Characterisation of gaseous iodine species detection using the multi-scheme chemical ionisation inlet-2 with bromide and nitrate chemical ionisation methods

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Abstract. The multi-scheme chemical ionisation inlet 1 (MION1) enables rapid switching between the measurement of atmospheric ions without chemical ionisation and neutral molecules using various atmospheric pressure chemical ionisation methods. In this study, we introduce the upgraded version, the multi-scheme chemical ionisation inlet 2 (MION2). The new design incorporates enhanced ion optics, resulting in increased reagent ion concentration, ensuring a robust operation, and
5 enabling the use of multiple chemical ionisation methods with the same ionisation time.

In order to simplify the regular calibration of MION2, we developed an open-source flow reactor chemistry model called MARFORCE. This model enables quantification of the chemical production of sulfuric acid (H_2SO_4), hypoiodous acid (HOI), and hydroperoxyl radical (HO_2). MARFORCE simulates the convection-diffusion-reaction processes occurring within typical cylindrical flow reactors with uniform inner diameters. The model also includes options to simulate chemical processes in two
10 scenarios: 1) when two flow reactors with different inner diameters are connected, and 2) when two flows are merged into one using a Y-shape tee, although with reduced accuracy. Furthermore, the chemical mechanism files in the model are compatible with the widely-used Master Chemical Mechanism (MCM), allowing for future adaptation to simulate other chemical processes in flow reactors.

Furthermore, we conducted a comprehensive characterisation of the bromide (Br^-) and nitrate (NO_3^-) chemical ionisation
15 methods with different ionisation times. We performed calibration experiments for H_2SO_4, HOI, and HO_2 by combining gas kinetic experiments with the MARFORCE model. The evaluation of sulfur dioxide (SO_2), water (H_2O), and molecular iodine

(I_2) involved dilution experiments from a gas cylinder (SO_2), dew point mirror measurements (H_2O), and a derivatization approach combined with high-performance liquid chromatography quantification (I_2), respectively.

Our findings indicate that the detection limit is inversely correlated with the fragmentation enthalpy of the analyte-reagent ion (Br^-) cluster. In other words, stronger binding (resulting in a larger fragmentation enthalpy) leads to a lower detection limit. Additionally, a moderately longer ionisation time enhances the detection sensitivity, thereby reducing the detection limit. For instance, when using the Br^- chemical ionisation method with a 300 ms ionisation time, the estimated detection limit for H_2SO_4 is 2.9×10^4 molec. cm^{-3}. Notably, this detection limit is even superior to that achieved by the widely-used Eisele-type chemical ionisation inlet $(7.6 \times 10^4$ molec. $cm^{-3})$, as revealed by direct comparisons.

While the NO_3^- chemical ionisation method remains stable in the presence of high humidity, we have observed that the Br^- chemical ionisation method (Br^--MION2) is significantly affected by air water content. Higher levels of air water lead to reduced sensitivity for HO_2 and SO_2 under the examined conditions. However, we have found that a sharp decline in sensitivity for H_2SO_4, HOI, and I_2 occurs only when the dew point exceeds 0.5-10.5 °C (equivalent to 20-40 % RH, calculated at 25 °C throughout this manuscript). For future studies utilising the atmospheric pressure Br^- chemical ionisation method, including Br^--MION2, it is crucial to carefully consider the molecular-level effects of humidity. By combining approaches such as the water-insensitive NO_3^--MION2 with Br^--MION2, MION2 can offer more comprehensive insights into the composition of atmospheric air than what can be achieved by either method alone.

By employing instrument voltage-scanning, chemical kinetic experiments, and quantum chemical calculations, we have conclusively established that the presence of iodine oxides does not interfere with the detection of HIO_3. Our comprehensive analysis reveals that the ions IO_3^-, $HIO_3 \cdot NO_3^-$, and $HIO_3 \cdot Br^-$, which are detected using the Br^- and NO_3^- chemical ionisation methods, are primarily, if not exclusively, generated from gaseous HIO_3 molecules within atmospherically relevant conditions.

1 Introduction

Chemical ionisation mass spectrometry (CIMS) has been widely used in atmospheric chemistry and aerosol formation studies due to its versatility and high sensitivity in measuring trace level gaseous species (see e.g., Eisele and Tanner (1993); Munson and Field (1966); Hansel et al. (1995); Huey (2007); Kirkby et al. (2011); Ehn et al. (2014); Lee et al. (2014); Berndt et al. (2016); Sipilä et al. (2016); Laskin et al. (2018); He et al. (2021b)). With chemical ionisation methods, an analyte is charged either by 1) receiving charge (proton, electron or ion) from the reagent ion or 2) forming a relatively stable cluster with the reagent ion. Mass spectrometers further measure the charged analyte-containing ions to obtain their molecular information.

Various chemicals have been employed as reagent ions in chemical ionisation methods. The commonly used reagent ions include nitrate (NO_3^-, Eisele and Tanner (1993)), acetate ($C_2H_3O_2^-$, Veres et al. (2008)), iodide (I^-, Caldwell et al. (1989)), hydronium (H_3O^+, Lagg et al. (1994)), and sporadically bromide (Br^-, Caldwell et al. (1989)) and ammonium (NH_4^+, Westmore and Alauddin (1986)). These reagent ions transfer charges to or form clusters with distinct subsets of trace gases. However,

the detection of an analyte-containing ion is influenced by its transmission through the mass spectrometer's ion optics, as collision-induced cluster fragmentation can diminish the analyte's signature. Analyte-reagent ion clusters with strong bonds have a higher likelihood of reaching the detector compared to weakly bonded clusters (Passananti et al., 2019). Hence, it is crucial to select a chemical ionisation method that preserves the analyte's signature. For example, the NO_3^--CIMS has been widely used to detect sulfuric acid (H_2SO_4) (Eisele and Tanner, 1993) and highly oxygenated organic molecules (Ehn et al., 2014). I^--CIMS is regularly used to detect semi-volatile organic compounds (Lee et al., 2014), bromine and chlorine-containing species (Liao et al., 2014; Wang et al., 2019) and e.g., dinitrogen pentoxide (N_2O_5) (Thornton et al., 2010). $C_2H_3O_2^-$-CIMS was used to detect small organic acids (Veres et al., 2008) and highly oxygenated organic compounds (Berndt et al., 2016). The bromide chemical ionisation method has recently been used to detect species such as HO_2 (Sanchez et al., 2016) and H_2SO_4 (Wang et al., 2021a). The detection of a series of halogenated species by the Br^- chemical ionisation method was first demonstrated by He (2017). Detailed characterisation of the Br^- chemical ionisation method utilising the multi-scheme chemical ionisation inlet 1 (MION1) was presented in several of our earlier studies (Wang et al., 2021a; Tham et al., 2021; He et al., 2021b). Multiple species were successfully calibrated using either analytical methods or inter-instrument comparison, including H_2SO_4, I_2, Cl_2 and HOI (Tham et al., 2021; Wang et al., 2021a). Among the calibrated species, H_2SO_4 and I_2 were shown to be detected at the collision limit (highest sensitivity). Although H_2SO_4 has been quantified using standard methods (Kürten et al., 2012), the quantification of the measured $I_2 \cdot Br^-$ signal remains challenging. This is primarily contributed by two factors: 1) the current Br^--MION1/2 have a detection upper limit of a few hundred pptv of I_2, beyond which the reagent ions get depleted and the measurement is non-linear, 2) on the contrary, spectroscopic and other methods could be limited by their high detection limits and may not be able to detect I_2 at appropriate levels. Therefore, the key is to find sensitive methods to quantify gaseous I_2 at tens to hundreds of pptv levels.

Ideally, simultaneous measurement of all the mentioned analytes could be achieved by employing their corresponding CIMS methods concurrently in ambient observations or complex laboratory experiments. However, CIMS instruments are costly, and research institutes often face limitations due to the availability of such instrumentation. As an alternative approach, chemical ionisation inlets that have the capability to switch between different reagent ions can be employed. Many switchable systems have been developed previously, such as for proton transfer reaction mass spectrometers (Jordan et al., 2009; Breitenlechner et al., 2017; Pan et al., 2017) and other chemical ionisation mass spectrometers (Hearn and Smith, 2004; Smith and Španěl, 2005; Agarwal et al., 2014; Brophy and Farmer, 2015). A common feature of these techniques is using a reduced-pressure ion-molecule reaction chamber, thus unavoidably diluting the gas molecules of interest by orders of magnitude. While the detection limit of an instrument is also affected by other factors, it is commonly observed that chemical ionisation inlets operating at reduced pressures have higher limits of detection compared to atmospheric pressure chemical ionisation inlets. For instance, reduced pressure inlets reported detection limits of various organic compounds from sub-pptv (parts per trillion by volume) to hundreds of pptv levels (Lee et al., 2014; Brophy and Farmer, 2015), while the best-performing atmospheric pressure chemical ionisation inlets regularly detect vapours at ppqv (parts per quadrillion by volume) levels for selected acids and highly oxygenated organic vapours using the same time-of-flight mass spectrometer (Jokinen et al., 2012; Ehn et al., 2014; He et al.,

85   2021b).

To reduce the detection limit of switchable reagent ion chemical ionisation systems, we developed the MION1 inlet, which allows for fast switching of reagent ion chemistry at atmospheric pressure (Rissanen et al., 2019). This technique has predominantly been employed for the detection of sulfuric acid and halogenated species using either the NO_3^- or Br^- chemical

90   ionisation methods (Rissanen et al., 2019; Tham et al., 2021; He et al., 2021b; Wang et al., 2021a; Finkenzeller et al., 2022). However, there are some remaining issues with the MION1. Firstly, its limit of detection is higher compared to another commonly used atmospheric pressure chemical ionisation inlet referred to as the "Eisele inlet" (Eisele and Tanner, 1993; Jokinen et al., 2012; Wang et al., 2021a). Secondly, the ionisation times for different chemical ionisation methods have to be different due to the design, which involves aligning and attaching the chemical ionisation units at varying distances on a cylindrical tube.

95   These challenges may restrict its suitability for detecting vapours at extremely low concentrations (e.g., at 10^5 to 10^6 molec. cm^{-3} or 4 to 40 ppqv) and interpreting the species detected by different chemical ionisation methods.

In this study, we introduce an upgraded version of the MION inlet, referred to as "MION2", which specifically addresses these issues. We conducted laboratory experiments to characterise the performance of this inlet using analytical methods and

100   a newly-developed open-source kinetic model. As chemical ionisation methods based on halogen anions (such as I^-) are commonly influenced by the water content in the air (Kercher et al., 2009; Mielke et al., 2011; Woodward-Massey et al., 2014; Lee et al., 2014), we also systematically examined the impact of air water content on the detection of Br^--MION2.

2   Methods

105   ### 2.1   Characterisation of the MION2 inlet

The ionisation inlet utilised in this study is the upgraded multi-scheme chemical ionisation inlet, MION2, developed by Karsa Ltd. This inlet is specifically designed to enable the measurement of neutral molecules using chemical ionisation methods while also facilitating the detection of atmospheric ions by disabling chemical ionisation. It offers the capability of rapid switching between two or more chemical ionisation methods, allowing for selective measurement of gaseous species at ambient pressure.

110   Currently, the MION2 inlet supports up to six ion sources.

Due to the geometric limitations of the previous MION1 inlet, the different ionisation sources in MION1 have to employ different ionisation times. The ionisation time is defined by the sample flow rate and the distance between the ion injection port and the instrument pinhole (refer to Figure 1). The improved geometry of the MION2 inlet overcomes this limitation,

115   allowing for the operation of three bipolar ion sources per ionisation time, all positioned at the same distance from the pinhole. Additionally, for the longer ionisation time, the length of the connecting pipe between the sources can be adjusted, providing

flexibility in modifying the ionisation time.

In this study, we employed the MION2 inlet with two chemical ionisation methods, namely NO_3^- and Br^-, along with two different ionisation times (35 and 300 milliseconds, ms, respectively). This configuration was chosen to investigate the characteristics of the inlet. To facilitate clear referencing, we designate the ion source positioned 3 cm away from the mass spectrometer as tower 1 (T1), while the source located 25 cm away from the mass spectrometer is referred to as tower 2 (T2) throughout this paper (see Figure 1).

Figure A1 illustrates the conceptual schematic of one of the ion sources, depicting the airflow and ion paths. The entire source is attached to a 24 mm inner diameter tube that is electrically grounded. The sample flow, which is provided by a mass flow controller (MFC) connected to a vacuum pump, is set at a rate of 22.5 standard litres per minute (slpm). The target molecules undergo ionisation by reacting with the reagent ions (NO_3^- or Br^-).

In this configuration, the ionisation time for the target molecules and charged reagent ions is approximately 35 ms for tower 1 and 300 ms for tower 2. A neutral reagent inflow is introduced, which consists of nitrogen or air enriched with reagent vapour. The reagent vapour is generated by passing nitrogen or air over liquid reagent (nitric acid, HNO_3, or dibromomethane, CH_2Br_2, in this study). The resulting mixture is then fed into the ion source, where it is ionised by a soft X-ray source (Hamamatsu L12535, 4.9 keV). The charged reagent ions are guided into the sample flow by an electric field within the ion source. This electric field is generated using concentric stainless steel electrode plates with orifices of different sizes (ranging from 5-10 mm in diameter), with resistances placed between each pair of plates. Two high voltages (approximately 2500 V and 250 V) are used in the inlet. The lower of the two voltages determines whether the reagent ions pass through the final orifice in the deflector electrode, effectively controlling the ionisation process and enabling the rapid switching between ion sources.

