

RC2:

This study presents a low-cost multi-parameter air quality monitoring system (LCS) that incorporates diverse machine learning algorithms. While the utilization of GA-BP techniques is emphasized, the paper falls short in clearly elucidating its novelty, thereby preventing a comprehensive understanding of its unique contribution. Additionally, the presence of several structural weaknesses within the paper necessitates significant revisions to enhance its coherence and overall quality.

Specific Comments:

Introduction

(1) The introduction would benefit from a more comprehensive review of recent literature. Additionally, the presence of unclear terms, such as "multi-dimensional multi-response" on line 16, requires clarification to ensure a precise and unambiguous understanding. In addition, the rationale behind the selection of the five specific algorithms used in the study remains unclear. Providing a clear justification for the choice of these algorithms would enhance the understanding of the research methodology and its relevance to the study's objectives.

The development along with increased population and urbanization brings disadvantages, such as decreasing air quality and impact on public and individual health (Khreis et al., 2022; Manisalidis et al., 2020; Singh et al., 2021). Among the atmospheric pollutants, the primary pollutant is fine particulate matter, which affects the respiratory system and cardiac activity of humans. The secondary pollutants are SO₂, CO, NO_x, and O₃, which also induce disease or chronic poisoning. To improve the understanding of air pollution exposure and predict future air quality trends (Zimmerman et al., 2018), air quality assessment and forecasting are the essentials. The conventional air quality monitoring instrumentations are high cost, which has limited the spatial coverage of the monitoring stations (Zimmerman et al., 2018). The development and applications of the low-cost commercially available sensor-based air quality monitoring system (LCS) would considerably reduce both installation and maintenance costs (Spinelle et al., 2017). The larger spatial density of the air quality grid monitoring network becomes possible, which would play an important role in monitoring pollution trend, locating of pollution source, supporting environmental management (Zhao et al., 2019) and support better epidemiological models (Khreis et al., 2022; Zimmerman et al., 2018). These demands promote the LCS growing gradually (Cui et al., 2021; Wang et al., 2016).

The LCS typically utilizes the electrochemical or light scattering sensors for gas-phase or particulate pollutants measurement, such as sulfur dioxide (SO₂), nitrogen oxide (NO₂), carbon monoxide (CO), ozone (O₃) and particulate matters (PM). These electrochemical sensors have intrinsic problems, such as temperature or humidity impacts, and gaseous cross-sensitivities (Spinelle et al., 2015, 2017; Wan et al., 2016; Zimmerman et al., 2018). For example, limited by the poor selection performance, the NO₂ electrochemical sensor also undergo redox reactions in the presence of O₃ gaseous pollutants. The diffusion coefficient of the electrochemical sensor can be affected by temperature and relative humidity (Hitchman et al., 1997; Masson et al., 2015). The reagent of the electrochemical sensor is consumed over time, which affects the stability of the sensor. These features of the sensors have historically been poorly addressed by laboratory calibrations, limiting the utility for air quality monitoring (Zimmerman et al., 2018).

The de-convolving of cross-sensitivity effect and stability on sensor performance is

