

1 **Realtime measurement of phase partitioning of organic**
2 **compounds using a Proton-Transfer-Reaction Time-of-Flight**
3 **Mass Spectrometer coupled to a CHARON inlet**

4 **Yarong Peng^{1,2}, Hongli Wang^{2,*}, Yaqin Gao², Shengao Jing², Shuhui Zhu², Dandan**
5 **Huang², Peizhi Hao³, Shengrong Lou², Tiantao Cheng^{4,5,*}, Cheng Huang², Xuan**
6 **Zhang^{3,*}**

7 ¹ Department of Environmental Science and Engineering, Fudan University, Shanghai, 200438, China

8 ² State Environmental Protection Key Laboratory of Formation and Prevention of Urban Air Pollution
9 Complex, Shanghai Academy of Environmental Sciences, Shanghai, 200233, China

10 ³ School of Natural Sciences, University of California, Merced, 95343, USA

11 ⁴ Department of Atmospheric and Oceanic Sciences, Fudan University, Shanghai, 200438, China

12 ⁵ Big Data Institute for Carbon Emission and Environmental Pollution, Fudan University, Shanghai,
13 200438, China

14

15 *Correspondence to:* Hongli Wang (wanghl@saes.sh.cn), Tiantao Cheng (ttcheng@fudan.edu.cn), Xuan
16 Zhang (xzhang87@ucmerced.edu)

17

18 **Abstract.** Understanding the gas-particle partitioning of semivolatile organic compounds
19 (SVOCs) is of crucial importance in the accurate representation of the global budget of
20 atmospheric organic aerosols. In this study, we quantified the gas- vs. particle-phase fractions
21 of a large number of SVOCs in real time in an urban area of East China with the use of a
22 CHemical Analysis of aeRosols ONline (CHARON) inlet coupled to a high resolution Proton
23 Transfer Reaction Time-of-Flight Mass Spectrometer (PTR-ToF-MS). We demonstrated the
24 use of the CHARON inlet for highly efficient collection of particulate SVOCs while
25 maintaining the intact molecular structures of these compounds. The collected month-long
26 dataset with hourly resolution allows us to examine the gas-particle partitioning of a variety of
27 SVOCs under ambient conditions. By comparing the measurements with model predictions
28 using the instantaneous equilibrium partitioning theory, we found that the dissociation of large
29 parent molecules during the PTR ionization process likely introduces large uncertainties to the
30 measured gas- vs. particle-phase fractions of less oxidized SVOCs, and therefore, caution
31 should be taken when linking the molecular composition to the particle volatility when
32 interpreting the PTR-ToF-MS data. Our analysis suggests that understanding the fragmentation
33 mechanism of SVOCs and accounting for the neutral losses of small moieties during the
34 molecular feature extraction from the raw PTR mass spectra could reduce, to a large extent, the
35 uncertainties associated with the gas-particle partitioning measurement of SVOCs in the
36 ambient atmosphere.

37

38 **1. Introduction**

39 Gas-particle partitioning of semivolatile organic compounds (SVOCs) is a critical process
40 involved in the formation and evolution of atmospheric organic aerosols (OA). Traditionally,

41 gas-particle equilibrium partitioning of organic substances is assumed to be established
42 instantaneously (Zhang and Seinfeld, 2013), this assumption is in question if particles are semi-
43 solid or glassy (Shiraiwa et al., 2013). Most studies to date addressing the kinetic limitations in
44 partitioning have used indirect and/or theoretical methods that are lack of chemical and
45 molecular specificity (Mai et al., 2015; Shiraiwa and Seinfeld, 2012). Direct measurements of
46 gas-particle partitioning of SVOCs are needed in order to develop accurate parameterizations
47 for the organic aerosol formation in climate models.

48 The major challenge in the characterization of gas-particle partitioning of SVOCs lies in
49 the realtime measurement of labile compounds while maintaining their intact molecular
50 structures with minimal fragmentation (Zhang et al., 2016a; Zhang et al., 2016b; Zhang et al.,
51 2019). In recent years, soft ionization techniques coupled to mass spectrometry have been
52 widely used for the measurement of gas-phase SVOCs at the molecular level (Veres et al., 2008;
53 Crouse et al., 2006; Heald and Kroll, 2020). Combined with thermal desorption methods, these
54 techniques have also been deployed to measure organic compounds in both gas and particle
55 phases nearly simultaneously (Krechmer et al., 2016). A notable example would be the use of
56 the Filter Inlet for Gases and AEROSols coupled with the Chemical Ionization Mass
57 Spectrometry (FIGAERO-CIMS) to quantify the gas-particle partitioning of a broad range of
58 organic compounds in real time (Lopez-Hilfiker et al., 2014; Ye et al., 2021; Voliotis et al., 2021;
59 Wang et al., 2020a; Lutz et al., 2019; Lee et al., 2018; Le Breton et al., 2018; Stark et al., 2017;
60 Lopez-Hilfiker et al., 2016; Lopez-Hilfiker et al., 2015; Palm et al., 2020). A number of studies
61 among those have reached a consensus that the thermograms method, i.e., using the calibrated
62 thermal desorption profiles vs. temperature to derive the volatility, likely provides the best
63 estimates of the actual phase distribution. In contrast, using the directly measured gas- and

64 particle-phase fractions of a given analyte will most likely introduce a significant positive or
65 negative bias to the volatility estimation due to the thermal decomposition of labile organic
66 compounds during the desorption process (Lopez-Hilfiker et al., 2015; Stark et al., 2017). Such
67 thermal decomposition (or ion fragmentation) artifacts, either positive or negative depending
68 on the molecular size, have been suggested to constitute the largest uncertainties in the
69 estimation of phase partitioning behaviors of SVOCs using the thermal desorption method at
70 ambient pressure (Thompson et al., 2016).

71 Along with the line of thermal desorption method development, an inlet designed for the
72 CHEMical Analysis of aeRosols ONline (CHARON) has been developed and coupled to the
73 Proton Transfer Reaction Time-of-Flight Mass Spectrometer (PTR-ToF-MS) in recent years.
74 As CHARON-PTR-ToF-MS does not rely on any form of pre-concentration on surfaces, it
75 could provide online and direct measurements of organic compounds in both phases, compared
76 with traditional thermal desorption instruments which still need to address artifacts during the
77 particle collection and desorption processes. Another potential advantage of CHARON-PTR-
78 ToF-MS is that the chemical information of the collected particles can be studied qualitatively
79 and quantitatively over a chemical composition level even at sub-nanogram mass
80 concentrations per molecule owing to the well studied ion-molecule reaction chemistry in PTR-
81 ToF-MS (Piel et al., 2019). CHARON has shown promising potential in the realtime analysis
82 of the chemical composition and spatiotemporal distributions of aerosols with laboratory-,
83 ground-, and aircraft-based platforms (Piel et al., 2019; Tan et al., 2018; Gkatzelis et al., 2018a;
84 Gkatzelis et al., 2018b; Muller et al., 2017; Eichler et al., 2017; Eichler et al., 2015; Antonsen
85 et al., 2017; Leglise et al., 2019; Piel et al., 2021). As a relatively new technique, the use of
86 CHARON-PTR-ToF-MS to investigate the gas-particle partitioning of organic compounds is

87 still quite limited. Only one study by Gkatzelis et al. (2018b) deployed CHARON, together
88 with two other aerosol sampling inlets, to measure the OA formation and aging from
89 monoterpenes and real plant emissions in chamber experiments. Whether the CHARON inlet
90 can be applied to the study of gas-particle partitioning of organic compounds under the actual
91 atmospheric conditions remains to be validated.

92 In this study, we assess the applicability of the CHARON inlet to the time-resolved
93 collection of organic compounds in their native molecular state using laboratory tests with a
94 series of authentic standards. We further employ the CHARON inlet coupled to a high
95 resolution PTR-ToF-MS instrument to measure an array of gaseous and particulate SVOCs in
96 an urban area of East China. The obtained month-long hourly dataset allows us to examine the
97 gas-particle partitioning of SVOCs spanning a range of volatilities. By comparing the
98 measurements with model predictions using the instantaneous equilibrium partitioning theory,
99 we found that fragmentation during the PTR ionization process may introduce large
100 uncertainties to the measured gas- vs. particle-phase fractions of less oxidized SVOCs.
101 Understanding the dissociation patterns of parent molecules and accounting for the
102 fragmentation losses when extracting the molecular features from the raw PTR mass spectra
103 are needed to improve the measured accuracy of SVOCs partitioning between the gas and
104 particle phase.

105

106 **2. Material and Methods**

107 **2.1. Sampling site**

108 The sampling site located in the campus of the Shanghai Academy of Environmental
109 Science (SAES) is representative of a typical urban setting surrounded by restaurants, shopping

110 malls, and residential and commercial buildings (Fig. S1). Two traffic-heavy streets in the area
111 (Caobao Road and Humin Highway) are located ~ 150 m and 450 m lateral distances to the east
112 of the sampling site. A few number of petrochemical and chemical industrial facilities are
113 located ~ 50 km to the south and southwest of the observation site, which likely bring certain
114 long-lived pollutants to the site at a typical wind speed of $\sim 1 - 3$ m/s. Major pollution sources
115 in the area include traffic, commercial and residential activities, and regional transport (Huang
116 et al., 2021; Peng et al., 2023). The sampling inlet of the PTR-ToF-MS instrument was installed
117 on the roof of an eight-story building ~ 24 m above the ground. A comprehensive measurement
118 of gas- and particle-phase compounds in the ambient air was performed from Oct 24 to Nov 22.
119 During the sampling period, the average temperature, relative humidity, and wind speed were
120 18.0 ± 3.0 °C, $61.0 \pm 15.0\%$, and 1.8 ± 0.7 m/s, respectively. The prevailing wind direction in
121 this region was from the northwest and north during the polluted period (Fig. S2).

122 **2.2. CHARON-PTR-ToF-MS**

123 **2.2.1. Operation protocols**

124 A Proton Transfer Reaction Time-of-Flight Mass Spectrometer (PTR-ToF-MS) coupled to
125 a CHEMICAL ANALYSIS OF aeRosols ONLINE (CHARON, Ionicon Analytik Inc, Innsbruck,
126 Austria) inlet was employed to measure the gas- and particle-phase concentrations of a series
127 of SVOCs. The PTR-ToF-MS instrument used here is equipped with a radio frequency (RF)-
128 only quadrupole ion guide that transmits ions more efficiently (PTR-QiTOF, Ionicon Analytik
129 Inc) but results in a low-mass cutoff (Fig. S4). The operating parameters of the PTR-ToF-MS
130 were held constant during the entire measurement period. The drift tube pressure, temperature,
131 and voltage were 2.9 mbar, 120 °C, and 500 V, respectively. These conditions correspond to an
132 E/N (E is the electric field, and N is the number density of the gas molecules in the drift tube)

133 value of ~ 100 Td ($1 \text{ Td} = 10^{-17} \text{ V cm}^2$) and a reaction time of 120 μs . Note that the E/N value
134 determines the collision energy of ions in the reactor and therefore the degree of fragmentation
135 and cluster formation. The operating conditions were selected for the purpose of relatively low
136 fragmentation intensities (compared to 120 – 140 Td) and limited production of water clusters
137 (compared to 60 – 80 Td). Low E/N enhances the degree of water clustering, which complicates
138 the analysis of analyte ions due to a complex interplay between cluster formation ($\text{RH}^+(\text{H}_2\text{O})_n$)
139 and proton transfer reactions (Holzinger et al., 2019). During the campaign, the sensitivity of
140 the PTR-ToF-MS was in the range of 300 – 1000 ncps ppb⁻¹ and the mass resolution was
141 maintained at $\sim 5000 \text{ m}/\Delta\text{m}$. Mass spectra were collected at a time resolution of 10 s.

