



# 1 Long-term Evaluation of Commercial Air Quality Sensors: An 2 Overview from the QUANT Study

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18 **Abstract.** In times of growing concern about the impacts of air pollution across the globe, lower-cost sensor  
19 technology is giving the first steps in helping to enhance our understanding and ability to manage air quality issues.  
20 While the benefits of greater spatial coverage and real-time measurements that these systems offer are evident,  
21 challenges still need to be addressed regarding sensor reliability and data quality. Given the limitations imposed by  
22 intellectual property, commercial implementations are often "black boxes", which represents an extra challenge as it  
23 limits end-users' understanding of the data production process. In this paper we present an overview of the QUANT  
24 (Quantification of Utility of Atmospheric Network Technologies) study, a comprehensive 3-year assessment across a  
25 range of urban environments in the United Kingdom. QUANT stands out as one of the most comprehensive studies  
26 of commercial air quality sensor systems carried out to date, encompassing a wide variety of companies in a single  
27 evaluation and including two generations of sensor technologies. Integrated into an extensive data set open to the  
28 public, it was designed to provide a long-term evaluation of the precision, accuracy, and stability of commercially  
29 available sensor systems. This overview discusses the assessment methodology, and key findings showcasing the  
30 significance of the study. The results shown here highlight the significant variation between systems, the incidence of  
31 corrections made by manufacturers, the effects of relocation to different environments and the long-term behaviour  
32 of the systems. Additionally, the importance of accounting for uncertainties associated with reference instruments in  
33 sensor evaluations is emphasised. Practical considerations in the application of these sensors in real-world scenarios



34 are also discussed, and potential solutions to end-users data challenges are presented. Offering key information about  
35 the sensor systems' capabilities the QUANT study will serve as a valuable resource for those seeking to implement  
36 commercial solutions as complementary tools to tackle air pollution.

37 **Keywords:** air pollution, commercial sensor systems, QUANT, long-term evaluation.

## 38 1. Introduction

39 Emerging lower-cost sensor systems<sup>1</sup> offer a promising alternative to the more expensive and complex monitoring  
40 equipment traditionally used for measuring air pollutants such as PM<sub>2.5</sub>, NO<sub>2</sub>, and O<sub>3</sub> (Okure et al., 2022). These  
41 innovative devices hold the potential to expand spatial coverage (Malings et al., 2020) and deliver real-time air  
42 pollution measurements (Tanzer-Gruener et al., 2020). However, concerns regarding the variable quality of the data  
43 they provide still hinder their acceptance as reliable measurement technologies (Karagulian et al., 2019; Zamora et  
44 al., 2020).

45 Sensors face key challenges such as cross-sensitivities (Cross et al., 2017; Levy Zamora et al., 2022; Pang et al., 2018),  
46 internal consistency (Feenstra et al., 2019; Ripoll et al., 2019), signal drift (A. Miech et al., 2023; Li et al., 2021;  
47 Sayahi et al., 2019), long term performance (Bulot et al., 2019; Liu et al., 2020) and data coverage (Brown & Martin,  
48 2023; Duvall et al., 2021; Feinberg et al., 2018). Additionally, environmental factors such as temperature (Bittner et  
49 al., 2022; Farquhar et al., 2021), and humidity (Crilley et al., 2018; Williams, 2020) can significantly influence sensor  
50 signals.

51 In recent years, manufacturers of both sensing elements (Han et al., 2021; Nazemi et al., 2019) and sensor systems  
52 have made significant technological advances (Chojer et al., 2020). For example, there are now commercial and non-  
53 commercial systems equipped with multiple sensing elements to measure distinct pollutants (Buehler et al., 2021;  
54 Hagan et al., 2019; Pang et al., 2021) helping mitigating the effects of cross-interferences. Additionally, enhancements  
55 in electrochemical sensors have been demonstrated in terms of their specificity (Baron & Saffell, 2017; Ouyang,  
56 2020).

57 However, the complex nature of their responses, coupled with their dependence on local conditions means sensor  
58 performance can be inconsistent (Bi et al., 2020). This complicates the comparison of results or anticipating sensor  
59 future performance across different studies. Moreover, assessments of sensor performance found in the academic  
60 literature often rely on a range of protocols (e.g., CEN (2021) and Duvall et al. (2021)) and data quality metrics (e.g.,  
61 Spinelle et al. (2017) and Zimmerman et al. (2018)), with many studies limited to a single-site co-location and/or of  
62 short-term evaluations that do not fully account for broader environmental variations (Karagulian et al., 2019).

63 The calibration of any instrument used to measure atmospheric composition is fundamental to guarantee their accuracy  
64 (Alam et al., 2020; Long et al., 2021; Wu et al., 2022). Using out-of-the-box sensor data without fit-for-purpose  
65 calibration can produce misleading results (Liang & Daniels, 2022). An effective calibration involves identifying and  
66 correcting systematic errors in the sensor readings. For standard air pollution measurement techniques, calibration is

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<sup>1</sup> The term "Sensor Systems" refers to sensors housed within a protective case, which includes a sampling and power system, electronic hardware and software for data acquisition, analog-to-digital conversion, data processing and their transfer (Karagulian et al., 2019). Unless specified otherwise, the term "sensor" will be used to refer to "Sensor Systems".



67 often performed in a controlled laboratory environment (Liang, 2021), or by sampling gas from a certified standard  
68 cylinder in the field.

69 Yet, the aforementioned challenges with lower-cost sensor-based devices suggest that such calibrations may not  
70 always accurately reflect real-world conditions (Giordano et al., 2021). A frequent approach involves co-locating  
71 sensors alongside regulatory instruments in their intended deployment areas and/or conditions and using data-driven  
72 methods to match the reference data (Liang & Daniels, 2022). Numerous studies have investigated the effectiveness  
73 of calibration methods for sensors e.g. (Bigi et al., 2018; Bittner et al., 2022; Malings et al., 2020; Spinelle et al., 2017;  
74 Zimmerman et al., 2018), including selecting appropriate reference instruments (Kelly et al., 2017), the need for  
75 regular calibration to maintain accuracy (Gamboa et al., 2023), the necessity of rigorous calibration protocols to ensure  
76 consistency (Kang et al., 2022), and transferability (Nowack et al., 2021) of results. Ultimately, the reliability and  
77 associated uncertainty of any applied calibration will influence the final sensor data quality.

78 For end-users to make informed decisions on the applicability of air pollution sensors, a realistic understanding of the  
79 expected performance in their chosen application is necessary (Rai et al., 2017). Despite this, there has been relatively  
80 little progress in clarifying the performance of sensors for air pollution measurements outside of the academic arena.  
81 This is largely due to the significant variability in both the number of sensors and the variety of applications tested,  
82 as well as the availability of highly accurate measurement instrumentation and/or regulatory networks to those outside  
83 of the atmospheric measurement academic field (e.g. Lewis and Edwards (2016) and Popoola et al. (2018)). From a  
84 UK clean air perspective, this ambiguity represents a major problem. The lack of a consistent message undermines  
85 the exploitation of these devices' unique strengths, notably their capability to form spatially dense networks with rapid  
86 time resolution. Consequently, there is potential for a mismatch in users' expectations of what sensor systems can  
87 deliver and their actual operating characteristics, eroding trust and reliability.