complex(Zimmerman et al., 2018). The linear or multivariate linear calibration models (Alexopoulos, 2010; Khreis et al., 2022; Zoest et al., 2019) have been developed. However the performance is poor on ambient data(Khreis et al., 2022). The accurate and precise calibration models for the low cost sensors are particularly critical to the success of dense sensor networks, as poor signal to noise ratios and cross-sensitivities hamper their ability to distinguish the pollutant concentrations. There has been increasing interest in multifarious algorithms for low-cost sensor calibration, and lots of studies using multi-input multi-output models(Alexopoulos, 2010) and neural networks(Spinelle et al., 2015) have been published. The artificial neural network (ANN) calibration model has the intelligence to process nonlinear data (Amuthadevi et al., 2021; Janabi et al., 2021), which has been used in calibration models for measuring ozone or nitrogen oxide(Esposito et al., 2016; Spinelle et al., 2015). For example, the ANN calibration model was used to calibrate O₃ and the uncertainty could meet the European data quality objectives; however, meeting these objectives for NO₂ remained a challenge(Spinelle et al., 2015). Dynamic neural network calibrations of NO₂ sensors were demonstrated with the mean absolute error less than 2 ppb; however, the performance for O₃ was not same(Esposito et al., 2016). High-dimensional multi-response model was used to calibrate CO, NO, NO₂, and O₃, with the 5 min average RMSE values of 39.2, 4.52, 4.56, and 9.71, respectively(Cross et al., 2017). Random-forest-based machine learning algorithm was used to improve the calibration strategies of low-cost sensors, with the mean absolute error values 38 ppb for CO, 10 ppm for CO₂, 3.5 ppb for NO₂, and 3.4 ppb for O₃, respectively(Zimmerman et al., 2018). Furthermore, multiple linear regression(Ionascu et al., 2021) based temperature and humidity correction and ANN based calibration shown the potential for significant further improvement for leave one out cross validation(Ali et al., 2021). With the 16 days process, the combined supervision calibration model was used to improve the R^2 of SO₂, NO₂ and O₃ by 75.8%, 38.6% and 4.7% to 0.58, 0.61, and 0.90, respectively(Cui et al., 2021). An integrated genetic programming dynamic neural network model was used to accurately estimate the carbon monoxide and nitrogen dioxide pollutant concentrations from the multi-sensor measurement data(Davut et al., 2022). A predictive model using multilayer perceptron, support vector regression, and linear regression was developed to analyze the CO₂ and particulate matter of in-vehicle, with the R^2 of 0.9981(Chew et al., 2021). The CNN, LSTM-CNN, and CNN-LSTM models were used to improve the prediction performance of the ozone by 3.58%, 1.68%, and 3.37%, respectively(Reza et al., 2023). However, these calibrations have only been tested utilizing fewer models with a short measurement period and small number of sensor matrix, each containing one sensor per pollutant (Cross et al., 2017; Esposito et al., 2016; Spinelle et al., 2015), not have been utilized to evaluate and predict the concentration values of multi pollutants simultaneously, such as PM_{2.5}, PM₁₀, SO₂, NO₂, CO and O₃.

The random-forest (RF)(Breiman, 2001; Liu et al., 2012), multivariate linear regression (MLR)(Alexopoulos, 2010), K Nearest Neighbor (KNN)(Zhao et al., 2021), BP neural network(Xu et al., 2021), and genetic-algorithm-back-propagation neural (GA-BP) network(Ning et al., 2019; Wang et al., 2019) are five commonly used machine learning algorithms with different characteristics. With the strong nonlinear mapping ability and adaptive ability, the RF, BP, and GA-BP are suitable for processing complex, high-dimensional, and nonlinear data with high prediction accuracy, such as the air quality monitoring. With the purpose to quantify the degree of influence of the independent variable, the MLR is suitable for evaluating the influence of multiple independent variables on the dependent variable, such as the cross-sensitivity effect between different factors.

The KNN is also a widely common algorithm to compare with RF, BP, GA-BP and MLR.

In this work, the LCS is developed to measure PM_{2.5}, PM₁₀, SO₂, NO₂, CO and O₃ simultaneously, and the performances of the calibration strategies based on the five machine learning algorithms are contrasted. Taking the original electronic signals of the sensors as input and measurements obtained by the reference instrumentations as output, five calibration strategies are applied and contrasted. The measurement is implemented in the real-world conditions almost a 12-month period (1 March 2021 and 28 February 2022) spanning multiple seasons and a wide range of meteorological conditions to ensure calibration model robustness. The performance of the different algorithms with the parameters, such as determination coefficient (R^2), root mean square error (RMSE) (Janabi et al., 2021), mean square error (MSE) and mean absolute error (MAE), are compared and discussed. The rest of this paper is organized as follows. The measurement setup is described in section 2. The principles of the calibration strategies are presented in section 3. The results and discussion are shown in section 4. The conclusion and discussion are drawn in section 5.

