

1 **Realtime measurement of phase partitioning of organic**  
2 **compounds using a Proton-Transfer-Reaction Time-of-Flight**  
3 **Mass Spectrometer coupled to a CHARON inlet**

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**Abstract.** Understanding the gas-particle partitioning of semivolatile organic compounds (SVOCs) is of crucial importance in the accurate representation of the global budget of atmospheric organic aerosols. In this study, we quantified the gas- vs. particle-phase fractions of a large number of SVOCs in real time in an urban area of East China with the use of a CHemical Analysis of aeRosols ONline (CHARON) inlet coupled to a high resolution Proton Transfer Reaction Time-of-Flight Mass Spectrometer (PTR-ToF-MS). We demonstrated the use of the CHARON inlet for highly efficient collection of particulate SVOCs while maintaining the intact molecular structures of these compounds. The collected month-long dataset with hourly resolution allows us to examine the gas-particle partitioning of a variety of SVOCs under ambient conditions. By comparing the measurements with model predictions using the instantaneous equilibrium partitioning theory, we found that the dissociation of large parent molecules during the PTR ionization process likely introduces large uncertainties to the measured gas- vs. particle-phase fractions of less oxidized SVOCs, and therefore, caution should be taken when linking the molecular composition to the particle volatility when interpreting the PTR-ToF-MS data. Our analysis suggests that understanding the fragmentation mechanism of SVOCs and accounting for the neutral losses of small moieties during the molecular feature extraction from the raw PTR mass spectra could reduce, to a large extent, the uncertainties associated with the gas-particle partitioning measurement of SVOCs in the ambient atmosphere.

## 1. Introduction

Gas-particle partitioning of semivolatile organic compounds (SVOCs) is a critical process involved in the formation and evolution of atmospheric organic aerosols (OA). Traditionally,

gas-particle equilibrium partitioning of organic substances is assumed to be established instantaneously (Zhang and Seinfeld, 2013), this assumption is in question if particles are semi-solid or glassy (Shiraiwa et al., 2013). Most studies to date addressing the kinetic limitations in partitioning have used indirect and/or theoretical methods that are lack of chemical and molecular specificity (Mai et al., 2015; Shiraiwa and Seinfeld, 2012). Direct measurements of gas-particle partitioning of SVOCs are needed in order to develop accurate parameterizations for the organic aerosol formation in climate models.

The major challenge in the characterization of gas-particle partitioning of SVOCs lies in the realtime measurement of labile compounds while maintaining their intact molecular structures with minimal fragmentation (Zhang et al., 2016a; Zhang et al., 2016b; Zhang et al., 2019). In recent years, soft ionization techniques coupled to mass spectrometry have been widely used for the measurement of gas-phase SVOCs at the molecular level (Veres et al., 2008; Crounse et al., 2006; Heald and Kroll, 2020). Combined with thermal desorption methods, these techniques have also been deployed to measure organic compounds in both gas and particle phases nearly simultaneously (Krechmer et al., 2016). A notable example would be the use of the Filter Inlet for Gases and AEROsols coupled with the Chemical Ionization Mass Spectrometry (FIGAERO-CIMS) to quantify the gas-particle partitioning of a broad range of organic compounds in real time (Lopez-Hilfiker et al., 2014; Ye et al., 2021; Voliotis et al., 2021; Wang et al., 2020a; Lutz et al., 2019; Lee et al., 2018; Le Breton et al., 2018; Stark et al., 2017; Lopez-Hilfiker et al., 2016; Lopez-Hilfiker et al., 2015; Palm et al., 2020). A number of studies among those have reached a consensus that the thermograms method, i.e., using the calibrated thermal desorption profiles vs. temperature to derive the volatility, likely provides the best estimates of the actual phase distribution. In contrast, using the directly measured gas- and

particle-phase fractions of a given analyte will most likely introduce a significant positive or negative bias to the volatility estimation due to the thermal decomposition of labile organic compounds during the desorption process (Lopez-Hilfiker et al., 2015; Stark et al., 2017). Such thermal decomposition (or ion fragmentation) artifacts, either positive or negative depending on the molecular size, have been suggested to constitute the largest uncertainties in the estimation of phase partitioning behaviors of SVOCs using the thermal desorption method at ambient pressure (Thompson et al., 2016).

Along with the line of thermal desorption method development, an inlet designed for the CHemical Analysis of aeRosols ONline (CHARON) has been developed and coupled to the Proton Transfer Reaction Time-of-Flight Mass Spectrometer (PTR-ToF-MS) in recent years. As CHARON-PTR-ToF-MS does not rely on any form of pre-concentration on surfaces, it could provide online and direct measurements of organic compounds in both phases, compared with traditional thermal desorption instruments which still need to address artifacts during the particle collection and desorption processes. Another potential advantage of CHARON-PTR-ToF-MS is that the chemical information of the collected particles can be studied qualitatively and quantitatively over a chemical composition level even at sub-nanogram mass concentrations per molecule owing to the well studied ion-molecule reaction chemistry in PTR-ToF-MS (Piel et al., 2019). CHARON has shown promising potential in the realtime analysis of the chemical composition and spatiotemporal distributions of aerosols with laboratory-, ground-, and aircraft-based platforms (Piel et al., 2019; Tan et al., 2018; Gkatzelis et al., 2018a; Gkatzelis et al., 2018b; Muller et al., 2017; Eichler et al., 2017; Eichler et al., 2015; Antonsen et al., 2017; Leglise et al., 2019; Piel et al., 2021). As a relatively new technique, the use of CHARON-PTR-ToF-MS to investigate the gas-particle partitioning of organic compounds is

87 still quite limited. Only one study by Gkatzelis et al. (2018b) deployed CHARON, together  
88 with two other aerosol sampling inlets, to measure the OA formation and aging from  
89 monoterpenes and real plant emissions in chamber experiments. Whether the CHARON inlet  
90 can be applied to the study of gas-particle partitioning of organic compounds under the actual  
91 atmospheric conditions remains to be validated.

92 In this study, we assess the applicability of the CHARON inlet to the time-resolved  
93 collection of organic compounds in their native molecular state using laboratory tests with a  
94 series of authentic standards. We further employ the CHARON inlet coupled to a high  
95 resolution PTR-ToF-MS instrument to measure an array of gaseous and particulate SVOCs in  
96 an urban area of East China. The obtained month-long hourly dataset allows us to examine the  
97 gas-particle partitioning of SVOCs spanning a range of volatilities. By comparing the  
98 measurements with model predictions using the instantaneous equilibrium partitioning theory,  
99 we found that fragmentation during the PTR ionization process may introduce large  
100 uncertainties to the measured gas- vs. particle-phase fractions of less oxidized SVOCs.  
101 Understanding the dissociation patterns of parent molecules and accounting for the  
102 fragmentation losses when extracting the molecular features from the raw PTR mass spectra  
103 are needed to improve the measured accuracy of SVOCs partitioning between the gas and  
104 particle phase.

## 106 **2. Material and Methods**

### 107 **2.1. Sampling site**

108 The sampling site located in the campus of the Shanghai Academy of Environmental  
109 Science (SAES) is representative of a typical urban setting surrounded by restaurants, shopping

110 malls, and residential and commercial buildings (Fig. S1). Two traffic-heavy streets in the area  
111 (Caobao Road and Humin Highway) are located  $\sim 150$  m and 450 m lateral distances to the east  
112 of the sampling site. A few number of petrochemical and chemical industrial facilities are  
113 located  $\sim 50$  km to the south and southwest of the observation site, which likely bring certain  
114 long-lived pollutants to the site at a typical wind speed of  $\sim 1 - 3$  m/s. Major pollution sources  
115 in the area include traffic, commercial and residential activities, and regional transport (Huang  
116 et al., 2021; Peng et al., 2023). The sampling inlet of the PTR-ToF-MS instrument was installed  
117 on the roof of an eight-story building  $\sim 24$  m above the ground. A comprehensive measurement  
118 of gas- and particle-phase compounds in the ambient air was performed from Oct 24 to Nov 22.  
119 During the sampling period, the average temperature, relative humidity, and wind speed were  
120  $18.0 \pm 3.0$  °C,  $61.0 \pm 15.0\%$ , and  $1.8 \pm 0.7$  m/s, respectively. The prevailing wind direction in  
121 this region was from the northwest and north during the polluted period (Fig. S2).

## 122 **2.2. CHARON-PTR-ToF-MS**

### 123 **2.2.1. Operation protocols**

124 A Proton Transfer Reaction Time-of-Flight Mass Spectrometer (PTR-ToF-MS) coupled to  
125 a CHEMical Analysis of aeRosols ONLINE (CHARON, Ionicon Analytik Inc, Innsbruck,  
126 Austria) inlet was employed to measure the gas- and particle-phase concentrations of a series  
127 of SVOCs. The PTR-ToF-MS instrument used here is equipped with a radio frequency (RF)-  
128 only quadrupole ion guide that transmits ions more efficiently (PTR-QiTOF, Ionicon Analytik  
129 Inc). The operating parameters of the PTR-ToF-MS were held constant during the entire  
130 measurement period. The drift tube pressure, temperature, and voltage were 2.9 mbar, 120 °C,  
131 and 500 V, respectively. These conditions correspond to an  $E/N$  ( $E$  is the electric field, and  $N$  is  
132 the number density of the gas molecules in the drift tube) value of  $\sim 100$  Td ( $1 \text{ Td} = 10^{-17} \text{ V}$

cm<sup>2</sup>) and a reaction time of 120  $\mu$ s. Note that the  $E/N$  value determines the collision energy of ions in the reactor and therefore the degree of fragmentation and cluster formation. The operating conditions were selected for the purpose of relatively low fragmentation intensities (compared to 120 – 140 Td) and limited production of water clusters (compared to 60 – 80 Td). Low  $E/N$  enhances the degree of water clustering, which complicates the analysis of analyte ions due to a complex interplay between cluster formation ( $RH^+(H_2O)_n$ ) and proton transfer reactions (Holzinger et al., 2019). During the campaign, the sensitivity of the PTR-ToF-MS was in the range of 300 – 1000 ncps ppb<sup>-1</sup> and the mass resolution was maintained at  $\sim 5000$  m/ $\Delta$ m. Mass spectra were collected at a time resolution of 10 s.

The CHARON inlet consists of 1) a gas-phase denuder (GPD) for stripping off gas-phase analytes, 2) an aerodynamic lens (ADL) for particle collimation which is combined with an inertial sampler for emanating the particle-enriched flow, and 3) a thermo-desorption unit (TDU) for particle volatilization. The CHARON inlet functionality has been described in great detail by Eichler et al. (2015). The inlet we used here had a particle enrichment factor of  $\sim 15$ , as discussed shortly. The vaporizer (TDU) was operated at 140 °C and  $\sim 8$  mbar absolute pressure. Measurements of organic compounds in the gas and particle phase were conducted using a parallel sampling system with two independent pumps, allowing for the selection of flow rates specifically adjusted for each phase, resulting in the overall residence time of less than 2 s (Fig. S3).