142 The CHARON inlet consists of 1) a gas-phase denuder (GPD) for stripping off gas-phase
143 analytes, 2) an aerodynamic lens (ADL) for particle collimation which is combined with an
144 inertial sampler for emanating the particle-enriched flow, and 3) a thermo-desorption unit (TDU)
145 for particle volatilization. The CHARON inlet functionality has been described in great detail
146 by Eichler et al. (2015). The inlet we used here had a particle enrichment factor of ~ 15 , as
147 discussed shortly. The vaporizer (TDU) was operated at 140 °C and ~ 8 mbar absolute pressure.
148 This temperature was chosen to ensure all the unknowns observed in the field can be evaporated
149 effectively while maintaining relatively intact molecular structures, see more details in Section
150 3.1. Measurements of organic compounds in the gas and particle phase were conducted using a
151 parallel sampling system with two independent pumps, allowing for the selection of flow rates
152 specifically adjusted for each phase, resulting in the overall residence time of less than 2 s (Fig.
153 S3).

154 **2.2.2. Sampling alternation between gas and particle phase**

155 Gas-phase compounds were measured by directly sampling the ambient air via a 2 m long

156 perfluoro-alkoxy (PFA) tube (1/4" OD) capped with a polytetrafluoroethylene (PTFE) filter
157 (Mitex™ PTFE membrane, 5 µm pore size, 47 mm diameter) to prevent the clogging of
158 particles in the PTR capillaries. The gas-phase inlet was independently connected to the PTR-
159 ToF-MS instrument upstream of the drift tube via a pressure-controlled subsampling PEEK
160 capillary (1/16" OD). Zero measurements were performed by overflowing catalytically
161 (platinum at 370°C) purified air through the inlet. Ambient particles were sampled through a
162 stainless steel tube (3/8" OD) with a flow rate of ~ 3 L min⁻¹, out of which a flow of ~ 500 ml
163 min⁻¹ was directed to the CHARON inlet. A PM_{2.5} cyclone was installed in front of the sampling
164 line to remove coarse particles (> 2.5 µm). The particle-phase background was measured by
165 placing a High-Efficiency Particulate Air filter (HEPA, HEPA-CAP 7, GE Healthcare UK
166 Limited, Buckinghamshire, UK) upstream of the CHARON inlet. Servo motor activated valves
167 made of passivated stainless steel were used for switching between the two inlet configurations.
168 During the campaign, CHARON-PTR-ToF-MS automatically switched between gas and
169 particle phase every 15 min. Detailed setup is given in Fig. S3 in the Supporting Information.

170 The built-in PTR-manager software (Ionicon Analytik GmbH, Innsbruck, Austria) offers
171 the possibility to program sequences by which the instrument switches between different
172 settings. It takes ~ 1 min for gases and particles to re-equilibrate when switching between these
173 two modes. Data generated during this transition period (~ 2 min) were not considered.
174 Instrument background was measured for 15 min every 5h. The limits of detection (LoD) at 1
175 min resolution were in the range of 5.6 ± 2.9 ng m⁻³ for gases and 0.7 ± 0.5 ng m⁻³ for particles,
176 respectively (Fig. S5). Concentrations of gaseous and particulate compounds shown here
177 included the last 5 min of every gas/particle-phase working mode, in order to minimize the
178 interferences carried over from the previous working mode by allowing for a sufficient amount

179 of equilibration time in the inlet (Piel et al., 2021). In order to synchronize the gas- and particle-
 180 phase data to calculate gas-particle partitioning, the average hourly data were then used for
 181 further analysis.

182 2.2.3. Sensitivity and calibration

183 Weekly calibrations were performed using a multicomponent calibration gas standard
 184 (Linde, USA) at five concentration levels from 0.5 to 10 ppb (Fig. S4a). The calibration mixture
 185 includes methanol, acetonitrile, acetaldehyde, acrolein, acetone, isoprene, methyl vinyl ketone,
 186 methyl ethyl ketone, 2-pentanone, toluene, styrene, p-xylene, 1,3,5-trimethylbenzene,
 187 naphthalene and α -pinene. Calibration standards with higher molecular weight were excluded
 188 because we only considered ion masses below 200 amu from the field measurement for the
 189 study of gas-particle partitioning, see discussions given in Text S1. Here the sensitivity of PTR-
 190 ToF-MS is defined as the normalized ion intensity of RH^+ (ncps) obtained at a mixing ratio of
 191 1 ppb. For a given species (R), its sensitivity (S , ncps ppb^{-1}) is a linear function of the rate
 192 constant of its reaction with H_3O^+ (k):

$$193 \quad S = \frac{\frac{I_{\text{RH}^+}}{I_{\text{H}_3\text{O}^+}} \times 10^6}{\frac{[R]}{N} \times 10^9} = k \times N \times 10^{-3} \times t \times \frac{T_{\text{RH}^+}}{T_{\text{H}_3\text{O}^+}} \times F_{\text{RH}^+} \quad (1)$$

$$194 \quad \text{corrected } S = \frac{S}{\frac{T_{\text{RH}^+}}{T_{\text{H}_3\text{O}^+}} \times F_{\text{RH}^+}} = a \times k \quad (2)$$

195 where the signals of H_3O^+ ($I_{\text{H}_3\text{O}^+}$) and RH^+ ions (I_{RH^+}) measured by the mass analyzer (in cps)
 196 can be related to the signals of H_3O^+ ($[\text{H}_3\text{O}^+]$) and RH^+ ($[\text{RH}^+]$) ions at the end of the drift tube,
 197 using their respective transmission efficiencies ($T_{\text{H}_3\text{O}^+}$ and T_{RH^+}) from the drift tube to the
 198 detector (De Gouw and Warneke, 2007). $[R]$ is the concentration of species R and N is the
 199 number density of gas in the drift tube. The reaction time (t) is determined by the ion drift

200 velocity. F_{RH^+} represents the fraction of product ions detected as RH^+ ions ($0 \leq F_{\text{RH}^+} \leq 1$). For
201 non-fragmenting compounds, $F_{\text{RH}^+} = 1$. The measured sensitivity is further corrected by
202 accounting for fragmentation and transmission efficiency, the value of which was derived from
203 laboratory experiments (Fig. S4a). Uncertainties associated with the addition of the low-mass
204 filter have been accounted for in the regression of individual transmission efficiency
205 measurements on corresponding mass to charge ratios. The overall relative standard deviations
206 were less than 15%. a is the slope of the linear regression of the corrected sensitivities on the
207 proton-transfer-reaction rate coefficients (k), as shown in Fig. S4b. Following the method by
208 Sekimoto et al. (2017), the linear regression result was used to determine the sensitivities of all
209 uncalibrated species. The overall uncertainty was less than 15% for compounds with standards
210 and around 50% for those without standards. Calculated sensitivity based on this method agrees
211 well with measurements of authentic standards (Fig. S4c).

212 **2.2.4. Enrichment factor**

213 The CHARON inlet was calibrated routinely with pure ammonium nitrate particles to
214 derive the enrichment factor as a function of the particle size following the procedures described
215 in Eichler et al. (2015). In addition, we used a selection of authentic standards (Table S1) to test
216 the effect of desorption temperature on the enrichment factor of labile compounds. Previous
217 studies with CHARON generally used a temperature of 140 °C to vaporize particles (Leglise et
218 al., 2019; Gkatzelis et al., 2018b; Tan et al., 2018). Herein, we tested the TDU temperature
219 ranging from 70 to 140°C. The selected chemical standards were individually dissolved in
220 distilled water (ethanol in the case of 2-Pentadecanone and 1-Pentadecanol) and nebulized by
221 an atomizer (TSI 3076, Shore-view, MN, USA) that was pressurized with ultrapure zero air.
222 The nebulizer outflow was diverted through two diffusion dryers to remove water vapor and an

223 activated charcoal denuder (NovaCarb F, Mast Carbon International Ltd., Guilford, UK) to
 224 remove organic vapors. The resulting flow of polydisperse particles was then delivered into a
 225 differential mobility analyzer (DMA, TSI 3080) for particle size selection. The transmitted
 226 particles at a given size bin (300 nm for organics and 100 – 450 nm range for ammonium nitrate)
 227 were introduced into the CHARON-PTR-ToF-MS analyzer and a condensation particle counter
 228 (CPC, TSI 3775), respectively. Particle mass concentrations were calculated based on the CPC
 229 number distribution measurements by assuming a shape factor of 0.8 for ammonium nitrate
 230 particles and 1 for organic particles, respectively.

231 The particle enrichment factor (EF) of a given analyte i was calculated as the ratio of the
 232 PTR-ToF-MS derived vs. CPC derived mass concentrations of analyte i at a given particle size
 233 bin:

$$234 \quad VMR_{(PTR)i} = \frac{I_i}{S_i} \quad (3)$$

$$235 \quad VMR_{(CPC)i} = \rho_i \times V \times N_i \times V_m / Mw_i \quad (4)$$

$$236 \quad EF = \frac{VMR_{(PTR)i}}{VMR_{(CPC)i}} \quad (5)$$

237 where I_i is the normalized signal of species i (ncps) by PTR-ToF-MS, S_i is the sensitivity
 238 (ncps ppb⁻¹), VMR is the volume mixing ratio (ppb), ρ is the density of species i (g cm⁻³),
 239 V is the volume of a particle sphere (m³), N_i is the number concentration of particles measured
 240 by CPC (cm⁻³), V_m is the molar volume of an ideal gas at 1 atm (22.4 L mol⁻¹), Mw_i is the
 241 molecular weight (g mol⁻¹). As the calculated sensitivities of most organics in the absence of
 242 authentic standards are subject to uncertainties (15% – 50%), we will herein use the
 243 multiplication of EF and S_i to evaluate the combined effect of CHARON enrichment and
 244 sensitivity on the measured concentrations of a given analyte i in the particle phase.