88 In this work, as part of the UK Clean Air program funded QUANT project, we deployed a variety of sensor  
89 technologies (43 commercial devices, 119 gas and 118 PM measurements) at 3 representative UK urban sites —  
90 Manchester, London and York— alongside extensive reference measurements, to generate the data for an extensive  
91 in-depth performance assessment. This project aims to not only evaluate the performance of sensor devices in a UK  
92 urban climatological context but also provide critical information for the successful application of these technologies  
93 in various environmental settings. To our knowledge, QUANT is the most extensive and longest-running evaluation  
94 of commercial sensor systems globally to date. Furthermore, we tested multiple manufacturers' data products for a  
95 significant number of these sensors to understand the implications of local calibration. This comprehensive approach  
96 offers unprecedented insights into the operational capabilities and limitations of these sensors in real-world conditions.  
97 Significantly, some of the insights gathered during QUANT have contributed to the development of the Publicly  
98 Available Specification (PAS 4023, 2023), which provides guidelines for the selection, deployment, maintenance, and  
99 quality assurance of air quality sensor systems.

100 In the following sections, we delve into the methodology and provide an overview of the QUANT dataset, as well as  
101 a discussion of some of the key findings and potential considerations for end-users.

## 102 **2. QUANT study design**

### 103 **2.1 Main study**



104 The Main QUANT assessment study aimed to perform a transparent long-term (19 Dec 2019 - 31 Oct 2022)  
 105 evaluation of commercially available sensor technologies for outdoor air pollution monitoring in UK urban  
 106 environments. Four duplicates of five different commercial sensor devices (Table 1) were purchased in Sept 2019  
 107 for inclusion in the study, with the selection criteria being: market penetration and/or previous performance  
 108 reported in the literature, ability to measure pollutants of interest (e.g. NO<sub>2</sub>, NO, O<sub>3</sub>, and PM<sub>2.5</sub>), and capacity to  
 109 run continuously reporting high time resolution data (1-15 min data) ideally in near real-time with data accessible  
 110 via an API.

111 **Table 1. Main QUANT devices description. The 20 units offered 56 gas and 56 PM measurements in total. For a detailed**  
 112 **description of the devices see Section S1 in the Supp.**

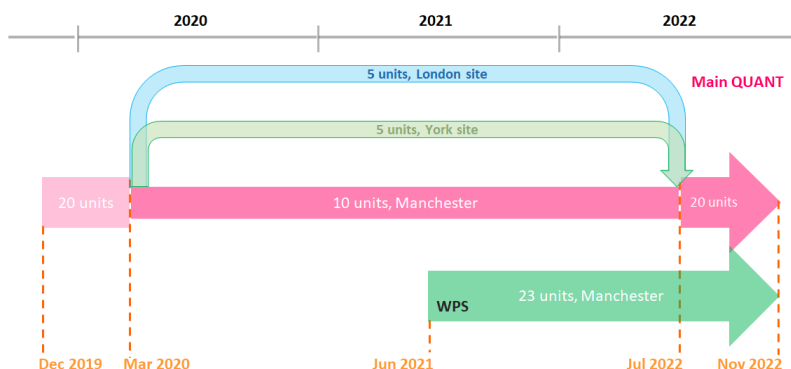
Product* (# units)	Company	Measurements								Cost (£)**
		NO	NO <sub>2</sub>	O <sub>3</sub>	CO	CO <sub>2</sub>	PM <sub>1</sub>	PM <sub>2.5</sub>	PM <sub>10</sub>	
AQY (4)	Aeroqual	-	✓	✓	-	-	-	✓	✓	~4.7K
AQM (4)	AQMesh	✓	✓	✓	-	✓	✓	✓	✓	~8.6K
Ari (4)	QuantAQ	✓	✓	✓	✓	✓	✓	✓	✓	~8.6K
PA (4)	PurpleAir	-	-	-	-	-	✓	✓	✓	~0.3K
Zep (4)	Earthsense	✓	✓	✓	-	-	✓	✓	✓	~7K

\*AQY: Aeroqual; AQM: AQMesh; Ari: Arisense; PA: PurpleAir; Zep: Zephyr.

\*\*Cost (Sep 2019) per unit including UK taxes and associated contractual costs (i.e., communication, data access, sensor replacement, etc.).

113 To capture the variability of UK urban environments, identical units were installed at three carefully selected field  
 114 sites. Two of these sites are highly instrumented urban background measurement supersites: the London Air  
 115 Quality Supersite (LAQS) and the Manchester Air Quality Supersite (MAQS), located in densely populated urban  
 116 areas with unique air quality challenges. The third site is a roadside monitoring site in York, which is part of the  
 117 Automatic Urban and Rural Network (AURN, <https://uk-air.defra.gov.uk/data/>), representing a urban environment  
 118 more influenced by traffic. This selection strategy ensures that the QUANT study's findings reflect the dynamics  
 119 of urban air quality across different UK settings, while providing comprehensive reference measurements. Further  
 120 details about each site can be found in Section S3 in the Supp., and the available reference instrumentation in  
 121 Section S4.

122 Initially, all the sensors were deployed in Manchester for approximately 3 months (mid-Dec 2019 to mid-Mar  
 123 2020) before being split up amongst the three sites (Fig. 1). At least one unit per brand was re-deployed to the  
 124 other two sites (mid-March 2020 to early-July 2022) leaving two devices per company in Manchester to assess  
 125 inter-device consistency. In the final 4 months of the study, all the sensor systems were relocated back to  
 126 Manchester (early July 2022 to the end of October 2022).



127

128 **Figure 1. Main Quant and Wider Participation Study (WPS) timeline.**

129 **2.2 Wider Participation Study**

130 The Wider Participation Study (WPS) was a no-cost complementary extension of the QUANT assessment,  
 131 specifically designed to foster innovation within the air pollution sensors domain. This segment of the study took  
 132 place entirely at the at MAQS from 10th June 2021 to 31st October 2022 (Fig. 1). It included a wider array of  
 133 commercial platforms (9 different sensor systems brands), and offered manufacturers the opportunity to engage in  
 134 a free-of-charge impartial evaluation process. Although participation criteria matched those of the Main QUANT  
 135 study, a key distinction lay in the voluntary nature of participation: vendors were invited to contribute multiple  
 136 sensor devices throughout the WPS study (see Table 2). Participants were able to demonstrate their systems’  
 137 performance against collocated high-resolution (1-minute) reference data at a state-of-the-art measurement site  
 138 such as the Manchester supersite.

139 **Table 2. The 23 WPS devices deployed at the Manchester supersite provided 63 gases and 62 PM measurements in total.**  
 140 **For a detailed description of the devices see the Section S2 in the Supp.**

Product* (# units)	Company	Measurements							
		NO	NO <sub>2</sub>	O <sub>3</sub>	CO	CO <sub>2</sub>	PM <sub>1</sub>	PM <sub>2.5</sub>	PM <sub>10</sub>
Mod (3)	QuantAQ	-	-	-	-	-	✓	✓	✓
AQM (3)	AQMesh	✓	✓	✓	✓	✓	✓	✓	✓
Atm (2)	RLS**	-	-	-	-	-	✓	✓	✓
IMB (2)	Bosch	-	✓	✓	-	-	-	✓	✓
Poll (2)	Oizom	✓	✓	✓	✓	✓	-	✓	✓
AP (3)	Kunak	✓	✓	✓	✓	✓	✓	✓	✓
SA (3)	Vortex IoT	-	✓	✓	-	-	-	✓	✓
NS (3)	Clarity	-	✓	-	-	-	✓	✓	✓
Prax (2)	SCS***	✓	✓	✓	✓	✓	✓	✓	✓

141

142 \*Mod: Modulair; AQM: AQMesh; Atm: Atmos, Polludrone; Poll; AP: Kunak Air Pro; SA: Silax Air, NS: Node-S, Prax:

143 Praxis. \*\*RLS: Respirer Living Sciences. \*\*\*SCS: South Coast Science.



### 144 **2.3 Data collection, co-located reference data and data products**

145 All sensor devices were installed at the measurement sites as per manufacturer recommendations. In addition to  
146 the device supplier's own cloud storage (accessed on-demand via each supplier's web portals), an automated daily  
147 scraping of each company's API was performed to save data onto a secure server at the University of York to  
148 ensure data integrity. PurpleAir units were exempt from this due to a lack of mobile data connection and poor  
149 internet signal at the sites; instead, readings were locally collected and manually uploaded. Minor pre-processing  
150 was applied at this stage to standardise the data format across all the devices. No outlier checks or data  
151 modifications were applied.