In contrast to the MION1 design (Rissanen et al., 2019), which relied on the reagent inflow and exhaust flow to define the source reagent flow, MION2 incorporates an additional purge flow to prevent the sample flow from entering the ion source. The purge flow consists of the same nitrogen or air used to generate the reagent flow. Upon entering the ion source, the purge flow splits into two streams: one stream prevents the sample flow from entering, while the other stream ensures that the neutral reagent does not enter the sample flow.

In MION2, the typical flow rates for the reagent, purge, and exhaust are 10, 100, and 50 standard cubic centimetres per minute (sccm), respectively. The reagent concentration in the ion source is estimated to be 2×10^{17} cm^{-3}. This design effectively addresses the challenges faced in MION1, where a compromise had to be made between the risk of contaminating the sample pipe with a neutral reagent or introducing sample air into the ion source, potentially leading to contamination or uncontrolled ion chemistry and resulting in detection biases. In MION2, the water vapour and other contaminants in the sample flow do not have the opportunity to oxidise the surfaces of the electrodes inside the ion source. Such oxidation would result

in reduced ion transmission from the ion source to the sample flow. Operational testing during ambient measurements has demonstrated that MION2 exhibits significantly improved stability compared to MION1. For example, recent measurements conducted at a coastal site in Finland involved the uninterrupted operation of MION2 for at least two months.

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Additionally, the upgraded ion optics inside the ion sources of MION2 have increased the transmission of reagent ions and the observed reagent ion concentration at the mass spectrometer by approximately one order of magnitude compared to MION1. This improvement was achieved by modifying the last electrode within the ion source to minimize ion residence time and reduce diffusion losses of ions.

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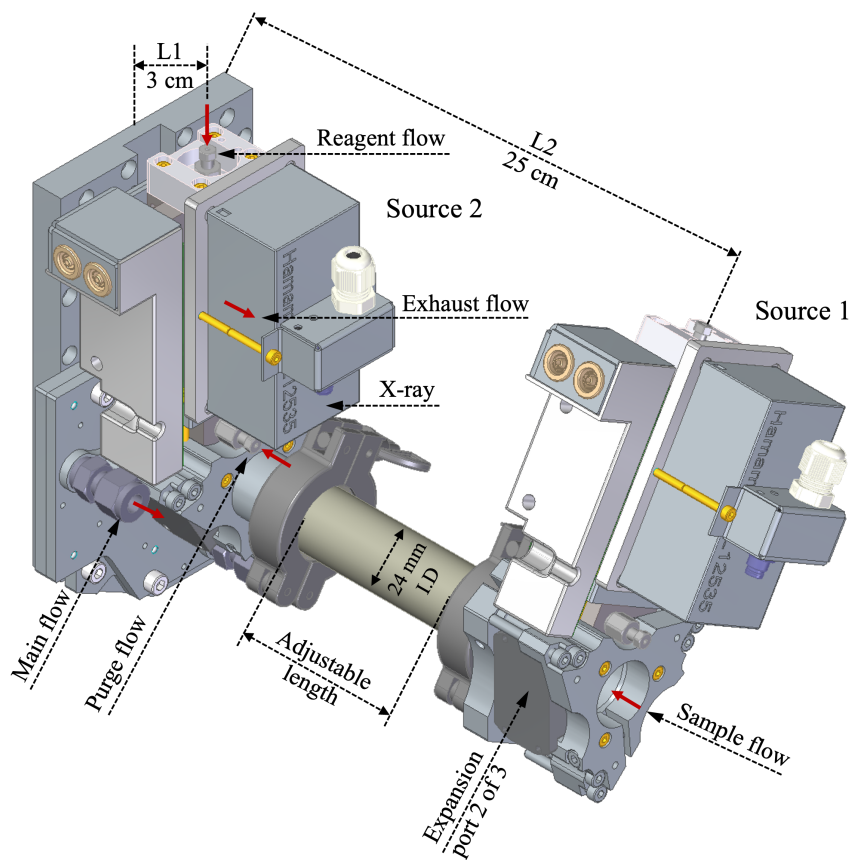
Figure 1. Schematic of the MION2 inlet illustrating its gas flows and ion paths. The new design increases the primary ion concentration and allows the operation of multiple chemical ionisation methods with the same ionisation time. L1 and L2 refer to the distances between the ion sources and the pinhole of the mass spectrometer. The exhaust flows are connected to two ports in the middle of the ion source.

2.2 Experimental setup

2.2.1 Calibration of inorganic species

The experimental setup used for characterising the MION2 inlet is illustrated in Figure A2. It consists of three main sections: the flow reactor section, the MION2 chemical ionisation inlet section, and an Atmospheric Pressure interface Time-of-Flight mass spectrometer (APi-TOF, Aerodyne Inc., Junninen et al. (2010)).

The flow reactor section includes a calibration source and several gas feeds. Synthetic air (Woikoski OY, Finland; purity $\geq$ 99.999 % with 20.9 % O_2), nitrogen (N_2, Woikoski OY, Finland; purity $\geq$ 99.999 %), and sulfur dioxide (SO_2, Air Products, USA; 99.5 % purity) were injected into the system using mass flow controllers connected to standard gas cylinders or tanks. These gases were pre-mixed before reaching the calibration source.

I_2 was generated either from a homemade permeation tube or a commercial permeation tube (VICI Metronics). A stream of nitrogen (50 sccm) was passed over the permeation tube at controlled temperatures ranging from 120 to 140 °C. The temperature of the permeation tubes was regulated using an electronically controlled heating mantle, allowing for adjustable yet stable iodine concentrations. Water vapour (H_2O) was controlled by adjusting the flow of nitrogen through a water bubbler, providing a controllable source of humidity.

The calibration source was mainly used to calibrate H_2SO_4, HO_2 and HOI. The H_2SO_4 calibration was detailed in Kürten et al. (2012) and the HOI calibration was presented in Tham et al. (2021) and Wang et al. (2021a). The calibration source in the experimental setup was constructed using an ~~aluminum~~ aluminium box that encloses a 3-quarter-inch quartz tube. The quartz tube was chosen for its high transmission properties for ultra-violet (UV) light emitted from a mercury lamp. Adjacent to the quartz tube, the mercury lamp is housed in an ~~aluminum~~ aluminium block that contains a filter-covered hole. The filter used in the ~~aluminum~~ aluminium block allows for high transmission of 185 nm light emitted from the lamp. This specific wavelength of light is effective in photolysing H_2O molecules, generating OH radicals.

Before conducting the calibration experiment, a mixed flow of nitrogen (N_2), oxygen (O_2), H_2O, and either SO_2 or I_2 was continuously passed through the calibration source. This flow ensures that the source is thoroughly flushed with the desired gases and vapours, creating a controlled environment for subsequent calibration measurements. The produced OH radicals from the calibration source then undergo reactions with an excess amount of SO_2 or a moderate amount of I_2 to produce H_2SO_4 or HOI as the final products, respectively. Additionally, the HO_2 radical is produced as a by-product of the H_2SO_4 calibration process.

To quantify the concentrations of H_2SO_4, HOI, and HO_2, an open-source Python library based on two-dimensional convection-diffusion-reaction equations was developed (Marine Atmospheric paRticle FORmation and ChEmistry, MARFORCE, Shen

and He (2023)). This library aims to provide a framework for performing similar calibration tasks. Furthermore, it also allows
users to simulate and predict concentrations of other chemical species by adopting different chemical reaction schemes. The
MARFORCE library can be used as a tool in future research endeavours involving flow reactor chemistry simulations.

The SO_2 calibration is straightforward. The SO_2 flow from the SO_2 gas cylinder was diluted with humidified nitrogen and
the mixed sample was fed into the inlet. The normalised $SO_2 \cdot Br^-$ signal was further compared with the estimated SO_2 con-
centration to derive a calibration factor.

To calibrate the absolute concentration of H_2O, a dew point mirror hygrometer (DewMaster Chilled Mirror Hygrometer, Ed-
geTec) was employed. The dew point mirror hygrometer drew a sample from a branch of the humidified flow before it entered
the MION2 inlet tube. By measuring the dew point temperature, the dew point mirror hygrometer provides an accurate and
reliable determination of the absolute H_2O concentration in the sample. This calibration method ensures precise measurement
of H_2O concentration, which is important for accurate analysis and interpretation of the experimental data.

2.2.2 ~~Calibration~~ Development of ~~molecular~~ an iodine source

To calibrate the measured signals of $I_2 \cdot Br^-$ in Br^--MION2, we acquired its stable signals by utilising I_2 emitted from a per-
meation tube, which was regulated at a constant temperature and subjected to a continuous nitrogen stream (50 sccm).

The key to this calibration is determining the quantities of I_2 emitted from the permeation tube. We have previously cali-
brated the I_2 measurement of Br^--MION1 using a cavity-enhanced differential optical absorption spectroscopy (CE-DOAS)
instrument (Wang et al., 2021a), an UV/Vis spectrophotometer and an inductively coupled plasma mass spectrometer (ICP-
MS) (Tham et al., 2021; Wang et al., 2021a). As none of these instruments was available for this study, we further adapted an
alternative method.

The collection of the I_2 sample followed exactly the same procedure as described in our previous studies (Tham et al., 2021;
Wang et al., 2021a). Briefly, 50 sccm nitrogen carrier gas flow was passed through an I_2 permeation tube for 300 minutes, under
120-140 °C. The nitrogen carrier stream containing the released I_2 was bubbled through ~~an~~ a Schlenk-type impinger charged
with 20 mL of hexane kept at -70 °C by a dry ice/acetone bath. After completion of the sampling process, the absorption flasks
were allowed to warm to the ambient temperature and sealed with a Teflon-coated glass stopper. The solution was stored at 4
°C until further processing.

Inspired by Mishra et al. (2000), I_2 was converted into a non-volatile and stable derivative, followed by quantification of the
latter using gas or liquid chromatography. Mishra et al. (2000) quantified I_2 in aqueous matrices by gas chromatography–mass

spectrometry(GC/MS) after I_2 reacting with 2,6-dimethylaniline to form the corresponding 4-iodo-derivative.

An adaptation of this method was required as the iodine to be determined was diluted in hexane. Specifically, the iodine derivatization reaction was conducted directly with the hexane solutions in the presence of an aqueous buffer, to reduce losses associated with a hexane-to-water transfer. To avoid any losses of the volatile I_2 through evaporation, the reaction was conducted in hermetically sealed headspace vials, with efficient phase mass transfer being facilitated by vigorous magnetic stirring.

Control of the pH of the buffer was crucial for achieving high derivatization yields, with pH at 7.00 providing the most favourable level of conversion after 2 hours. Attempts to perform the derivatization reaction under homogeneous conditions in hexane in the presence of a variety of soluble organic bases (e.g., tertiary amines) returned poor yields and led to the formation of several side products, most probably due to iodine oxidation. Experiments using 1.00 mL aliquots of the I_2 sample solutions under investigation produced the derivative at the limit of detection, precluding a reliable quantification of the derivative by the reverse phase high-performance liquid chromatography (RP-HPLC/UV).

To improve the analytical sensitivity, 10 mL aliquots of the iodine sample solution were employed for derivatization. To boost the sensitivity further, a high volume (15 μL) of the concentrated derivatization solution was injected into the HPLC system. Unfortunately, the hexane in the injection solution and the high injection volume gave rise to retention time instability and peak distortion. Subsequent optimisation of the chromatographic method provided robust reverse phase chromatographic conditions. Specifically, this was achieved by using relatively weakly eluting isocratic conditions for sample elution, followed by strongly eluting conditions for column cleaning and reconditioning. Using the fully optimised protocol, the derivative could be readily quantified for 0.17 to 11.05 μg mL^{-1} initial iodine concentrations, with the LOD (limit of detection) and LOQ (limit of quantification) being 0.012 μg mL^{-1} and 0.035 μg mL^{-1}. Using this method, the hexane solution obtained by absorption of iodine from the permeation tube was found to contain 0.26 μg iodine mL^{-1}. Considering a total sample volume of 20 mL, the iodine output rate of the permeation tube under the employed conditions was calculated to be 17.3 ng min^{-1}.

It is worth noting that the sensitivity of the current method can be further improved by employing more sensitive separation and/or detection techniques, e.g., liquid ~~chromatography–mass~~ chromatography-mass spectrometry (LC/MS) or GC/MS.

2.2.3 Humidity dependence of analyte detection

An integral aspect of the characterisation involves investigating the influence of water on the detection of MION2 when employing the bromide chemical ionisation method. As water is essential in the calibration source to generate OH radicals, which subsequently yield either H_2SO_4 or HOI, we incorporated an additional dilution flow that merges with the calibration source through a Y piece (refer to Figure A3). This experimental configuration allows for the alteration of the absolute humidity of the sample, independent of the OH production rate in the calibration source. During the experiments with varying humidity,

the total flows of the dilution part and the flow reactor section were kept constant, while the relative humidity of the dilution flow was varied by mixing different combinations of dry and humidified flows. By employing this approach, we maintained a consistent level of systematic errors arising from the blending of the dilution and sample flows. By comparing the relative signal intensities of analyte-containing ions, we could examine the influence of water on the detection of different analytes (e.g., H_2SO_4, HOI and HO_2).

2.2.4 Quartz flow reactor setup

In order to study the sensitivity of Br^--MION2 to other oxidised iodine species, e.g., IO, OIO, HIO_3, I_2O_3, I_2O_4 and HIO_2, a quartz flow reactor with an inner diameter of 2.4 cm and a length of 94 cm was used. The residence time inside the quartz tube was 8.5 s. A green LED, with a wavelength of 528 nm, was hung on top and in parallel to the quartz flow reactor to initiate iodine photochemistry. In order to keep the temperature and light uniformity in the quartz flow reactor, the flow reactor was wrapped together with the green LED light by aluminium foil. The schematics of the setup are shown in Figure A4.