### **2.2.2. Sampling alternation between gas and particle phase**

Gas-phase compounds were measured by directly sampling the ambient air via a 2 m long perfluoro-alkoxy (PFA) tube (1/4" OD) capped with a polytetrafluoroethylene (PTFE) filter (Mitex<sup>TM</sup> PTFE membrane, 5  $\mu$ m pore size, 47 mm diameter) to prevent the clogging of

particles in the PTR capillaries. The gas-phase inlet was independently connected to the PTR-ToF-MS instrument upstream of the drift tube via a pressure-controlled subsampling PEEK capillary (1/16" OD). Zero measurements were performed by overflowing catalytically (platinum at 370°C) purified air through the inlet. Ambient particles were sampled through a stainless steel tube (3/8" OD) with a flow rate of  $\sim 3 \text{ L min}^{-1}$ , out of which a flow of  $\sim 500 \text{ ml min}^{-1}$  was directed to the CHARON inlet. A PM<sub>2.5</sub> cyclone was installed in front of the sampling line to remove coarse particles ( $> 2.5 \mu\text{m}$ ). The particle-phase background was measured by placing a High-Efficiency Particulate Air filter (HEPA, HEPA-CAP 7, GE Healthcare UK Limited, Buckinghamshire, UK) upstream of the CHARON inlet. Servo motor activated valves made of passivated stainless steel were used for switching between the two inlet configurations. During the campaign, CHARON-PTR-ToF-MS automatically switched between gas and particle phase every 15 min. Detailed setup is given in Fig. S3 in the Supporting Information.

The built-in PTR-manager software (Ionicon Analytik GmbH, Innsbruck, Austria) offers the possibility to program sequences by which the instrument switches between different settings. It takes  $\sim 1 \text{ min}$  for gases and particles to re-equilibrate when switching between these two modes. Data generated during this transition period ( $\sim 2 \text{ min}$ ) were not considered. Instrument background was measured for 15 min every 5h. The limits of detection (LoD) at 1 min resolution were in the range of  $5.6 \pm 2.9 \text{ ng m}^{-3}$  for gases and  $0.7 \pm 0.5 \text{ ng m}^{-3}$  for particles, respectively (Fig. S5). Concentrations of gaseous and particulate compounds shown here included the last 5 min of every gas/particle-phase working mode, in order to minimize the interferences carried over from the previous working mode by allowing for a sufficient amount of equilibration time in the inlet (Piel et al., 2021). In order to synchronize the gas- and particle-phase data to calculate gas-particle partitioning, the average hourly data were then used for



further analysis.

### 2.2.3. Sensitivity and calibration

Weekly calibrations were performed using a multicomponent calibration gas standard (Linde, USA) at five concentration levels from 0.5 to 10 ppb (Fig. S4a). The calibration mixture includes methanol, acetonitrile, acetaldehyde, acrolein, acetone, isoprene, methyl vinyl ketone, methyl ethyl ketone, 2-pentanone, toluene, styrene, p-xylene, 1,3,5-trimethylbenzene, naphthalene and  $\alpha$ -pinene. Here the sensitivity of PTR-ToF-MS is defined as the normalized ion intensity of  $\text{RH}^+$  (ncps) obtained at a mixing ratio of 1 ppb. For a given species ( $R$ ), its sensitivity ( $S$ , ncps  $\text{ppb}^{-1}$ ) is a linear function of the rate constant of its reaction with  $\text{H}_3\text{O}^+$  ( $k$ ) :

$$S = \frac{\frac{I_{\text{RH}^+}}{I_{\text{H}_3\text{O}^+}} \times 10^6}{\frac{[R]}{N} \times 10^9} = k \times N \times 10^{-3} \times t \times \frac{T_{\text{RH}^+}}{T_{\text{H}_3\text{O}^+}} \times F_{\text{RH}^+} \quad (1)$$

$$\text{corrected } S = \frac{S}{\frac{T_{\text{RH}^+}}{T_{\text{H}_3\text{O}^+}} \times F_{\text{RH}^+}} = a \times k \quad (2)$$

where the signals of  $\text{H}_3\text{O}^+$  ( $I_{\text{H}_3\text{O}^+}$ ) and  $\text{RH}^+$  ions ( $I_{\text{RH}^+}$ ) measured by the mass analyzer (in cps) can be related to the signals of  $\text{H}_3\text{O}^+$  ( $[\text{H}_3\text{O}^+]$ ) and  $\text{RH}^+$  ( $[\text{RH}^+]$ ) ions at the end of the drift tube, using their respective transmission efficiencies ( $T_{\text{H}_3\text{O}^+}$  and  $T_{\text{RH}^+}$ ) from the drift tube to the detector (De Gouw and Warneke, 2007).  $[R]$  is the concentration of species  $R$  and  $N$  is the number density of gas in the drift tube. The reaction time ( $t$ ) is determined by the ion drift velocity.  $F_{\text{RH}^+}$  represents the fraction of product ions detected as  $\text{RH}^+$  ions ( $0 \leq F_{\text{RH}^+} \leq 1$ ). For non-fragmenting compounds,  $F_{\text{RH}^+} = 1$ . The measured sensitivity is further corrected by accounting for fragmentation and transmission efficiency. The relative transmission efficiency of ions was derived from laboratory experiments (Fig. S4a).  $a$  is the slope of the linear regression of the corrected sensitivities on the proton-transfer-reaction rate coefficients ( $k$ ), as

shown in Fig. S4b. Following the method by Sekimoto et al. (2017), the linear regression result was used to determine the sensitivities of all uncalibrated species. The overall uncertainty was less than 15% for compounds with standards and around 50% for those without standards. Calculated sensitivity based on this method agrees well with measurements of authentic standards (Fig. S4c).

#### **2.2.4. Enrichment factor**

The CHARON inlet was calibrated routinely with pure ammonium nitrate particles to derive the enrichment factor as a function of the particle size following the procedures described in Eichler et al. (2015). In addition, we used a selection of authentic standards (Table S1) to test the effect of desorption temperature on the enrichment factor of labile compounds. Previous studies with CHARON generally used a temperature of 140 °C to vaporize particles (Leglise et al., 2019; Gkatzelis et al., 2018b; Tan et al., 2018). Herein, we tested the TDU temperature ranging from 70 to 140°C. The selected chemical standards were individually dissolved in distilled water (ethanol in the case of 2-Pentadecanone and 1-Pentadecanol) and nebulized by an atomizer (TSI 3076, Shore-view, MN, USA) that was pressurized with ultrapure zero air. The nebulizer outflow was diverted through two diffusion dryers to remove water vapor and an activated charcoal denuder (NovaCarb F, Mast Carbon International Ltd., Guilford, UK) to remove organic vapors. The resulting flow of polydisperse particles was then delivered into a differential mobility analyzer (DMA, TSI 3080) for particle size selection. The transmitted particles at a given size bin (300 nm for organics and 100 – 450 nm range for ammonium nitrate) were introduced into the CHARON-PTR-ToF-MS analyzer and a condensation particle counter (CPC, TSI 3775), respectively. Particle mass concentrations were calculated based on the CPC number distribution measurements by assuming a shape factor of 0.8 for ammonium nitrate

particles and 1 for organic particles, respectively.

The particle enrichment factor ( $EF$ ) of a given analyte  $i$  was calculated as the ratio of the PTR-ToF-MS derived vs. CPC derived mass concentrations of analyte  $i$  at a given particle size bin:

$$VMR_{(PTR)i} = \frac{I_i}{S_i} \quad (3)$$

$$VMR_{(CPC)i} = \rho_i \times V \times N_i \times V_m / Mw_i \quad (4)$$

$$EF = \frac{VMR_{(PTR)i}}{VMR_{(CPC)i}} \quad (5)$$

where  $I_i$  is the normalized signal of species  $i$  (ncps) by PTR-ToF-MS,  $S_i$  is the sensitivity (ncps ppb<sup>-1</sup>),  $VMR$  is the volume mixing ratio (ppb),  $\rho$  is the density of species  $i$  (g cm<sup>-3</sup>),  $V$  is the volume of a particle sphere (m<sup>3</sup>),  $N_i$  is the number concentration of particles measured by CPC (cm<sup>-3</sup>),  $V_m$  is the molar volume of an ideal gas at 1 atm (22.4 L mol<sup>-1</sup>),  $Mw_i$  is the molecular weight (g mol<sup>-1</sup>). As the calculated sensitivities of most organics in the absence of authentic standards are subject to uncertainties (15% – 50%), we will herein use the multiplication of  $EF$  and  $S_i$  to evaluate the combined effect of CHARON enrichment and sensitivity on the measured concentrations of a given analyte  $i$  in the particle phase.

### 2.2.5. Data processing

Data were analyzed using the Tofware package (v3.2.0, Tofwerk Inc), within the Igor Pro software (v7.0, Wavemetrics). Using this package, time-dependent mass calibrations were performed using four ions (H<sub>3</sub><sup>18</sup>O<sup>+</sup>, NO<sup>+</sup>, C<sub>6</sub>H<sub>5</sub>I<sup>+</sup> and C<sub>6</sub>H<sub>5</sub>I<sub>2</sub><sup>+</sup>), where C<sub>6</sub>H<sub>5</sub>I<sup>+</sup> and C<sub>6</sub>H<sub>5</sub>I<sub>2</sub><sup>+</sup> were produced from the internal standard di-iodobenzene. The relative mass deviation was within 6 – 8 ppm across the mass spectra. Considering the humidity dependence of reagent ions (H<sub>3</sub>O<sup>+</sup> and H<sub>3</sub>O<sup>+</sup>(H<sub>2</sub>O)), the fitted product ion signals (RH<sup>+</sup>) were normalized to a standard reagent ion

of  $10^6$  cps (counts per second). Elemental composition was determined based on the accurate  $m/z$  (mass to charge ratio) and isotopic pattern analysis. A list of  $\sim 1600$  ions was extracted, including both gas- and particle-phase ions. Molecular formulas including only C, H, and O atoms were assigned to the detected ions by the addition of one proton in cases where the elemental composition analysis returned multiple options. About 85% of the signals were elementally resolved by the  $C_xH_yO_z$  formula in ambient air mass spectra. A small number of nitrogen containing compounds, such as nitroaromatics, were also identified but not included in the following analysis. Throughout of the context, we use the word “species” to refer to all compounds with assigned molecular formula, which may include multiple isomers.