152 In addition to providing an independent assessment of sensor performance, QUANT also aimed to contribute to  
153 device manufacturers to help advance the field of air pollution sensors. To this end, three separate 1-month periods  
154 of reference data, spaced every 6 months, were shared with each supplier, provisional data soon after each period,  
155 and ratified data when available. For an overview of reference instrumentation at each site refer Table S1, and for  
156 details on the quality assurance procedures applied to the reference instruments see Table S2.

157 Access to colocated reference data allowed the companies to assess sensors' performance and, if they chose, to  
158 generate and provide additional calibrated data products. These products are distinct data versions provided by  
159 manufacturers throughout QUANT, before and/or after sharing reference data—for instance, “out-of-box”, “cal1”,  
160 “cal2”, etc. Figures S1 and S2 show a time-line of the different data products. To see the dates and periods of the  
161 shared reference data refer to Table S3. All reference data was embargoed until it was released to all manufacturers  
162 simultaneously to ensure consistency across manufacturers. Not every manufacturer opted to use this data to apply  
163 corrections or improve calibrations, but if they chose to do so, the updated measurements were treated as a separate  
164 data product. Device calibrations were performed solely at the discretion of the manufacturers without any  
165 intervention from our team, thus limiting the involvement of vendors/manufacturers in the provision of standard  
166 sensor outputs and unit maintenance as would be required by any standard customer.

### 167 **3. Results and discussion**

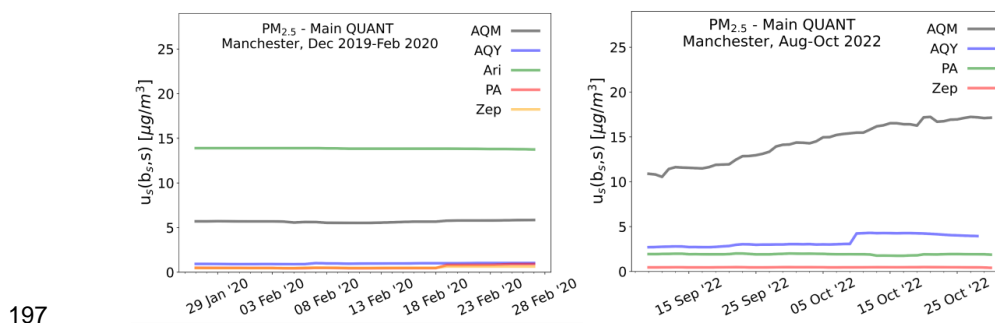
168 A key challenge in sensor performance evaluation is the high spatial and temporal variability errors that impact  
169 the accuracy of their readings, making the application of laboratory corrections more challenging. Furthermore,  
170 the overreliance on global performance metrics, such as  $R^2$  (i.e., the Coefficient of Determination), RMSE (i.e.,  
171 the Root Mean Squared Error), and MAE (i.e., the Mean Absolute Error) is an important issue when assessing  
172 sensors. While these metrics provide a general understanding of sensor performance, they can be limiting or even  
173 misleading, restricting a comprehensive understanding of the error structure and the measurement information  
174 content (Diez et al., 2022).

175 In response to these challenges, the QUANT assessment represents the most extensive independent appraisal of  
176 air pollution sensors in UK urban atmospheres. As the results presented here illustrate, QUANT is dedicated to  
177 examining sensor performance through multiple complementary perspectives and metrics, aiming to integrate  
178 these to accurately reflect the complexity of this dataset. By making the dataset open-access, it enables other  
179 stakeholders to evaluate it based on criteria that align with their specific needs and contexts. The following sections  
180 aim to provide an overview of the data and provide initial findings, with a focus on those that are most relevant to  
181 end-users of these technologies.



### 182 3.1 Inter-device precision

183 Inter-device precision refers to the consistency of measurements across multiple devices of the same type, an  
184 important characteristic to ensure the reliability of sensor outputs over time (Moreno-Rangel et al., 2018). During  
185 QUANT, all the devices were collocated for the first 3 months and the final 3 months of the deployment to assess  
186 inter-device precision and its changes over time. Fig. 2 shows the inter-device precision (as defined by the CEN/TS  
187 17660-1:2021, i.e., the "between sensor system uncertainty" metric:  $u_s(b_s, s)$ ) of  $PM_{2.5}$  measurements during these  
188 periods. While most of the companies display a certain level of inter-device precision stability in each period  
189 (except for one, with a seemingly upward trend in the final period), there are evident long-term changes. Notably,  
190 out of the four manufacturers assessed in the final period (each having 3 devices running simultaneously), three  
191 experienced a decline in their inter-device precision compared to two years earlier. This is likely due to both  
192 hardware degradation but also drift in the calibration, which at this point had been applied between 16 and 34  
193 months prior (depending on the manufacturer). For extended periods, inconsistencies among devices from the  
194 same manufacturer might emerge, leading to varying readings under similar conditions. Consequently, data  
195 collected from different devices may not be directly comparable, which could result in inaccuracies or  
196 misinterpretations when analyzing air quality trends or making decisions.



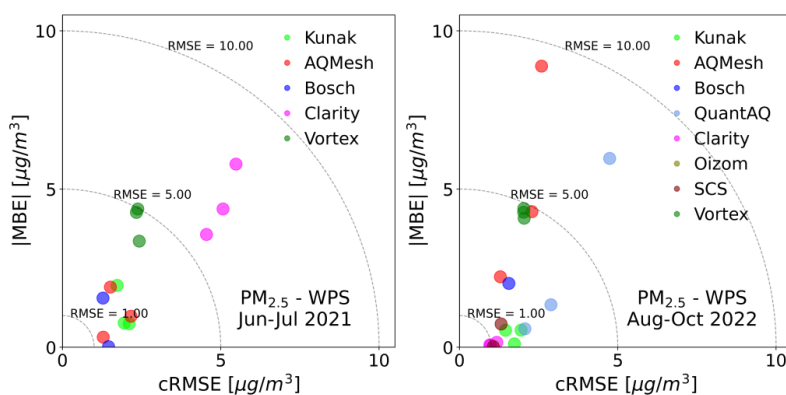
197  
198 **Figure 2. The inter-device precision of  $PM_{2.5}$  measurements from "identical" devices across the 5 companies**  
199 **participating in QUANT is assessed using the "between sensor system uncertainty" metric (defined by the CEN/TS**  
200 **17660-1:2021 as  $u_s(b_s, s)$ ). Each line represents this metric as a composite of all sensors per brand (excluding units with**  
201 **less than 70% data) within a 40-day sliding window.**

202 It is worth noting that the inter-device precision provides no information on the accuracy of the sensor  
203 measurements; a batch of devices may provide a highly consistent, but also highly inaccurate measurement of the  
204 target pollutant.