245 **2.2.5. Data processing**

246 Data were analyzed using the Tofware package (v3.2.0, Tofwerk Inc), within the Igor Pro
247 software (v7.0, Wavemetrics). Using this package, time-dependent mass calibrations were
248 performed using four ions ($\text{H}_3^{18}\text{O}^+$, NO^+ , $\text{C}_6\text{H}_5\text{I}^+$ and $\text{C}_6\text{H}_5\text{I}_2^+$), where $\text{C}_6\text{H}_5\text{I}^+$ and $\text{C}_6\text{H}_5\text{I}_2^+$ were
249 produced from the internal standard di-iodobenzene. The relative mass deviation was within 6
250 – 8 ppm across the mass spectra. Considering the humidity dependence of reagent ions (H_3O^+
251 and $\text{H}_3\text{O}^+(\text{H}_2\text{O})$), the fitted product ion signals (RH^+) were normalized to a standard reagent ion
252 of 10^6 cps (counts per second). Elemental composition was determined based on the accurate
253 m/z (mass to charge ratio) and isotopic pattern analysis. A list of ~ 1600 ions was extracted,
254 including both gas- and particle-phase ions. Molecular formulas including only C, H, and O
255 atoms were assigned to the detected ions by the addition of one proton in cases where the
256 elemental composition analysis returned multiple options. About 85% of the signals were
257 elementally resolved by the $\text{C}_x\text{H}_y\text{O}_z$ formula in ambient air mass spectra. A small number of
258 nitrogen containing compounds, such as nitroaromatics, were also identified but not included
259 in the following analysis. Throughout of the context, we use the word “species” to refer to all
260 compounds with assigned molecular formula, which may include multiple isomers.

261 **2.3. Complementary measurements**

262 In addition to CHARON-PTR-ToF-MS, a Thermal desorption Aerosol Gas chromatograph
263 (TAG) was also employed to measure a series of particle-phase organic species. Details of the
264 TAG operation and data analysis protocols can be found in previous studies (He et al., 2020;
265 Wang et al., 2020b; Zhu et al., 2021). The elemental composition and mass concentration of
266 particles were measured by an Aerodyne high-resolution time-of-flight Aerosol Mass
267 Spectrometer (AMS), with details of operation and quality control protocols given by our recent

268 study (Huang et al., 2021). Volatile organic compounds (VOCs, C₂ – C₁₂) were analyzed by a
 269 custom-built online gas chromatography system equipped with a mass spectrometer and a flame
 270 ionization detector (GC-MS/FID). The performance of this system can be found in our previous
 271 publications (Zhu et al., 2018; Wang et al., 2014). Meteorological parameters (ambient
 272 temperature, wind speed, wind direction, and relative humidity) were collected by an automatic
 273 weather station (Metone 590 series) mounted on the roof top of the campaign site.

274 **2.4. Gas-particle partitioning measurements vs. modeling**

275 The CHARON-PTR-ToF-MS measured gas- and particle-phase concentrations of a given
 276 species i can be used to calculate its particle-phase fraction ($F_{p,i}$).

$$277 \quad P_i = \frac{I_{p,i} \times \left(\frac{m}{z_i} - 1\right)}{V_m \times S_i \times EF} \quad (6)$$

$$278 \quad G_i = \frac{I_{g,i} \times \left(\frac{m}{z_i} - 1\right)}{V_m \times S_i} \quad (7)$$

$$279 \quad F_{p,i} = \frac{P_i}{P_i + G_i} \quad (8)$$

280 where P_i and G_i are the mass concentrations (ng m⁻³) of species i in the particle and gas
 281 phase, respectively. $I_{p,i}$ and $I_{g,i}$ are the normalized signal (ncps) of the PTR-ToF-MS detected
 282 ion i in the particle and gas phase, respectively. V_m is taken as 22.4 L/mol. S_i is calculated
 283 or measured sensitivity (ncps ppb⁻¹), see details in Section 2.2.3. As structural isomers cannot
 284 be resolved in the mass spectra, the calculation here assumes that all isomers with the same
 285 molecular formula have the same chemical properties, i.e., saturation vapor pressures.
 286 Substitution of Equations (6) and (7) to Equation (8) yields the final expression of the particle-
 287 phase fraction of species i ($F_{p,i}$), which is a function of the observed PTR-MS raw signals of
 288 species i in the gas and particle phase (in total ion counts), as well as the particle enrichment

289 factor (EF) of species i .

$$290 \quad F_{p,i} = \frac{I_{p,i}/EF}{I_{p,i}/EF + I_{g,i}} \quad (9)$$

291 Gas-particle partitioning of a given analyte i was also modeled using the equilibrium
292 partitioning theory (Pankow, 1994):

$$293 \quad F_{p,i} = \frac{1}{1 + C_i^*/C_{OA}} \quad (10)$$

$$294 \quad C_i^* = \frac{10^6 M w_i \zeta_i p_i}{RT} \quad (11)$$

295 where C_{OA} is the organic aerosol concentration measured by AMS ($\mu\text{g m}^{-3}$), C_i^* is the
296 saturation mass concentration ($\mu\text{g m}^{-3}$), $M w_i$ is the molecular weight (g mol^{-1}), ζ is the
297 activity coefficient (assumed as unity), p_i is the pure component liquid vapor pressure (Pa), R
298 is the universal gas constant ($8.2 \times 10^{-5} \text{ m}^3 \text{ atm K}^{-1} \text{ mol}^{-1}$), and T is the ambient temperature
299 (K). As detailed chemical information is lacking for all species detected by PTR-ToF-MS, here
300 we use the expression given by Donahue et al. (2011) to approximate the value of C_i^* :

$$301 \quad \log_{10} C_i^* = (n_C^0 - n_C^i) b_C - n_O^i b_O - 2 \frac{n_C^i n_O^i}{n_C^i + n_O^i} b_{CO} \quad (12)$$

302 where $n_C^0 = 25$, $b_C = 0.475$, $b_O = 2.3$, and $b_{CO} = -0.3$.

303 **2.5. Uncertainties in the measured and modeled gas-particle partitioning**

304 The uncertainty associated with the PTR-MS measured concentrations in both gas and
305 particle phases is less than 15% for compounds with chemical standards based on the optimally
306 fitted transmission efficiency curve. For those in the absence of standards, their PTR
307 sensitivities were calculated theoretically using Equations (1-2), and the uncertainty in the
308 calculation mainly arises from the estimation of polarizability and dipole moment of the target
309 molecule, which has been estimated to be within $\sim 50\%$ when only the elemental composition

310 of that molecule is given (Sekimoto et al., 2017). It is important to note, however, that the
311 uncertainty associated with the estimated PTR sensitivity has zero influence on the measured
312 particle-phase fraction of any given compound because the sensitivity term is essentially
313 canceled in the divisor function in Equation (9). The uncertainty associated with the particle
314 fraction of a given species i derived from the PTR-MS measurements arises predominantly
315 from the “ EF ” term. Since the uncertainty of the measured EF depends on the uncertainty of
316 $I_{p,i}$, we thus express the overall uncertainties of the measured gas-particle partitioning as:

$$317 \quad \text{Unc}(F_{p,i}) = \sqrt{\text{Unc}(EF)^2 + \text{Unc}(I_{g,i})^2} \quad (13)$$

318 with the calibration standards used in this study, the enrichment factor is calculated to be within
319 25% error ($\text{Unc}(EF)$) (see detailed calculations listed in Table S2 and S3), including the effect
320 of wall loss inside the inlet tubing and the precision of the measurement. Huang et al. (2019)
321 has tested the uncertainty of wall loss ($\text{Unc}(I_{g,i})$) in this PTR-MS instrument as 28%. Therefore,
322 the overall uncertainty in the measured particle-phase fraction ($\text{Unc}(F_{p,i})$) was estimated as
323 38%.

324 The uncertainty associated with the modeled gas-particle partitioning arises primarily from
325 the uncertainty in the estimation of the saturation mass concentration (C_i^*) based on the method
326 developed by Donahue et al. (2011). In this method, the saturation mass concentration of
327 species i is a non-linear function of the numbers of carbon and oxygen atoms in that particular
328 species, see Equation (12). For each generic molecular formula, i.e., C_xH_y , C_xH_yO , $C_xH_yO_2$, and
329 $C_xH_yO_4$, Donahue et al. (2011) have used a total of 25, 48, 18, and 10 chemical standards with
330 known volatilities to validate the estimated saturation concentrations, and the estimated errors
331 were taken as 34%, 16%, 25%, and 54%, respectively (see Table S6). As the $C_xH_yO_3$ group was

332 not tested, we tentatively assumed the associated errors as the same as the $C_xH_yO_4$ group. The
333 extent to which these uncertainties may affect the difference between measurements and model
334 results was discussed in detail in the Supporting Information (Fig. S10).

335

336 **3. Results and Discussion**

337 **3.1. Particle enrichment: effect of desorption temperature**

338 Thermal desorption as a common procedure used in the chemical characterization of
339 organic aerosols is often susceptible to fragmentation of non-refractory compounds. Due to the
340 high temperature used to evaporate particles collected, labile and large molecules are inevitably
341 subject to fragmentation, thereby introducing large uncertainties to the measured mass and
342 composition of the particulate organic compounds (Lopez-Hilfiker et al., 2015; Yatavelli et al.,
343 2012; Zhao et al., 2013). Thermal decomposition of oxidized organic compounds has been
344 observed at vaporizer temperature as low as 200 °C, the lowest temperature required to vaporize
345 OA as reported (Stark et al., 2017). While decreasing the vaporizer temperature is necessary to
346 maintain the intact structure of labile molecules, low temperature (e.g., 85 °C), however, might
347 fail to completely evaporate the collected particles into vapors, resulting in an underestimation
348 of the collected OA mass (Inomata et al., 2014). Here we performed a series of sensitivity tests
349 to identify the optimal vaporizer temperature in the CHARON inlet for the measurements of
350 organic compounds in the particle phase.

351 Prior to the temperature sensitivity test, we have validated that the particle enrichment
352 factor, also known as collection efficiency and defined as the ratio of the particle mass
353 concentration upstream to downstream of the aerodynamic lens, does not depend on the particle
354 size. As shown in Fig. 1a, the measured *EF* value for ammonium nitrate particles, detected as

355 NO_2^+ produced from the nitric acid vapor, remains constant as ~ 15 in the 150 – 450 nm particle
356 size range. The lower values in the 100 – 150 nm size range can be explained by the lower
357 particle transmission efficiency in the gas phase denuder, e.g., 75% – 80% for 100 nm particles
358 (Eichler et al., 2015). Also, particles below 150 nm are less efficiently concentrated in the
359 subsampling flow after the aerodynamic lens. Therefore, we used the monodisperse particles
360 generated from selected organic standards at 300 nm for the temperature sensitivity test.