2.3 MARFORCE model description

As described above, calibration of H_2SO_4, HO_2, and HOI requires a numerical model to simulate the ~~chemical dynamics~~ radial diffusion, chemical reactions and transport in the calibration source and inlet tube. ~~This process~~ These processes determine the concentration of the analyte and can be simplified into a two-dimensional convection-diffusion-reaction problem. The concept of such a model was illustrated elsewhere (Kürten et al., 2012), specifically for the calibration of H_2SO_4. Our earlier studies also presented a numerical model for HOI calibration with similar principles but a simplified iodine chemistry scheme was instead implemented (Tham et al., 2021; Wang et al., 2021a). Nevertheless, neither of these studies made their calibration scripts publicly accessible, and the scripts lack adaptability for different chemistry schemes. Consequently, we have developed an open-source two-dimensional flow reactor model named MARFORCE to address these limitations. MARFORCE is built in Python and hosted on GitHub (Shen and He, 2023), allowing free access to interested users. The model comprises two main components: 1) the fluid dynamics simulation module and 2) the gas-phase photochemistry module.

2.3.1 Convection-diffusion-reaction equation

The convection-diffusion-reaction equation has been derived and discussed extensively in the literature (Gormley and Kennedy, 1948; Kürten et al., 2012) and is only briefly discussed here:

$$\frac{\partial c_i}{\partial t} = D_i\left(\frac{1}{r}\frac{\partial c_i}{\partial r} + \frac{\partial^2 c_i}{\partial r^2} + \frac{\partial^2 c_i}{\partial z^2}\right) - \frac{2Q}{\pi R^2}\left(1 - \frac{r^2}{R^2}\right)\frac{\partial c_i}{\partial z} + P \tag{1}$$

where i corresponds to a specific chemical (e.g., H_2SO_4), the c_i is the concentration, D_i is the diffusion coefficient, r is the distance in the radial direction, R is the radius of the flow reactor, Q is the total flow in the flow reactor, z is the distance in

the axial direction and P shows the production (positive values) or loss (negative values) rate due to chemical reactions. As the flow in tangential direction is symmetrical, the $\frac{1}{r^2}\frac{\partial^2 c_i}{\partial \theta^2}$ term has been ignored.

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The diffusion coefficient in the model can be defined in three ways: 1) manually defined using experimental values, 2) calculated by kinetic theory or 3) calculated based on elemental composition using Fuller's method (Fuller et al., 1966).

The convection and diffusion processes were validated against a theoretical prediction by Alonso et al. (2016). A fixed
300 amount of H_2SO_4 was set at the first cross-section of the MARFORCE simulation and H_2SO_4 was further carried to the outlet of a cylinder only by convection and diffusion processes. Comparing the H_2SO_4 profiles at the outlet yields on average a 0.4 % difference between the MARFORCE model and the theoretical prediction by Alonso et al. (2016) (Figure A5). This suggests that the convection and diffusion processes in the MARFORCE model are well simulated.

305 **2.3.2   Gas-phase photochemistry**

The photolysis and chemical reactions in the H_2SO_4, HO_2 and HOI calibrations can be simulated by a set of differential equations which describe the production and loss of various species. To make the MARFORCE model more versatile, the model was designed to accommodate the input file format from the Master Chemical Mechanism (MCM) (Jenkin et al., 1997; Saunders et al., 2003), a near-explicit chemistry mechanism for numerous organic precursors. The scripts used to compile and interpret
310 MCM mechanisms were adapted from O'Meara et al. (2021). The input file extracted from MCM is reshaped and the reaction equations, reaction rate coefficients, reactants, products, their indices, and stoichiometric numbers are generated accordingly. The temperature and pressure dependence of reaction rate coefficients are taken into consideration. Finally, differential equations for each species based on its production and loss processes are produced and solved. Additionally, the MARFORCE model leaves an option to set abundant species as constants, so their concentrations are assumed uniform and homogeneous in
315 the flow reactor. These species include, for example, O_2, N_2, SO_2, I_2, and H_2O in the H_2SO_4 and HOI calibration experiments. With its flexibility, the MARFORCE model can be readily adapted to simulate organic oxidation or any other experiments using a laminar flow reactor.

There are two default chemistry schemes provided in the MARFORCE model and they are used for the H_2SO_4 calibration
320 and HOI calibration, respectively. The reaction rate coefficients utilised in these two schemes are tabulated in Table A1. The most important procedure of these calibration experiments is to obtain the OH concentration. The OH concentration is determined by the photon intensity produced by the calibration source (It-product) and the absolute water content in the air passing through the calibration source. It-product refers to the product of UV light intensity at 185 nm and effective illumination time. In this study, we derived the It-product from the N_2O experiment, which was conducted under the same conditions as the
325 H_2SO_4 calibration experiments. The details of the It-product determination can be found in Kürten et al. (2012). In brief, the chemical actinometry method was employed, which involves the conversion of N_2O to NO_x (primarily NO), to determine the

light intensity. Since NO exhibits lower reactivity compared to OH and can be conveniently measured using commercial NO$_x$ monitors, the "It-product" can be derived accordingly. Considering that the calibration experiment duration is relatively short (a few hours) compared to the potential lifetime of the mercury lamp, it is reasonable to assume that the attenuation of the It-product over time is negligible.

The OH initial concentration is further defined as

$$[\text{OH}] = \text{It} \times \sigma_{\text{H}_2\text{O}} \times \Phi_{\text{H}_2\text{O}} \times [\text{H}_2\text{O}] \tag{2}$$

where $\sigma_{\text{H}_2\text{O}}$ is the absorption cross-section of water vapour, 7.22×10^{-22} cm^2 (Creasey et al., 2000), and $\Phi_{\text{H}_2\text{O}}$ is the quantum yield (unity in this case).

2.3.3 Flow mixture

In addition to its ability to simulate a standard cylindrical flow reactor with ~~a~~ uniform size, the MARFORCE model also possesses limited capabilities in two specific conditions: 1) simulating two interconnected flow reactors with varying sizes: the model is capable of simulating scenarios where two flow reactors of different sizes are connected. 2) simulating reactions when a dilution flow is merged with the sample flow through a Y-shaped tee. These additional features enhance the versatility of the MARFORCE model, allowing for a more comprehensive analysis of complex flow reactor systems.

The first design aims to cope with the different sizes of the chemical ionisation inlet and the calibration source itself. For example, the MION2 inlet utilises a KF25 connector with an inner diameter of 24 mm while the calibration source utilises a 3/4" tube with an inner diameter of 15.6 mm. Our model considers an instantaneous transition between the tubes of different sizes, i.e., the chemical distributions at the last cross-section of the first cylinder are copied into the first cross-section of the second cylinder while the axial flow speed is adjusted to the cross-section of the second cylinder. As this simplification ignores the convective transport of species to the walls at the transition region, it likely gives the concentration upper limit at the pinhole of the mass spectrometer. Since the inner diameter difference between the calibration source and the MION2 inlet is relatively small in this study, we expect that the resulting uncertainty is well within the overall systematic uncertainty of -50/+100 %.

The second design considers that an additional dilution flow is utilised to reduce the sample water content when entering the Br$^-$-MION2 inlet. Similarly, we assume an instantaneous transition at the spot where the dilution flow is added. In this case, both the chemical distribution and axial flow speed are changed since a new branch of flow is added. The simulation is carried out with a two-process procedure: before and after the dilution. First, we carry out a standard simulation before adding the dilution flow. Once the flow is fully developed and the chemical distribution reaches a steady state in the simulation, the last cross-section at the grid right before adding the dilution flow is stored and recalculated into the first cross-section of the next simulation. The second simulation is further carried out after considering the dilution flow, together with the changes in

360 chemical distribution and axial flow speed.

It should be noted that the fluid dynamics processes are overly simplified in the second design and therefore, this option should be used with caution. In this study, this option is necessary only because investigating the detection humidity effect of e.g., H_2SO_4, HO_2 and HOI requires adding a dilution flow after the calibration source. In order to estimate the magnitude of
365 error caused by the simplification of fluid dynamics, we carried out experiments comparing calibration results obtained with the first design (straight tube) and the second design (Y piece) and the results are shown in Figure A6. We find that the second design additionally introduces a 12 % higher calibration factor in the H_2SO_4 calibration and a 27 % higher calibration factor in the HOI calibration, compared with the calibrations using the first design. Therefore, the application of the second design for the purpose of this study is reasonable and does not introduce an excess amount of uncertainties. This mainly concerns the
370 H_2SO_4, HO_2 and HOI calibration experiments.

2.4 Quantum chemical calculations

Cluster fragmentation enthalpies were calculated using quantum chemical methods. The initial conformational sampling was performed using the Spartan'18 program. The cluster geometry was then optimised using density function theory (DFT)
375 methods at the ωB97X-D/aug-cc-pVTZ-PP level of theory (Chai and Head-Gordon, 2008; Kendall et al., 1992). Iodine and bromine pseudopotential definitions were taken from the Environmental Molecular Sciences Laboratory (EMSL) basis set library (Feller, 1996; Peterson et al., 2003). Calculations were carried out using the Gaussian 16 program (Frisch et al., 2016). An additional coupled-cluster single-point energy correction at the DLPNO-CCSD(T)/def2-QZVPP (Riplinger and Neese, 2013; Riplinger et al., 2013; Weigend and Ahlrichs, 2005) level of theory was carried out on the lowest energy conformers to refine
380 the DFT calculated enthalpies. The coupled-cluster calculation was performed using the ORCA program version 4.2.1 (Neese, 2012).

The master equation solver for multi-energy well reactions (MESMER) program was used to investigate the ionisation chemistry of $I_2O_3 \cdot HNO_3NO_3^-$. For the $I_2O_3 \cdot HNO_3NO_3^-$ complex, Lennard-Jones potentials of $\sigma = 6.5$ Å and $\epsilon = 300$ K were
385 used, which are identical to those used previously for similar iodine systems (Gálvez et al., 2013). The MesmerILT method was used with a pre-exponential value of 1.26×10^{-9} cm^3 molec^{-1} s^{-1}, which is equal to the $I_2O_3 + HNO_3NO_3^-$ collision rate calculated using the average dipole orientation (ADO) method. Varying the collision rate by a factor of 3 has no effect on the MESMER results, indicating that the reported final fragmentation rate coefficients of $I_2O_3 \cdot HNO_3NO_3^-$ are not sensitive to the accuracy of the computed collision rate.

390

3 Results and Discussion

3.1 Calibration of H_2SO_4, HOI and HO_2 using MARFORCE

Gaseous H_2SO_4 concentration is regularly measured around the globe using the nitrate chemical ionisation method. In this study, a direct H_2SO_4 calibration has been carried out for the MION2 inlet at tower 1 using either Br^- (Br^--MION2-T1, Figure A3) or NO_3^- (NO_3^--MION2-T1) chemical ionisation methods, and additionally at tower 2 with Br^- (Br^--MION2-T2, Figure A7) chemical ionisation method. The MARFORCE model is utilised to simulate the evolution of various species at the cross-section of the inlet tube as shown in Figure 2. The actual H_2SO_4 concentrations can be calculated by correlating the count rates, which represent the ratio of the measured H_2SO_4 signals to primary ions. Subsequently, the predicted H_2SO_4 concentrations are ~~further~~ compared with the measured normalised ~~ratios~~ signals to derive calibration factors (Table 1).

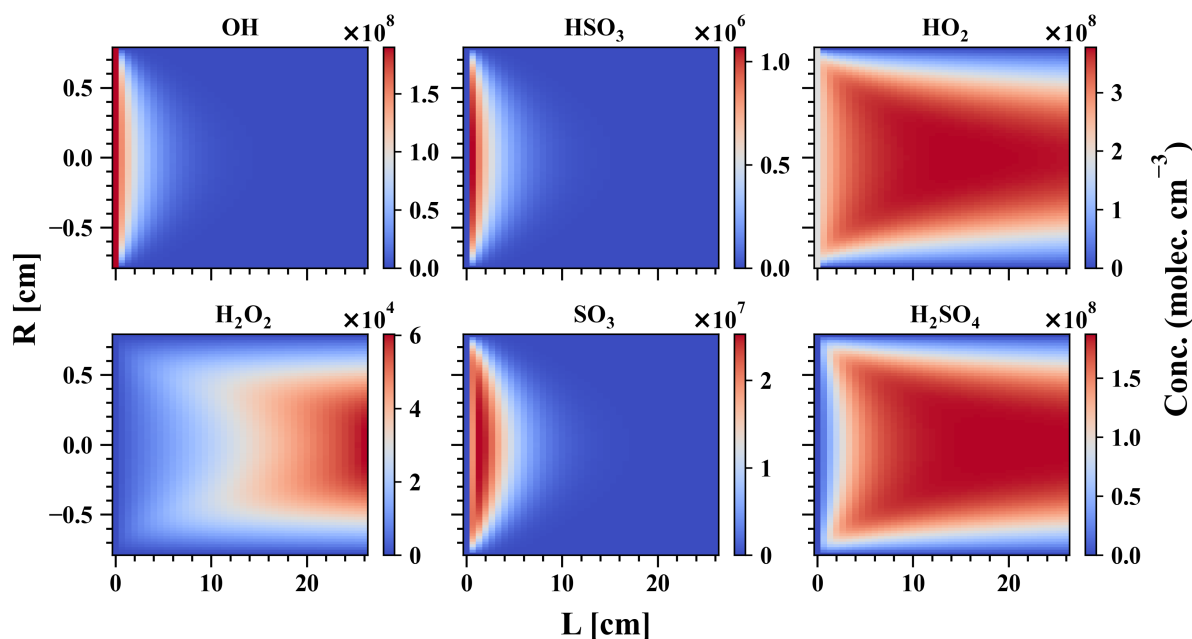



Figure 2. MARFORCE simulation results of a H_2SO_4 calibration experiment. The x-axis shows the distance from the UVP pen-ray lamp to the entrance of the chemical ionisation chamber and the y-axis shows the distance in the radial direction. Conditions for the simulation: R = 0.78 cm, L = 26 cm, sample flow = 10.6 slpm, $[SO_2] = 5.78 \times 10^{13}$ cm^{-3}, $[O_2] = 2.42 \times 10^{16}$ cm^{-3} and $[H_2O] = 2.8 \times 10^{16}$ cm^{-3}.