205 The "target plot" (as shown in Fig. 3) is a tool commonly used to depict the bias/variance decomposition of an  
206 instrument's error relative to a reference (for more details see Jolliff et al. (2009)). The mean error bias (MBE) is  
207 used to characterise accuracy and precision is quantified by the centered Root Mean Squared Error (cRMSE, e.g.  
208 Kim et al. (2022) also called unbiased Root Mean Squared Error (uRMSE, e.g. Guimarães et al. (2018)). Fig. 3  
209 visualises the performance of a set of  $PM_{2.5}$  sensors of the WPS deployment for the first 2 months (out-of-box  
210 data) and the last 3 months of collocation (manufacturer-supplied calibrations). In addition to highlighting which  
211 devices are most accurate, Fig. 3 also provides an additional perspective of inter-device precision.



212



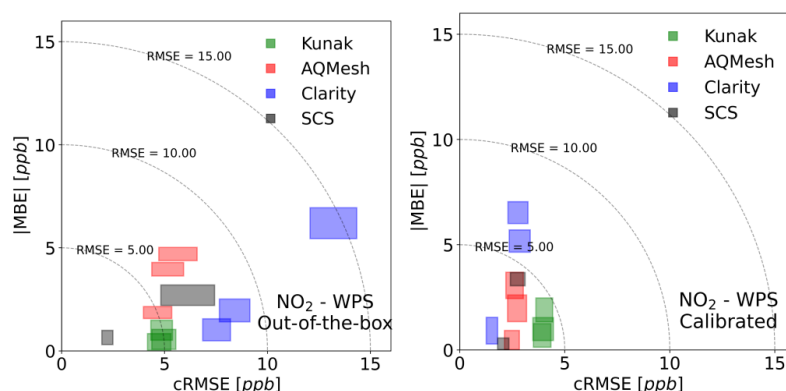
213

214 **Figure 3.** Target diagrams for the WPS  $PM_{2.5}$  measurements during the initial co-location period (Jun-Jul 2021, left)  
215 and final co-location period (Aug-Oct 2022, right). The error (RMSE) for each instrument is decomposed into the MBE  
216 (y-axis) and cRMSE (x-axis). Each point represents an individual sensor device, with duplicate devices having the same  
217 colour. Since only units with more than 75% of the data were considered, the plot on the right shows fewer units than  
218 the plot on the left.

### 219 3.2 Device accuracy and collocation calibrations

220 Sensor measurement accuracy denotes how close a sensor's readings are to reference values (Wang et al., 2015).  
221 Characterizing this feature is imperative for establishing sensor reliability and making informed decisions based on  
222 its data. Fig. 4 shows that collocation calibration can greatly impact observed  $NO_2$  sensor performance in a number  
223 of ways. Firstly, measurement bias is often, but not always, reduced following calibration, as evidenced by a general  
224 trend for devices to migrate towards the origin (RMSE = 0 ppb). Secondly, it can help to improve within-manufacturer  
225 precision by grouping sensor systems from the same company closer together. The figure also highlights a  
226 fundamental challenge with evaluating sensor systems: the measured performance can vary dramatically over time —  
227 and space— as the surrounding environmental conditions change. To quantify this, 95% Confidence Intervals (CIs)  
228 were estimated for each device using bootstrap simulation and are visualised as a shaded region. For the out-of-box  
229 data, these regions are noticeably larger than in the calibrated results for most manufacturers, suggesting that  
230 collocation calibration has helped to tailor the response of each device to the specific site conditions. This is reinforced  
231 by the cRMSE component reducing by a greater extent than the MBE, in the terminology of machine learning the  
232 calibration has helped reduce the variance portion of the bias-variance trade-off.

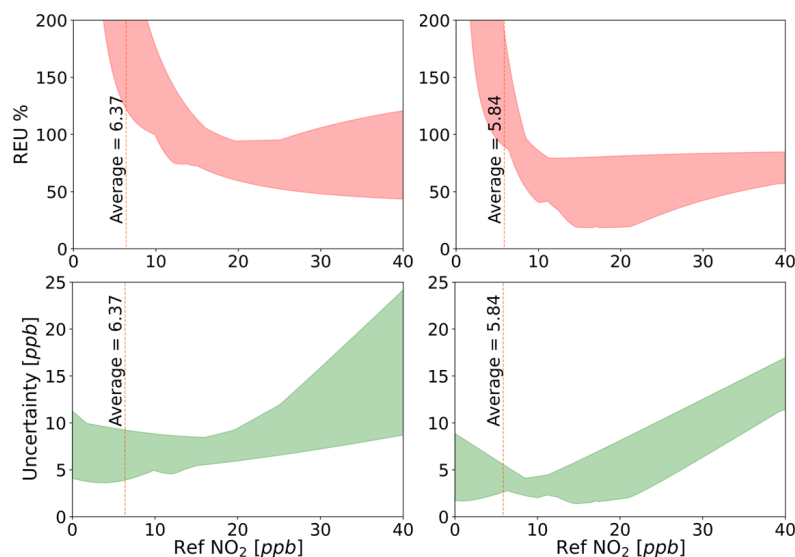




233

234 **Figure 4. Effect of colocation calibration on NO<sub>2</sub> sensor accuracy. The accuracy is quantified using RMSE, which is**  
 235 **decomposed into MBE (y-axis) and cRMSE (x-axis). 95% confidence regions were estimated using bootstrap sampling.**  
 236 **The left panel displays results from the period Jun - Jul 2021 (‘out-of-the-box’ data), while the right-hand panel**  
 237 **summarises Aug 2021 when calibrations were applied for all the WPS manufacturers.**

238 However, it is important to note a limitation of Target Plots: they primarily focus on sensor behaviour around the  
 239 mean. Therefore, the collective improvement evidenced by Fig. 4 might be only partial. For applications where it is  
 240 important to understand how calibrations impact lower or higher percentiles, considering other metrics or visual tools  
 241 would be advisable. An example of this is the absolute and Relative Expanded Uncertainty (REU, defined by the  
 242 Technical Specification CEN/TS 17660-1:202). Fig. 5 illustrates how NO<sub>2</sub> calibrations might not only improve  
 243 collective performance around the mean (as indicated by the dotted red line in Fig. 5 and previously displayed in the  
 244 target plot) but across the entire concentration range.



245

246 **Figure 5. The top plots display the REU (%) across the concentration range, while the bottom plots depict the Absolute**  
 247 **Uncertainty (ppb) —both before (left plots) and after (right plots) calibrating NO<sub>2</sub> WPS systems. The shaded areas**



248 **represent the collective variability evolution (all sensors from all companies) of both metrics. These plots were constructed**  
249 **using the minimum and maximum value of the REU and the Absolute Uncertainty for the entire concentration range.**

250 However, a note of caution when interpreting results from observational studies such as these is that it is impossible  
251 to ascertain a direct causal relationship between calibration and sensor performance as there are numerous other  
252 confounding factors at play (Diez et al., 2022). Notably these two data products are being assessed over different  
253 periods when many other factors will have changed, for example, the local meteorological conditions as well as  
254 human-made factors such as reduced traffic levels following the COVID-19 lockdown that commenced in March  
255 2020.