(2) Measurement setup

This section would greatly benefit from expansion to ensure a comprehensive understanding of the study. Specifically, there is a need for more clarity regarding the data collection process, including details on the quantity of data collected for each pollutant and any procedures employed for outlier removal.

Thanks for question.

The time taken for one set of data collection was one minute and repeated 4 times. The outlier data of the 4 sets of data was eliminated by using the Dixon principle. The remained data was used to get the mean values for each experiment. The values of the LCS and reference instruments were separately logged to the server with the interval of 5 minutes.

(3) Additionally, in Section 2.1, the inclusion of a map illustrating the data collection site would provide crucial contextual information.



Figure 1 Location of the air quality monitoring station during the measurement period

Thanks for question.

Measurements for gas-phase pollutants and particle pollutants were made continuously between 1 March 2021 and 28 February 2022, which were used as the start and end dates for the analyses. The

location, shown in Figure 1, was 30 Yaochang Street, Zhongyuan District, Zhengzhou City, Henan Province.

(4) In Section 2.2, it is essential to specify the precise names of the sensors used or provide access to datasheets, particularly for the Alphasense sensors. Clarifying whether the PM sensor used is named PM300S, for instance, would enhance the transparency of the study.

Thanks for your question. The LCS uses the commercially available particulate matter sensor (PM3006, Cubic sensor and Instrument Co., China) and electrochemical SO₂, NO₂, O₃, CO sensors (B4, Alphasense, UK), respectively.

(5) Moreover, the paragraph discussing laboratory tests requires expansion. Given the apparent linearity of the sensor response to concentrations, it is necessary to explicate the rationale behind testing non-linear methods. Exploring concentration curves at various temperatures and humidity levels would contribute to a more thorough analysis.

Thanks for your question. The rationale behind testing non-linear methods is as follows.

1) The linear or multivariate linear calibration models have been developed. However the performance is poor on ambient data, because the output voltages of the four type of gaseous sensors were nonlinearly fluctuated with the linearly increasing temperature and the relative humidity (RH) (Cui et al., 2021).

2) The cross-interference between the four types of gaseous sensors is another problem that needs to be addressed. The cross-interference effect is nonlinear.

The laboratory test is conducted with the purpose to check whether the sensors work effectively before installed into the LCS, calibrated with the different models and used in real-world conditions. The linearity of the gas sensors is tested under steadily increased concentration, which is from 0 - 5 mgm⁻³ for CO sensor, 0 - 0.2 mgm⁻³ for NO₂, 0 - 1.1 mgm⁻³ for O₃ and 0 - 1.4 mgm⁻³ for SO₂ with five more test points, shown in Figure 3. Since the units of outputs of the reference instruments and the sensors were different, the slope was not expected to be 1 (Cui et al., 2021). From the Figure 3, we can tell that the R² for the gas sensors were more than 0.93, which indicated that these gas sensors had good linear responses before calibration, and verified the sensor working properly and effectively and could be applied to the LCS.

(6) Lastly, directly citing the manufacturer of the reference monitors mentioned on line 23 of page 5, as well as providing information on the methodology employed for the weekly calibrations, would significantly strengthen the study's transparency.

Thanks for your question.

According to the technical specifications for operation and quality control of ambient air quality continuous automated monitoring system for SO₂, NO₂, O₃ and CO of China (China, 2018), and the technical guide for automatic monitoring by beta ray method for particulate matter in ambient air (PM₁₀ and PM_{2.5}) (China, 2020), the reference gas and particulate analyzers are checked and calibrated weekly and monthly, respectively.

(7) Calibration method

The equation (2) seems unclear; there might be a typographical error with X' instead of X.

Thanks for your question. The error is revised in the new version.

(8) Furthermore, in section 3.2, it would be beneficial to include additional statistics such as Mean Squared Error (MSE) or Mean Absolute Error (MAE) to provide a more comprehensive evaluation of the model's performance.

Thanks for your question. The revised version is as follows.