The derived calibration factor for Br^--MION2-T1 (8.1×10^9) is approximately eight times higher than that of Br^--MION2-T2 (9.8×10^8). This observation is consistent with the fact that the ionisation time from tower 2 to the pinhole (around 300 ms) is roughly 8.6 times longer than that of tower 1 (35 ms). A longer ionisation time leads to a greater conversion of Br^- and $H_2O \cdot Br^-$ into $H_2SO_4 \cdot Br^-$ or HSO_4^-, resulting in a lower calibration factor.

Table 1. Calibration ~~coefficients~~ factors and detection limits for selected species measured by the MION2 inlet and Eisele-type inlet. It should be noted that the reported numbers are specific to the experimental conditions and instrument tuning in our experiments. Different instrument tuning can result in different calibration ~~coefficients~~ factors and detection limits. Undesired impurities may result in elevated detection limits despite the calibration ~~coefficients~~ factors being the same for the analytes. Therefore, these numbers should not be applied to another study without carrying out the same calibration experiments described in this study.

Species	calibration factors (MION2)			Detection limit		
	Tower 1 (Ionisation Time = 35 ms)		Tower 2 (300 ms)	MION2 (Br^-)		Eisele inlet (NO_3^-) (160 ms)
				Tower 1	Tower 2	
	NO_3^-	Br^-	Br^-	(RH = 3.7%)	(RH < 0.1%)	(RH = 5.6%)
H_2SO_4	1.3×10^{10}	8.1×10^9 (RH = 0.2-23.3%)	9.8×10^8 (RH = 0.3-11.6%)	$^a 1 \times 10^5$	$^f 2.9 \times 10^4$	$^i 7.6 \times 10^4$
HOI	n/a	1.8×10^{10} (RH = 3-17%)	5.1×10^9 (RH = 3-17%)	$^b 5.9 \times 10^5$	$^g 1.6 \times 10^5$	n/a
HIO_3	n/a	n/a	n/a	$^a 1.3 \times 10^5$	n/a	$^i 9 \times 10^4$
HO_2	n/a	2.8×10^{11} (RH = 2.5%)	1.2×10^{11} (RH = 2.7%)	$^c 3.3 \times 10^6$ (RH = 2.7%)	$^h 5.7 \times 10^5$ (RH = 0.3%)	n/a
SO_2	n/a	2.6×10^{16} (RH = 10%)	2.1×10^{16} (RH = 9.9%)	$^d 1.8 \times 10^9$ (RH = 0.5%)	n/a	n/a
I_2	n/a	8.2×10^9 (RH = 26-37%)	n/a	$^e 3.3 \times 10^5$	n/a	n/a
IO	n/a	n/a	n/a	$^a 1.6 \times 10^5$	$^f 2.5 \times 10^4$	n/a
OIO	n/a	n/a	n/a	$^a 2.0 \times 10^5$	$^f 3.1 \times 10^4$	n/a
I_2O_2	n/a	n/a	n/a	$^a 1.9 \times 10^5$	$^f 3.5 \times 10^4$	n/a
I_2O_3	n/a	n/a	n/a	$^a 1.9 \times 10^5$	$^f 4.2 \times 10^4$	n/a
I_2O_4	n/a	n/a	n/a	$^a 1.9 \times 10^5$	$^f 3.0 \times 10^4$	n/a
I_2O_5	n/a	n/a	n/a	$^a 2 \times 10^5$	$^f 3.7 \times 10^4$	n/a

Unit: molec. cm^{-3}; "n/a" refers to "not available"; The experiments were conducted at room temperature. H_2SO_4 calibration factor is applied to estimate the detection limits of iodine oxides. Since iodine oxides may not be detected at the kinetic limit, their LODs are mere estimations and can be higher than the reported values in this study. The detection limits are estimated with 1-min data and one-hour data collection time. The RH reported in this table is calculated at 25 °C. The calibration factors and LODs have a systematic error of a factor of two (-50 %/+100 %).

Calibration factors used for the LOD calculation are: a. 8.1×10^9, b. 1.8×10^{10}, c. 2.8×10^{11}, d. 1.03×10^{14}, e. 8.2×10^9, f. 9.8×10^8, g. 5.1×10^9, h. 4.1×10^{10}, i. 3.5×10^9. H_2SO_4 calibration factor is applied to estimate the detection limits for iodine oxides. Since iodine oxides may not be detected at the kinetic limit, their LODs are mere estimations and can be higher than the reported values in this study. The detection limits are estimated with 1-min data and one-hour data collection time.

In the case of NO$_3^-$-MION2-T1, it exhibits a similar sensitivity to H$_2$SO$_4$ detection as Br$^-$-MION2-T1. This similarity is likely due to the consistent ionisation time (using tower 1) for both methods, since both the NO$_3^-$ and Br$^-$ chemical ionisation methods measure H$_2$SO$_4$ at the collision limit, as mentioned in previous studies (Kürten et al., 2012; Wang et al., 2021a).

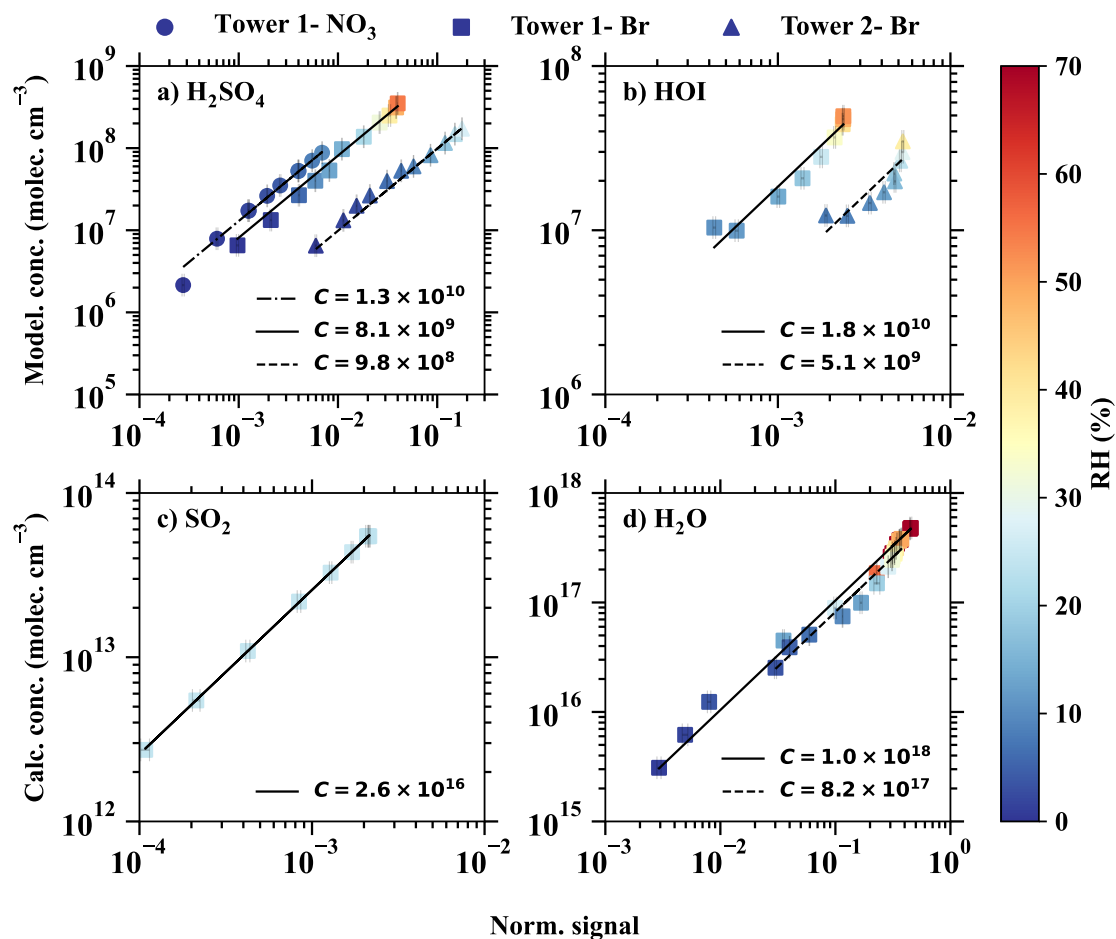



Figure 3. The modelled or calculated vapour concentrations vs. the normalised signals for a) H$_2$SO$_4$, b) HOI, c) SO$_2$, d) H$_2$O. The dashed-dotted, solid, and dashed lines are the linear fits of the results from different inlet modes: 1) tower 1 with the NO$_3^-$ chemical ionisation method, 2) tower 1 with the Br$^-$ chemical ionisation method, and tower 2 with the Br$^-$ chemical ionisation method. The slopes of the fitted lines represent the calibration ~~coefficients~~factors, shown in the legend. The colour bar shows the relative humidity in the calibration experiments.

410   Sanchez et al. (2016) has reported that the bromide chemical ionisation mass spectrometer (Br$^-$-CIMS) is capable of detecting HO$_2$ radicals at ambient relevant concentrations. In this study, we calibrated HO$_2$ together with H$_2$SO$_4$, as HO$_2$ is a by-product in the chemical production of H$_2$SO$_4$ (see Table A1). As the binding of HO$_2$ with Br$^-$ is significantly weaker

than that of H_2SO_4 with Br^-, the collision induced fragmentation of $HO_2 \cdot Br^-$ in the ion-optics of the mass spectrometer is larger (Passananti et al., 2019). Additionally, as the humidity effect of HO_2 will be shown to be strong in section 3.3, the

415 calibration ~~coefficient~~ factor of HO_2 has to be derived with respect to a specific humidity level. The derived HO_2 calibration factors at 2.5 - 2.7 % RH (25 °C) are 2.8×10^{11} and 1.2×10^{11}, respectively, for Br^--MION2-T1 and Br^--MION2-T2 (Table 1).

The HOI calibration was also carried out using the H_2SO_4 calibration source, except that the SO_2 source was replaced with an I_2 source. As can be seen in Figure 3 and Table 1, the calibration factor for HOI is roughly two times that of H_2SO_4. This

420 suggests that HOI is detected at close to the collision limit. It is worth noting that we find instrument setting affects HOI detection significantly since HOI is not strongly bonded to Br^-. The preferred fragmentation pathway is $HOI \cdot Br^- \longrightarrow HOI + Br^-$ (Table 2), and thus a fraction of $HOI \cdot Br^-$ dissociates into HOI and Br^- after passing the ion optics of the mass spectrometer. A more fragmentation-oriented setting can result in a higher fraction of $HOI \cdot Br^-$ getting lost in the ion optics, thus resulting in a higher calibration factor, i.e., lower sensitivity. As an example, in our earlier studies (Tham et al., 2021; Wang et al., 2021a),

425 we used a relatively fragmenting setting compared to the one used in this study in an attempt to reduce $(H_2O)_n \cdot Br^-$ clusters and other water-associated clusters. This experimental setup led to a calibration factor for HOI that was eight times higher than the calibration factor for H_2SO_4.

3.2 Calibration of H_2O and SO_2

430 $H_2O \cdot Br^-$ is a regular peak and one of the primary ions measured by the Br^--MION2. Br^--MION2 is, therefore, able to measure absolute water content if the $H_2O \cdot Br^-$ signal is calibrated against a dew point mirror instrument. Such a calibration has at least two purposes: 1) the calibrated $H_2O \cdot Br^-$:Br^- can be used as an indicator of the fragmentation level of the Br^--MION2 and 2) compared to regular relative humidity sensors and dew point mirrors, Br^--MION2 exhibits higher sensitivity towards H_2O. In this study, the calibration of H_2O was performed for both Br^--MION-T1 and Br^--MION-T2, as illustrated in Figure 3.

435 Interestingly, the calibration factors for both towers did not show significant differences. This can be attributed to the presence of an excess amount of H_2O, which establishes a rapid equilibrium with Br^- and $H_2O \cdot Br^-$ irrespective of the ionisation time.

As a reasonable binding enthalpy of $SO_2 \cdot Br^-$ was predicted using quantum chemical calculations (Table 2), we continued to check whether the Br^--MION2 allows us to detect SO_2. A variable amount of SO_2 was mixed with a fixed amount of dilu-

440 tion flow at a constant relative humidity (RH, 10 %) which was measured by the Br^--MION2. Clear $SO_2 \cdot Br^-$ was measured and it increased linearly with the SO_2 concentration in the sample flow (Figure 3). However, the calibration factor of SO_2 is roughly six orders of magnitude higher than that of H_2SO_4 at 10 % RH. This is consistent with the weaker binding of $SO_2 \cdot Br^-$ compared with $H_2SO_4 \cdot Br^-$. Additionally, SO_2 calibration is extremely sensitive to RH changes as can be seen in Figure 4. In this study, the best achieved detection limit was 9.4×10^7 cm^{-3} at an RH below 0.1 %. Theoretically, it is possible to enhance

445 the sensitivity even further by reducing the absolute water content.