### 256 **3.3 Reference instrumentation is key**

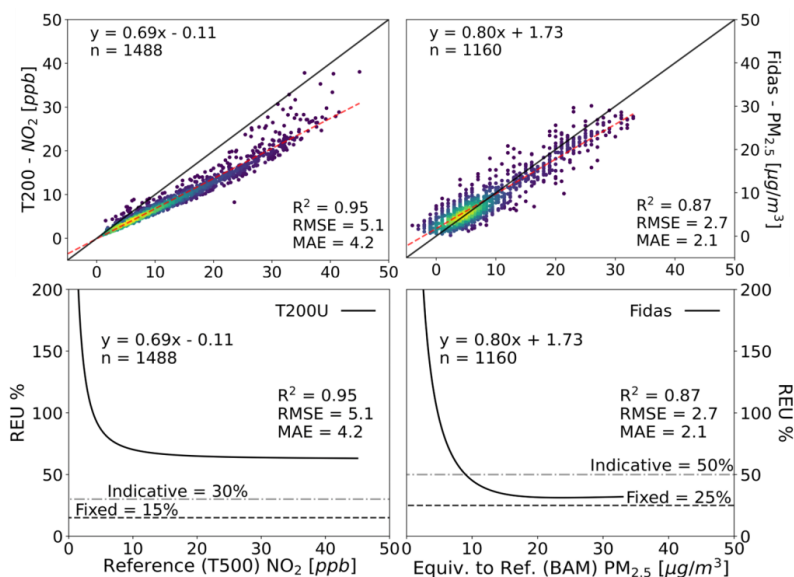
257 A common assumption when evaluating the performance of sensors is that the metrological characteristics of the  
258 sensor predominantly influence discrepancies detected in colocations. While this presumption can often be justified  
259 due to both devices' relative scales of measurement errors, it is not always the case. Since every measurement is subject  
260 to uncertainties, it is crucial to consider those associated with the reference when deriving the calibration factors of  
261 placement.

262 Fig. 6 (left plots) displays the performance of a NO<sub>2</sub> reference instrument (Teledyne T200U) specifically installed for  
263 QUANT, located next to the usual instrument at the Manchester supersite (Teledyne T500). Although they use  
264 different analytical techniques (chemiluminescence for the T200U and Cavity Attenuated Phase Shift Spectroscopy  
265 for the T500), their measurements are highly correlated ( $R^2 \sim 0.95$ ). However, it's possible to identify a proportional  
266 bias (slope=0.69), attributed to retaining the initial calibration (conducted in York) without subsequent adjustments,  
267 a situation exacerbated by an unnoticed mechanical failure of one of the instrument's components. The REU  
268 demonstrates that, under these circumstances, an instrument designated as a reference does not meet the minimum  
269 requirements set out by the Data Quality Objectives (DQOs) of the EU Air Quality Directive 2008/50/EC. Figure S3  
270 shows a unique sensor evaluated against both the T500 and the T200U. The comparison against the T200U yields  
271 better results, suggesting that, in a hypothetical scenario where it was the only instrument at the site, this could lead  
272 to misleading conclusions. This situation reinforces the idea that instruments should not only be adequately  
273 characterised but also undergo rigorous quality assurance and data quality control programs, as well as receive  
274 appropriate maintenance (Pinder et al., 2019). All of this must be performed before and during the use of any  
275 instrument.

276 For PM monitoring the current EU reference method is the gravimetric technique (CEN EN 12341, 2023), which is  
277 a non-continuous monitoring method that requires weighing the sampled filters and off-line processing of the results.  
278 Techniques that have proven to be equivalent to the reference method (called "equivalent to reference" in the EU Air  
279 Quality Directive) are very often used in practice. In the UK context, the Beta Attenuated Monitor (BAM) and FIDAS  
280 (optical aerosol spectrometer) are equivalent-to-reference methods commonly used as part of the Urban AURN  
281 Network (Allan et al., 2022). To illustrate these differences in practice, Fig. 6 compares these two equivalent-to-  
282 reference PM<sub>2.5</sub> measurements obtained with a BAM (AURN York site, located on a busy avenue), and a FIDAS unit  
283 specifically installed for QUANT. During this specific period, they do not fully agree ( $R^2 = 0.87$ ). Despite a not very  
284 pronounced bias (slope=0.80), the dispersion of points around the best-fit line is noticeable, limiting the linearity of  
285 the FIDAS compared to the BAM.



286 In the hypothetical case that the BAM were to be considered the reference method (arbitrarily chosen for this example  
287 as it is the current instrument at the AURN York site) when assessing the FIDAS under these test conditions, it would  
288 only meet the criterion stipulated by the EU DQOs for indicative measurements, but not for fixed (i.e., reference)  
289 measurements. Of course, this example is primarily intended to illustrate the magnitude of differences between both  
290 methods for this particular application, and by no means does this observation imply that the FIDAS measurements  
291 are inherently problematic.



292

293 **Figure 6.** The left plots depict the comparison between the Teledyne T200U (chemiluminescence analyzer) and the  
294 reference method (Teledyne T500 CAPS analyzer) at the Manchester supersite. The plots to the right illustrate PM<sub>2.5</sub>  
295 measurements in York, taken with a FIDAS instrument (optical aerosol spectrometer) and a BAM 1020 (beta  
296 attenuation monitor), both equivalent-to-reference methods. While the top plots show the regression (including some  
297 typical single-value metrics), those on the bottom present the REU alongside the DQOs defined by the European  
298 Directive 2008/50/EC.

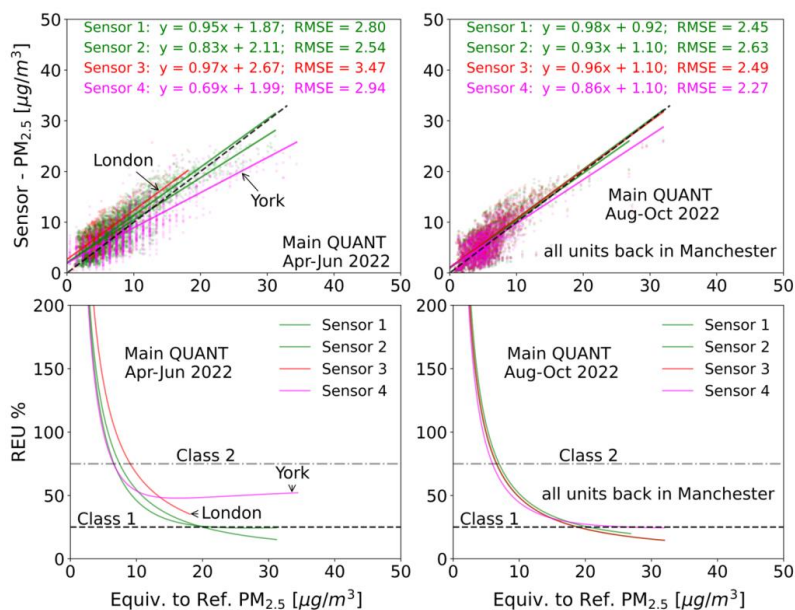
299 Although these two instruments (BAM and Fidas) show a greater concordance between themselves than with sensors  
300 (for the comparison of two sensor systems against the BAM and the Fidas, refer to Fig. S4), the choice of the  
301 measurement method can have a considerable impact on evaluations of this type. This underscores the importance of  
302 adequately characterising the uncertainties of the reference monitor when evaluating sensors.