### **2.3. Complementary measurements**

In addition to CHARON-PTR-ToF-MS, a Thermal desorption Aerosol Gas chromatograph (TAG) was also employed to measure a series of particle-phase organic species. Details of the TAG operation and data analysis protocols can be found in previous studies (He et al., 2020; Wang et al., 2020b; Zhu et al., 2021). The elemental composition and mass concentration of particles were measured by an Aerodyne high-resolution time-of-flight Aerosol Mass Spectrometer (AMS), with details of operation and quality control protocols given by our recent study (Huang et al., 2021). Volatile organic compounds (VOCs,  $C_2 - C_{12}$ ) were analyzed by a custom-built online gas chromatography system equipped with a mass spectrometer and a flame ionization detector (GC-MS/FID). The performance of this system can be found in our previous publications (Zhu et al., 2018; Wang et al., 2014). Meteorological parameters (ambient temperature, wind speed, wind direction, and relative humidity) were collected by an automatic weather station (Metone 590 series) mounted on the roof top of the campaign site.

## 2.4. Gas-particle partitioning measurements vs. modeling

The CHARON-PTR-ToF-MS measured gas- and particle-phase concentrations of a given species  $i$  can be used to calculate its particle-phase fraction ( $F_{p,i}$ ).

$$P_i = \frac{I_{p,i} \times \left(\frac{m}{z_i} - 1\right)}{V_m \times S_i \times EF} \quad (6)$$

$$G_i = \frac{I_{g,i} \times \left(\frac{m}{z_i} - 1\right)}{V_m \times S_i} \quad (7)$$

$$F_{p,i} = \frac{P_i}{P_i + G_i} \quad (8)$$

where  $P_i$  and  $G_i$  are the mass concentrations ( $\text{ng m}^{-3}$ ) of species  $i$  in the particle and gas phase, respectively.  $I_{p,i}$  and  $I_{g,i}$  are the normalized signal (ncps) of the PTR-ToF-MS detected ion  $i$  in the particle and gas phase, respectively.  $V_m$  is taken as 22.4 L/mol.  $S_i$  is calculated or measured sensitivity (ncps ppb<sup>-1</sup>), see details in Section 2.2.3. As structural isomers cannot be resolved in the mass spectra, the calculation here assumes that all isomers with the same molecular formula have the same chemical properties, i.e., saturation vapor pressures. Substitution of Equations (6) and (7) to Equation (8) yields the final expression of the particle-phase fraction of species  $i$  ( $F_{p,i}$ ), which is a function of the observed PTR-MS raw signals of species  $i$  in the gas and particle phase (in total ion counts), as well as the particle enrichment factor ( $EF$ ) of species  $i$ .

$$F_{p,i} = \frac{I_{p,i}/EF}{I_{p,i}/EF + I_{g,i}} \quad (9)$$

Gas-particle partitioning of a given analyte  $i$  was also modeled using the equilibrium partitioning theory (Pankow, 1994):

$$F_{p,i} = \frac{1}{1 + C_i^* / C_{OA}} \quad (10)$$

$$C_i^* = \frac{10^6 M_{w_i} \zeta_i p_i}{RT} \quad (11)$$

where  $C_{OA}$  is the organic aerosol concentration measured by AMS ( $\mu\text{g m}^{-3}$ ),  $C_i^*$  is the saturation mass concentration ( $\mu\text{g m}^{-3}$ ),  $M_{w_i}$  is the molecular weight ( $\text{g mol}^{-1}$ ),  $\zeta$  is the activity coefficient (assumed as unity),  $p_i$  is the pure component liquid vapor pressure (Pa),  $R$  is the universal gas constant ( $8.2 \times 10^{-5} \text{ m}^3 \text{ atm K}^{-1} \text{ mol}^{-1}$ ), and  $T$  is the ambient temperature (K). As detailed chemical information is lacking for all species detected by PTR-ToF-MS, here we use the expression given by Donahue et al. (2011) to approximate the value of  $C_i^*$ :

$$\log_{10} C_i^* = (n_C^0 - n_C^i) b_C - n_O^i b_O - 2 \frac{n_C^i n_O^i}{n_C^i + n_O^i} b_{CO} \quad (12)$$

where  $n_C^0 = 25$ ,  $b_C = 0.475$ ,  $b_O = 2.3$ , and  $b_{CO} = -0.3$ .

## 2.5. Uncertainties in the measured and modeled gas-particle partitioning

The uncertainty associated with the PTR-MS measured concentrations in both gas and particle phases is less than 15% for compounds with chemical standards based on the optimally fitted transmission efficiency curve. For those in the absence of standards, their PTR sensitivities were calculated theoretically using Equations (1-2), and the uncertainty in the calculation mainly arises from the estimation of polarizability and dipole moment of the target molecule, which has been estimated to be within  $\sim 50\%$  when only the elemental composition of that molecule is given (Sekimoto et al., 2017). It is important to note, however, that the uncertainty associated with the estimated PTR sensitivity has zero influence on the measured particle-phase fraction of any given compound because the sensitivity term is essentially canceled in the divisor function in Equation (9). The uncertainty associated with the particle fraction of a given species  $i$  derived from the PTR-MS measurements arises predominantly

from the “ $EF$ ” term. Since the uncertainty of the measured  $EF$  depends on the uncertainty of  $I_{p,i}$ , we thus express the overall uncertainties of the measured gas-particle partitioning as:

$$Unc(F_{p,i}) = \sqrt{Unc(EF)^2 + Unc(I_{g,i})^2} \quad (13)$$

with the calibration standards used in this study, the enrichment factor is calculated to be within 25% error ( $Unc(EF)$ ) (see detailed calculations listed in Table S2 and S3), including the effect of wall loss inside the inlet tubing and the precision of the measurement. Huang et al. (2019) has tested the uncertainty of wall loss ( $Unc(I_{g,i})$ ) in this PTR-MS instrument as 28%. Therefore, the overall uncertainty in the measured particle-phase fraction ( $Unc(F_{p,i})$ ) was estimated as 38%.

The uncertainty associated with the modeled gas-particle partitioning arises primarily from the uncertainty in the estimation of the saturation mass concentration ( $C_i^*$ ) based on the method developed by Donahue et al. (2011). In this method, the saturation mass concentration of species  $i$  is a non-linear function of the numbers of carbon and oxygen atoms in that particular species, see Equation (12). For each generic molecular formula, i.e.,  $C_xH_y$ ,  $C_xH_yO$ ,  $C_xH_yO_2$ , and  $C_xH_yO_4$ , Donahue et al. (2011) have used a total of 25, 48, 18, and 10 chemical standards with known volatilities to validate the estimated saturation concentrations, and the estimated errors were taken as 34%, 16%, 25%, and 54%, respectively (see Table S6). As the  $C_xH_yO_3$  group was not tested, we tentatively assumed the associated errors as the same as the  $C_xH_yO_4$  group. The extent to which these uncertainties may affect the difference between measurements and model results was discussed in detail in the Supporting Information (Fig. S10).

### 3. Results and Discussion

#### 3.1. Particle enrichment: effect of desorption temperature

Thermal desorption as a common procedure used in the chemical characterization of organic aerosols is often susceptible to fragmentation of non-refractory compounds. Due to the high temperature used to evaporate particles collected, labile and large molecules are inevitably subject to fragmentation, thereby introducing large uncertainties to the measured mass and composition of the particulate organic compounds (Lopez-Hilfiker et al., 2015; Yatavelli et al., 2012; Zhao et al., 2013). Thermal decomposition of oxidized organic compounds has been observed at vaporizer temperature as low as 200 °C, the lowest temperature required to vaporize OA as reported (Stark et al., 2017). While decreasing the vaporizer temperature is necessary to maintain the intact structure of labile molecules, low temperature (e.g., 85 °C), however, might fail to completely evaporate the collected particles into vapors, resulting in an underestimation of the collected OA mass (Inomata et al., 2014). Here we performed a series of sensitivity tests to identify the optimal vaporizer temperature in the CHARON inlet for the measurements of organic compounds in the particle phase.

Prior to the temperature sensitivity test, we have validated that the particle enrichment factor, also known as collection efficiency and defined as the ratio of the particle mass concentration upstream to downstream of the aerodynamic lens, does not depend on the particle size. As shown in Fig. 1a, the measured *EF* value for ammonium nitrate particles, detected as  $\text{NO}_2^+$  produced from the nitric acid vapor, remains constant as  $\sim 15$  in the 150 – 450 nm particle size range. The lower values in the 100 – 150 nm size range can be explained by the lower particle transmission efficiency in the gas phase denuder, e.g., 75% – 80% for 100 nm particles (Eichler et al., 2015). Also, particles below 150 nm are less efficiently concentrated in the



subsampling flow after the aerodynamic lens. Therefore, we used the monodisperse particles generated from selected organic standards at 300 nm for the temperature sensitivity test.

A number of chemical standards that are representative of alcohols, carbonyls, and carboxylic acids and with the vapor pressure ranging from  $10^{-14}$  to  $10^{-1}$  Pa at 25 °C (taken from EPA EPI Suite (2012), see values given in Table S1) were used to generate organic aerosols, which, upon size selection at 300 nm, were directed to the CHARON inlet. Particle evaporation occurs downstream of the aerodynamic lens in the gas phase and on the tube and orifice surface to which submicron particles rapidly diffuse at  $\sim 8$  mbar operating pressure. The thermal desorption unit was designed to ensure that ammonium sulfate particles ( $10^{-20}$  Pa) can be completely evaporated (Piel et al., 2019; Eichler et al., 2015). As the desorption temperature was varied from 70 °C to 140 °C, the intensities of all detected ions (including both parent and fragment ions) for each organic standard analyzed were stable within 15%, as shown in Fig. 1b. Also note that we did not observe any ions produced from decarboxylation and/or dehydration during the particle evaporation process. This is because the relative low operation temperature and the short heat exposure time could effectively limit any thermal dissociation of organic molecules. This demonstrates that the parent molecule fragmentation, if any, does not occur under the range of desorption temperature used in the CHARON inlet, but rather results from the ionic dissociation process in the PTR ionization chamber, see more discussions in Section 3.3. We therefore used the sum total of intensities of all major ions detected as the PTR-ToF-MS response to a given organic standard analyzed. Fig. 1c shows that the derived enrichment factors stay constant for all compounds investigated ( $M_w \sim 160 - 230$  g/mol). The relative signals of all fragment ions were stable over the range of the desorption temperature as shown in Fig. 2. This suggests that the desorption temperature used here, even as low as 70 °C, is

sufficient to evaporate SVOCs (volatility  $> 10^{-14}$  Pa at 25 °C), due to the low operating pressure ( $\sim 8$  mbar) that significantly enhances the partitioning shift to the gas phase. One low volatility compound, sucrose ( $Mw$  is 342 g/mol and vapor pressure is  $4.69 \times 10^{-14}$  Pa), has a slightly lower enhancement factor compared with all the other organic standards tested. This is mainly due to the intensive dehydration of the parent compound in the ionization chamber, and as a result, only a few fragment ions were captured, resulting in a lower PTR response and thereby lower  $EF$  value calculated from Equations (3-5). The  $EF$  values of sucrose also have much higher standard deviations at all temperatures due to fragmentation (Table S3).