Table 2. Fragmentation enthalpies (the opposite of binding enthalpies) of analytes with the Br^-. The cluster geometry was optimised at the ωB97X-D/aug-cc-pVTZ-PP level of theory at 298.15 K (Chai and Head-Gordon, 2008; Kendall et al., 1992). The enthalpies were calculated at the DLPNO-CCSD(T)/def2-QZVPP at 298.15 K

Cluster fragmentation pathway	Fragmentation enthalpies (kcal mol^{-1})
$I_2 \cdot Br^- \longrightarrow I_2 + Br^-$	33.3
$I_2 \cdot H_2OBr^- \longrightarrow I_2 \cdot Br^- + H_2O$	8.0
$IO \cdot Br^- \longrightarrow IO + Br^-$	24.5
$IO \cdot H_2OBr^- \longrightarrow IO + H_2O \cdot Br^-$	21.3
$IO \cdot H_2OBr^- \longrightarrow IO \cdot Br^- + H_2O$	9.9
$OIO \cdot Br^- \longrightarrow OIO + Br^-$	23.2
$OIO \cdot H_2OBr^- \longrightarrow OIO + H_2O \cdot Br^-$	22.1
$OIO \cdot H_2OBr^- \longrightarrow OIO \cdot Br^- + H_2O$	11.9
$I_2O_3 \cdot Br^- \longrightarrow IO_3^- + IBr$	24.6
$I_2O_4 \cdot Br^- \longrightarrow I_2O_4 + Br^-$	42.6
$I_2O_4 \cdot H_2OBr^- \longrightarrow I_2O_4 + H_2O \cdot Br^-$	48.8
$I_2O_4 \cdot H_2OBr^- \longrightarrow I_2O_4 \cdot Br^- + H_2O$	10.5
$HIO_3 \cdot Br^- \longrightarrow IO_3^- + HBr$	[a]29.9
$HIO_3 \cdot Br^- \longrightarrow HIO_3 + Br^-$	[a]35.7
$HIO_3 \cdot H_2OBr^- \longrightarrow HIO_3 + H_2O \cdot Br^-$	33.1
$HIO_3 \cdot H_2OBr^- \longrightarrow HIO_3 \cdot Br^- + H_2O$	11.2
$HIO_3 \cdot H_2OBr^- \longrightarrow IO_3 \cdot H_2O^- + HBr$	26.7
$HIO_2 \cdot Br^- \longrightarrow HIO_2 + Br^-$	[b]29.2
$HIO_2 \cdot H_2OBr^- \longrightarrow HIO_2 + H_2O \cdot Br^-$	15.5
$HIO_2 \cdot H_2OBr^- \longrightarrow HIO_2 \cdot Br^- + H_2O$	1.3
$HIO_2 \cdot H_2OBr^- \longrightarrow IO_2 \cdot H_2O^- + HBr$	27.4
$HOI \cdot Br^- \longrightarrow HOI + Br^-$	[b]26.9
$HOI \cdot H_2OBr^- \longrightarrow HOI + H_2O \cdot Br^-$	22.9
$HOI \cdot H_2OBr^- \longrightarrow HOI \cdot Br^- + H_2O$	9.6
$HOI \cdot H_2OBr^- \longrightarrow IO \cdot H_2O^- + HBr$	48.4
$H_2O \cdot Br^- \longrightarrow H_2O + Br^-$	13.2
$HO_2 \cdot Br^- \longrightarrow HO_2 + Br^-$	23.1
$H_2SO_4 \cdot Br^- \longrightarrow HSO_4^- + HBr$	[b]27.9
$H_2SO_4 \cdot H_2OBr^- \longrightarrow H_2SO_4 + H_2O \cdot Br^-$	36.1
$H_2SO_4 \cdot H_2OBr^- \longrightarrow H_2SO_4 \cdot Br^- + H_2O$	8.2
$H_2SO_4 \cdot H_2OBr^- \longrightarrow HSO_4 \cdot H_2O^- + HBr$	22.0
$SO_2 \cdot Br^- \longrightarrow SO_2 + Br^-$	19.4

[a] the fragmentation enthalpy is updated from Wang et al. (2021a) as a lower energy $HIO_3 \cdot Br^-$ cluster geometry, which has an additional Br-I interaction, has been located in this study (see Figure **??**). [b] Value adopted from Wang et al. (2021a).

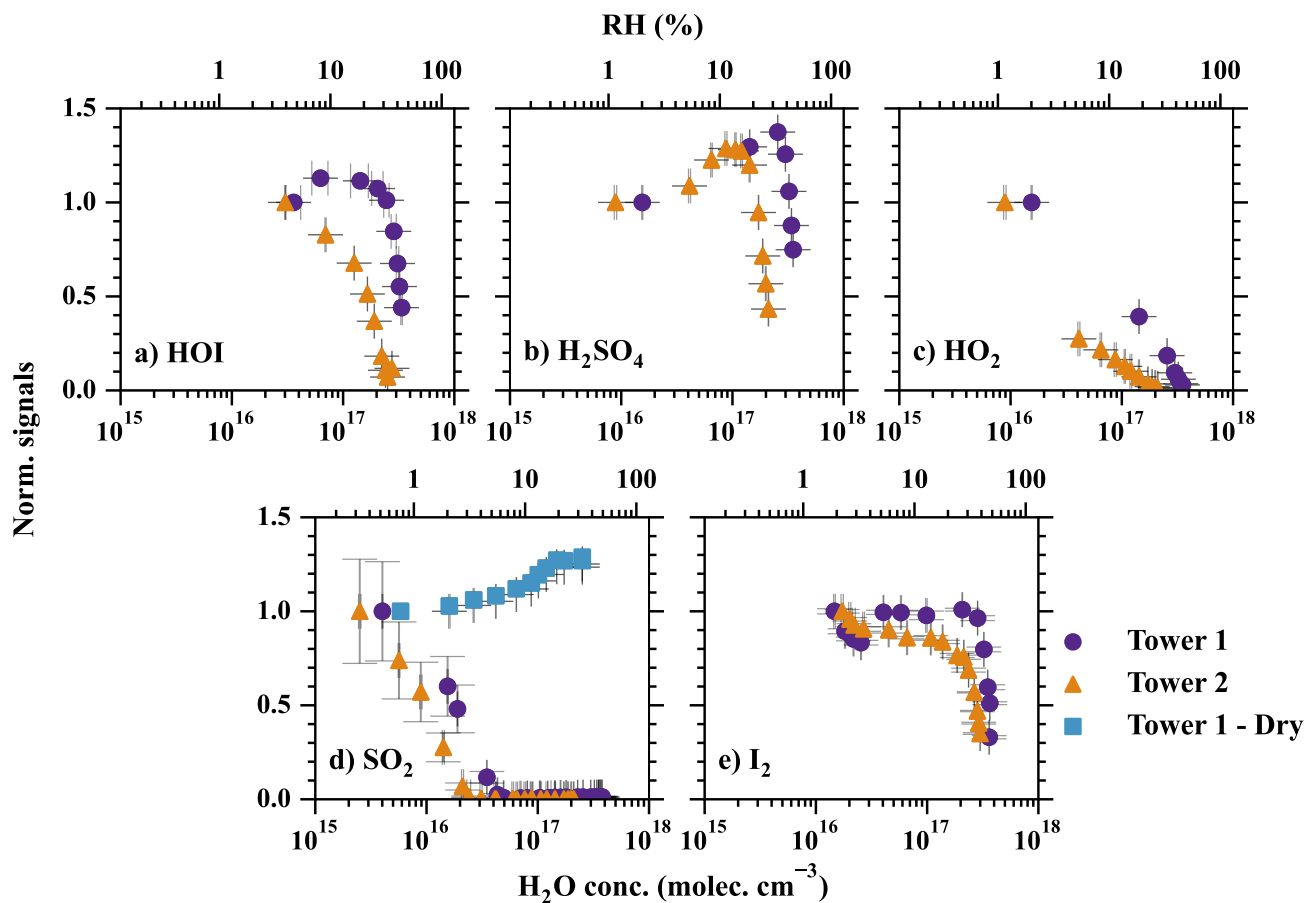



Figure 4. The effect of humidity on the detection efficiency of a) HOI, b) H_2SO_4, c) HO_2, d) SO_2, and e) I_2. The measured signals in each set of experiments are normalised by the signal at the lowest RH. Therefore, the normalised signals represent how the increasing RH is affecting the detection limit compared with the initial point. The purple circles and orange triangles show the detection humidity effects of tower 1 and tower 2, respectively. The blue squares refer to the experiments conducted with a dry flow added before the MION2 inlet. The RH is converted from absolute H_2O concentrations at 25 °C. Error bars represent one standard deviation.

3.3 Detection humidity effect

The measurement sensitivity of ~~halide anion based~~ the halide anion-based chemical ionisation method was regularly reported to be affected by atmospheric water content (Kercher et al., 2009; Mielke et al., 2011; Woodward-Massey et al., 2014; Lee et al., 2014). The humidity effect of atmospheric pressure Br^--MION2 could be amplified due to the higher water content present in air samples. Although Wang et al. (2021a) has demonstrated that the detection of I_2 by Br^--MION1 was not affected within a limited humidity variation (40 - 80 % RH at -10 °C), characterisation under a wider range of humidity conditions is needed. As the detection humidity effect in this study exclusively refers to the effect of absolute humidity instead of relative humidity,

absolute humidity parameters such as dew/frost point or H_2O molecule concentration are commonly presented together with the relative humidity (at 25 °C, otherwise notified).

In this study, we examine the detection humidity effect of H_2SO_4, HOI, HO_2, SO_2 and I_2 with RH from below 1 % to 60 % at 25 °C. Unlike H_2SO_4, HO_2 and HOI, which require generation from a calibration source, both SO_2 and I_2 have their own standardised sources. This simplifies their control during the characterisation of the detection humidity effect. Therefore, a straight flow reactor is used to premix the analyte containing air sample to the Br^--MION2 (Figure A2). It is worth noting that we do not account for the wall loss of SO_2 and I_2 in the analysis. The wall loss of SO_2 is negligible at the time scale of the calibration processes (few seconds). Despite I_2 vapour can both condense on and evaporate from the walls of the flow reactor, equilibrium can be achieved given a sufficient amount of time. In our experiments, it took up to 24 hours to reach equilibrium. Once equilibrium is established, the condensation and evaporation of I_2 balance each other out, making the estimation of I_2 concentration straightforward.

On the other hand, the characterisation of the detection humidity effect of H_2SO_4, HOI and HO_2 is more challenging as the production of these species is nearly proportional to the amount of H_2O passing the calibration source. Therefore, an experimental apparatus was built which enabled humidifying the air sample after the calibration source, thus without disturbing HO_x production processes in the calibration source (Figure A3).

The results of the humidity characterisation are shown in Figure 4. Although only five species were characterised and observed for their distinct humidity sensitivity, a general conclusion can be drawn that applies to essentially all of the species: an excessive amount of water content leads to a decrease in detection sensitivity. The species with stronger binding with Br^- exhibits less sensitivity to changes in humidity (e.g, H_2SO_4 and I_2), while the weakly bonded ones (HOI, SO_2 and HO_2) are strongly affected. The humidity tolerance of the measured species can be ordered as I_2 > HOI > HO_2 > SO_2 which is the same order as the strength of their bindings with Br^- (Table 2).

Interestingly, the detection humidity effect of H_2SO_4 is observed to be non-linear, i.e., the detection sensitivity of H_2SO_4 first increases with higher RH but eventually has a sharp drop at around 40 % RH. The enhancement of H_2SO_4 detection at below ca. 33 % RH could be contributed by two mechanisms. First, the diffusivity of H_2SO_4 is lower at higher RH (Hanson and Eisele, 2000). A higher RH, therefore, reduces the wall deposition of H_2SO_4 in the inlet tube, thus effectively increasing the detected H_2SO_4. This is a universal factor that influences all H_2SO_4 detection techniques with appreciable sampling line residence time. The second possibility is that at low RH regime, H_2O does enhance H_2SO_4 detection by offering more modes through which the excess energy of the cluster can dissipate in the formation of $H_2SO_4 \cdot Br^-$, thus resulting in a relatively more stable cluster (Iyer et al., 2017). Regardless of the sources of the detection humidity effect at the low water content regime, the maximum systematic error is measured to be 37 % by comparing the experiment carried out at 2 % RH (frost point of -25 °C) and the experiment carried out at 33 % (dew point of 7.6 °C) in Figure 4b. Based on our findings, we anticipate that

the detection humidity effect of H_2SO_4 would be moderate when the dew point is below approximately 7.6 °C. However, it is important to exercise caution when conducting measurements under higher absolute humidity conditions.

Additionally, a longer ionisation time by utilising the Br^--MION2-T2 results in a stronger detection humidity effect as shown in Figure 4. This phenomenon is the most significant for HOI, i.e., the detection of HOI is more humidity dependent using Br^--MION2-T2 than Br^--MION2-T1. This phenomenon also elucidates the curvature observed in the HOI calibration when employing the MION2-T2 (Figure 3): the diminished detection sensitivity of HOI counterbalances the augmented HOI production at elevated water content. Although this effect is difficult to quantify, it practically suggests that the Br^- chemical ionisation method should employ a shorter ionisation time (i.e., using the ~~tower 1~~T1) when operating MION2 with multiple chemical ionisation methods.

In summary, we find that the detection of Br^--MION2 is strongly affected by air water content. The atmospheric pressure Br^- chemical ionisation method is suitable for laboratory experiments where water content is controlled and atmospheric observations in the cryosphere where air water content is low. Nevertheless, the humidity effect should be considered individually for different analytes and the binding enthalpy between the analyte and Br^- is likely a good indicator. As the NO_3^--MION2 (or the NO_3^- chemical ionisation in general) is known to have minimal detection humidity sensitivity, it is commonly operated together with the Br^--MION2. Performing a cross-check of mutually measured species, such as H_2SO_4, HIO_3, and oxidised organic species, will provide crucial insights into whether and when the detection capability of Br^--MION2 is compromised by the water content in the air. In this context, the new design of Br^--MION2, which enables three chemical ionisation methods to have the same ionisation time, is essential.