361 A number of chemical standards that are representative of alcohols, carbonyls, and
362 carboxylic acids and with the vapor pressure ranging from 10^{-14} to 10^{-1} Pa at 25 °C (taken from
363 EPA EPI Suite (2012), see values given in Table S1) were used to generate organic aerosols,
364 which, upon size selection at 300 nm, were directed to the CHARON inlet. Particle evaporation
365 occurs downstream of the aerodynamic lens in the gas phase and on the tube and orifice surface
366 to which submicron particles rapidly diffuse at ~ 8 mbar operating pressure. The thermal
367 desorption unit was designed to ensure that ammonium sulfate particles (10^{-20} Pa) can be
368 completely evaporated (Piel et al., 2019; Eichler et al., 2015). As the desorption temperature
369 was varied from 70 °C to 140 °C, the intensities of all detected ions (including both parent and
370 fragment ions) for each organic standard analyzed were stable within 15%, as shown in Fig. 1b.
371 Also note that we did not observe any ions produced from decarboxylation and/or dehydration
372 during the particle evaporation process. This is because the relative low operation temperature
373 and the short heat exposure time could effectively limit any thermal dissociation of organic
374 molecules. This demonstrates that the parent molecule fragmentation, if any, does not occur
375 under the range of desorption temperature used in the CHARON inlet, but rather results from
376 the ionic dissociation process in the PTR ionization chamber, see more discussions in Section
377 3.3. We therefore used the sum total of intensities of all major ions detected as the PTR-ToF-

378 MS response to a given organic standard analyzed. Fig. 1c shows that the derived enrichment
379 factors stay constant for all compounds investigated ($M_w \sim 160 - 230$ g/mol). The relative
380 signals of all fragment ions were stable over the range of the desorption temperature as shown
381 in Fig. 2. This suggests that the desorption temperature used here, even as low as 70 °C, is
382 sufficient to evaporate SVOCs (volatility $> 10^{-14}$ Pa at 25 °C), due to the low operating pressure
383 (~ 8 mbar) that significantly enhances the partitioning shift to the gas phase. One low volatility
384 compound, sucrose (M_w is 342 g/mol and vapor pressure is 4.69×10^{-14} Pa), has a slightly lower
385 enhancement factor compared with all the other organic standards tested. This is mainly due to
386 the intensive dehydration of the parent compound in the ionization chamber, and as a result,
387 only a few fragment ions were captured, resulting in a lower PTR response and thereby lower
388 *EF* value calculated from Equations (3-5). The *EF* values of sucrose also have much higher
389 standard deviations at all temperatures due to fragmentation (Table S3).

390 It has been recognized that species with one functional group follow certain fragmentation
391 patterns during the PTR ionization process (Pagonis et al., 2019; Francis et al., 2007; Spanel et
392 al., 1997; Tani et al., 2003; Spanel and Smith, 1997), such as dehydration of acids and alcohols.
393 The observed dissociation of carboxylic acid standards used in this study, e.g., phthalic acid
394 and 3-methylbutane-1,2,3-tricarboxylic acid, can be explained by this common fragmentation
395 pattern. The fragmentation mechanism of multi-functionalized species is rather complicated and
396 a number of fragments can be produced upon PTR ionization. Nevertheless, the identity and
397 abundance of fragments from a given multi-functionalized species have been found comparable
398 under the same PTR operation protocols (Leglise et al., 2019; Gkatzelis et al., 2018a). For
399 example, *cis*-pinonic acid yields the following fragments (main ions only and relative
400 abundance in parentheses): m/z 71.049 ($\sim 28\%$), 115.075 ($\sim 36\%$), 167.108 ($\sim 19\%$), and 185.117

401 (~11%), which is comparable with an earlier study (Leglise et al., 2019): m/z 71.049 (~27%),
402 115.075 (~33%), 167.108 (~26%), and 185.117 (~14%) at 100 Td settings.

403 **3.2. Molecular features of detected organic species**

404 A month-long field dataset of particle- and gas-phase organic species was collected at
405 hourly resolution using the CHARON-PTR-ToF-MS instrument. A comparison of PTR-ToF-
406 MS measurements with other techniques available on site was performed for both gas and
407 particle phases. For the gas phase, quantitative measurements of a suite of VOCs by GC-
408 MS/FID, including benzene, toluene, styrene, C₈ and C₉ aromatics, acrolein, and C₄, C₅, and C₆
409 ketones, agree well with corresponding PTR-ToF-MS measurements, as shown in Fig. S6. For
410 the particle phase, the time series of a group of C_xH_yO₄ species (including C₄H₆O₄, C₅H₈O₄,
411 C₆H₁₀O₄, and C₈H₆O₄) are in reasonable agreement with corresponding measurements taken by
412 TAG ($r \sim 0.60 - 0.80$), although the CHARON-PTR-ToF-MS measured total molecular mass
413 is generally lower than the TAG measurements by a factor of 2 to 6 (Fig. S7). This is likely
414 caused by the fragmentation (e.g., loss of H₂O, see Fig. S8) of the parent compounds during the
415 ionization process, as discussed in detail in Section 3.3. The time series of total OA mass
416 characterized by CHARON-PTR-ToF-MS also agree with the AMS measurements ($r \sim 0.91$,
417 Fig. S9). Previous studies have reported the particulate organics measured by PTR-MS with a
418 thermal desorption inlet account for 25% – 60% mass of the total organic aerosols measured
419 by AMS (Holzinger et al., 2013). Direct comparison of the total OA mass loading is not
420 applicable here since the CHARON-PTR-ToF-MS measurement only focused on compounds
421 with the mass to charge ratio below 300 Th. That the majority of ions detected by PTR are
422 present in the lower mass range is primarily due to the fragmentation of larger masses during
423 the ionization process, as discussed extensively in Section 3.3.

424 Fig. 3 (a-b) shows the PTR-ToF-MS spectra of dominant ions averaged over the entire
425 campaign in both gas and particle phases. The mass concentrations of individual ions are in the
426 range of 7.9 – 7179.3 ng m⁻³ in the gas phase and 0.6 – 82.7 ng m⁻³ in the particle phase. A total
427 of 152 species (with > 60% data points above the PTR-ToF-MS detection limits) are identified,
428 contributing to ~ 69% and ~ 44%, respectively, of the total organic mass measured in the gas
429 and particle phase. The molecular distribution characterized by the carbon and oxygen of these
430 species is given in Fig. 3 (c-d). The most abundant species are characterized by a generic
431 formula of C_xH_yO and C_xH_yO₂, resolving ~ 64% and ~ 46% in total of all identified species in
432 the gas and particle phase, respectively. Another dominant component in the gas phase is
433 hydrocarbon-like compounds (C_xH_y) (~ 27%), which contribute ~ 12% of the organic mass in
434 the particle phase. Species with higher oxygen numbers (> 2) contribute to a large fraction (~
435 42%) of the total particulate mass. These C_xH_yO₁₋₄ groups exhibit different diurnal cycles, as
436 shown in Fig. 4, reflecting their unique formation chemistry. The C_xH_y group peaks in the early
437 morning rush hour and likely originates from primary traffic emissions. On the contrary, both
438 C_xH_yO₃ and C_xH_yO₄ groups peak at noon, suggesting a strong secondary formation source. The
439 diurnal trends for C_xH_yO and C_xH_yO₂ groups are relatively flat during the day, likely indicative
440 of an intertwined primary emission and secondary formation processes.

441 **3.3. Measured vs. modeled gas-particle partitioning**

442 Fig. 5 shows the calculated particle-phase fraction (F_p) of the identified 152 species using
443 the CHARON-PTR-ToF-MS measurements in both phases. It is important to note that an
444 authentic standard is not required for the calculation of F_p for any given species, because the
445 PTR sensitivity term is essentially canceled in the divisor function in Equations (6-8). Also
446 given in Fig. 5 is the simulated F_p of the derived molecular formulas of all species identified

447 using the equilibrium partitioning theory, see method described in Section 2.4. Interestingly,
448 for oxidized species such as $C_xH_yO_4$, their measured F_p agree reasonably with the simulations,
449 as shown in Fig. 5a. The $C_xH_yO_4$ group resides in the relatively higher mass range, and species
450 identified in this group are more likely actual compounds rather than fragments from larger
451 parent molecules. Even these species dissociate into lower mass ions during PTR ionization,
452 their calculated particle-phase fractions are unaffected by such fragmentation processes because
453 signals of parent ions decrease by the same extent upon fragmentation in both gas and particle
454 phases. As the oxygen number decreases, the measured F_p values tend to deviate from the
455 simulations by up to several orders of magnitude. Note that these less oxidized compounds (e.g.,
456 C_xH_y) are mostly small molecules and they are highly unlikely present in the condensed phase
457 as closed-shell monomers (Pankow and Asher, 2008; Holzinger et al., 2010). Instead, they are
458 more likely fragments produced from the decomposition of larger molecules, which no
459 surprisingly favor partitioning in the particle phase. It is worth noting that the uptake of small
460 oxidized compounds on the aerosol aqueous phase does not significantly affect the overall
461 particle phase fraction of these compounds, see detailed calculations in Text S3. The
462 discrepancy in the measurement-model comparison underscores the importance of
463 understanding the fragmentation mechanism during PTR ionization when extracting molecular
464 features from the raw mass spectra.

465 Parent ion fragmentation has been widely observed in PTR-MS instruments (Pagonis et al.,
466 2019). Oxygenates exhibit trends in neutral losses of water or saturated alcohols. Here, we
467 apply a correction to the molecular formula of the 152 identified species by assuming that these
468 species are fragments produced from their parent precursors through the neutral losses of a
469 carboxyl group ($-CO_2$), a carbonyl group ($-CO$), a hydroxyl group ($-H_2O$), or an alcohol group

470 (-C₂H₆O). By applying this correction, the modeled F_p of a given species C_xH_yO_z would actually
471 represent the particle-phase fraction of the its parent species C_xH_yO_z•CO₂, C_xH_yO_z•CO,
472 C_xH_yO_z•H₂O, or C_xH_yO_z•C₂H₆O. As shown in Fig. 5 (b-e), such a correction could significantly
473 increase the modeled F_p values by several orders of magnitude. The assumption of neutral
474 losses of CO₂ or C₂H₆O allows for much improved agreement between modeled vs. measured
475 F_p values for less oxidized species. This implies that these small and less oxidized species are
476 likely fragments resulting from the decomposition of larger parent precursors. As our particle
477 enrichment test (details given in Section 3.1) has confirmed that the thermal desorption
478 temperature employed for particle evaporation does not lead to any intensive fragmentation,
479 therefore the collision-induced dissociation during the proton transfer reaction process becomes
480 the predominant process that produces fragments (Lindinger et al., 1998; Gueneron et al., 2015;
481 Gkatzelis et al., 2018b). Although the electric field applied to the drift tube is considered low
482 to moderate compared with most previous PTR-MS measurements ($E/N \sim 100$ Td in this study
483 vs. $E/N \sim 120 - 140$ Td commonly found in PTR-MS measurements) (Pagonis et al., 2019),
484 parent ion fragmentation was still widely observed here and complicated the mass spectra
485 interpretation and molecular feature extraction. While some recent CHARON measurements
486 employed lower electric field in the drift tube ($E/N \sim 60$ Td) (Leglise et al., 2019; Gkatzelis et
487 al., 2018a; Piel et al., 2019), such conditions could promote the formation of water cluster ions,
488 which increase with humidity and reduce the PTR sensitivity, and therefore are not ideally
489 suitable for our field measurements. A compromise solution would be using a moderate electric
490 field in the drift tube (e.g., ~ 100 Td) and meanwhile applying appropriate molecular corrections
491 to all ions detected in the mass spectra by considering possible neutral losses of small moieties.
492 Since in this study only the information of molecular formula is derived from the PTR-MS

493 spectra, we thus provide the lower and upper bound of the gas-particle partitioning corrections
494 owing to neutral losses of H₂O and CO₂, respectively. In general, lower masses with higher
495 volatilities are subject to notable changes in the particle-phase fraction as a result of neutral
496 losses during the PTR ionization process, see detailed discussions in Text S2.