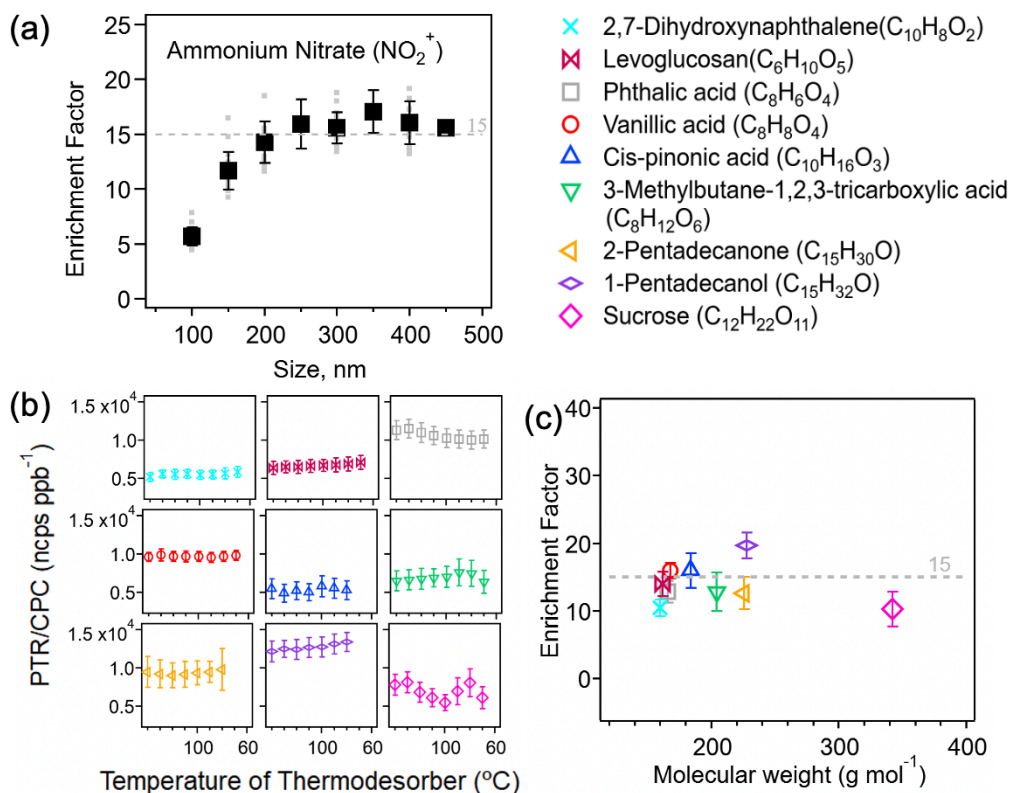


Figure 1. (a) Measured unitless enrichment factor ( $EF$ ) of ammonium nitrate particles as a

function of particle size in the 100 – 450 nm range. Grey markers represent all replicating measurements. The error bar denotes one standard deviation ( $1\sigma$ ) of the average. (b) Ratios of PTR-ToF-MS signals (including both parent and fragment ions) to CPC counts ( $\pm 1\sigma$ ) at 300 nm for all organic standards studied. (c)  $EF$  ( $\pm 1\sigma$ ) of selected organic standards based on the calculated sensitivity.

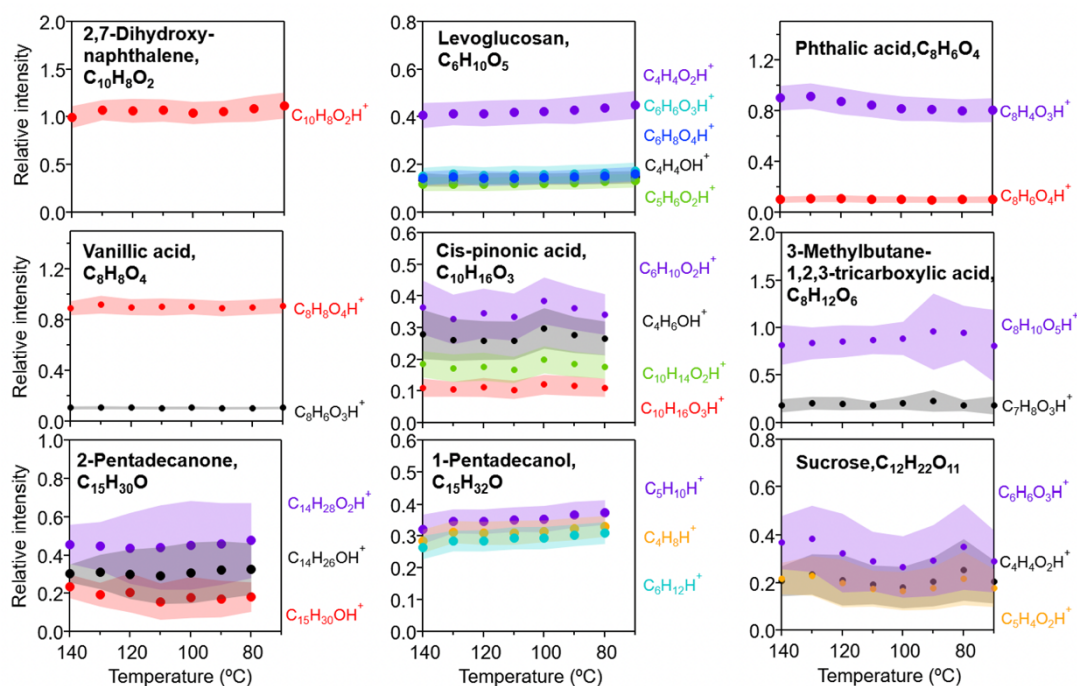


Figure 2. Ratios of CHARON-PTR-ToF-MS signals (ncps) to CPC measurements (ppb) of all detected ions (including both parents and fragments) for a given pure organic standard tested under different desorption temperatures (70 – 140 °C), normalized to the corresponding ratios obtained at 140 °C. Ions with relative intensities less than 10% are excluded. Red markers represent the parent peaks. Colored shades represent the relative standard deviations at different

temperatures (exact values are given in Table S4).

It has been recognized that species with one functional group follow certain fragmentation patterns during the PTR ionization process (Pagonis et al., 2019; Francis et al., 2007; Spanel et al., 1997; Tani et al., 2003; Spanel and Smith, 1997), such as dehydration of acids and alcohols. The observed dissociation of carboxylic acid standards used in this study, e.g., phthalic acid and 3-methylbutane-1,2,3-tricarboxylic acid, can be explained by this common fragmentation pattern. The fragmentation mechanism of multi-functionalized species is rather complicated and a number of fragments can be produced upon PTR ionization. Nevertheless, the identity and abundance of fragments from a given multi-functionalized species have been found comparable under the same PTR operation protocols (Leglise et al., 2019; Gkatzelis et al., 2018a). For example, *cis*-pinonic acid yields the following fragments (main ions only and relative abundance in parentheses):  $m/z$  71.049 (~28%), 115.075 (~36%), 167.108 (~19%), and 185.117 (~11%), which is comparable with an earlier study (Leglise et al., 2019):  $m/z$  71.049 (~27%), 115.075 (~33%), 167.108 (~26%), and 185.117 (~14%) at 100 Td settings.

### 3.2. Molecular features of detected organic species

A month-long field dataset of particle- and gas-phase organic species was collected at hourly resolution using the CHARON-PTR-ToF-MS instrument. A comparison of PTR-ToF-MS measurements with other techniques available on site was performed for both gas and particle phases. For the gas phase, quantitative measurements of a suite of VOCs by GC-MS/FID, including benzene, toluene, styrene, C<sub>8</sub> and C<sub>9</sub> aromatics, acrolein, and C<sub>4</sub>, C<sub>5</sub>, and C<sub>6</sub> ketones, agree well with corresponding PTR-ToF-MS measurements, as shown in Fig. S6. For the particle phase, the time series of a group of C<sub>x</sub>H<sub>y</sub>O<sub>4</sub> species (including C<sub>4</sub>H<sub>6</sub>O<sub>4</sub>, C<sub>5</sub>H<sub>8</sub>O<sub>4</sub>,

C<sub>6</sub>H<sub>10</sub>O<sub>4</sub>, and C<sub>8</sub>H<sub>6</sub>O<sub>4</sub>) are in reasonable agreement with corresponding measurements taken by TAG ( $r \sim 0.60 - 0.80$ ), although the CHARON-PTR-ToF-MS measured total molecular mass is generally lower than the TAG measurements by a factor of 2 to 6 (Fig. S7). This is likely caused by the fragmentation (e.g., loss of H<sub>2</sub>O, see Fig. S8) of the parent compounds during the ionization process, as discussed in detail in Section 3.3. The time series of total OA mass characterized by CHARON-PTR-ToF-MS also agree with the AMS measurements ( $r \sim 0.91$ , Fig. S9). Previous studies have reported the particulate organics measured by PTR-MS with a thermal desorption inlet account for 25% – 60% mass of the total organic aerosols measured by AMS (Holzinger et al., 2013). Direct comparison of the total OA mass loading is not applicable here since the CHARON-PTR-ToF-MS measurement only focused on compounds with the mass to charge ratio below 300 Th. That the majority of ions detected by PTR are present in the lower mass range is primarily due to the fragmentation of larger masses during the ionization process, as discussed extensively in Section 3.3.