3.4 Attempts to reduce the detection humidity effect

Various approaches were explored to mitigate the detection humidity effect. One commonly used method is to employ a low-pressure chemical ~~ionization~~ ionisation system, which has been successfully implemented in iodide chemical ~~ionization~~ ionisation systems (Lee et al., 2014) and bromide chemical ~~ionization~~ ionisation systems (Wang et al., 2021a). However, reducing the relative humidity (RH) of the air sample comes at the expense of reducing the measurement sensitivity for species detected at the collision limit, such as H_2SO_4, HIO_3 and I_2, as the air sample unavoidably undergoes dilution in this process. We estimated previously that the Br^--FIGAERO inlet had more than 10 times higher detection limit compared to the Br^--MION1 inlet (Wang et al., 2021a). For example, the Br^--FIGAERO had an HIO_3 detection limit of 5.1×10^6 cm^{-3} which struggles to detect atmospheric levels of HIO_3 (commonly below 10^7 cm^{-3}) (He et al., 2021b). The lower level of detection limit provided by the Br^--MION2 inlet is therefore essential in the detection of iodine species. Another important factor is the reaction of halogen radicals with analytes. Besides halogen anions, halogen radicals can also be produced by chemical ionisation processes. While iodine radical (I·) mostly reacts with halogen species and a minimal number of organic species, bromide radical (Br·) reacts with a wider range of organic species as it has a larger reactivity. Conventional low-pressure systems that involve

mixing analytes with reagent gases, such as the FIGAERO inlet, can introduce additional complexities when interpreting mass spectra. As a result, alternative approaches were pursued to effectively reduce the detection humidity effect.

525

The first method is the dilution method. Instead of measuring the air sample directly, a dry dilution flow was mixed with the air sample at the entrance of the Br^--MION2 inlet (see Figure A4). We tested this method for the SO_2 detection with an air sample flow of 1.8 slpm and a dilution flow of 20.7 slpm (Figure 4). The x-axis for this set of experiments represents humidity in the air sample instead of the humidity after the dilution to compare with the experiments without adding the dilution flow.

530 We observe a significantly reduced detection humidity effect compared to the case without dilution. It is noteworthy that as the air sample was diluted by a factor of 12.5, the detection limit of the instrument is likely enhanced by the same factor. However, since the detection humidity effect for SO_2 is significantly higher than other species (e.g., H_2SO_4, HOI and I_2), the dilution is still effective for SO_2 measurement. For example, no $SO_2 \cdot Br^-$ signal would not be measured at 40 % RH (25 °C) if the air sample is not diluted but a noticeable signal would be measured if the air sample is diluted. A similar conclusion is likely

535 applicable to other species but with a different optimal humidity cut-off.

The second method is additionally introducing a core-sampling device that uses the air sample as the core flow and a dry synthetic air flow as the sheath flow (Figure A8). This takes advantage of the fact that H_2O diffuses into the sheath flow faster than other analytes with larger molecular weight, thus effectively reducing the RH in the core flow from which the instrument

540 pinhole collects the most sample. Nevertheless, it is important to note that the core-sampling method, while helping to mitigate the detection humidity effect, also leads to a reduction in the $SO_2 \cdot Br^-$ signal. This is because the SO_2 itself gets diluted, partially countering the benefits of the reduced detection humidity effect.

Various sample-to-sheath flow combinations were tested as presented in Figure 5. The measured $SO_2 \cdot Br^-$ signal from all

545 sets of experiments was normalised by the experiment with the sample-to-sheath ratio of 21:1 at 0.21 % RH (25 °C). The results indicate that reducing the sample-to-sheath ratio effectively alleviates the SO_2 detection humidity effect. It is observed that different mixing ratios have only a moderate impact on the measured $SO_2 \cdot Br^-$ when the H_2O concentration is below 10^{16} cm^{-3} (1 % RH), indicating a low detection humidity effect in such conditions. However, the core-sampling device clearly enhances the SO_2 detection efficiency when the H_2O concentration is larger than 10^{16} cm^{-3}. The sample-to-sheath ratio of

550 1:21 enables effective detection of SO_2 at around 4.5×10^{17} cm^{-3} (60 % RH) of H_2O while the sample-to-sheath ratio of 21:1 is not able to detect SO_2 after around 4.3×10^{16} cm^{-3} of H_2O (6 % RH). Overall, the sample-to-sheath ratio of 1:21 is at least two orders of magnitude more effective in detecting SO_2 when H_2O is greater than 2×10^{16} cm^{-3}. Therefore, the core-sampling method is an effective method for reducing the detection humidity effect of species which are weakly bonded with Br^-. Despite the reduced detection humidity effect, it is important to note that the sample water content still impacts the

555 detection limit of SO_2. Therefore, dedicated experiments need to be conducted to accurately determine the concentration of SO_2.

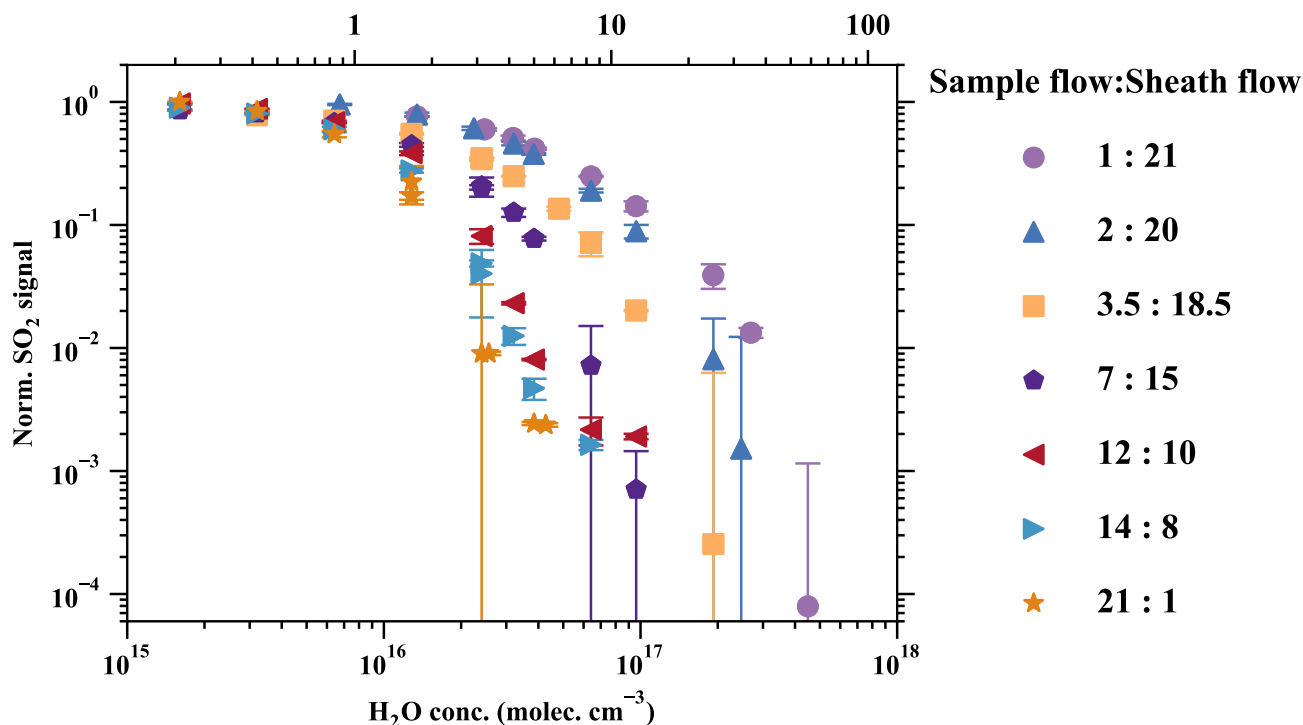



Figure 5. Reducing the detection humidity effect with the core sampling method (Figure A8). This design takes ~~the~~ advantage of the faster diffusion of H_2O than SO_2 from the sample flow to the sheath flow and effectively reduces the RH in the sample flow. Various sample-to-sheath ratios were tested at different H_2O concentrations to find the optimal setting. All the data are normalised to the lowest RH data point in the sample-to-sheath = 21:1 experiment. Due to experimental constraints, the sample-to-sheath ratios = a) 3.5:18.5, b) 2:20 and c) 1:21 experiments started only from the second, third and fourth lowest RH points, respectively. All other experiments collected data in all humidity conditions. The error bars represent the standard deviation of the normalised SO_2 signals.

3.5 Limit of detection

The limit of detection (LOD) is an essential parameter for a chemical ionisation inlet system. For the convenience of inter-comparison, we define the LOD in this study as:

$$LOD = \mu + 3 \times \sigma \tag{3}$$

where μ is the mean value of one-hour mass spectrometric data with a one-minute time resolution and σ is the standard variation of the same data. Both μ and σ include the experimentally derived calibration ~~coefficient~~ factor. The species without direct calibration utilise the calibration ~~coefficient~~ factor of H_2SO_4, thus the LODs for these species generally represent the lower limit. The LOD is determined by introducing pure nitrogen or synthetic air into the chemical ionisation inlet, where none of the species listed in Table 1 are expected to be present. It is important to emphasise that this LOD definition is specifically

suitable for distinguishing trace gas concentrations from background levels in long-term observations.

The reported LODs can be affected by many factors. Some of these factors are 1) the purity of the reagent source (e.g., HNO$_3$ or CH$_2$Br$_2$ solution), 2) the purity of the sample air used at the LOD determination experiment, 3) the signal-to-noise (electronic background noise) ratio of the instrument, 4) the fragmentation level (controlled by the tuning of the instrument) of the mass spectrometer, 5) the humidity of the sample air used at the LOD determination experiment (for Br$^-$ chemical ionisation method) and 6) different ways of estimating LODs.

Therefore, comparing the LODs derived in this study with earlier studies may not be meaningful. Hence, we additionally compared the H$_2$SO$_4$ LOD of the MION2 inlet with the widely-used Eisele-type inlet, both attached to the same mass spectrometer (Table 1). The direct comparison suggests that the Br$^-$-MION2-T1 LOD is roughly 30 % higher than the LOD of the Eisele inlet, thus a comparable performance. When we increased the ionisation time from 35 ms (Br$^-$-MION2-T1) to 300 ms (Br$^-$-MION2-T2), the LOD of Br$^-$-MION2 for H$_2$SO$_4$ is further reduced by a factor of three, thus Br$^-$-MION2-T2 performs better than the Eisele inlet. This suggests that the MION2 inlet can achieve comparable (Br$^-$-MION2-T1) or even better (Br$^-$-MION2-T2) LOD than the Eisele inlet. Additionally, the Eisele-type inlet was regularly shown to have a LOD as low as 10^4 cm^{-3} (Jokinen et al., 2012), a well-performing mass spectrometer may further reduce the LOD of MION2. Nevertheless, the attained levels of LOD are sufficiently low for atmospheric measurements. The molecules in question typically require concentrations above 10^6 cm^{-3} to exert a significant influence on atmospheric chemistry and aerosol formation.

3.6 Voltage scanning and cluster formation enthalpy

Collision induced cluster fragmentation is an unavoidable issue which affects the detection of analytes that are weakly bonded with the reagent ion. Since if a charged cluster is loosely bonded, collisions between charged clusters and air molecules in the atmospheric pressure interface may break a large portion of the charged clusters apart prior to reaching the detector (Passananti et al., 2019). Therefore, charged cluster binding strength is an important factor determining whether an analyte-reagent ion cluster can be measured by the mass spectrometer (Iyer et al., 2016; Lopez-Hilfiker et al., 2016; Wang et al., 2021a). Lopez-Hilfiker et al. (2016) has shown that the level of collision induced cluster fragmentation is associated with the voltage differences between the first and second quadrupoles in the atmospheric pressure interface of the mass spectrometer. The voltage difference was shown to be indicative of the fragmentation level of the CIMS and it positively correlates with the cluster formation enthalpy (Iyer et al., 2016).

In this study, we carried out voltage scan experiments with the same procedures as described in Lopez-Hilfiker et al. (2016). Briefly, we kept the voltage differences inside two individual quadruples constant while changing the voltage difference between these two quadruples to modulate energies in the collision processes and the results are shown in Figure 6. Generally, a higher voltage difference indicates a higher fragmentation level which in turn results in a lower remaining fraction of charged

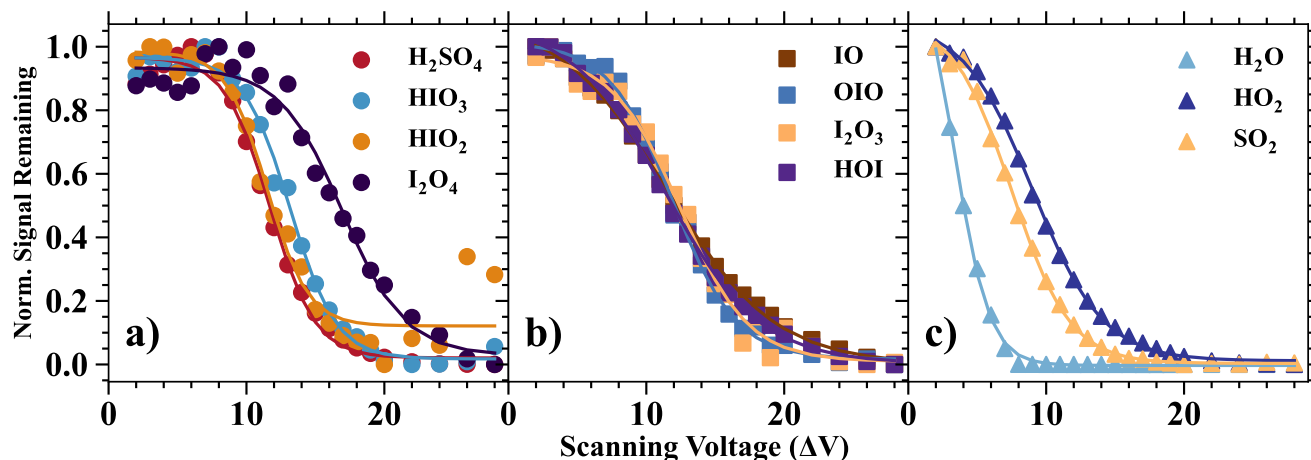



Figure 6. Normalised signal remaining vs. the scanning voltage (ΔV). The normalised signal remaining of each species is normalised by the maximum and minimum values of its values with different ΔV (or dV). The ΔV describes the voltage difference between the skimmer and the second quadruple and can be considered an indicator of the softness of the instrument tuning (Lopez-Hilfiker et al., 2016). A higher ΔV commonly indicates a more fragmenting setting.

clusters. Charged clusters that are less sensitive to voltage changes, especially in the low voltage difference regime are more stable.