To quantitatively compare the performance of the five calibration models applied to the LCS, and balance the disadvantages of the different metrics, the determination coefficient (R^2), root mean square error (RMSE) (Janabi et al., 2021), mean square error (MSE) and mean absolute error (MAE) are utilized. The R^2 reflects the fit degree between the model output data and the reference monitor measurement. The measurement results should meet the requirements of environmental standards of China (Jiao et al., 2016). The RMSE measures how much error there is between the predicted values and the reference measurements, and is sensitive to extreme values (Chai & Draxler, 2014). The MAE is a disadvantage against RMSE and a good choice to evaluate the error when the distribution is not Gaussian (Reza, Behzad, & Gulen, 2023). The formula for each of the evaluation metrics are presented as equations (5)-(8), respectively.

$$R^2 = 1 - \frac{\sum_{i=1}^n (y_i - \hat{y}_i)^2}{\sum_{i=1}^n (y_i - \bar{y})^2}, \quad (5)$$

$$\text{RMSE} = \sqrt{\frac{1}{n} \sum_{i=1}^n (y_i - \hat{y}_i)^2}, \quad (6)$$

$$\text{MSE} = \frac{1}{n} \sum_{i=1}^n (y_i - \hat{y}_i)^2, \quad (7)$$

$$\text{MAE} = \frac{1}{n} \sum_{i=1}^n |y_i - \hat{y}_i| \quad (8)$$

Where \hat{y}_i , y_i and \bar{y} represent the i th model output data from the algorithm-based LCS system, the reference data from the reference instrumentations, and the mean value of the reference instrumentations, respectively. The n is the number of the measurement data in the dataset.

(9) Results and discussion

Paragraph 4.1 presents intriguing insights; however, it could benefit from a clearer presentation. For instance, the method of determining the number of trees in the random forest is not explicitly elucidated.

Thanks for your question. The revised version is as follows.

For the RF model, the number of trees is determined by using grid search method, which will search the optimal hyper-parameter by traversing a given hyper-parameter combination (Zhu, Zhu, Zhou, Zhu, & Zhang, 2022). A total of 11 kinds of tree numbers are set between 2 and 22. By using grid search to traverse these 11 kinds of tree numbers to obtain different R^2 .

(10) Additionally, while it is evident that a sub-period was chosen for testing, the rationale behind this selection remains unexplained. Clarifying these aspects would enhance the overall coherence and understanding of the paragraph.

Thanks for your question. For instance, the R^2 for different gas pollutants within 1 March 2021 and 30 June 2021 are shown in Figure 7. The R^2 is improved as the number of trees increasing. The rate of increase and the variation of R^2 is negligible beyond 20. The terminal node is specified using a maximum number of sub-node points per node. The R^2 is also improved as the number of sub-nodes increasing under the same tree number. The rate of increase and the variation of R^2 is negligible beyond 100. More number of the tree or the sub-node incur higher computational cost and time for the training and small performance improvement. Using this method, the same number of trees can be obtained with the different gas pollutants within 1 July 2021 and 31 October 2021, 1 November 2021 and 28 February 2022.

(11) In paragraph 4.2, including the size of each segment, as well as the reference temperature, humidity, and concentration range, would enhance the comprehensiveness of the experimental setup and contribute to a more detailed understanding of the study.

Thanks for your question. The revised version is as follows. The temperature and humidity results are also provided.