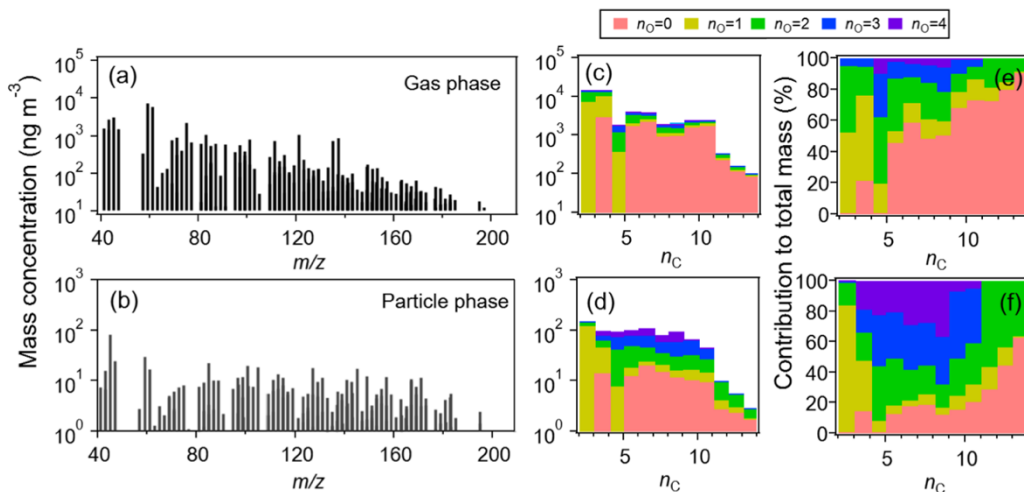


Figure 3. Background subtracted monthly-average PTR-ToF-MS mass spectra in the (a) gas phase and (b) particle phase. Mass distributions of all identified species resolved by the carbon

and oxygen numbers ( $n_C$  and  $n_O$ ) in the (c) gas phase and (d) particle phase, as well as their relative contribution to the total organic mass in the (e) gas phase and (f) particle phase.

Fig. 3 (a-b) shows the PTR-ToF-MS spectra of dominant ions averaged over the entire campaign in both gas and particle phases. The mass concentrations of individual ions are in the range of  $7.9 - 7179.3 \text{ ng m}^{-3}$  in the gas phase and  $0.6 - 82.7 \text{ ng m}^{-3}$  in the particle phase. A total of 152 species (with  $> 60\%$  data points above the PTR-ToF-MS detection limits) are identified, contributing to  $\sim 69\%$  and  $\sim 44\%$ , respectively, of the total organic mass measured in the gas and particle phase. The molecular distribution characterized by the carbon and oxygen of these species is given in Fig. 3 (c-d). The most abundant species are characterized by a generic formula of  $C_xH_yO$  and  $C_xH_yO_2$ , resolving  $\sim 64\%$  and  $\sim 46\%$  in total of all identified species in the gas and particle phase, respectively. Another dominant component in the gas phase is hydrocarbon-like compounds ( $C_xH_y$ ) ( $\sim 27\%$ ), which contribute  $\sim 12\%$  of the organic mass in the particle phase. Species with higher oxygen numbers ( $> 2$ ) contribute to a large fraction ( $\sim 42\%$ ) of the total particulate mass. These  $C_xH_yO_{1-4}$  groups exhibit different diurnal cycles, as shown in Fig. 4, reflecting their unique formation chemistry. The  $C_xH_y$  group peaks in the early morning rush hour and likely originates from primary traffic emissions. On the contrary, both  $C_xH_yO_3$  and  $C_xH_yO_4$  groups peak at noon, suggesting a strong secondary formation source. The diurnal trends for  $C_xH_yO$  and  $C_xH_yO_2$  groups are relatively flat during the day, likely indicative of an intertwined primary emission and secondary formation processes.

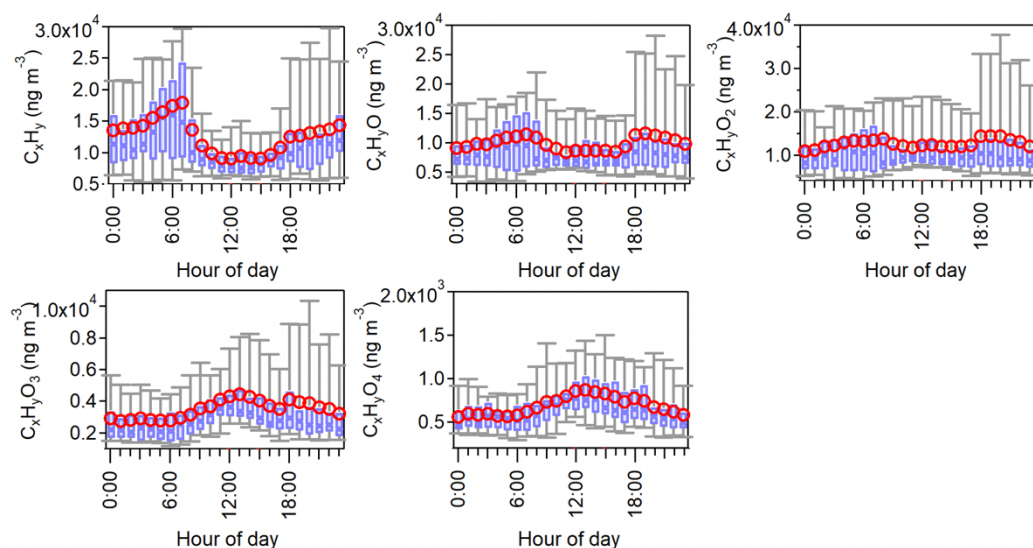


Figure 4. Diurnal variations of observed gas-phase species with a generic formula of  $C_{2-13}H_{2-22}O_{0-4}$ . Hourly average values ( $\text{ng m}^{-3}$ ), together with 10<sup>th</sup>, 25<sup>th</sup>, 75<sup>th</sup>, and 90<sup>th</sup> percentiles, are also plotted.

### 3.3. Measured vs. modeled gas-particle partitioning

Fig. 5 shows the calculated particle-phase fraction ( $F_p$ ) of the identified 152 species using the CHARON-PTR-ToF-MS measurements in both phases. It is important to note that an authentic standard is not required for the calculation of  $F_p$  for any given species, because the PTR sensitivity term is essentially canceled in the divisor function in Equations (6-8). Also given in Fig. 5 is the simulated  $F_p$  of the derived molecular formulas of all species identified using the equilibrium partitioning theory, see method described in Section 2.4. Interestingly, for oxidized species such as  $C_xH_yO_4$ , their measured  $F_p$  agree reasonably with the simulations, as shown in Fig. 5a. The  $C_xH_yO_4$  group resides in the relatively higher mass range, and species

identified in this group are more likely actual compounds rather than fragments from larger parent molecules. Even these species dissociate into lower mass ions during PTR ionization, their calculated particle-phase fractions are unaffected by such fragmentation processes because signals of parent ions decrease by the same extent upon fragmentation in both gas and particle phases. As the oxygen number decreases, the measured  $F_p$  values tend to deviate from the simulations by up to several orders of magnitude. Note that these less oxidized compounds (e.g.,  $C_xH_y$ ) are mostly small molecules and they are highly unlikely present in the condensed phase as closed-shell monomers (Pankow and Asher, 2008; Holzinger et al., 2010). Instead, they are more likely fragments produced from the decomposition of larger molecules, which no surprisingly favor partitioning in the particle phase. The discrepancy in the measurement-model comparison underscores the importance of understanding the fragmentation mechanism during PTR ionization when extracting molecular features from the raw mass spectra.

Parent ion fragmentation has been widely observed in PTR-MS instruments (Pagonis et al., 2019). Oxygenates exhibit trends in neutral losses of water or saturated alcohols. Here, we apply a correction to the molecular formula of the 152 identified species by assuming that these species are fragments produced from their parent precursors through the neutral losses of a carboxyl group ( $-CO_2$ ), a carbonyl group ( $-CO$ ), a hydroxyl group ( $-H_2O$ ), or an alcohol group ( $-C_2H_6O$ ). By applying this correction, the modeled  $F_p$  of a given species  $C_xH_yO_z$  would actually represent the particle-phase fraction of the its parent species  $C_xH_yO_z \cdot CO_2$ ,  $C_xH_yO_z \cdot CO$ ,  $C_xH_yO_z \cdot H_2O$ , or  $C_xH_yO_z \cdot C_2H_6O$ . As shown in Fig. 5 (b-e), such a correction could significantly increase the modeled  $F_p$  values by several orders of magnitude. The assumption of neutral losses of  $CO_2$  or  $C_2H_6O$  allows for much improved agreement between modeled vs. measured  $F_p$  values for less oxidized species. This implies that these small and less oxidized species are



likely fragments resulting from the decomposition of larger parent precursors. As our particle enrichment test (details given in Section 3.1) has confirmed that the thermal desorption temperature employed for particle evaporation does not lead to any intensive fragmentation, therefore the collision-induced dissociation during the proton transfer reaction process becomes the predominant process that produces fragments (Lindinger et al., 1998; Gueneron et al., 2015; Gkatzelis et al., 2018b). Although the electric field applied to the drift tube is considered low to moderate compared with most previous PTR-MS measurements ( $E/N \sim 100$  Td in this study vs.  $E/N \sim 120 - 140$  Td commonly found in PTR-MS measurements) (Pagonis et al., 2019), parent ion fragmentation was still widely observed here and complicated the mass spectra interpretation and molecular feature extraction. While some recent CHARON measurements employed lower electric field in the drift tube ( $E/N \sim 60$  Td) (Leglise et al., 2019; Gkatzelis et al., 2018a; Piel et al., 2019), such conditions could promote the formation of water cluster ions, which increase with humidity and reduce the PTR sensitivity, and therefore are not ideally suitable for our field measurements. A compromise solution would be using a moderate electric field in the drift tube (e.g.,  $\sim 100$  Td) and meanwhile applying appropriate molecular corrections to all ions detected in the mass spectra by considering possible neutral losses of small moieties. Since in this study only the information of molecular formula is derived from the PTR-MS spectra, we thus provide the lower and upper bound of the gas-particle partitioning corrections owing to neutral losses of  $H_2O$  and  $CO_2$ , respectively. In general, lower masses with higher volatilities are subject to notable changes in the particle-phase fraction as a result of neutral losses during the PTR ionization process, see detailed discussions in Text S1.

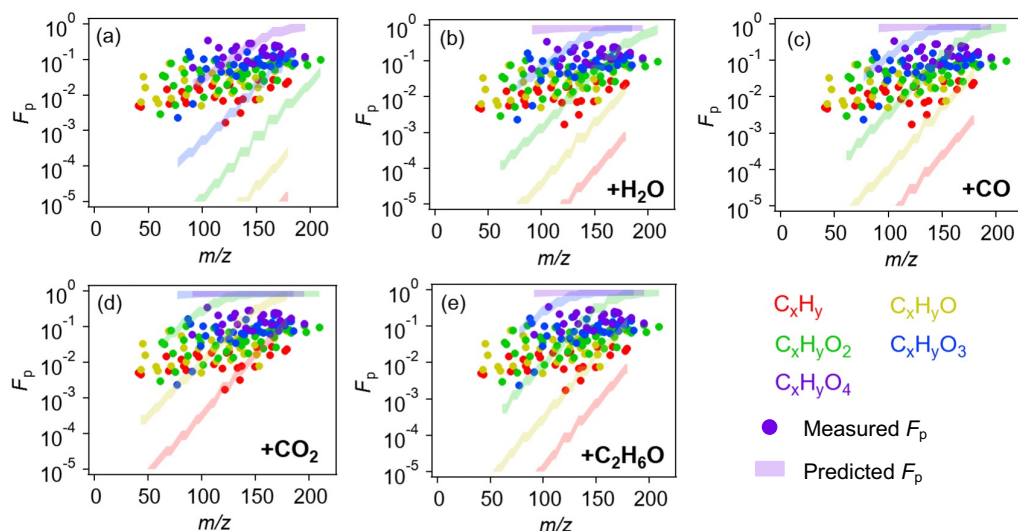


Figure 5. (a) Campaign average fraction of organic species in the particle phase ( $F_p$ ) grouped by the oxygen number. Solid markers represent the calculated  $F_p$  based on the CHARON-PTR-ToF-MS measurements. Colored shades represent the predicted  $F_p$  of corresponding molecular formulas. (b-e) Measured vs. predicted  $F_p$  assuming the identified species are fragments of corresponding parent compounds through neutral losses of  $H_2O$ ,  $CO$ ,  $CO_2$ , and  $C_2H_6O$ , respectively (values are given in Table S5).