A series of iodine oxides and oxoacids is evaluated together with other inorganic species such as H_2O, HO_2, SO_2 and H_2SO_4
605 (Figure 6). Based on the results, we categorise the analytes into three categories: 1) analytes which are strongly bonded with Br^-, 2) analytes which are moderately bonded with Br^- and 3) analytes which are weakly bonded with Br^-. The species H_2SO_4, HIO_3, HIO_2, and I_2O_4 can be classified into the first category since the initial change in voltage difference does not have a significant impact on the normalised signal. This indicates that these species are detected at the collision limit. It is also apparent that H_2O, HO_2 and SO_2 belong to the third category since a small increase in the voltage difference leads to
610 substantially reduced normalised ratios. Finally, IO, OIO, I_2O_3 and HOI are moderately bonded with Br^-. These moderately bonded charged clusters can reach a close to collision limit detection if the instrument is softly tuned (the voltage difference is small) but their detection sensitivity can change dramatically if the instrument fragmentation level is high. Lopez-Hilfiker et al. (2016) defined a parameter ΔV_{50} (dV$_{50}$, i.e., the dV value at half the maximum of the signal remaining) to describe the analyte and reagent ion binding strength. In this study, the dV$_{50}$ is defined by the following equation:

615 $$\text{NSR} = \frac{\text{SR}}{1 + e^{-k \times (\text{dV} - \text{dV}_{50})}} + \text{SR}_{\text{max,pred}} \qquad (4)$$

where NSR is the normalised signal remaining, SR is the signal remaining, dV$_{50}$ is the desired fitted value as represented in Figure 7 and SR~~max,pred~~ max,pred is the fitted value that represents the maximum SR when a compound does not undergo

fragmentation while passing through the ion optics.

620   Additionally, formation free enthalpies of various charged clusters are calculated using quantum chemical calculations (see Methods) and are compared with dV_{50} as shown in Figure 7. The two sets of parameters, consisting of theoretical predictions and measurements of the binding strength, provide a consistent understanding as demonstrated in previous studies (Lopez-Hilfiker et al., 2016; Iyer et al., 2016). In summary, strongly bonded charged clusters exhibit larger fragmentation-free enthalpies, larger dV_{50} values, and lower calibration factors. Examples of such species include $H_2SO_4 \cdot Br^-$ and $I_2 \cdot Br^-$. On

625   the other hand, weakly bonded charged clusters exhibit opposite properties, including species like $HO_2 \cdot Br^-$, $H_2O \cdot Br^-$ and $SO_2 \cdot Br^-$.

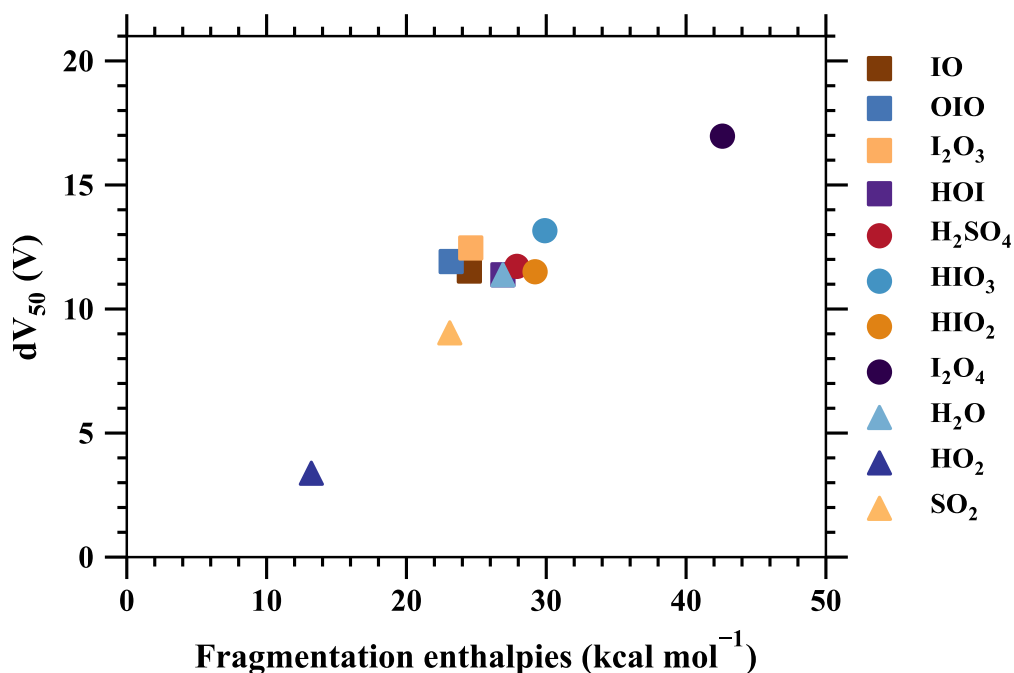
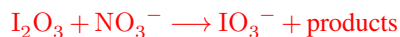



Figure 7. The voltage at which 50 % of analyte-bromide adducts have dissociated (dV_{50}) vs. the fragmentation free enthalpies of the adducts (Table 2).

3.7   Validation of the iodic acid (HIO_3) measurement ~~of iodine-containing species~~

Oxidised iodine vapours have been shown to influence atmospheric oxidation capacity (Saiz-Lopez et al., 2014; Sherwen et al.,
630   2016; Wang et al., 2021b) and particle formation processes (Hoffmann et al., 2001; O'Dowd et al., 2002). Recent publications have proposed iodine oxoacids as the critical driver for iodine particle formation processes (Sipilä et al., 2016; Baccarini et al., 2020; He et al., 2021b, a; Zhang et al., 2022; Liu et al., 2023). However, active debate remains concerning the presence of

gaseous HIO_3 and whether HIO_3 plays an important role in atmospheric aerosol nucleation. For example, a recent laboratory study shed doubts on the existence of gaseous HIO_3 as the authors only managed to measure HIO_3 in the particle phase with a photoionisation mass spectrometer but not in the gas phase. They concluded that the particle phase HIO_3 was formed from higher iodine oxides instead of from gaseous HIO_3 (Gómez Martín et al., 2020). Furthermore, they proposed a hypothesis that the IO_3^- signal, previously attributed to gaseous HIO_3 measurements using NO_3^--CIMS (Sipilä et al., 2016), could also originate from ~~(where y = 2-4)~~ I_2O_{2-4} species. Their evidence is primarily the exothermicity of the ~~reactions with~~ reactions from I_2O_{2-4} + NO_3^- ~~which forms~~ to IO_3^- ~~as part of the products~~. However, it should be noted that exothermic reactions do not guarantee that the reactions occur at significant rates ~~. For example, reactions such as~~

$$I_2O_3 + NO_3^- \longrightarrow IO_3^- + products$$

~~involves breaking multiple strong I-O and N-O bonds, which are likely associated with high kinetic barriers . As a result, it can be anticipated that this reaction does not occur as rapidly as the reaction + ⟶ + , where only a single proton transfer reaction takes place.~~

as various transition states and barriers exist in these reactions. Therefore, direct instrument validation is desired. In a more recent study, Gómez Martín et al. (2022) alternatively used the nitrate chemical ionisation method and detected gaseous HIO_3, consistent with our earlier studies (Sipilä et al., 2016; He et al., 2021b, a; Wang et al., 2021a; Finkenzeller et al., 2022) regarding the existence of gaseous HIO_3. ~~However, the authors also~~ The authors suggested that the measured $HIO_3 \cdot NO_3^-$ ion, previously interpreted as HIO_3, could potentially be formed through reactions such as the following:

$$I_2O_3 \cdot HNO_3 NO_3^- \longrightarrow IONO_2 + HIO_3 \cdot NO_3^- \tag{5}$$

due to the reaction being exothermic. ~~Besides the same reasons noted above, this~~ This hypothesis is challenged by the fact that the reaction

$$I_2O_3 \cdot HNO_3 NO_3^- \longrightarrow I_2O_3 \cdot NO_3^- + HNO_3 \tag{6}$$

is a favoured pathway compared to the reaction 5 as shown in Figure 8. ~~We further estimate~~ The authors of Finkenzeller et al. (2022) additionally further estimated that the MESMER derived overall rate coefficients at 298 K, 1 atm for reactions 5 and 6 are 2.3×10^{-12} cm^3 molec^{-1} s^{-1}, and 1.26×10^{-9} cm^3 molec^{-1} s^{-1}, respectively. Therefore, the yield of the reaction 6 is close to unity and cannot affect the HIO_3 detection.

It is essential to highlight that our previous studies (He et al., 2021b; Finkenzeller et al., 2022) as well as the studies by Gómez Martín et al. (2020, 2022) have consistently concluded that I_2O_4 is the predominant form of I_2O_y. Fortunately, the gaseous I_2O_4 species can be effectively measured using both the NO_3^- and Br^- chemical ionisation methods. Finkenzeller et al. (2022) calculated the cluster formation enthalpy of $I_2O_4 \cdot NO_3^-$ as -45.6 kcal mol^{-1}, which indicates that the $I_2O_4 \cdot NO_3^-$ cluster is extremely stable. Gómez Martín et al. (2020) found that the $I_2O_4 + NO_3^- \longrightarrow$ products $+ IO_3^-$ reaction is endothermic thus less likely to occur. The same principle applies to the Br^- chemical ionisation method as well. As mentioned earlier

in the previous section, voltage scan experiments have shown that the $I_2O_4 \cdot Br^-$ cluster is the most stable among the clusters investigated (refer to Figure 7). Consequently, I_2O_4 is detected at the collision limit using the Br^- chemical ionisation method, and it does not fragment into species such as IO_3^-. Given that the measured concentration of I_2O_4 is more than one order of magnitude lower than that of HIO_3 according to previous studies (Wang et al., 2021a; He et al., 2021b; Finkenzeller et al., 2022), it is unlikely that ~~species have~~ I_2O_{3-4} has a significant impact on the detection of HIO_3.

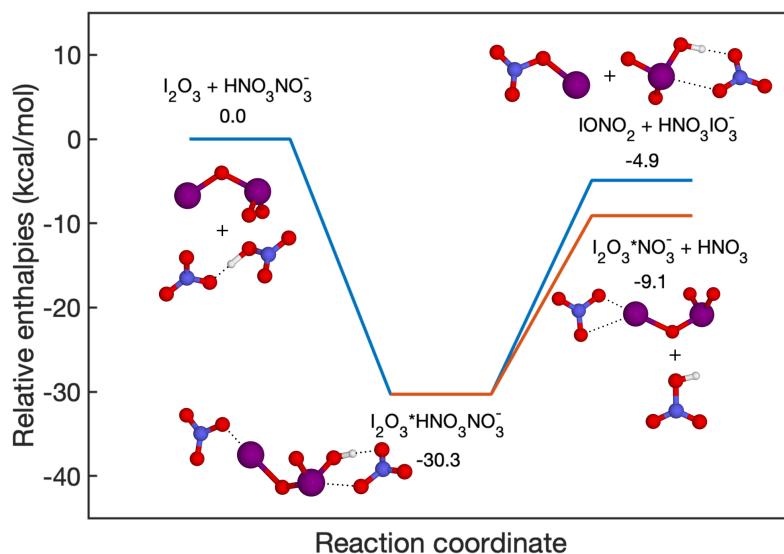



Figure 8. Fragmentation pathways of $I_2O_3 \cdot HNO_3NO_3^-$. The enthalpies are calculated at the DLPNO-CCSD(T)/def2-QZVPP//wB97X-D/aug-cc-pVTZ-PP level of theory.