### 303 3.4 Systems performance after location transfer

304 An extreme example of sensor performance varying due to environmental conditions is when sensors are moved  
305 between locations, as their apparent performance may vary drastically. Fig. 7 displays the REU and regression  
306 plots for four of the same PM<sub>2.5</sub> sensor system in two periods: April-June 2022 when the devices were working  
307 across the 3 sites (York, Manchester and London), and August-October 2022 when they were all reunited in  
308 Manchester. The RMSE remains reasonably consistent between the devices across the periods and locations.  
309 However, the device that moved from York to Manchester saw its slope change from 0.69 to 0.86. Because this  
310 device's slope is consistent with the other units when running in Manchester, this is likely due to the different

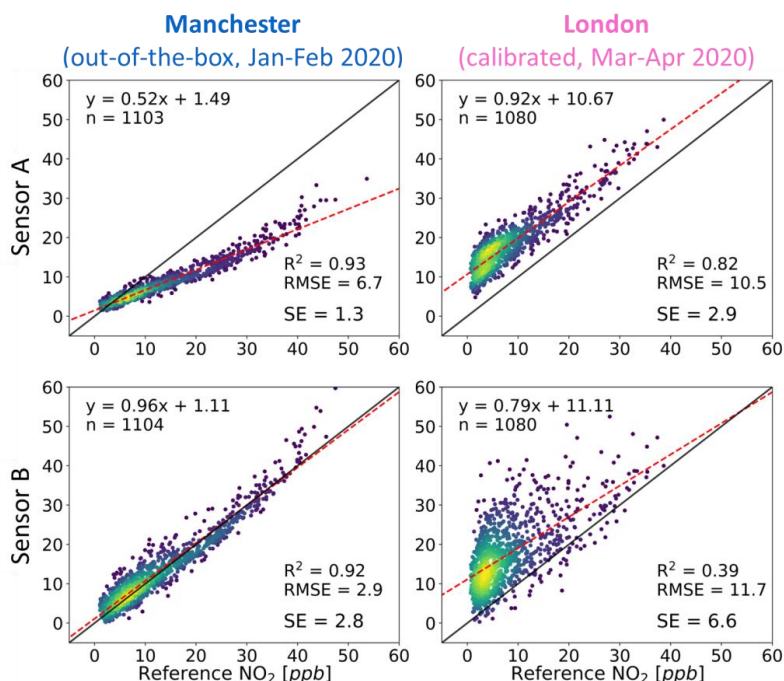


311 sensor responses in the specific environments. The precise cause of this change is not immediately evident and  
 312 will be the focus of a follow-up study, but could be due to changes in local conditions impacting sensor calibration  
 313 and/or differences in actual PM<sub>2.5</sub> sources and particle characteristics at the sites (Raheja et al., 2022).



314  
 315 **Figure 7. Regression (top) and REU (bottom) plots showing data from four PM<sub>2.5</sub> sensors (same manufacturer) over 2 time**  
 316 **periods: Apr-Jun 2022 and Aug-Oct 2022. The four devices were in separate locations in the first period, but all deployed**  
 317 **in Manchester in the second.**

318 A second example of performance changing between locations is presented in Fig. 8, showing NO<sub>2</sub> data from two  
 319 sensor systems (different brands, one shown on top of the other) before (left plots) and after (right plots) they were  
 320 moved from Manchester to London in March 2020. Both sensors saw a reduction in agreement with the reference  
 321 instrument at the London site compared to Manchester, despite both these sites being classified as urban-background  
 322 with reference instrument performance regularly audited by the UK National Physical Laboratory.



323

324 **Figure 8.** Comparison of NO<sub>2</sub> measurements for two systems (A and B) that were moved between Manchester (left plots)  
325 and London (right plots). The Manchester deployment was from January - February 2020, and the London data was  
326 recorded from April - May 2020.

327 The primary distinction between both systems' behaviour lies in the fact that the sensor located in the top row, even  
328 after being relocated to London, maintains a linear response (albeit slightly more degraded than that observed in  
329 Manchester, as the R<sup>2</sup> and RMSE show). In contrast, in the second system (bottom row), the response is notably  
330 noisier as the Standard Error (SE) —which is the dispersion of the data around the best-line fit line, i.e., the remaining  
331 error after bias correction. In scenarios akin to this latter, where there is a high variance in the residuals, a linear  
332 correction will not provide a significant improvement. While more sophisticated corrections could be applied, these  
333 will be limited by domain knowledge of the end-user, and potentially by other complex data sources that might be  
334 available. However, it is important to remember that additional post-processing could increase the risk of overfitting  
335 (Aula et al., 2022). On the other hand, for cases like the top plots, users might benefit from trying to correct them  
336 using simple linear correction (e.g. using reference instruments if available) or other approaches that could provide  
337 means for zero and span correction. A straightforward and cost-effective example could be the use of diffusion tubes  
338 for the case of NO<sub>2</sub>, as discussed in Section 3.6.

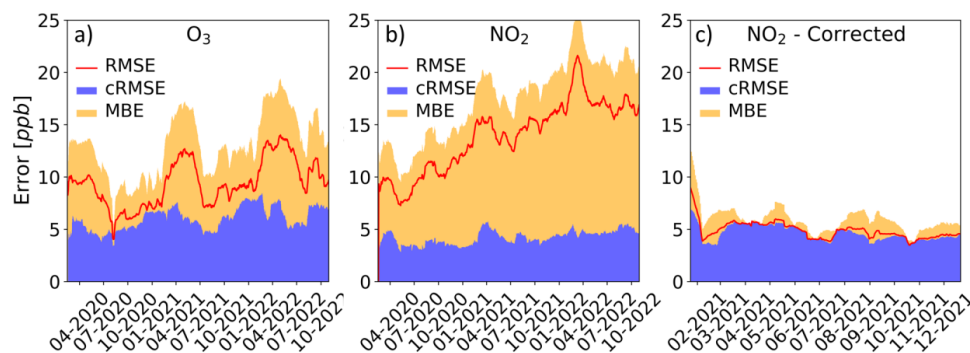
### 339 3.5 Long-term stability

340 The long-term stability of sensor response is also an important facet of its performance, especially for certain use  
341 cases such as multi-year network deployments. There can be multiple causes of long-term changes to sensor  
342 response, for example, particles settling inside the sampling chamber in optical-based sensors(e.g. Hofman et al.  
343 (2022)), or the gradually changing composition of electrochemical cells (e.g. Williams (2020)). How these changes  
344 manifest themselves in the data must be identified if ways to account for them are to be implemented.



345 Fig. 9 shows the temporal nature of the O<sub>3</sub> and NO<sub>2</sub> errors (MBE, cRMSE and RMSE) from a sensor system  
346 between February 2020 and October 2022. The O<sub>3</sub> shows (Fig. 9a) a gradual increase in the overall measurement  
347 error, largely due to an increase in the MBE. It also shows a distinct seasonality MBE, increasing by a factor of 3-  
348 4 between March and July compared to the August-February period. The cRMSE component shows fluctuations  
349 during the study but only has a small increasing trend. The NO<sub>2</sub> system (Fig. 9b) demonstrates a consistently  
350 increasing overall error, with a less pronounced seasonal influence. The bias contributes greatly to the total error  
351 (see Section 3.6 for NO<sub>2</sub> sensor correction, Fig. 9c).