497

498 **4. Conclusions**

499 Recent studies have suggested that some of the model-measurement discrepancies in the
500 representation of ambient organic aerosol budget might be due to the nonequilibrium
501 gas/particle partitioning caused by kinetic limitations in the presence of glassy or semi-solid
502 phase (Perraud et al., 2012; Mai et al., 2015; Shiraiwa et al., 2013). It is therefore necessary to
503 validate whether the equilibrium partitioning theory could adequately describe the
504 condensation of semivolatile organic vapors onto atmospheric aerosols under ambient
505 conditions, and the accurate measurement of these SVOCs in both gas and particle phases is
506 the crucial prerequisite. In this study, we have employed the PTR-ToF-MS instrument coupled
507 to a CHARON inlet, together with a suite of complementary measurements, to characterize the
508 atmospheric partitioning behaviors of an array of SVOCs in an urban environment of East
509 China. Prior to the application to the field measurements, we first performed a series of
510 laboratory experiments to test whether the CHARON inlet is capable of sampling organic
511 molecules (including alcohols, carbonyls, and carboxylic acids) in their native states. With the
512 low pressure condition used in the CHARON inlet, a thermal desorption temperature less than
513 140 °C could adequately evaporate organic compounds with vapor pressure higher than 10⁻¹⁴
514 Pa while minimizing the thermal decomposition of labile functionalities. The auto-switch
515 function between the gas and particle mode with one single PTR-ToF-MS instrument could

516 monitor gaseous and particulate organic compounds in real time, thereby providing important
517 information of their partitioning behaviors in the ambient atmosphere. Particle-phase fractions
518 of a total of 152 organic species were derived from the CHARON-PTR-ToF-MS measurements
519 and further compared with model predictions using the instantaneous equilibrium partitioning
520 theory. While the model captured the particle-phase fraction of oxidized compounds (e.g.,
521 $C_xH_yO_{3-4}$), predictions of less oxidized compounds, notably the C_xH_y family, differ from the
522 corresponding measurements by several orders of magnitude. Such a large discrepancy is very
523 likely caused by the intensive fragmentation of the parent organic compounds during the PTR
524 ionization process. Accounting for common fragmentation patterns in the simulations of gas-
525 particle partitioning, for example, neutral losses of $-CO_2$, $-CO$, $-H_2O$, or $-C_2H_6O$, could largely
526 improve the model-measurement agreement. Such corrections are very necessary towards an
527 accurate measurement of both particle- and gas-phase SVOCs using the CHARON-PTR-ToF-
528 MS instrument. Our study suggests the crucial importance of optimizing operation conditions
529 and understanding the fragmentation mechanism in the particle collection, vaporization, and
530 ionization processes in understanding the gas-particle partitioning of organic compounds using
531 any thermal desorption based aerosol measurement method.

532

533 *Data availability.* The data shown in the paper are available upon request from the
534 corresponding author.

535 *Author Contributions.* YP carried out experiments and measurements and drafted the
536 manuscript. HW and XZ designed the experimental studies, supervised the laboratory work and
537 wrote the manuscript. YG and SJ supported the ambient measurements. SZ, DH, PH, and SL
538 supported the data analysis. TC and CH supervised the scientific work. All authors have given

539 approval to the final version of the manuscript.

540 *Competing interests.* The authors declare no competing financial interest.

541 *Acknowledgements.* This research has been supported by the National Natural Science
542 Foundation of China (No. 42175135, No. 42175179), the National Key R&D Programm of
543 China (No. 2022YFE0136200) and the Shanghai Science and Technology Commission of the
544 Shanghai Municipality (No. 20ZR1447800).

545

546 **References**

547 Antonsen, S., Bunkan, A. J. C., D'Anna, B., Eichler, P., Farren, N., Hallquist, M., Hamilton, J.
548 F., Kvarnliden, H., Mikoviny, T., Muller, M., Nielsen, C. J., Stenstrom, Y., Tan, W., Wisthaler,
549 A., and Zhu, L.: Atmospheric chemistry of tert-butylamine and AMP, in: *Enrgy Proced*, edited
550 by: Dixon, T., Laloui, L., and Twinning, S., *Energy Procedia*, Elsevier Science Bv, Amsterdam,
551 1026-1032, 10.1016/j.egypro.2017.03.1248, 2017.

552 Crounse, J. D., McKinney, K. A., Kwan, A. J., and Wennberg, P. O.: Measurement of gas-phase
553 hydroperoxides by chemical ionization mass spectrometry, *Anal. Chem.*, 78, 6726-6732,
554 10.1021/ac0604235, 2006.

555 de Gouw, J. and Warneke, C.: Measurements of volatile organic compounds in the earths
556 atmosphere using proton-transfer-reaction mass spectrometry, *Mass Spectrometry Reviews*, 26,
557 223-257, 10.1002/mas.20119, 2007.

558 Donahue, N. M., Epstein, S. A., Pandis, S. N., and Robinson, A. L.: A two-dimensional
559 volatility basis set: 1. organic-aerosol mixing thermodynamics, *Atmos. Chem. Phys.*, 11, 3303-
560 3318, 10.5194/acp-11-3303-2011, 2011.

561 Eichler, P., Muller, M., D'Anna, B., and Wisthaler, A.: A novel inlet system for online chemical
562 analysis of semi-volatile submicron particulate matter, *Atmos. Meas. Tech.*, 8, 1353-1360,
563 10.5194/amt-8-1353-2015, 2015.

564 Eichler, P., Muller, M., Rohmann, C., Stengel, B., Orasche, J., Zimmermann, R., and Wisthaler,
565 A.: Lubricating Oil as a Major Constituent of Ship Exhaust Particles, *Environ. Sci. Technol.*

566 Lett., 4, 54-58, 10.1021/acs.estlett.6b00488, 2017.

567 EPA: Estimation Program Interface (EPI) Suite (v4.11), US [code], 2012.

568 Francis, G. J., Milligan, D. B., and McEwan, M. J.: Gas-Phase Reactions and Rearrangements
569 of Alkyl Esters with H₃O⁺, NO⁺, and O₂^{•+}: A Selected Ion Flow Tube Study, *J. Phys. Chem.*
570 *A*, 111, 9670-9679, 10.1021/jp0731304, 2007.

571 Gkatzelis, G. I., Tillmann, R., Hohaus, T., Muller, M., Eichler, P., Xu, K. M., Schlag, P., Schmitt,
572 S. H., Wegener, R., Kaminski, M., Holzinger, R., Wisthaler, A., and Kiendler-Scharr, A.:
573 Comparison of three aerosol chemical characterization techniques utilizing PTR-ToF-MS: a
574 study on freshly formed and aged biogenic SOA, *Atmos. Meas. Tech.*, 11, 1481-1500, 2018a.

575 Gkatzelis, G. I., Hohaus, T., Tillmann, R., Gensch, I., Muller, M., Eichler, P., Xu, K. M., Schlag,
576 P., Schmitt, S. H., Yu, Z. J., Wegener, R., Kaminski, M., Holzinger, R., Wisthaler, A., and
577 Kiendler-Scharr, A.: Gas-to-particle partitioning of major biogenic oxidation products: a study
578 on freshly formed and aged biogenic SOA, *Atmos. Chem. Phys.*, 18, 12969-12989,
579 10.5194/acp-18-12969-2018, 2018b.

580 Gueneron, M., Erickson, M. H., VanderSchelden, G. S., and Jobson, B. T.: PTR-MS
581 fragmentation patterns of gasoline hydrocarbons, *International Journal of Mass Spectrometry*,
582 379, 97-109, 10.1016/j.ijms.2015.01.001, 2015.

583 He, X., Wang, Q., Huang, X. H. H., Huang, D. D., Zhou, M., Qiao, L., Zhu, S., Ma, Y.-g., Wang,
584 H.-l., Li, L., Huang, C., Xu, W., Worsnop, D. R., Goldstein, A. H., and Yu, J. Z.: Hourly
585 measurements of organic molecular markers in urban Shanghai, China: Observation of
586 enhanced formation of secondary organic aerosol during particulate matter episodic periods,
587 *Atmos. Environ.*, 240, 10.1016/j.atmosenv.2020.117807, 2020.

588 Heald, C. L. and Kroll, J. H.: The fuel of atmospheric chemistry: Toward a complete description
589 of reactive organic carbon, *Sci. Adv.*, 6, 10.1126/sciadv.aay8967, 2020.

590 Holzinger, R., Goldstein, A. H., Hayes, P. L., Jimenez, J. L., and Timkovsky, J.: Chemical
591 evolution of organic aerosol in Los Angeles during the CalNex 2010 study, *Atmos. Chem. Phys.*,
592 13, 10125-10141, 10.5194/acp-13-10125-2013, 2013.

593 Holzinger, R., Williams, J., Herrmann, F., Lelieveld, J., Donahue, N. M., and Rockmann, T.:
594 Aerosol analysis using a Thermal-Desorption Proton-Transfer-Reaction Mass Spectrometer

595 (TD-PTR-MS): a new approach to study processing of organic aerosols, *Atmos. Chem. Phys.*,
596 10, 2257-2267, 10.5194/acp-10-2257-2010, 2010.

597 Holzinger, R., Acton, W. J. F., Bloss, W. J., Breitenlechner, M., Crilley, L. R., Dusanter, S.,
598 Gonin, M., Gros, V., Keutsch, F. N., Kiendler-Scharr, A., Kramer, L. J., Krechmer, J. E.,
599 Languille, B., Locoge, N., Lopez-Hilfiker, F., Materić, D., Moreno, S., Nemitz, E., Quéléver,
600 L. L. J., Sarda Esteve, R., Sauvage, S., Schallhart, S., Sommariva, R., Tillmann, R., Wedel, S.,
601 Worton, D. R., Xu, K., and Zaytsev, A.: Validity and limitations of simple reaction kinetics to
602 calculate concentrations of organic compounds from ion counts in PTR-MS, *Atmos. Meas.*
603 *Tech.*, 12, 6193-6208, 10.5194/amt-12-6193-2019, 2019.

604 Huang, D. D., Zhu, S., An, J., Wang, Q., Qiao, L., Zhou, M., He, X., Ma, Y., Sun, Y., Huang,
605 C., Yu, J. Z., and Zhang, Q.: Comparative Assessment of Cooking Emission Contributions to
606 Urban Organic Aerosol Using Online Molecular Tracers and Aerosol Mass Spectrometry
607 Measurements, *Environ. Sci. Technol.*, 10.1021/acs.est.1c03280, 2021.

