize-resolved times fBC. Zhao et al. (2021b) used fBC value of 0.17 based on SP2 measurement. However, SP2-derived 271 $f_{\rm BC}$ could not represent the bulk $f_{\rm BC}$ over the whole particle size spectrum due to the detection limit of SP2. And different regions 272 might have different f_{BC} . In this study, f_{BC} was varied and the resulting $m_{eBC,bulk,AAC-AE33,200-720}$ was compared with $m_{rBC,bulk,DMA-SP2,200-720}$ 273 720. f_{BC} was set as 0.35 in this study because b_1 was 1.0 between $m_{eBC,bulk,AAC-AE33,200-720}$ and $m_{rBC,bulk,DMA-SP2,200-720}$ when f_{BC} was 0.35. 274 275 f_{BC} was varied from 0.25 (- 30 % of 0.35) to 0.46 (+ 30 % of 0.35) to analyze the influence of f_{BC} on the eBCMSD_{AAC-AE33}, as shown in Fig. 4c. eBCMSD_{AAC-AE33} was more influenced around 870 nm. The variation of f_{BC} led to uncertainty of ± 3 % in 276 $m_{\rm eBC, bulk, AAC-AE33, 200-720}$, indicating that simplification of $f_{\rm BC}$ was a minor uncertainty source compared to MAC and transfer function 277 278 inversion.

279 **4.4 Uncertainty from instrumental noise**

280 The instrumental noise stemmed from the fluctuation of the light source and flow of AE33, which was reflected as fluctuation in

I, I_0 and *F*, further leading to the fluctuation in ATN, σ_{ATN} and σ_{ab} . The instrumental noise was defined as the std of $\sigma_{ab,binned}$ and was shown in Fig. S7b. It could be seen that the instrumental noise did not exhibit significant dependence on D_p . Comparing Fig. S7a and Fig. S7b, the instrumental noise was irrelevant to the value of $\sigma_{ab,binned}$. Figure S7c illustrated that the instrumental noise was also not correlated to $\sigma_{ab,bulk}$ with R^2 , b_1 and b_0 of 0.0, 0.0 and 0.1 Mm⁻¹, respectively, implying that the instrumental noise was not dependent on the pollution level.

The std of instrumental noise was added to (subtracted from) $\sigma_{ab,binned}$ and the derived eBCMSD_{AAC-AE33} was used to study the influence of instrumental noise on eBCMSD_{AAC-AE33}. The mean result was shown in Fig. 4d. eBCMSD_{AAC-AE33} larger than 1000 nm was more influenced by the instrumental noise than eBCMSD_{AAC-AE33} smaller than 500 nm. $\sigma_{ab,binned}$ larger than 1000 nm was relatively small (about 0.3 Mm⁻¹) compared to $\sigma_{ab,binned}$ less than 870 nm (about 0.9 Mm⁻¹). The mean instrumental noise was 0.1 Mm⁻¹ and exhibited no significant dependence on D_p . Therefore, $\sigma_{ab,binned}$ larger than 1000 nm was more affected by the instrumental noise, resulting in distinct variation in eBCMSD_{AAC-AE33}. The uncertainty in $m_{eBC,bulk,AAC-AE33,200-720}$ was not obvious, which varied from -2% to -1%, also minor compared to MAC and transfer function inversion.

293 5 Summary

294 Knowledge of bulk black carbon (BC) characteristics, such as bulk BC mass concentration ($m_{BC,bulk}$), is not sufficient for deeper 295 understanding the influence of BC on radiation and health. BC mass size distribution (BCMSD) was one of the BC microphysical 296 characteristics that could indicate emission source, reflect atmospheric aging and effectively reduce uncertainty related to BC 297 radiative effect. However, current BCMSD measurement ranging from upper micron to larger than 1 um was insufficient in time 298 resolution and sophisiticated for routine measurement. In this study, a new method to determine eBCMSD was proposed characterized by wide size range of measurement up to 1.5 µm, high time resolution up to 1 hour and convenience for extensive 299 300 measurement. The eBCMSD was retrieved by size-resolved absorption coefficient ($\sigma_{ab,size-resolved}$) measured by an aerodynamic aerosol classifer in tandem with an aethaelometer model AE33 (AAC - AE33), denoted as eBCMSD_{AAC-AE33}. Size-resolved number 301 302 concentration (Nsize-resolved) was measured concurrently by scanning mobility particle sizer (SMPS) and an aerodynamic particle sizer (APS) to model the influence of size on mass absorption cross-section (MAC). Uncertainty analysis was performed with respect to 303 MAC, transfer function inversion, number fraction of BC-containing particle (f_{BC}) and instrumental noise, which led to led to 24 %, 304 11 %, 3 % and 0.5 % variation on the derived $m_{\text{BC,bulk}}$, respectively. 305