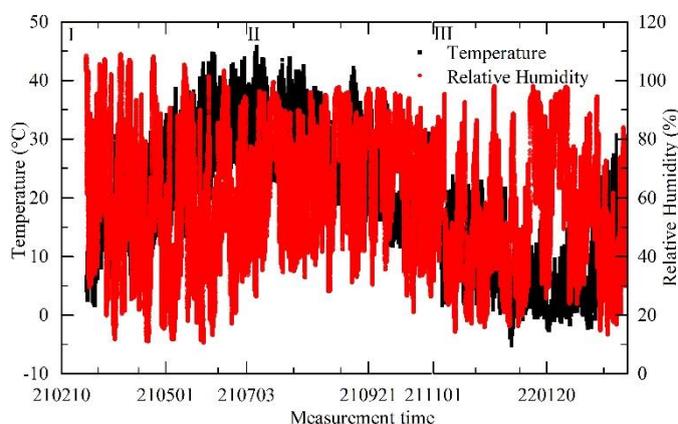


Figure 9 Temperature/relative humidity ranges during the measurement period

During the measurement period, the ranges of the ambient temperature and relative humidity separately were -5°C to 50°C and 10% to 98%, shown in Figure 9. The ambient temperature increased, decreased and fluctuated separately within 1 March 2021 and 30 June 2021, 1 July 2021 and 31 October 2021, 1 November 2021 and 28 February 2022, dividing the whole measurement period into three segments.

(12) In Figure 9 (b), including a normalized version of the Root Mean Square Error (RMSE) would be beneficial to enable an accurate comparison among the three periods. The same principle applies to Table 3; including a normalized version of the Root Mean Square Error (RMSE) would facilitate an accurate comparison among the different parameters.

Thanks for your question. The RMSE in table is obtained using equation (6). Where \hat{y}_i and y_i represent the i th model output data form the algorithm-based LCS system, and the reference data from the reference instrumentations, respectively. Thus, the result is the normalized version. The wrong expression is revised in the new version.

$$\text{RMSE} = \sqrt{\frac{1}{n} \sum_{i=1}^n (y_i - \hat{y}_i)^2}, \quad (6)$$

(13) Furthermore, the text mentions a division into train and test sets. It would be valuable to clarify whether a cross-validation was also conducted to ensure the robustness of the model evaluation.

Thanks for your question. To ensure the robustness of the model evaluation, the 5-fold cross validation is also conducted. The dataset is divided into 5 mutually exclusive subsets with same size, the 4 subset is randomly selected as the training set each time, and the remaining 1 subset is used as the test set. After completing each round of validation, the 4 copies are selected again to train the model and the remaining 1 copy is used for validation. After several rounds (less than 5), the loss function is selected to evaluate the optimal model and parameters (Mahesh et al., 2023; Zimmerman et al., 2018).

(14) These consideration are also valid for the results of gas measurements. Moreover it would be insightful to include a discussion of the varying results obtained for each segment. For instance, a better detailed analysis of why the performance of SO₂ is consistently good for period II but considerably poorer for the other periods would enrich the understanding of the data and provide valuable insights into the underlying factors influencing the results.

Thanks for your question. The revised version is as follows.

For the data of SO₂, the results of RF are better than the ones of other methods, and have little difference among the three periods. However, the performances of other methods (MLR, KNN, BP, GA-BP) are poorer for the first and third periods. There may be some reasons for this phenomenon. The first one is the cross interference effect from NO₂ and O₃, which have the wide range of fluctuations (from about 20 μgm^{-3} to 125 μgm^{-3}) and increasing tendency in period I, respectively. The NO₂ and SO₂ can react chemically under certain conditions to produce sulfuric acid (H₂SO₄) and nitric acid (HNO₃), which will affect the reading of SO₂ sensor. The O₃, highly oxidizing gas, may react with SO₂ to form H₂SO₄ or sulfite (H₂SO₃), resulting in inaccurate sensor readings. The second one is the ambient temperature has a wide range of fluctuations (from about minus 5 °C to plus 45 °C) during the first and third periods, which will affect the stability of electrode material and the readings of the sensor. The last one is the concentration of ambient SO₂ is high (more than 30 μgm^{-3}) in period I and period III, beyond the actual measurement range of the SO₂ sensor, which will be researched in future.

(15) Finally, there is a typo on line 6 of page 14 (dada instead of data).

Thanks for your question. The typo is revised in the new version with the red color.

(16) Conclusion

The conclusion paragraph would benefit from a more explicit discussion on the presence of a recommended algorithm for calibration and a thorough examination of its potential limitations. By addressing the challenges associated with generalizing black box models, notably random forests, the conclusion could provide a more nuanced understanding of the practical implications and constraints that may arise from the study's findings.