608 Huang, G., Liu, Y., Shao, M., Li, Y., Chen, Q., Zheng, Y., Wu, Z., Liu, Y., Wu, Y., Hu, M., Li,
609 X., Lu, S., Wang, C., Liu, J., Zheng, M., and Zhu, T.: Potentially Important Contribution of
610 Gas-Phase Oxidation of Naphthalene and Methyl-naphthalene to Secondary Organic Aerosol
611 during Haze Events in Beijing, *Environ. Sci. Technol.*, 53, 1235-1244, 10.1021/acs.est.8b04523,
612 2019.

613 Inomata, S., Sato, K., Hirokawa, J., Sakamoto, Y., Tanimoto, H., Okumura, M., Tohno, S., and
614 Imamura, T.: Analysis of secondary organic aerosols from ozonolysis of isoprene by proton
615 transfer reaction mass spectrometry, *Atmos. Environ.*, 97, 397-405,
616 10.1016/j.atmosenv.2014.03.045, 2014.

617 Krechmer, J. E., Groessl, M., Zhang, X., Junninen, H., Massoli, P., Lambe, A. T., Kimmel, J.
618 R., Cubison, M. J., Graf, S., Lin, Y.-H., Budisulistiorini, S. H., Zhang, H., Surratt, J. D.,
619 Knochenmuss, R., Jayne, J. T., Worsnop, D. R., Jimenez, J.-L., and Canagaratna, M. R.: Ion
620 mobility spectrometry–mass spectrometry (IMS–MS) for on- and offline analysis of
621 atmospheric gas and aerosol species, *Atmos. Meas. Tech.*, 9, 3245-3262, 10.5194/amt-9-3245-
622 2016, 2016.

623 Le Breton, M., Wang, Y. J., Hallquist, A. M., Pathak, R. K., Zheng, J., Yang, Y. D., Shang, D.
624 J., Glasius, M., Bannan, T. J., Liu, Q. Y., Chan, C. K., Percival, C. J., Zhu, W. F., Lou, S. R.,
625 Topping, D., Wang, Y. C., Yu, J. Z., Lu, K. D., Guo, S., Hu, M., and Hallquist, M.: Online gas-

626 and particle-phase measurements of organosulfates, organosulfonates and nitrooxy
627 organosulfates in Beijing utilizing a FIGAERO ToF-CIMS, *Atmos. Chem. Phys.*, 18, 10355-
628 10371, 2018.

629 Lee, B. H., Lopez-Hilfiker, F. D., D'Ambro, E. L., Zhou, P., Boy, M., Petäjä, T., Hao, L.,
630 Virtanen, A., and Thornton, J. A.: Semi-volatile and highly oxygenated gaseous and particulate
631 organic compounds observed above a boreal forest canopy, *Atmos. Chem. Phys.*, 18, 11547-
632 11562, 10.5194/acp-18-11547-2018, 2018.

633 Leglise, J., Muller, M., Piel, F., Otto, T., and Wisthaler, A.: Bulk Organic Aerosol Analysis by
634 Proton-Transfer-Reaction Mass Spectrometry: An Improved Methodology for the
635 Determination of Total Organic Mass, O:C and H:C Elemental Ratios, and the Average
636 Molecular Formula, *Anal. Chem.*, 91, 12619-12624, 10.1021/acs.analchem.9b02949, 2019.

637 Lindinger, W., Hansel, A., and Jordan, A.: Proton-transfer-reaction mass spectrometry (PTR-
638 MS): on-line monitoring of volatile organic compounds at pptv levels, *Chem. Soc. Rev.*, 27,
639 347-354, 10.1039/a827347z, 1998.

640 Lopez-Hilfiker, F. D., Mohr, C., Ehn, M., Rubach, F., Kleist, E., Wildt, J., Mentel, T. F., Lutz,
641 A., Hallquist, M., Worsnop, D., and Thornton, J. A.: A novel method for online analysis of gas
642 and particle composition: description and evaluation of a Filter Inlet for Gases and AEROSols
643 (FIGAERO), *Atmos. Meas. Tech.*, 7, 983-1001, 10.5194/amt-7-983-2014, 2014.

644 Lopez-Hilfiker, F. D., Mohr, C., Ehn, M., Rubach, F., Kleist, E., Wildt, J., Mentel, T. F.,
645 Carrasquillo, A. J., Daumit, K. E., Hunter, J. F., Kroll, J. H., Worsnop, D. R., and Thornton, J.
646 A.: Phase partitioning and volatility of secondary organic aerosol components formed from α -
647 pinene ozonolysis and OH oxidation: the importance of accretion products and other low
648 volatility compounds, *Atmos. Chem. Phys.*, 15, 7765-7776, 10.5194/acp-15-7765-2015, 2015.

649 Lopez-Hilfiker, F. D., Mohr, C., D'Ambro, E. L., Lutz, A., Riedel, T. P., Gaston, C. J., Iyer, S.,
650 Zhang, Z., Gold, A., Surratt, J. D., Lee, B. H., Kurten, T., Hu, W. W., Jimenez, J., Hallquist, M.,
651 and Thornton, J. A.: Molecular Composition and Volatility of Organic Aerosol in the
652 Southeastern U.S.: Implications for IEPOX Derived SOA, *Environ. Sci. Technol.*, 50, 2200-
653 2209, 10.1021/acs.est.5b04769, 2016.

654 Lutz, A., Mohr, C., Le Breton, M., Lopez-Hilfiker, F. D., Priestley, M., Thornton, J. A., and
655 Hallquist, M.: Gas to Particle Partitioning of Organic Acids in the Boreal Atmosphere, *ACS*

656 Earth Space Chem., 3, 1279-1287, 10.1021/acsearthspacechem.9b00041, 2019.

657 Mai, H., Shiraiwa, M., Flagan, R. C., and Seinfeld, J. H.: Under What Conditions Can
658 Equilibrium Gas-Particle Partitioning Be Expected to Hold in the Atmosphere?, Environ. Sci.
659 Technol., 49, 11485-11491, 10.1021/acs.est.5b02587, 2015.

660 Muller, M., Eichler, P., D'Anna, B., Tan, W., and Wisthaler, A.: Direct Sampling and Analysis
661 of Atmospheric Particulate Organic Matter by Proton-Transfer-Reaction Mass Spectrometry,
662 Anal. Chem., 89, 10889-10897, 10.1021/acs.analchem.7b02582, 2017.

663 Pagonis, D., Sekimoto, K., and de Gouw, J.: A Library of Proton-Transfer Reactions of H₃O⁺
664 Ions Used for Trace Gas Detection, Journal of the American Society for Mass Spectrometry, 30,
665 1330-1335, 10.1007/s13361-019-02209-3, 2019.

666 Palm, B. B., Peng, Q. Y., Fredrickson, C. D., Lee, B., Garofalo, L. A., Pothier, M. A.,
667 Kreidenweis, S. M., Farmer, D. K., Pokhrel, R. P., Shen, Y. J., Murphy, S. M., Permar, W., Hu,
668 L., Campos, T. L., Hall, S. R., Ullmann, K., Zhang, X., Flocke, F., Fischer, E. V., and Thornton,
669 J. A.: Quantification of organic aerosol and brown carbon evolution in fresh wildfire plumes,
670 Proc Natl Acad Sci U S A, 117, 29469-29477, 10.1073/pnas.2012218117, 2020.

671 Pankow, J. F.: An absorption-model of the gas aerosol partitioning involved in the formation of
672 secondary organic aerosol, Atmos. Environ., 28, 189-193, 10.1016/1352-2310(94)90094-9,
673 1994.

674 Pankow, J. F. and Asher, W. E.: SIMPOL.1: a simple group contribution method for predicting
675 vapor pressures and enthalpies of vaporization of multifunctional organic compounds, Atmos.
676 Chem. Phys., 8, 2773-2796, 10.5194/acp-8-2773-2008, 2008.

677 Peng, Y., Wang, H., Wang, Q., Jing, S., An, J., Gao, Y., Huang, C., Yan, R., Dai, H., Cheng, T.,
678 Zhang, Q., Li, M., Hu, J., Shi, Z., Li, L., Lou, S., Tao, S., Hu, Q., Lu, J., and Chen, C.:
679 Observation-based sources evolution of non-methane hydrocarbons (NMHCs) in a megacity of
680 China, J. Environ. Sci., 124, 794-805, 10.1016/j.jes.2022.01.040, 2023.

681 Perraud, V., Bruns, E. A., Ezell, M. J., Johnson, S. N., Yu, Y., Alexander, M. L., Zelenyuk, A.,
682 Imre, D., Chang, W. L., Dabdub, D., Pankow, J. F., and Finlayson-Pitts, B. J.: Nonequilibrium
683 atmospheric secondary organic aerosol formation and growth, Proc Natl Acad Sci U S A, 109,
684 2836-2841, 10.1073/pnas.1119909109, 2012.

685 Piel, F., Muller, M., Mikoviny, T., Pusede, S. E., and Wisthaler, A.: Airborne measurements of
686 particulate organic matter by proton-transfer-reaction mass spectrometry (PTR-MS): a pilot
687 study, *Atmos. Meas. Tech.*, 12, 5947-5958, 10.5194/amt-12-5947-2019, 2019.

688 Piel, F., Müller, M., Winkler, K., Skytte af Sättra, J., and Wisthaler, A.: Introducing the extended
689 volatility range proton-transfer-reaction mass spectrometer (EVR PTR-MS), *Atmos. Meas.*
690 *Tech.*, 14, 1355-1363, 10.5194/amt-14-1355-2021, 2021.

691 Sekimoto, K., Li, S. M., Yuan, B., Koss, A., Coggon, M., Warneke, C., and de Gouw, J.:
692 Calculation of the sensitivity of proton-transfer-reaction mass spectrometry (PTR-MS) for
693 organic trace gases using molecular properties, *International Journal of Mass Spectrometry*, 421,
694 71-94, 10.1016/j.ijms.2017.04.006, 2017.

695 Shiraiwa, M. and Seinfeld, J. H.: Equilibration timescale of atmospheric secondary organic
696 aerosol partitioning, *Geophysical Research Letters*, 39, n/a-n/a, 10.1029/2012gl054008, 2012.

697 Shiraiwa, M., Zuend, A., Bertram, A. K., and Seinfeld, J. H.: Gas-particle partitioning of
698 atmospheric aerosols: interplay of physical state, non-ideal mixing and morphology, *Phys.*
699 *Chem. Chem. Phys.*, 15, 11441-11453, 10.1039/c3cp51595h, 2013.

700 Spanel, P. and Smith, D.: SIFT studies of the reactions of H₃O⁺, NO⁺ and O₂⁺ with a series
701 of alcohols, *Int. J. Mass Spectrom. Ion Processes*, 167-168, 375-388,
702 [https://doi.org/10.1016/S0168-1176\(97\)00085-2](https://doi.org/10.1016/S0168-1176(97)00085-2), 1997.

703 Spanel, P., Ji, Y. F., and Smith, D.: SIFT studies of the reactions of H₃O⁺, NO⁺ and O₂⁽⁺⁾
704 with a series of aldehydes and ketones, *International Journal of Mass Spectrometry*, 165, 25-
705 37, 10.1016/s0168-1176(97)00166-3, 1997.