352



353

354 **Figure 9.** Error (as RMSE, red line) of one of the systems belonging to the Main QUANT, decomposed into cRMSE (in  
355 blue) and MBE (in yellow) estimated based on a 40-day (1-day slide) moving window. Panel a) is for O<sub>3</sub> measurements,  
356 and panel b) is for NO<sub>2</sub> (April 2020-Oct 2022). Panel c) is also for NO<sub>2</sub>, this time showing the effect of a linear correction  
357 using diffusion tubes (see next section for more details).

### 358 3.6 Informing end-use applications

359 Ultimately, for any air pollution monitoring application, the requirements of the task should dictate the  
360 measurement technology options available. For example, if the requirement for a particular measurement is to  
361 assess legal compliance, then lower measurement uncertainty must be a key consideration as the reported values  
362 need to be compared to a limit value. In contrast, if an application aimed to look at long-term trends in pollutants,  
363 then absolute accuracy may not be as important as the long-term stability of sensor response. In order to realise  
364 the potential of air pollution sensor technologies, end users need to be provided with the information required to  
365 critically assess the strengths and weaknesses of potential candidate sensor devices, ideally in an easy-to-access  
366 and interpret manner.

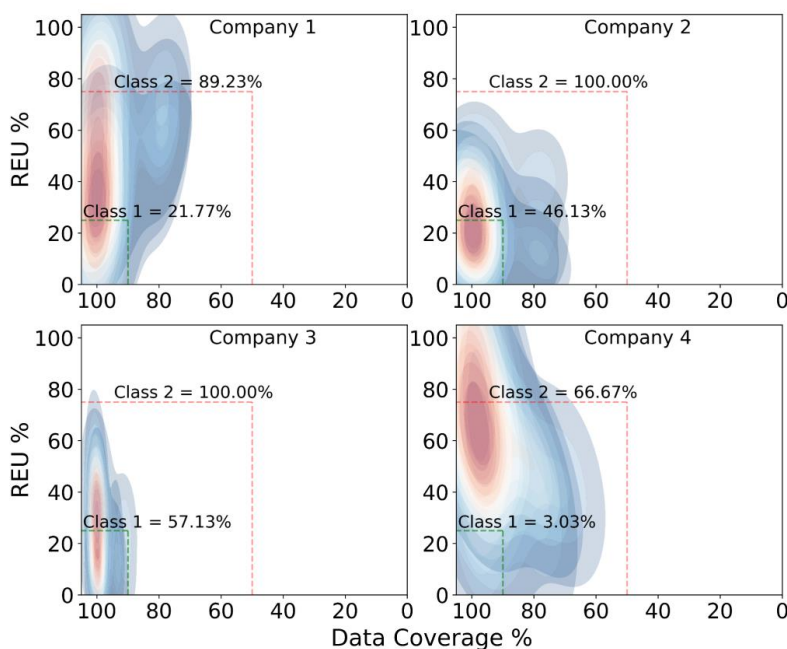
367

368 Understanding the uncertainty associated with a measurement instrument is essential for recognizing its  
369 capabilities and limitations. Accurate instruments are crucial, especially in areas like public health decision-  
370 making, where inaccurate data can have profound implications (Molina Rueda et al., 2023). Furthermore,  
371 instruments that operate autonomously ensure consistent, uninterrupted data collection, making them more  
372 efficient and cost-effective in terms of maintenance and calibration. Figure 10 shows the REU (y-axis) and Data  
373 Coverage (DC, x-axis) of companies measuring NO<sub>2</sub> with more than 2 systems running to avoid ambiguity in the





374 results. Using multiple systems, not only avoids ambiguity in results but also enhances the robustness of the data  
 375 collected. Both REU and DC are key criteria within the EU scheme (EU 2008/50/EC) for evaluating the  
 376 performance of measurement methods, and are complemented by the CEN/TS 17660-1:2021 specifically for  
 377 sensors. This document defines three different sensor system tiers. Class 1 sensors, bounded by the green rectangle,  
 378 offer higher accuracy than Class 2 sensors, highlighted by the red rectangle (Class 3 sensors have no set  
 379 requirements). Presenting the data like this helps users anticipate the performance of sensor systems —under the  
 380 assumption that all sensors from the same brand will behave similarly in equivalent environmental conditions—  
 381 providing more insight into selecting the appropriate instrument for a given project or study.



382

383 **Figure 10. The REU vs. Data Coverage (DC) for 4 systems companies was evaluated during the WPS for the period Nov**  
 384 **2021-Oct 2022 (after all companies had at least one calibrated product). Both the REU and the DC were estimated**  
 385 **based on a 40-day size (which is the number of days used by CEN/TS 17660-1:2021 for on-field tests) moving window**  
 386 **(1-day slide). While the green rectangle represents the DQOs for Class 1 sensors, the red one limits the DQOs for Class**  
 387 **2 sensors (Class 3 sensors have no requirements).**

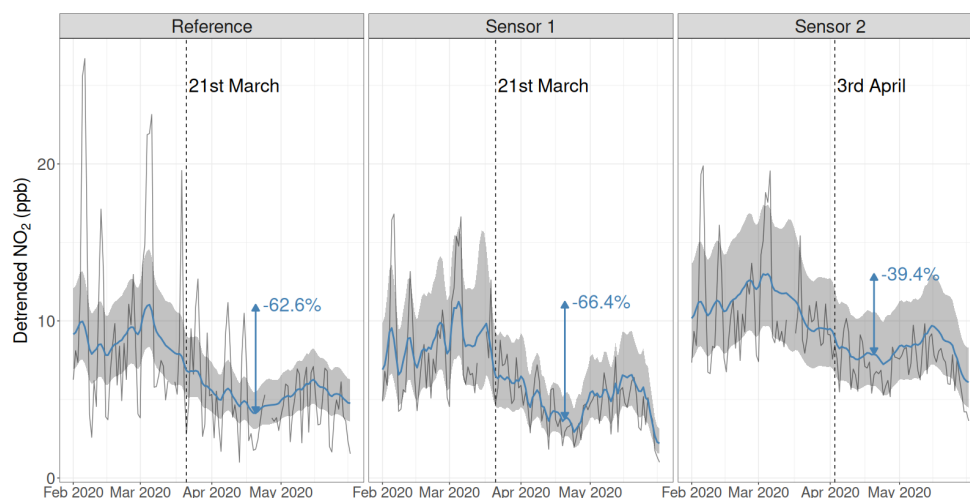
388 Depending on the nature of the sensor data uncertainty, methods can be implemented to improve certain aspects  
 389 of the data quality for a particular application. One such example is the use of distributed networks to estimate  
 390 sensor measurement errors, such as that described by (J. Kim et al., 2018). Depending on the application, simpler  
 391 methods could also be available to reduce the magnitude of the changing bias, and thus significantly improve the  
 392 accuracy of an individual sensor system, but also that of broader sensor networks. For the case shown in Fig.9b,  
 393 one possible way to do this would be using supporting observations of NO<sub>2</sub> made via diffusion tubes. These  
 394 measurements are widely used to monitor NO<sub>2</sub> concentrations in UK urban environments, due to their lower cost  
 395 (~£5 per tube) and ease of deployment, but only provide average concentrations over periods of weeks to months  
 396 (Butterfield et al., 2021). During QUANT, NO<sub>2</sub> diffusion tubes were deployed at the 3 colocation sites (see Section  
 397 S8 at the Supp. for more details). Combining these measurements offers the possibility of quantifying the average



398 sensor bias, thus reducing the error on the sensor measurement whilst maintaining the benefits of its high-time  
399 resolution observations. It is important to note that while bias correction has been applied to the sensor data, the  
400 NO<sub>2</sub> diffusion tube concentrations used for comparison purposes must also be adjusted (e.g. following DEFRA  
401 (2022)). Fig. 9c shows the accuracy of the same NO<sub>2</sub> sensor data shown in Fig. 9b but applies a monthly offset  
402 calculated as the difference between its monthly average measurement and that from the diffusion tube (see Figure  
403 S5). This shows a dramatic reduction in overall error largely driven by its bias correction. What remains largely  
404 resulting from the cRMSE, i.e. the error variance that might arise from limitations from the sensing technology  
405 itself and/or the conversion algorithms used to transform the raw signals into the concentration output. To validate  
406 the efficacy and reliability of this bias correction method, further long-term studies are warranted.