Thanks for your question. The conclusion is revised in the new version.

A low-cost air quality monitoring system (LCS) based on RF, MLR, KNN, BP, GA-BP algorithms were proposed. The system can measure gas-phase pollutants (SO_2 , NO_2 , CO and O_3) and particle pollutants ($\text{PM}_{2.5}$ and PM_{10}), simultaneously. With the purpose to estimate the performance of the five algorithms, the LCS was mounted at the same location (Zhengzhou City, China) and consistent height with the reference monitoring system. The measurement was made continuously from 1 March 2021 to 28 February 2022, with the ranges of the ambient temperature and relative humidity separately minus 5°C to plus 50°C and 10% to 98%. The values of the LCS and reference instruments were separately logged to the server for further comparative analysis.

With the pretreated and individual particle counters, T and RH as input, and the concentrations of $\text{PM}_{2.5}$ and PM_{10} measured by the reference instrumentation separately as output, the multi-input one-output evaluation models based on RF, MLR, KNN, BP, GA-BP algorithms can be obtained. With the four types of electro-chemical sensors raw data, T and RH as input, and the measurements from the reference monitors as output, the multi-input multi-output evaluation models based on the five algorithms can be obtained. The performance of the calibration models are quantitatively compared by utilizing R^2 , RMSE, MSE and MAE.

The experimental results show that the R^2 of RF for the PM is better than 0.98; the R^2 of MLR for the PM is less than 0.91; the R^2 of the other three model are within 0.86 and 0.98. The R^2 of RF for the gas pollutants (SO_2 , NO_2 , CO and O_3) is better than 0.93; the R^2 of KNN, BP and GA-BP for the gas pollutants (SO_2 , NO_2 , CO and O_3) is within 0.27 to 0.97; the R^2 of MLR for the NO_2 , CO and O_3 is within 0.46 to 0.90, but for SO_2 less than 0.40, and even less than 0.1.

The maximum RMSE values of $\text{PM}_{2.5}$, PM_{10} , O_3 , CO , NO_2 , and SO_2 between the reference data and the RF, MLR, KNN, BP, GA-BP-based algorithms data are 5.49, 18.68, 13.05, 14.35, and 14.35; 10.37, 45.05, 27.08, 23.10, and 23.65; 4.08, 17.79, 10.57, 14.67, and 14.40; 0.06, 0.23, 0.16, 0.18, and 0.18; 3.99, 14.54, 9.61, 11.07, and 11.21; 2.84, 28.80, 16.44, 21.39, and 21.16, respectively. The maximum MAE values of $\text{PM}_{2.5}$, PM_{10} , O_3 , CO , NO_2 , and SO_2 between the reference data and the RF, MLR, KNN, BP, GA-BP-based algorithms data are 3.45, 12.80, 8.31, 9.55, and 9.54; 5.28, 23.20, 13.35, 15.26, and 15.43; 2.88, 13.46, 7.33, 11.14, and 10.90; 0.05, 0.19, 0.11, 0.14, and 0.14; 2.80, 11.08, 6.85, 8.27, and 8.41; 1.16, 4.24, 2.84, 3.43, and 3.40, respectively.

It should be noted that the results of RF are better than the ones of other methods, have very good agreement with the reference monitors, and have little difference among the three periods. However, the performances of other methods (MLR, KNN, BP, GA-BP) have poor agreement, especially during the first and third periods, which are the winter and spring. There may be some reasons, such as the cross interference effect, the wide range of fluctuation of the climatic factors, and the limitation of the actual measurement range and precision.

Overall, we conclude that, with careful data management and calibration using the machine learning

algorithms, especially the RF method, these measurements are consistent with the national environmental protection standard requirement of China, the LCS may significantly improve our ability to spatial heterogeneity in air pollutant concentrations. The air pollutant maps will assist researchers, policymakers, and communities in developing new policies or mitigation strategies to enhance human health. In the next research, we will focus on improving the matching of the measurement precise and range, the generalization of the algorithms in more applications, and the performance of the SO₂ sensor.