706 Stark, H., Yatavelli, R. L. N., Thompson, S. L., Kang, H., Krechmer, J. E., Kimmel, J. R., Palm,
707 B. B., Hu, W. W., Hayes, P. L., Day, D. A., Campuzano-Jost, P., Canagaratna, M. R., Jayne, J.
708 T., Worsnop, D. R., and Jimenez, J. L.: Impact of Thermal Decomposition on Thermal
709 Desorption Instruments: Advantage of Thermogram Analysis for Quantifying Volatility
710 Distributions of Organic Species, *Environ. Sci. Technol.*, 51, 8491-8500,
711 10.1021/acs.est.7b00160, 2017.

712 Tan, W., Zhu, L., Mikoviny, T., Nielsen, C. J., Wisthaler, A., Eichler, P., Muller, M., D'Anna,
713 B., Farren, N. J., Hamilton, J. F., Pettersson, J. B. C., Hallquist, M., Antonsen, S., and Stenstrom,
714 Y.: Theoretical and Experimental Study on the Reaction of tert-Butylamine with OH Radicals

715 in the Atmosphere, *J. Phys. Chem. A*, 122, 4470-4480, 10.1021/acs.jpca.8b01862, 2018.

716 Tani, A., Hayward, S., and Hewitta, C. N.: Measurement of monoterpenes and related
717 compounds by proton transfer reaction-mass spectrometry (PTR-MS), *International Journal of*
718 *Mass Spectrometry*, 223, 561-578, 10.1016/s1387-3806(02)00880-1, 2003.

719 Thompson, S. L., Yatavelli, R. L. N., Stark, H., Kimmel, J. R., Krechmer, J. E., Day, D. A., Hu,
720 W., Isaacman-VanWertz, G., Yee, L., Goldstein, A. H., Khan, M. A. H., Holzinger, R., Kreisberg,
721 N., Lopez-Hilfiker, F. D., Mohr, C., Thornton, J. A., Jayne, J. T., Canagaratna, M., Worsnop, D.
722 R., and Jimenez, J. L.: Field intercomparison of the gas/particle partitioning of oxygenated
723 organics during the Southern Oxidant and Aerosol Study (SOAS) in 2013, *Aerosol Sci. Technol.*,
724 51, 30-56, 10.1080/02786826.2016.1254719, 2016.

725 Veres, P., Roberts, J. M., Warneke, C., Welsh-Bon, D., Zahniser, M., Herndon, S., Fall, R., and
726 de Gouw, J.: Development of negative-ion proton-transfer chemical-ionization mass
727 spectrometry (NI-PT-CIMS) for the measurement of gas-phase organic acids in the atmosphere,
728 *International Journal of Mass Spectrometry*, 274, 48-55, 10.1016/j.ijms.2008.04.032, 2008.

729 Voliotis, A., Wang, Y., Shao, Y., Du, M., Bannan, T. J., Percival, C. J., Pandis, S. N., Alfarra, M.
730 R., and McFiggans, G.: Exploring the composition and volatility of secondary organic aerosols
731 in mixed anthropogenic and biogenic precursor systems, *Atmos. Chem. Phys.*, 21, 14251-14273,
732 10.5194/acp-21-14251-2021, 2021.

733 Wang, H. L., Lou, S. R., Huang, C., Qiao, L. P., Tang, X. B., Chen, C. H., Zeng, L. M., Wang,
734 Q., Zhou, M., Lu, S. H., and Yu, X. N.: Source Profiles of Volatile Organic Compounds from
735 Biomass Burning in Yangtze River Delta, China, *Aerosol Air Qual. Res.*, 14, 818-828,
736 10.4209/aaqr.2013.05.0174, 2014.

737 Wang, M. Y., Chen, D. X., Xiao, M., Ye, Q., Stolzenburg, D., Hofbauer, V., Ye, P. L., Vogel, A.
738 L., Mauldin, R. L., Amorim, A., Baccharini, A., Baumgartner, B., Brilke, S., Dada, L., Dias, A.,
739 Duplissy, J., Finkenzeller, H., Garmash, O., He, X. C., Hoyle, C. R., Kim, C., Kvashnin, A.,
740 Lehtipalo, K., Fischer, L., Molteni, U., Petaja, T., Pospisilova, V., Quelever, L. L. J., Rissanen,
741 M., Simon, M., Tauber, C., Tome, A., Wagner, A. C., Weitz, L., Volkamer, R., Winkler, P. M.,
742 Kirkby, J., Worsnop, D. R., Kulmala, M., Baltensperger, U., Dommen, J., El Haddad, I., and
743 Donahue, N. M.: Photo-oxidation of Aromatic Hydrocarbons Produces Low-Volatility Organic
744 Compounds, *Environ. Sci. Technol.*, 54, 7911-7921, 2020a.

745 Wang, Q. Q., He, X., Zhou, M., Huang, D. D., Qiao, L. P., Zhu, S. H., Ma, Y. G., Wang, H. L.,
746 Li, L., Huang, C., Huang, X. H. H., Xu, W., Worsnop, D., Goldstein, A. H., Guo, H., and Yu, J.
747 Z.: Hourly Measurements of Organic Molecular Markers in Urban Shanghai, China: Primary
748 Organic Aerosol Source Identification and Observation of Cooking Aerosol Aging, *ACS Earth*
749 *Space Chem.*, 4, 1670-1685, 2020b.

750 Yatavelli, R. L. N., Lopez-Hilfiker, F., Wargo, J. D., Kimmel, J. R., Cubison, M. J., Bertram, T.
751 H., Jimenez, J. L., Gonin, M., Worsnop, D. R., and Thornton, J. A.: A Chemical Ionization
752 High-Resolution Time-of-Flight Mass Spectrometer Coupled to a Micro Orifice Volatilization
753 Impactor (MOVI-HRToF-CIMS) for Analysis of Gas and Particle-Phase Organic Species,
754 *Aerosol Sci. Technol.*, 46, 1313-1327, 2012.

755 Ye, C., Yuan, B., Lin, Y., Wang, Z., Hu, W., Li, T., Chen, W., Wu, C., Wang, C., Huang, S., Qi,
756 J., Wang, B., Wang, C., Song, W., Wang, X., Zheng, E., Krechmer, J. E., Ye, P., Zhang, Z., Wang,
757 X., Worsnop, D. R., and Shao, M.: Chemical characterization of oxygenated organic
758 compounds in the gas phase and particle phase using iodide CIMS with FIGAERO in urban air,
759 *Atmos. Chem. Phys.*, 21, 8455-8478, 10.5194/acp-21-8455-2021, 2021.

760 Zhang, X. and Seinfeld, J. H.: A functional group oxidation model (FGOM) for SOA formation
761 and aging, *Atmos. Chem. Phys.*, 13, 5907-5926, 10.5194/acp-13-5907-2013, 2013.

762 Zhang, X., Dalleska, N. F., Huang, D. D., Bates, K. H., Sorooshian, A., Flagan, R. C., and
763 Seinfeld, J. H.: Time-resolved molecular characterization of organic aerosols by PILS plus
764 UPLC/ESI-Q-TOFMS, *Atmos. Environ.*, 130, 180-189, 10.1016/j.atmosenv.2015.08.049,
765 2016a.

766 Zhang, X., Zhang, H. F., Xu, W., Wu, X. K., Tyndall, G. S., Orlando, J. J., Jayne, J. T., Worsnop,
767 D. R., and Canagaratna, M. R.: Molecular characterization of alkyl nitrates in atmospheric
768 aerosols by ion mobility mass spectrometry, *Atmos. Meas. Tech.*, 12, 5535-5545, 10.5194/amt-
769 12-5535-2019, 2019.

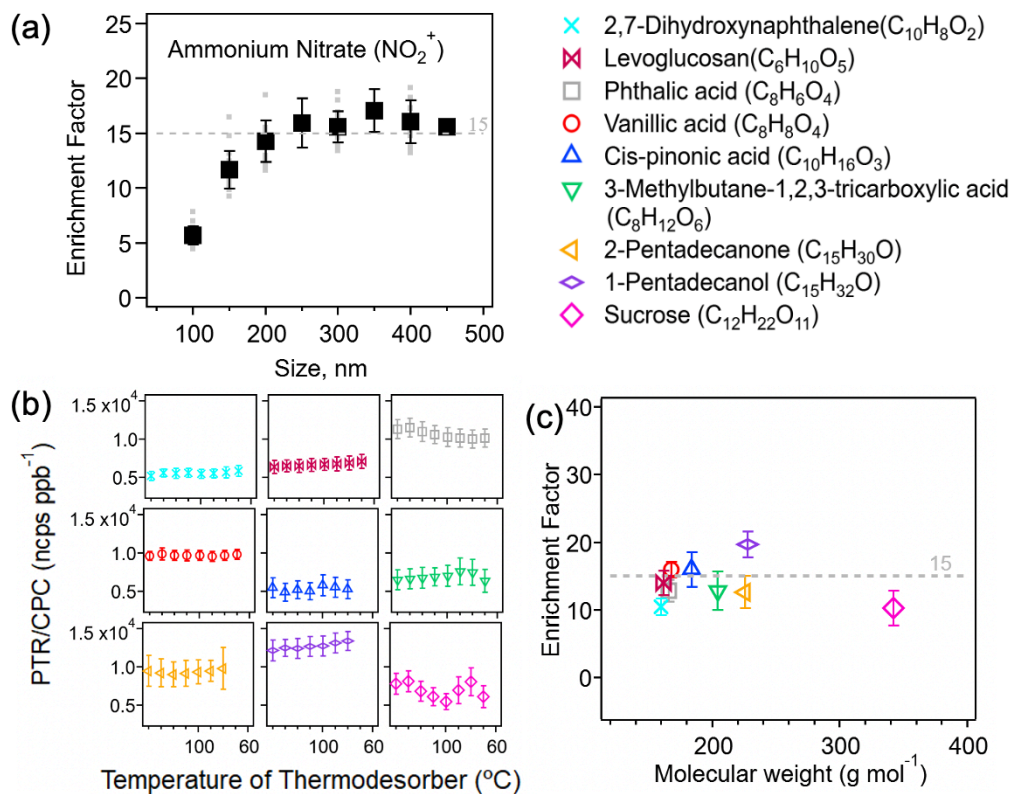
770 Zhang, X., Krechmer, J. E., Groessl, M., Xu, W., Graf, S., Cubison, M., Jayne, J. T., Jimenez,
771 J. L., Worsnop, D. R., and Canagaratna, M. R.: A novel framework for molecular
772 characterization of atmospherically relevant organic compounds based on collision cross
773 section and mass-to-charge ratio, *Atmos. Chem. Phys.*, 16, 12945-12959, 10.5194/acp-16-
774 12945-2016, 2016b.

775 Zhao, Y. L., Kreisberg, N. M., Worton, D. R., Isaacman, G., Weber, R. J., Liu, S., Day, D. A.,
776 Russell, L. M., Markovic, M. Z., VandenBoer, T. C., Murphy, J. G., Hering, S. V., and Goldstein,
777 A. H.: Insights into Secondary Organic Aerosol Formation Mechanisms from Measured
778 Gas/Particle Partitioning of Specific Organic Tracer Compounds, *Environ. Sci. Technol.*, 47,
779 3781-3787, 2013.