The proposed method was applied to a field measurement in Yangtze River Delta and validated by comparing the rBCMSD with that measured by an differential mobility analyzer in tandem with a single-particle soot photometer (DMA – SP2), denoted as rBCMSD_{DMA-SP2}. The results showed that for particle diameter (D_p) less than 300 nm, eBCMSD_{AAC-AE33} was higher than rBCMSD_{DMA-SP2}, which might be caused by underestimated MAC by the method proposed by Zhao et al. (2021b) or incomplete parameterization of spectral broadening of AAC. eBCMSD_{AAC-AE33} was consistent with rBCMSD_{DMA-SP2} for D_p larger than 300 nm. $m_{BC,bulk}$ integrated over the size range that AAC – AE33 and DMA – SP2 both measured (200 nm – 720 nm), denoted as $m_{eBC,bulk,AAC-AE33}$ AE33,200-720 and $m_{rBC,bulk,DMA-SP2,200-720}$, compared well with each other with determination coefficient (R^2), slope (b_1), and intercept (b_0)

- of 0.8, 1.0 and 0.1 μ g m⁻³, respectively. However, DMA SP2 did not measure D_p larger than 700 nm, leading to 0.2 μ g m⁻³ (33 %) underestimation of $m_{BC,bulk}$, highlighting the necessity to measure BCMSD with a wider size range.
- As mentioned above, MAC was the largest uncertainty source, leading to significant uncertainty for D_p less than 300 nm. Transfer function inversion was the second largest uncertainty source, which was induced by incomplete parameterization of spectral broadening. The uncertainty in transfer function inversion led to systematic overestimation of $m_{eBC,bulk,AAC-AE33,200-720}$, which was corrected in this study. Both f_{BC} and instrumental noise were minor uncertainty sources and one order of magnitude less than MAC and transfer function inversion. f_{BC} was simplification of complicated BC mixing states, leading to relatively big uncertainty in eBCMSD at 870 nm, around the boundary between condensation mode and droplet mode. The eBCMSD for D_p larger than 1000 nm was more sensitive to instrumental noise.

322 Code/Data availability

323 The code and measurement data involved in this study are available upon request to the authors.

324 Author contribution

325 CZ determined the main goal of this study. WZ and GZ designed the methods. WZ carried them out and prepared the paper with 326 contributions from all co-authors.

327 Competing interests

328 The authors declare that they have no conflict of interest.

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438 Figure 1: Schematic diagram of the measurement (green) and the validation (blue) setup.

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Figure 2: Time series of (a) $m_{BC,bulk}$ from BCMSD integrated from 200 nm to 720 nm ($m_{BC,bulk,200-720}$) determined by the proposed method (red dot, denoted as "eBC, AAC – AE33") and DMA – SP2 (green cross, denoted as "rBC, DMA – SP2"), BCMSD determined by (b) the proposed method (eBC) and (c) DMA – SP2 (rBC). The red line in (b) marked particle size of 720 nm. The pollution episodes were shaded with orange in (a).



Figure 3: Mean and std of eBCMSD_{AAC-AE33} (green) and rBCMSD_{DMA-SP2} (red) during the measurement period. Mean
BCMSD was denoted by the solid line. Std of BCMSD was denoted by the dotted line.



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Figure 4: Uncertainty in eBCMSD_{AAC-AE33} arising from (a) MAC lookup table, (b) transfer function inversion, (c) f_{BC} and (d) instrumental noise. The solid lines in each panel were the same and were the mean eBCMSD_{AAC-AE33} during the measurement period. The dotted lines and dashed lines in each panel were mean eBCMSD_{AAC-AE33} ± standard deviation (std) calculated from (a) MAC + std of MAC and MAC – std of MAC, (b) $\mu_{\Omega,Johnson}$ and 1.3 times $\mu_{\Omega,Johnson}$, (c) f_{BC} of 35% times 0.7 and 1.3, (d) eBCMSD_{AAC-AE33} + std of instrumental noise and – std of instrumental noise.

457 Table 1: The Uncertainty in the *m*_{eBC,bulk,AAC-AE33,200-720} resulted from MAC lookup table, transfer function inversion, *f*_{BC} and

458 instrumental noise.

Uncertainty source	MAC	Transfer function inversion	$f_{ m BC}$	Instrumental noise
Uncertainty	-20% ~ + 28 %	- 1 % ~ + 21 %	- 3 % ~ + 3 %	- 2 % ~ - 1 %