#### 4. Conclusions

Recent studies have suggested that some of the model-measurement discrepancies in the representation of ambient organic aerosol budget might be due to the nonequilibrium gas/particle partitioning caused by kinetic limitations in the presence of glassy or semi-solid phase (Perraud et al., 2012; Mai et al., 2015; Shiraiwa et al., 2013). It is therefore necessary to validate whether the equilibrium partitioning theory could adequately describe the condensation of semivolatile organic vapors onto atmospheric aerosols under ambient

conditions, and the accurate measurement of these SVOCs in both gas and particle phases is the crucial prerequisite. In this study, we have employed the PTR-ToF-MS instrument coupled to a CHARON inlet, together with a suite of complementary measurements, to characterize the atmospheric partitioning behaviors of an array of SVOCs in an urban environment of East China. Prior to the application to the field measurements, we first performed a series of laboratory experiments to test whether the CHARON inlet is capable of sampling organic molecules (including alcohols, carbonyls, and carboxylic acids) in their native states. With the low pressure condition used in the CHARON inlet, a thermal desorption temperature less than 140 °C could adequately evaporate organic compounds with vapor pressure higher than  $10^{-14}$  Pa while minimizing the thermal decomposition of labile functionalities. The auto-switch function between the gas and particle mode with one single PTR-ToF-MS instrument could monitor gaseous and particulate organic compounds in real time, thereby providing important information of their partitioning behaviors in the ambient atmosphere. Particle-phase fractions of a total of 152 organic species were derived from the CHARON-PTR-ToF-MS measurements and further compared with model predictions using the instantaneous equilibrium partitioning theory. While the model captured the particle-phase fraction of oxidized compounds (e.g.,  $C_xH_yO_{3-4}$ ), predictions of less oxidized compounds, notably the  $C_xH_y$  family, differ from the corresponding measurements by several orders of magnitude. Such a large discrepancy is very likely caused by the intensive fragmentation of the parent organic compounds during the PTR ionization process. Accounting for common fragmentation patterns in the simulations of gas-particle partitioning, for example, neutral losses of  $-CO_2$ ,  $-CO$ ,  $-H_2O$ , or  $-C_2H_6O$ , could largely improve the model-measurement agreement. Such corrections are very necessary towards an accurate measurement of both particle- and gas-phase SVOCs using the CHARON-PTR-ToF-

MS instrument. Our study suggests the crucial importance of optimizing operation conditions and understanding the fragmentation mechanism in the particle collection, vaporization, and ionization processes in understanding the gas-particle partitioning of organic compounds using any thermal desorption based aerosol measurement method.

*Data availability.* The data shown in the paper are available upon request from the corresponding author.

*Author Contributions.* YP carried out experiments and measurements and drafted the manuscript. HW and XZ designed the experimental studies, supervised the laboratory work and wrote the manuscript. YG and SJ supported the ambient measurements. SZ, DH, PH, and SL supported the data analysis. TC and CH supervised the scientific work. All authors have given approval to the final version of the manuscript.

*Competing interests.* The authors declare no competing financial interest.

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## References

Antonsen, S., Bunkan, A. J. C., D'Anna, B., Eichler, P., Farren, N., Hallquist, M., Hamilton, J. F., Kvarnliden, H., Mikoviny, T., Muller, M., Nielsen, C. J., Stenstrom, Y., Tan, W., Wisthaler, A., and Zhu, L.: Atmospheric chemistry of tert-butylamine and AMP, in: Enrgy Procecd, edited by: Dixon, T., Laloui, L., and Twinning, S., Energy Procedia, Elsevier Science Bv, Amsterdam, 1026-1032, 10.1016/j.egypro.2017.03.1248, 2017.

Crounse, J. D., McKinney, K. A., Kwan, A. J., and Wennberg, P. O.: Measurement of gas-phase

580 hydroperoxides by chemical ionization mass spectrometry, *Anal. Chem.*, 78, 6726-6732,  
581 10.1021/ac0604235, 2006.

582 de Gouw, J. and Warneke, C.: Measurements of volatile organic compounds in the earth's  
583 atmosphere using proton-transfer-reaction mass spectrometry, *Mass Spectrometry Reviews*, 26,  
584 223-257, 10.1002/mas.20119, 2007.

585 Donahue, N. M., Epstein, S. A., Pandis, S. N., and Robinson, A. L.: A two-dimensional  
586 volatility basis set: 1. organic-aerosol mixing thermodynamics, *Atmos. Chem. Phys.*, 11, 3303-  
587 3318, 10.5194/acp-11-3303-2011, 2011.

588 Eichler, P., Muller, M., D'Anna, B., and Wisthaler, A.: A novel inlet system for online chemical  
589 analysis of semi-volatile submicron particulate matter, *Atmos. Meas. Tech.*, 8, 1353-1360,  
590 10.5194/amt-8-1353-2015, 2015.

591 Eichler, P., Muller, M., Rohmann, C., Stengel, B., Orasche, J., Zimmermann, R., and Wisthaler,  
592 A.: Lubricating Oil as a Major Constituent of Ship Exhaust Particles, *Environ. Sci. Technol.*  
593 *Lett.*, 4, 54-58, 10.1021/acs.estlett.6b00488, 2017.

594 EPA: Estimation Program Interface (EPI) Suite (v4.11), US [code], 2012.

595 Francis, G. J., Milligan, D. B., and McEwan, M. J.: Gas-Phase Reactions and Rearrangements  
596 of Alkyl Esters with  $\text{H}_3\text{O}^+$ ,  $\text{NO}^+$ , and  $\text{O}_2^+\bullet$ : A Selected Ion Flow Tube Study, *J. Phys. Chem.*  
597 *A*, 111, 9670-9679, 10.1021/jp0731304, 2007.

598 Gkatzelis, G. I., Tillmann, R., Hohaus, T., Muller, M., Eichler, P., Xu, K. M., Schlag, P., Schmitt,  
599 S. H., Wegener, R., Kaminski, M., Holzinger, R., Wisthaler, A., and Kiendler-Scharr, A.:  
600 Comparison of three aerosol chemical characterization techniques utilizing PTR-ToF-MS: a  
601 study on freshly formed and aged biogenic SOA, *Atmos. Meas. Tech.*, 11, 1481-1500, 2018a.

602 Gkatzelis, G. I., Hohaus, T., Tillmann, R., Gensch, I., Muller, M., Eichler, P., Xu, K. M., Schlag,  
603 P., Schmitt, S. H., Yu, Z. J., Wegener, R., Kaminski, M., Holzinger, R., Wisthaler, A., and  
604 Kiendler-Scharr, A.: Gas-to-particle partitioning of major biogenic oxidation products: a study  
605 on freshly formed and aged biogenic SOA, *Atmos. Chem. Phys.*, 18, 12969-12989,  
606 10.5194/acp-18-12969-2018, 2018b.

607 Gueneron, M., Erickson, M. H., VanderSchelden, G. S., and Jobson, B. T.: PTR-MS  
608 fragmentation patterns of gasoline hydrocarbons, *International Journal of Mass Spectrometry*,

609 379, 97-109, 10.1016/j.ijms.2015.01.001, 2015.

610 He, X., Wang, Q., Huang, X. H. H., Huang, D. D., Zhou, M., Qiao, L., Zhu, S., Ma, Y.-g., Wang,  
 611 H.-l., Li, L., Huang, C., Xu, W., Worsnop, D. R., Goldstein, A. H., and Yu, J. Z.: Hourly  
 612 measurements of organic molecular markers in urban Shanghai, China: Observation of  
 613 enhanced formation of secondary organic aerosol during particulate matter episodic periods,  
 614 *Atmos. Environ.*, 240, 10.1016/j.atmosenv.2020.117807, 2020.

615 Heald, C. L. and Kroll, J. H.: The fuel of atmospheric chemistry: Toward a complete description  
 616 of reactive organic carbon, *Sci. Adv.*, 6, 10.1126/sciadv.aay8967, 2020.

617 Holzinger, R., Goldstein, A. H., Hayes, P. L., Jimenez, J. L., and Timkovsky, J.: Chemical  
 618 evolution of organic aerosol in Los Angeles during the CalNex 2010 study, *Atmos. Chem. Phys.*,  
 619 13, 10125-10141, 10.5194/acp-13-10125-2013, 2013.

620 Holzinger, R., Williams, J., Herrmann, F., Lelieveld, J., Donahue, N. M., and Rockmann, T.:  
 621 Aerosol analysis using a Thermal-Desorption Proton-Transfer-Reaction Mass Spectrometer  
 622 (TD-PTR-MS): a new approach to study processing of organic aerosols, *Atmos. Chem. Phys.*,  
 623 10, 2257-2267, 10.5194/acp-10-2257-2010, 2010.

624 Holzinger, R., Acton, W. J. F., Bloss, W. J., Breitenlechner, M., Crilley, L. R., Dusanter, S.,  
 625 Gonin, M., Gros, V., Keutsch, F. N., Kiendler-Scharr, A., Kramer, L. J., Krechmer, J. E.,  
 626 Languille, B., Locoge, N., Lopez-Hilfiker, F., Materić, D., Moreno, S., Nemitz, E., Quéléver,  
 627 L. L. J., Sarda Esteve, R., Sauvage, S., Schallhart, S., Sommariva, R., Tillmann, R., Wedel, S.,  
 628 Worton, D. R., Xu, K., and Zaytsev, A.: Validity and limitations of simple reaction kinetics to  
 629 calculate concentrations of organic compounds from ion counts in PTR-MS, *Atmos. Meas.*  
 630 *Tech.*, 12, 6193-6208, 10.5194/amt-12-6193-2019, 2019.