780 Zhu, H., Wang, H., Jing, S., Wang, Y., Cheng, T., Tao, S., Lou, S., Qiao, L., Li, L., and Chen,
781 J.: Characteristics and sources of atmospheric volatile organic compounds (VOCs) along the
782 mid-lower Yangtze River in China, *Atmos. Environ.*, 190, 232-240,
783 10.1016/j.atmosenv.2018.07.026, 2018.

784 Zhu, S., Wang, Q., Qiao, L., Zhou, M., Wang, S., Lou, S., Huang, D., Wang, Q., Jing, S., Wang,
785 H., Chen, C., Huang, C., and Yu, J. Z.: Tracer-based characterization of source variations of
786 PM_{2.5} and organic carbon in Shanghai influenced by the COVID-19 lockdown, *Faraday*
787 *Discuss*, 226, 112-137, 10.1039/d0fd00091d, 2021.

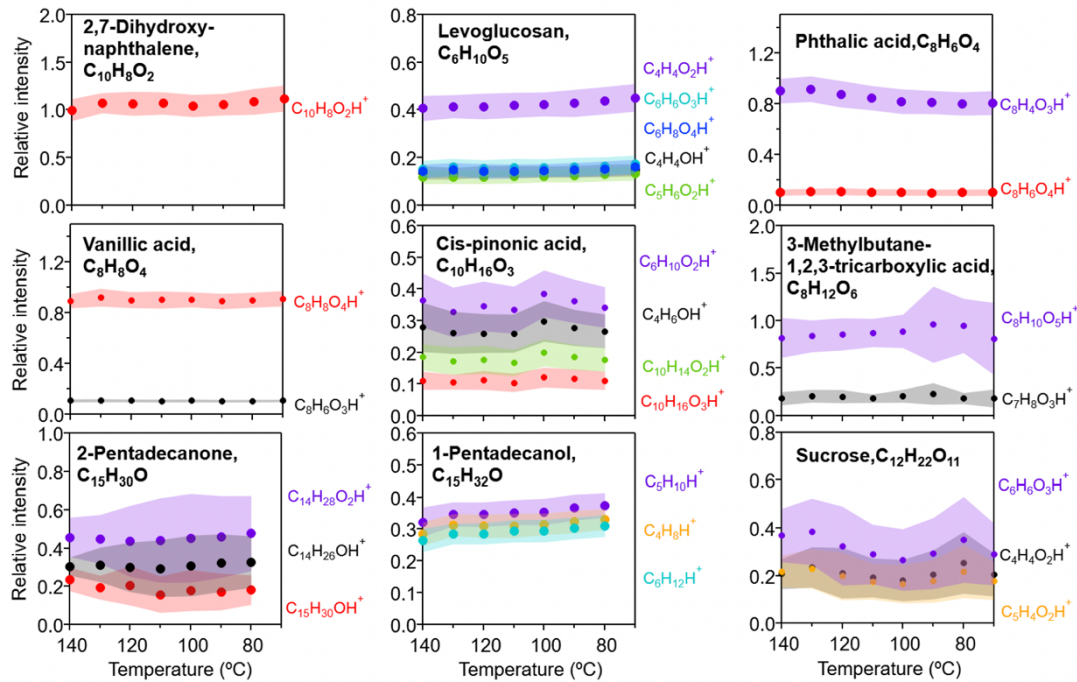
788



790

791 Figure 1. (a) Measured unitless enrichment factor (EF) of ammonium nitrate particles as a
 792 function of particle size in the 100 – 450 nm range. Grey markers represent all replicating
 793 measurements. The error bar denotes one standard deviation (1σ) of the average. (b) Ratios of
 794 PTR-ToF-MS signals (including both parent and fragment ions) to CPC counts ($\pm 1\sigma$) at 300
 795 nm for all organic standards studied. (c) EF ($\pm 1\sigma$) of selected organic standards based on the
 796 calculated sensitivity.

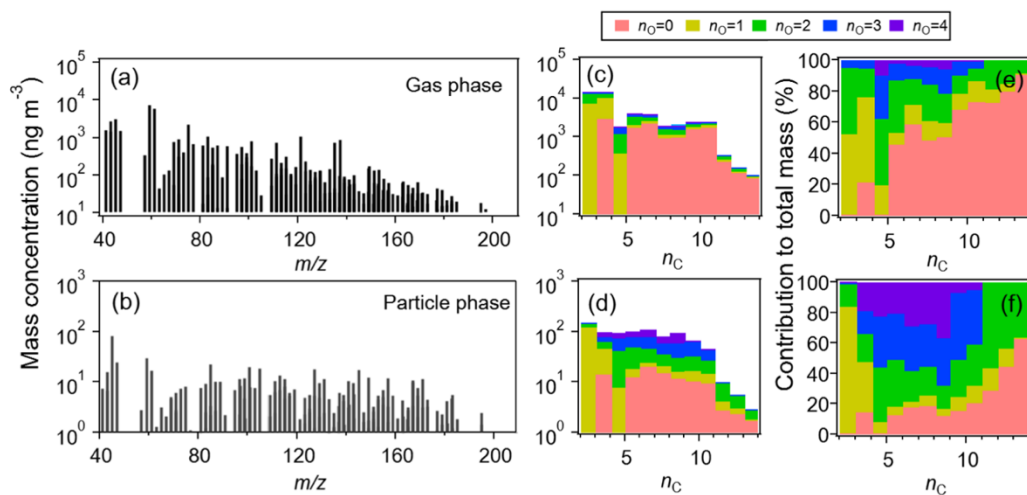
797



798

799 Figure 2. Ratios of CHARON-PTR-ToF-MS signals (ncps) to CPC measurements (ppb) of all
 800 detected ions (including both parents and fragments) for a given pure organic standard tested
 801 under different desorption temperatures (70 – 140 °C), normalized to the corresponding ratios
 802 obtained at 140 °C. Ions with relative intensities less than 10% are excluded. Red markers
 803 represent the parent peaks. Colored shades represent the relative standard deviations at different
 804 temperatures (exact values are given in Table S4).

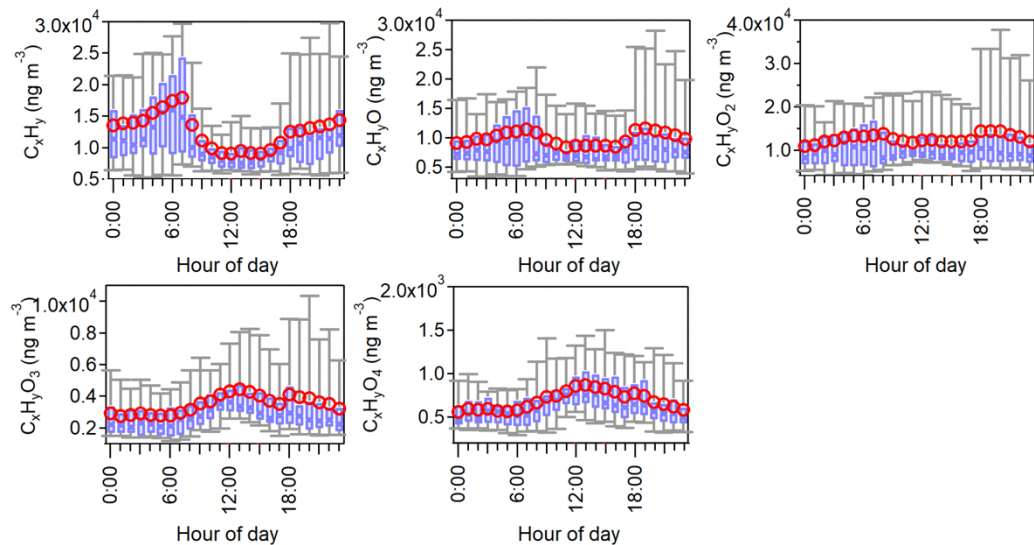
805



806

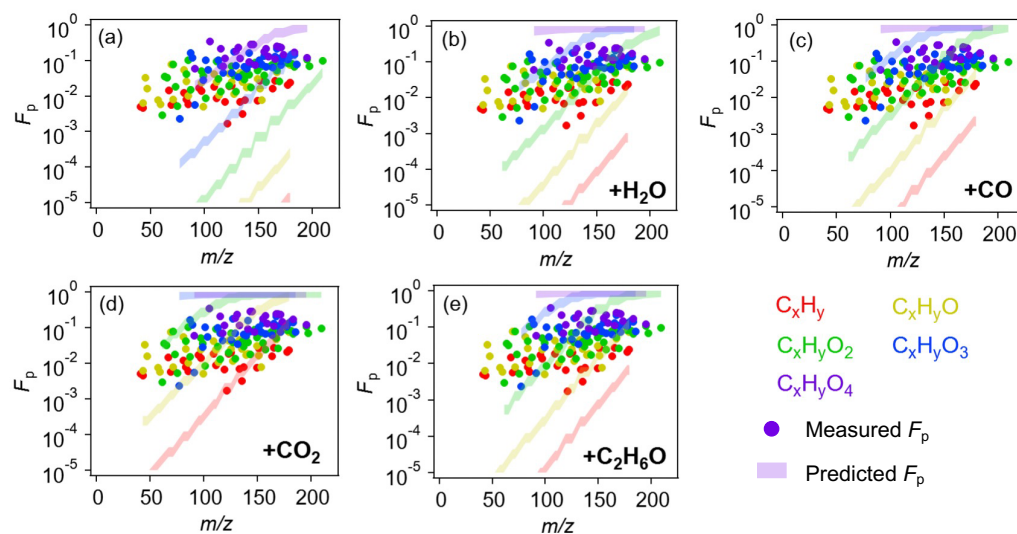
807 Figure 3. Background subtracted monthly-average PTR-ToF-MS mass spectra in the (a) gas
 808 phase and (b) particle phase. Mass distributions of all identified species resolved by the carbon
 809 and oxygen numbers (n_C and n_O) in the (c) gas phase and (d) particle phase, as well as their
 810 relative contribution to the total organic mass in the (e) gas phase and (f) particle phase.

811



812

813 Figure 4. Diurnal variations of observed gas-phase species with a generic formula of $C_{2-13}H_{2-22}O_{0-4}$. Hourly average values ($ng\ m^{-3}$), together with 10th, 25th, 75th, and 90th percentiles, are
 814 also plotted.
 815
 816



817
 818 Figure 5. (a) Campaign average fraction of organic species in the particle phase (F_p) grouped
 819 by the oxygen number. Solid markers represent the calculated F_p based on the CHARON-PTR-
 820 ToF-MS measurements. Colored shades represent the predicted F_p of corresponding molecular
 821 formulas. (b-e) Measured vs. predicted F_p assuming the identified species are fragments of
 822 corresponding parent compounds through neutral losses of H_2O , CO , CO_2 , and C_2H_6O ,
 823 respectively (values are given in Table S5).
 824