407 The development and communication of methods that improve sensor data quality, ideally in digestible case  
408 studies, would likely increase the successful application of sensor devices for local air quality management. There  
409 is also a need for similar case studies showcasing the successful application of sensor devices for particular  
410 monitoring tasks. An example of this from the QUANT dataset is the use of sensor devices to successfully identify  
411 change points in a pollutant's concentration profile. Determining when a specific pollutant has changed its  
412 temporal nature is a challenging task as there are a large number of confounding factors that influence a pollutant's  
413 concentration at a specific point in time, including but not limited to seasonal factors, environmental conditions  
414 (both natural and arising from human behaviour), and meteorological factors. A novel statistical approach to  
415 smoothing air quality measurements was applied, accounting for these external factors (Lacy & Moller). This  
416 method was applied to NO<sub>2</sub> concentrations determined from the sensor systems that had remained in Manchester  
417 throughout 2020, aiming to identify whether the well-documented reduction in ambient NO<sub>2</sub> concentrations could  
418 be observed due to changes in travel patterns associated with COVID-19 restrictions. To provide an objective  
419 quantification of whether a change-point had occurred, the Bayesian online change-point detection (Adams &  
420 MacKay, 2007) was applied. Of the 8 devices that measured NO<sub>2</sub>, clear changepoints corresponding to the  
421 introduction of a lockdown were identified in 2 (Fig.11). While this is an unsupervised analysis, it demonstrates  
422 the potential of these devices to identify long-term trends with appropriate processing, even with only having had  
423 3 months of training data to fit the model to. This is especially aided by the given algorithm's ability to use  
424 reference data as a prior allowing sensor systems to fine-tune the model.





425

426 **Figure 11.** NO<sub>2</sub> measurements (black solid line) and detrended estimates (blue solid line with 95% confidence interval  
427 in the shaded grey region) from the reference instrument (left panel) and 2 sensor systems (middle and right panels)  
428 from Manchester in 2020. Vertical dashed lines and their corresponding dates indicate identified change points, which  
429 correspond to the introduction of the first national lockdown due to COVID-19 on the 23rd of March 2020. The  
430 percentage in blue represents the relative peak-trough decrease from 5th March to 20th April.

#### 431 4. Conclusions

432 Lower-cost air pollution sensor technologies have significant potential to improve our understanding and ability  
433 to manage air pollution issues. Large-scale uptake in the use of these devices has been primarily limited by  
434 concerns over data quality and a general lack of a realistic characterisation of the measurement uncertainties  
435 making it difficult to design end uses that make the most of the data information content. Developments in the  
436 field of air pollution sensor technology are also developing rapidly, with advances in both the measurement  
437 technology and particularly in the data post-processing and calibration. A challenge with the use of sensor-based  
438 devices is that many of the end-use communities do not have access to extensive reference-grade air pollution  
439 measurement capability (Lewis & Edwards, 2016), or in many cases expertise in making atmospheric  
440 measurements. For this reason, reliable information on expected sensor performance needs to be available to aid  
441 effective end-use applications. Large-scale independent assessments of air sensor technologies are non-trivial and  
442 costly, however, making it difficult for end users to find relevant performance information on current sensor  
443 technologies. The QUANT assessment is a multi-year study across multiple locations, that aims to provide relevant  
444 information on the strengths and weaknesses of commercial air pollution sensors in UK urban environments.

445 The QUANT sensor systems were installed at two highly instrumented urban background measurement sites, in  
446 Manchester and London, and one roadside monitoring station in York. The study design ensured that multiple  
447 devices were collocated to assess inter-device precision, and devices were also moved between locations and able  
448 to test additional calibration data products to assess and enable developments in sensor performance under realistic  
449 end-use scenarios. A wider participation component of the Main QUANT assessment was also run at the  
450 Manchester site to expand the market representation of devices included in the study, and also to assess recent  
451 developments in the field.



452 A high-level analysis of the dataset has highlighted multiple facets of air pollution sensor performance that will  
453 help inform their future usage. Inter-device precision has been shown to vary, both between different device types  
454 and over different periods of time, with the most accurate devices generally showing the highest levels of inter-  
455 device precision. The accuracy of the reported data for a particular device can be impacted by a variety of factors,  
456 from the calibrations applied to its location or seasonality. This has important implications for the way sensor-  
457 based technologies are deployed and supports the case made by others (Bittner et al., 2022; Farquhar et al., 2021;  
458 Crilley et al., 2018; Williams, 2020; Bi et al., 2020) that practical methods to monitor sensor bias will be crucial  
459 in uses where data accuracy is paramount.

460 In addition to these findings, this overview lays the groundwork for more detailed research to be presented in  
461 future publications. Subsequent analyses will focus on providing a more nuanced understanding of the  
462 uncertainty in air pollution sensor measurements, thus equipping end-users with better insights into of the  
463 capability of sensor data. Future studies will delve into specific aspects of air pollution sensor performance: 1) a  
464 comprehensive performance evaluation of PM<sub>2.5</sub> data, assessing their accuracy and reliability under different  
465 environmental conditions; 2) an in-depth analysis of NO<sub>2</sub> measurements, examining their sensitivity and response  
466 in various urban environments; and 3) a detailed investigation into the detection limits of these sensor technologies,  
467 targeting their optimized application in low concentration scenarios. These focused studies are basic steps needed  
468 to further advance our understanding of sensors' capabilities and limitations, ensuring informed and effective  
469 application in air quality monitoring.

#### 470 **Supplementary**

471 The supplement related to this article is available online at:

#### 472 **Data availability**

473 The data for this study can be found at the Centre for Environmental Data Analysis (CEDA): Lacy, S.; Diez, S.;  
474 Edwards, P. (2023): Quantification of Utility of Atmospheric Network Technologies: (QUANT): Low-cost air  
475 quality measurements from 52 commercial devices at three UK urban monitoring sites. NERC EDS Centre for  
476 Environmental Data Analysis, date of citation.  
477 <https://catalogue.ceda.ac.uk/uuid/ae1df3ef736f4248927984b7aa079d2e>.

#### 478 **Author contributions**

479 The initial draft of the manuscript was created by SD, PME, and SL. The research was conceptualized, designed,  
480 and conducted by PME and SD. Methodological framework and conceptualization were developed by SD, PME,  
481 and SL. Data analysis was primarily conducted by SD and SL. The software tools for data visualization and  
482 analysis were developed by SD and enhanced by SL. MF, MP and NM supplied the reference data critical for the  
483 study. TB, HC, DH, SG, NAM and JU made substantive revisions to the manuscript, enriching the final  
484 submission.

#### 485 **Competing interests**

486 The authors declare that they have no conflict of interest.

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