631 Huang, D. D., Zhu, S., An, J., Wang, Q., Qiao, L., Zhou, M., He, X., Ma, Y., Sun, Y., Huang,  
 632 C., Yu, J. Z., and Zhang, Q.: Comparative Assessment of Cooking Emission Contributions to  
 633 Urban Organic Aerosol Using Online Molecular Tracers and Aerosol Mass Spectrometry  
 634 Measurements, *Environ. Sci. Technol.*, 10.1021/acs.est.1c03280, 2021.

635 Huang, G., Liu, Y., Shao, M., Li, Y., Chen, Q., Zheng, Y., Wu, Z., Liu, Y., Wu, Y., Hu, M., Li,  
 636 X., Lu, S., Wang, C., Liu, J., Zheng, M., and Zhu, T.: Potentially Important Contribution of  
 637 Gas-Phase Oxidation of Naphthalene and Methyl-naphthalene to Secondary Organic Aerosol  
 638 during Haze Events in Beijing, *Environ. Sci. Technol.*, 53, 1235-1244, 10.1021/acs.est.8b04523,

639 2019.

640 Inomata, S., Sato, K., Hirokawa, J., Sakamoto, Y., Tanimoto, H., Okumura, M., Tohno, S., and  
 641 Imamura, T.: Analysis of secondary organic aerosols from ozonolysis of isoprene by proton  
 642 transfer reaction mass spectrometry, *Atmos. Environ.*, 97, 397-405,  
 643 10.1016/j.atmosenv.2014.03.045, 2014.

644 Krechmer, J. E., Groessl, M., Zhang, X., Junninen, H., Massoli, P., Lambe, A. T., Kimmel, J.  
 645 R., Cubison, M. J., Graf, S., Lin, Y.-H., Budisulistiorini, S. H., Zhang, H., Surratt, J. D.,  
 646 Knochenmuss, R., Jayne, J. T., Worsnop, D. R., Jimenez, J.-L., and Canagaratna, M. R.: Ion  
 647 mobility spectrometry–mass spectrometry (IMS–MS) for on- and offline analysis of  
 648 atmospheric gas and aerosol species, *Atmos. Meas. Tech.*, 9, 3245-3262, 10.5194/amt-9-3245-  
 649 2016, 2016.

650 Le Breton, M., Wang, Y. J., Hallquist, A. M., Pathak, R. K., Zheng, J., Yang, Y. D., Shang, D.  
 651 J., Glasius, M., Bannan, T. J., Liu, Q. Y., Chan, C. K., Percival, C. J., Zhu, W. F., Lou, S. R.,  
 652 Topping, D., Wang, Y. C., Yu, J. Z., Lu, K. D., Guo, S., Hu, M., and Hallquist, M.: Online gas-  
 653 and particle-phase measurements of organosulfates, organosulfonates and nitrooxy  
 654 organosulfates in Beijing utilizing a FIGAERO ToF-CIMS, *Atmos. Chem. Phys.*, 18, 10355-  
 655 10371, 2018.

656 Lee, B. H., Lopez-Hilfiker, F. D., D'Ambro, E. L., Zhou, P., Boy, M., Petäjä, T., Hao, L.,  
 657 Virtanen, A., and Thornton, J. A.: Semi-volatile and highly oxygenated gaseous and particulate  
 658 organic compounds observed above a boreal forest canopy, *Atmos. Chem. Phys.*, 18, 11547-  
 659 11562, 10.5194/acp-18-11547-2018, 2018.

660 Leglise, J., Muller, M., Piel, F., Otto, T., and Wisthaler, A.: Bulk Organic Aerosol Analysis by  
 661 Proton-Transfer-Reaction Mass Spectrometry: An Improved Methodology for the  
 662 Determination of Total Organic Mass, O:C and H:C Elemental Ratios, and the Average  
 663 Molecular Formula, *Anal. Chem.*, 91, 12619-12624, 10.1021/acs.analchem.9b02949, 2019.

664 Lindinger, W., Hansel, A., and Jordan, A.: Proton-transfer-reaction mass spectrometry (PTR-  
 665 MS): on-line monitoring of volatile organic compounds at pptv levels, *Chem. Soc. Rev.*, 27,  
 666 347-354, 10.1039/a827347z, 1998.

667 Lopez-Hilfiker, F. D., Mohr, C., Ehn, M., Rubach, F., Kleist, E., Wildt, J., Mentel, T. F., Lutz,  
 668 A., Hallquist, M., Worsnop, D., and Thornton, J. A.: A novel method for online analysis of gas

669 and particle composition: description and evaluation of a Filter Inlet for Gases and AEROsols  
670 (FIGAERO), *Atmos. Meas. Tech.*, 7, 983-1001, 10.5194/amt-7-983-2014, 2014.

671 Lopez-Hilfiker, F. D., Mohr, C., Ehn, M., Rubach, F., Kleist, E., Wildt, J., Mentel, T. F.,  
672 Carrasquillo, A. J., Daumit, K. E., Hunter, J. F., Kroll, J. H., Worsnop, D. R., and Thornton, J.  
673 A.: Phase partitioning and volatility of secondary organic aerosol components formed from  $\alpha$ -  
674 pinene ozonolysis and OH oxidation: the importance of accretion products and other low  
675 volatility compounds, *Atmos. Chem. Phys.*, 15, 7765-7776, 10.5194/acp-15-7765-2015, 2015.

676 Lopez-Hilfiker, F. D., Mohr, C., D'Ambro, E. L., Lutz, A., Riedel, T. P., Gaston, C. J., Iyer, S.,  
677 Zhang, Z., Gold, A., Surratt, J. D., Lee, B. H., Kurten, T., Hu, W. W., Jimenez, J., Hallquist, M.,  
678 and Thornton, J. A.: Molecular Composition and Volatility of Organic Aerosol in the  
679 Southeastern U.S.: Implications for IEPOX Derived SOA, *Environ. Sci. Technol.*, 50, 2200-  
680 2209, 10.1021/acs.est.5b04769, 2016.

681 Lutz, A., Mohr, C., Le Breton, M., Lopez-Hilfiker, F. D., Priestley, M., Thornton, J. A., and  
682 Hallquist, M.: Gas to Particle Partitioning of Organic Acids in the Boreal Atmosphere, *ACS*  
683 *Earth Space Chem.*, 3, 1279-1287, 10.1021/acsearthspacechem.9b00041, 2019.

684 Mai, H., Shiraiwa, M., Flagan, R. C., and Seinfeld, J. H.: Under What Conditions Can  
685 Equilibrium Gas-Particle Partitioning Be Expected to Hold in the Atmosphere?, *Environ. Sci.*  
686 *Technol.*, 49, 11485-11491, 10.1021/acs.est.5b02587, 2015.

687 Muller, M., Eichler, P., D'Anna, B., Tan, W., and Wisthaler, A.: Direct Sampling and Analysis  
688 of Atmospheric Particulate Organic Matter by Proton-Transfer-Reaction Mass Spectrometry,  
689 *Anal. Chem.*, 89, 10889-10897, 10.1021/acs.analchem.7b02582, 2017.

690 Pagonis, D., Sekimoto, K., and de Gouw, J.: A Library of Proton-Transfer Reactions of  $\text{H}_3\text{O}^+$   
691 Ions Used for Trace Gas Detection, *Journal of the American Society for Mass Spectrometry*, 30,  
692 1330-1335, 10.1007/s13361-019-02209-3, 2019.

693 Palm, B. B., Peng, Q. Y., Fredrickson, C. D., Lee, B., Garofalo, L. A., Pothier, M. A.,  
694 Kreidenweis, S. M., Farmer, D. K., Pokhrel, R. P., Shen, Y. J., Murphy, S. M., Permar, W., Hu,  
695 L., Campos, T. L., Hall, S. R., Ullmann, K., Zhang, X., Flocke, F., Fischer, E. V., and Thornton,  
696 J. A.: Quantification of organic aerosol and brown carbon evolution in fresh wildfire plumes,  
697 *Proc Natl Acad Sci U S A*, 117, 29469-29477, 10.1073/pnas.2012218117, 2020.

698 Pankow, J. F.: An absorption-model of the gas aerosol partitioning involved in the formation of



699 secondary organic aerosol, *Atmos. Environ.*, 28, 189-193, 10.1016/1352-2310(94)90094-9,  
700 1994.

701 Pankow, J. F. and Asher, W. E.: SIMPOL.1: a simple group contribution method for predicting  
702 vapor pressures and enthalpies of vaporization of multifunctional organic compounds, *Atmos.*  
703 *Chem. Phys.*, 8, 2773-2796, 10.5194/acp-8-2773-2008, 2008.

704 Peng, Y., Wang, H., Wang, Q., Jing, S., An, J., Gao, Y., Huang, C., Yan, R., Dai, H., Cheng, T.,  
705 Zhang, Q., Li, M., Hu, J., Shi, Z., Li, L., Lou, S., Tao, S., Hu, Q., Lu, J., and Chen, C.:  
706 Observation-based sources evolution of non-methane hydrocarbons (NMHCs) in a megacity of  
707 China, *J. Environ. Sci.*, 124, 794-805, 10.1016/j.jes.2022.01.040, 2023.

708 Perraud, V., Bruns, E. A., Ezell, M. J., Johnson, S. N., Yu, Y., Alexander, M. L., Zelenyuk, A.,  
709 Imre, D., Chang, W. L., Dabdub, D., Pankow, J. F., and Finlayson-Pitts, B. J.: Nonequilibrium  
710 atmospheric secondary organic aerosol formation and growth, *Proc Natl Acad Sci U S A*, 109,  
711 2836-2841, 10.1073/pnas.1119909109, 2012.

712 Piel, F., Muller, M., Mikoviny, T., Pusede, S. E., and Wisthaler, A.: Airborne measurements of  
713 particulate organic matter by proton-transfer-reaction mass spectrometry (PTR-MS): a pilot  
714 study, *Atmos. Meas. Tech.*, 12, 5947-5958, 10.5194/amt-12-5947-2019, 2019.

715 Piel, F., Müller, M., Winkler, K., Skytte af Sætra, J., and Wisthaler, A.: Introducing the extended  
716 volatility range proton-transfer-reaction mass spectrometer (EVR PTR-MS), *Atmos. Meas.*  
717 *Tech.*, 14, 1355-1363, 10.5194/amt-14-1355-2021, 2021.

718 Sekimoto, K., Li, S. M., Yuan, B., Koss, A., Coggon, M., Warneke, C., and de Gouw, J.:  
719 Calculation of the sensitivity of proton-transfer-reaction mass spectrometry (PTR-MS) for  
720 organic trace gases using molecular properties, *International Journal of Mass Spectrometry*, 421,  
721 71-94, 10.1016/j.ijms.2017.04.006, 2017.

722 Shiraiwa, M. and Seinfeld, J. H.: Equilibration timescale of atmospheric secondary organic  
723 aerosol partitioning, *Geophysical Research Letters*, 39, n/a-n/a, 10.1029/2012gl054008, 2012.

724 Shiraiwa, M., Zuend, A., Bertram, A. K., and Seinfeld, J. H.: Gas-particle partitioning of  
725 atmospheric aerosols: interplay of physical state, non-ideal mixing and morphology, *Phys.*  
726 *Chem. Chem. Phys.*, 15, 11441-11453, 10.1039/c3cp51595h, 2013.

727 Spanel, P. and Smith, D.: SIFT studies of the reactions of  $\text{H}_3\text{O}^+$ ,  $\text{NO}^+$  and  $\text{O}_2^+$  with a series

of alcohols, *Int. J. Mass Spectrom. Ion Processes*, 167-168, 375-388,  
[https://doi.org/10.1016/S0168-1176\(97\)00085-2](https://doi.org/10.1016/S0168-1176(97)00085-2), 1997.

Spanel, P., Ji, Y. F., and Smith, D.: SIFT studies of the reactions of  $\text{H}_3\text{O}^+$ ,  $\text{NO}^+$  and  $\text{O}_2^+$  with a series of aldehydes and ketones, *International Journal of Mass Spectrometry*, 165, 25-37, 10.1016/s0168-1176(97)00166-3, 1997.

Stark, H., Yatavelli, R. L. N., Thompson, S. L., Kang, H., Krechmer, J. E., Kimmel, J. R., Palm, B. B., Hu, W. W., Hayes, P. L., Day, D. A., Campuzano-Jost, P., Canagaratna, M. R., Jayne, J. T., Worsnop, D. R., and Jimenez, J. L.: Impact of Thermal Decomposition on Thermal Desorption Instruments: Advantage of Thermogram Analysis for Quantifying Volatility Distributions of Organic Species, *Environ. Sci. Technol.*, 51, 8491-8500, 10.1021/acs.est.7b00160, 2017.

Tan, W., Zhu, L., Mikoviny, T., Nielsen, C. J., Wisthaler, A., Eichler, P., Muller, M., D'Anna, B., Farren, N. J., Hamilton, J. F., Pettersson, J. B. C., Hallquist, M., Antonsen, S., and Stenstrom, Y.: Theoretical and Experimental Study on the Reaction of tert-Butylamine with OH Radicals in the Atmosphere, *J. Phys. Chem. A*, 122, 4470-4480, 10.1021/acs.jpca.8b01862, 2018.

Tani, A., Hayward, S., and Hewitta, C. N.: Measurement of monoterpenes and related compounds by proton transfer reaction-mass spectrometry (PTR-MS), *International Journal of Mass Spectrometry*, 223, 561-578, 10.1016/s1387-3806(02)00880-1, 2003.

Thompson, S. L., Yatavelli, R. L. N., Stark, H., Kimmel, J. R., Krechmer, J. E., Day, D. A., Hu, W., Isaacman-VanWertz, G., Yee, L., Goldstein, A. H., Khan, M. A. H., Holzinger, R., Kreisberg, N., Lopez-Hilfiker, F. D., Mohr, C., Thornton, J. A., Jayne, J. T., Canagaratna, M., Worsnop, D. R., and Jimenez, J. L.: Field intercomparison of the gas/particle partitioning of oxygenated organics during the Southern Oxidant and Aerosol Study (SOAS) in 2013, *Aerosol Sci. Technol.*, 51, 30-56, 10.1080/02786826.2016.1254719, 2016.

Veres, P., Roberts, J. M., Warneke, C., Welsh-Bon, D., Zahniser, M., Herndon, S., Fall, R., and de Gouw, J.: Development of negative-ion proton-transfer chemical-ionization mass spectrometry (NI-PT-CIMS) for the measurement of gas-phase organic acids in the atmosphere, *International Journal of Mass Spectrometry*, 274, 48-55, 10.1016/j.ijms.2008.04.032, 2008.

Voliotis, A., Wang, Y., Shao, Y., Du, M., Bannan, T. J., Percival, C. J., Pandis, S. N., Alfarra, M. R., and McFiggans, G.: Exploring the composition and volatility of secondary organic aerosols

758 in mixed anthropogenic and biogenic precursor systems, *Atmos. Chem. Phys.*, 21, 14251-14273,  
 759 10.5194/acp-21-14251-2021, 2021.

760 Wang, H. L., Lou, S. R., Huang, C., Qiao, L. P., Tang, X. B., Chen, C. H., Zeng, L. M., Wang,  
 761 Q., Zhou, M., Lu, S. H., and Yu, X. N.: Source Profiles of Volatile Organic Compounds from  
 762 Biomass Burning in Yangtze River Delta, China, *Aerosol Air Qual. Res.*, 14, 818-828,  
 763 10.4209/aaqr.2013.05.0174, 2014.

764 Wang, M. Y., Chen, D. X., Xiao, M., Ye, Q., Stolzenburg, D., Hofbauer, V., Ye, P. L., Vogel, A.  
 765 L., Mauldin, R. L., Amorim, A., Baccharini, A., Baumgartner, B., Brilke, S., Dada, L., Dias, A.,  
 766 Duplissy, J., Finkenzeller, H., Garmash, O., He, X. C., Hoyle, C. R., Kim, C., Kvashnin, A.,  
 767 Lehtipalo, K., Fischer, L., Molteni, U., Petaja, T., Pospisilova, V., Quelever, L. L. J., Rissanen,  
 768 M., Simon, M., Tauber, C., Tome, A., Wagner, A. C., Weitz, L., Volkamer, R., Winkler, P. M.,  
 769 Kirkby, J., Worsnop, D. R., Kulmala, M., Baltensperger, U., Dommen, J., El Haddad, I., and  
 770 Donahue, N. M.: Photo-oxidation of Aromatic Hydrocarbons Produces Low-Volatility Organic  
 771 Compounds, *Environ. Sci. Technol.*, 54, 7911-7921, 2020a.

772 Wang, Q. Q., He, X., Zhou, M., Huang, D. D., Qiao, L. P., Zhu, S. H., Ma, Y. G., Wang, H. L.,  
 773 Li, L., Huang, C., Huang, X. H. H., Xu, W., Worsnop, D., Goldstein, A. H., Guo, H., and Yu, J.  
 774 Z.: Hourly Measurements of Organic Molecular Markers in Urban Shanghai, China: Primary  
 775 Organic Aerosol Source Identification and Observation of Cooking Aerosol Aging, *ACS Earth*  
 776 *Space Chem.*, 4, 1670-1685, 2020b.

777 Yatavelli, R. L. N., Lopez-Hilfiker, F., Wargo, J. D., Kimmel, J. R., Cubison, M. J., Bertram, T.  
 778 H., Jimenez, J. L., Gonin, M., Worsnop, D. R., and Thornton, J. A.: A Chemical Ionization  
 779 High-Resolution Time-of-Flight Mass Spectrometer Coupled to a Micro Orifice Volatilization  
 780 Impactor (MOVI-HRToF-CIMS) for Analysis of Gas and Particle-Phase Organic Species,  
 781 *Aerosol Sci. Technol.*, 46, 1313-1327, 2012.

782 Ye, C., Yuan, B., Lin, Y., Wang, Z., Hu, W., Li, T., Chen, W., Wu, C., Wang, C., Huang, S., Qi,  
 783 J., Wang, B., Wang, C., Song, W., Wang, X., Zheng, E., Krechmer, J. E., Ye, P., Zhang, Z., Wang,  
 784 X., Worsnop, D. R., and Shao, M.: Chemical characterization of oxygenated organic  
 785 compounds in the gas phase and particle phase using iodide CIMS with FIGAERO in urban air,  
 786 *Atmos. Chem. Phys.*, 21, 8455-8478, 10.5194/acp-21-8455-2021, 2021.

787 Zhang, X. and Seinfeld, J. H.: A functional group oxidation model (FGOM) for SOA formation  
 788 and aging, *Atmos. Chem. Phys.*, 13, 5907-5926, 10.5194/acp-13-5907-2013, 2013.

Zhang, X., Dalleska, N. F., Huang, D. D., Bates, K. H., Sorooshian, A., Flagan, R. C., and Seinfeld, J. H.: Time-resolved molecular characterization of organic aerosols by PILS plus UPLC/ESI-Q-TOFMS, *Atmos. Environ.*, 130, 180-189, 10.1016/j.atmosenv.2015.08.049, 2016a.

Zhang, X., Zhang, H. F., Xu, W., Wu, X. K., Tyndall, G. S., Orlando, J. J., Jayne, J. T., Worsnop, D. R., and Canagaratna, M. R.: Molecular characterization of alkyl nitrates in atmospheric aerosols by ion mobility mass spectrometry, *Atmos. Meas. Tech.*, 12, 5535-5545, 10.5194/amt-12-5535-2019, 2019.

Zhang, X., Krechmer, J. E., Groessl, M., Xu, W., Graf, S., Cubison, M., Jayne, J. T., Jimenez, J. L., Worsnop, D. R., and Canagaratna, M. R.: A novel framework for molecular characterization of atmospherically relevant organic compounds based on collision cross section and mass-to-charge ratio, *Atmos. Chem. Phys.*, 16, 12945-12959, 10.5194/acp-16-12945-2016, 2016b.

Zhao, Y. L., Kreisberg, N. M., Worton, D. R., Isaacman, G., Weber, R. J., Liu, S., Day, D. A., Russell, L. M., Markovic, M. Z., VandenBoer, T. C., Murphy, J. G., Hering, S. V., and Goldstein, A. H.: Insights into Secondary Organic Aerosol Formation Mechanisms from Measured Gas/Particle Partitioning of Specific Organic Tracer Compounds, *Environ. Sci. Technol.*, 47, 3781-3787, 2013.

Zhu, H., Wang, H., Jing, S., Wang, Y., Cheng, T., Tao, S., Lou, S., Qiao, L., Li, L., and Chen, J.: Characteristics and sources of atmospheric volatile organic compounds (VOCs) along the mid-lower Yangtze River in China, *Atmos. Environ.*, 190, 232-240, 10.1016/j.atmosenv.2018.07.026, 2018.

Zhu, S., Wang, Q., Qiao, L., Zhou, M., Wang, S., Lou, S., Huang, D., Wang, Q., Jing, S., Wang, H., Chen, C., Huang, C., and Yu, J. Z.: Tracer-based characterization of source variations of PM<sub>2.5</sub> and organic carbon in Shanghai influenced by the COVID-19 lockdown, *Faraday Discuss*, 226, 112-137, 10.1039/d0fd00091d, 2021.