Most importantly, complex and distinct chain reactions lead to the formation of higher iodine oxides and iodine oxoacids~~are formed through complex and distinct chain reactions. Laboratory~~. Conducting laboratory experiments with elevated iodine concentrations could inevitably ~~disturb the ratio of iodine oxides to~~ disrupt the balance between iodine oxides and iodine oxoacids. The concentration of iodine monoxide (IO) is ~~commonly considered a good indicator of the intensity of atmospheric~~

~~iodine activities and was~~ widely regarded as a reliable indicator of atmospheric iodine chemistry intensity and has been shown to influence the ratio of iodine oxides ~~and~~ to iodine oxoacids (Finkenzeller et al., 2022). ~~We capitalised on this phenomenonand conducted~~ Leveraging this phenomenon, we executed chemical perturbation experiments by varying the concentration of ozone (O_3) ~~, while keeping the concentration~~ while maintaining constant concentrations of iodine (I_2) and ~~the light intensity constant in~~ light intensity within a laminar flow reactor. The concentrations of IO were carefully controlled to span from levels below to

a few parts per trillion by volume levels. These experiments were replicated for both the Br^--MION2-T1 and NO_3^--MION2-T1, as ~~illustrated~~ depicted in Figure 9. The measured IO_3^- signal ~~is~~ was compared with $HIO_3 \cdot NO_3^-$ and $IONO_2 \cdot NO_3^-$

signals from the NO_3^--MION2-T1~~and to~~, and with $HIO_3 \cdot Br^-$, $I_2O_3 \cdot Br^-$, and $I_2O_4 \cdot Br^-$ signals from the Br^--MION2-T1~~to find out the origin~~, in order to ascertain the source of IO_3^-. It is ~~worth noting~~ noteworthy that the gaseous signals of HIO_3 ($HIO_3 \cdot NO_3^-$ and $HIO_3 \cdot Br^-$) exhibit a perfectly linear relationship with the ~~signals of~~ IO_3^- signals. However, the signals of

685     $IONO_2 \cdot NO_3^-$, $I_2O_3 \cdot Br^-$, and $I_2O_4 \cdot Br^-$ demonstrate a non-linear dependence on IO_3^-. This ~~suggests~~ observation implies that the primary source of IO_3^- is gaseous HIO_3, ~~since if does contribute to~~, as a non-linear correlation between $HIO_3 \cdot NO_3^-$ ~~and~~, $HIO_3 \cdot Br^-$ ~~with~~, and IO_3^- would be expected ~~.~~ if I_2O_{3-4} significantly contributed to IO_3^-. Furthermore, if the proposed reaction 5 was to occur at a substantial rate, one would anticipate the $IONO_2 \cdot NO_3^-$ signal to display a non-linear dependence on IO_3^- and $HIO_3 \cdot Br^-$. However, this is not observed.

690

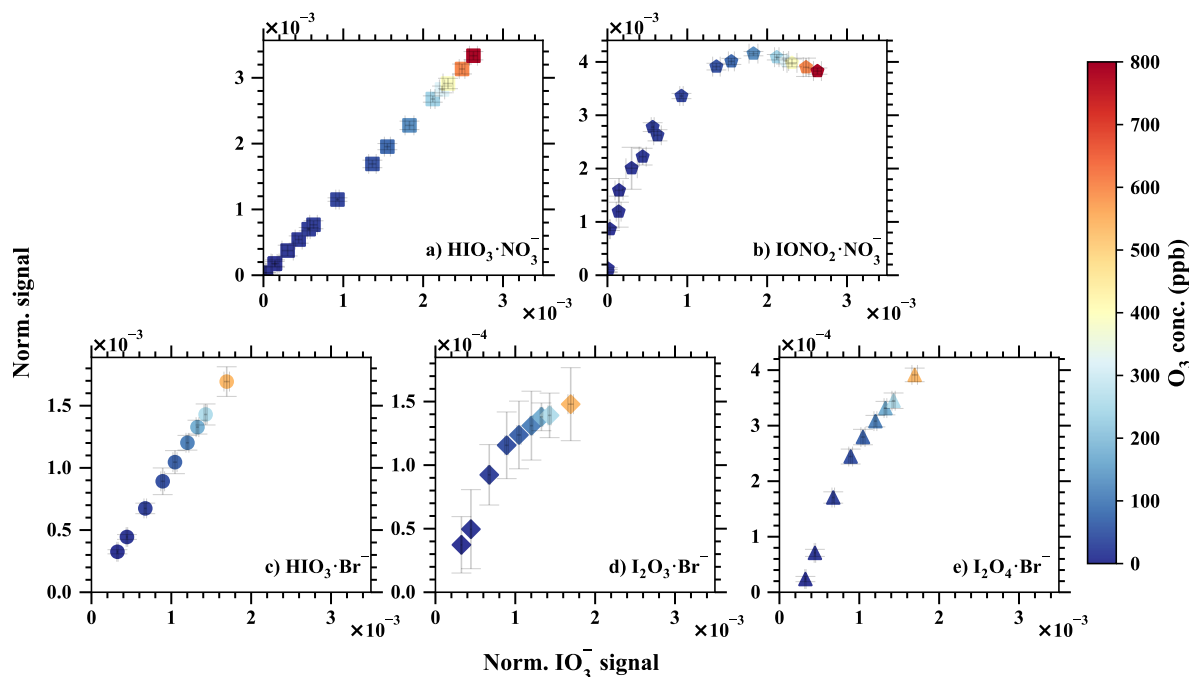



Figure 9. The normalised IO_3^- signal vs. the normalised signals of a) $HIO_3 \cdot NO_3^-$, b) $IONO_2 \cdot NO_3^-$, c) $HIO_3 \cdot Br^-$, d) $I_2O_3 \cdot Br^-$, and ~~d~~e) $I_2O_4 \cdot Br^-$. The iodine injection and light intensity were kept constant but the O_3 concentration was varied to modulate the ratio of iodine oxides to oxoacids. Error bars show one standard deviation. Notice the different y-axis scales.

~~Therefore, we conclude that is unlikely to significantly contribute to~~ Hence, we deduce that the contribution of I_2O_{3-4} to the IO_3^- ~~signal under atmospheric conditions . Experiments conducted with ambient-level precursors consistently demonstrate that~~ and $HIO_3 \cdot NO_3^-$ signals within the marine boundary layer conditions is improbable. Consistent experimentation with precursor concentrations at ambient levels consistently reveals a notable scarcity of gaseous I_2O_4 ~~is considerably less abundant~~

695   when compared to HIO_3 (He et al., 2021b, a; Finkenzeller et al., 2022). ~~Model~~ Furthermore, simulations of iodine chemistry ~~at~~ conducted for the Maïdo observatory ~~further~~ indicate that the ~~combined~~ collective concentration of I_2O_3 and I_2O_4 ~~is only~~

~~around~~ accounts for a mere 1 % of HIO$_3$, thus making it unlikely to ~~affect~~ exert any substantial influence on HIO$_3$ measurements ~~and the formation~~ or the generation of iodine particles in marine boundary layer conditions (Finkenzeller et al., 2022).

4   Summary and conclusion

In this study, we present an upgraded version of the multi-scheme chemical ionisation inlet, known as MION2. It is capable of simultaneously operating in atmospheric ion measurement mode and employing multiple chemical ionisation methods. ~~Although~~ While the fundamental concept of this inlet remains the same as MION1 (Rissanen et al., 2019), ~~the~~ MION2 enhances operational stability and enables the concurrent use of multiple chemical ionisation methods with the same ionisation time. Additionally, we find that the new version significantly improves performance by effectively focusing reagent ions, resulting in lower limits of detection (LODs). ~~Moreover, it enhances operational stability and enables the concurrent use of multiple chemical ionisation methods with the same ionisation time.~~

We further developed a Python open-source flow reactor kinetic model (MARFORCE, see Shen and He (2023)) to simulate convection-diffusion-reaction equations in cylindrical flow reactors to calibrate gaseous species such as H$_2$SO$_4$, HOI, and HO$_2$. The model is also compatible with the widely-used Master Chemical Mechanism, thus allowing future implementation of other chemical mechanisms.

~~The~~ Furthermore, we undertook a comprehensive characterisation of the MION2 inlet~~was further characterised for the detection of various inorganic speciesusing the~~'s capabilities in detecting an array of inorganic species, employing both Br$^-$ and NO$_3^-$ chemical ionisation ~~methods at two different~~ techniques with distinct ionisation times. By combining the analytical calibration with the MARFORCE model, we quantified the photochemical production of H$_2$SO$_4$, HOI, and HO$_2$ ~~were calibrated by employing the MARFORCE model, which quantifies their photochemical production in~~ within a flow reactor. ~~Based on our estimations, the limits of detection (LODs) are approximately~~ We reveal that the LODs hover around 10^5 molec. cm^{-3} (averaged over a 1-minute ~~data averaging~~interval) for species ~~like~~ such as H$_2$SO$_4$ and HIO$_3$~~when the ionisation time is set to~~, with an ionisation time of 35 ms. ~~By using~~ With a longer ionisation time (300 ms), the LOD for H$_2$SO$_4$ ~~is further reduced~~ experiences further reduction to 2.9 × 10^4 molec. cm^{-3} (approximately 1 ~~ppqv). A direct comparisondemonstrates that~~ part per quadrillion by volume). Upon direct comparison, the MION2 inlet ~~exhibits comparable or even better LODs compared to the widely-used~~ demonstrates equivalent or superior LODs in contrast to the widely-employed Eisele inlet (Jokinen et al., 2012). ~~Hence, this upgraded version~~ Thus, this enhanced iteration of the inlet ~~offers exceptional sensitivityfor the~~ showcases exceptional sensitivity, making it a formidable asset for the precise measurement of trace gases relevant to atmospheric particle formation.

Furthermore, we conducted an assessment of the detection capabilities of SO$_2$ and I$_2$ since they serve as crucial precursors for H$_2$SO$_4$ and HIO$_3$ respectively. We found that the Br$^-$-MION2 inlet is capable of detecting SO$_2$ by diluting a gas cylinder

containing a known quantity of SO_2. In addition to our previously established methods for calibrating gaseous I_2 (Wang et al., 2021a; Tham et al., 2021), we successfully employed a derivatization approach in conjunction with high-performance liquid chromatography to quantify the iodine permeation rate, which was found to be as low as 17.3 ng min^{-1}. The I_2 calibration using the Br^--MION2 inlet further confirms that I_2 is detected at the collision limit, similar to H_2SO_4, and aligns with our previous estimations (Wang et al., 2021a).

As the Br^--MION2 measures H_2O in the form of $H_2O \cdot Br^-$, we quantified the H_2O detection with a dew point mirror instrument by running them side by side. As a large portion of Br^- is converted to $H_2O \cdot Br^-$ in the ion-molecule reaction chamber, we predicted the fragmentation pathways of analyte-$H_2O \cdot Br^-$ clusters using quantum chemical calculations. We show that H_2O evaporates from the analyte-$H_2O \cdot Br^-$ clusters when passing the ion optics of our mass spectrometer due to the weak attachment of H_2O to the charged clusters. However, the chemical signature of the analyte is commonly preserved as the analyte-Br^- cluster or deprotonated analyte anion.

~~The detection using the~~ We have observed that the application of the Br^- chemical ionisation method at atmospheric pressure is susceptible to the ~~presence of excessive moisture in the air. Analytes that are detected at the collision limit, such as~~ influence of elevated moisture levels in the ambient air despite its enhanced detection sensitivity. We find that analytes detected at collision limits, including H_2SO_4, HIO_3, and I_2, ~~exhibit a significant decline~~ experience a noticeable reduction in measurement sensitivity when the dew point ~~exceeds~~ rises beyond 0.5 - 10.5 °C (equivalent to 20 - 40 % RH). ~~Furthermore~~ Moreover, the detection of ~~weakly bonded analytes~~ analytes with weaker bonds, such as HO_2 and SO_2, is ~~more profoundly affected by~~ notably more profoundly impacted by variations in water content, even ~~when~~ in situations where the dew point ~~is~~ remains below 0 °C. ~~For instance, the limit of detection (LOD)~~ To illustrate, LOD for HO_2 is approximately one order of magnitude higher than that of H_2SO_4 at ~~a relative humidity (RH) of~~ 2.7 %~~RH~~, while the LOD for SO_2 ~~is~~ surpasses that of H_2SO_4 by approximately three orders of magnitude ~~higher than that of at~~ when the RH falls below 0.1 %~~RH.~~. These outcomes show the critical role played by atmospheric water content in the effectiveness of Br^- chemical ionisation, particularly for species characterised by weak bonding interactions. Such insights contribute to our understanding of the method's limitations and provide valuable considerations for optimising analytical conditions in future atmospheric studies.

In order to mitigate the impact of humidity on detection, two methods, namely the dilution method and the core-sampling method, were tested in this study. We found that both methods effectively reduce the influence of humidity on detection. By employing these methods, it becomes possible to detect ambient levels of SO_2 (below 1 part per billion by volume) even at RH levels of up to 50 %, which would otherwise be challenging. However, it should be noted that the use of these methods inevitably results in sample dilution, thereby affecting the detection of species that are less affected by air water content, such as H_2SO_4, HOI, and I_2. Therefore, these methods should be employed selectively, when there is a specific objective, such as detecting extremely low levels of SO_2 or when the sample's dew point is higher than 10 °C (40 % RH). This implies that atmospheric pressure Br^- chemical ionisation is suitable for laboratory experiments with controlled RH and for ambient mea-

765 surements in relatively cold environments. When interpreting data obtained through the atmospheric pressure Br⁻ chemical ionisation method, it is crucial to carefully account for the influence of water by employing analytical characterisation or predicting fragmentation enthalpy. Despite these considerations, the MION2 inlet, which allows for the concurrent operation of the water-insensitive NO_3^- chemical ionisation method and the water-sensitive yet more versatile Br⁻ chemical ionisation method, provides a more comprehensive understanding of atmospheric conditions compared to using either of these methods

770 in isolation.

Finally, we validated the measurement of gaseous HIO_3 using both the NO_3^- and Br⁻ chemical ionisation methods. The signal of HIO_3 typically consists of IO_3^- and either $HIO_3 \cdot NO_3^-$ or $HIO_3 \cdot Br^-$, depending on the chemical ionisation method employed. Through experimental and theoretical validation, we confirmed that all three ions primarily originate from genuine

775 gaseous iodic acid and that iodine oxides do not contribute to the formation of these ions under ~~atmospherically relevant conditions.~~ marine boundary layer conditions.

Code availability. The MARFORCE model is shared through GitHub repository (https://github.com/momo-catcat/MARFORCE-flowtube). Other data analysis codes can be requested from the corresponding authors.

Data availability. Data is available upon request from the corresponding authors.

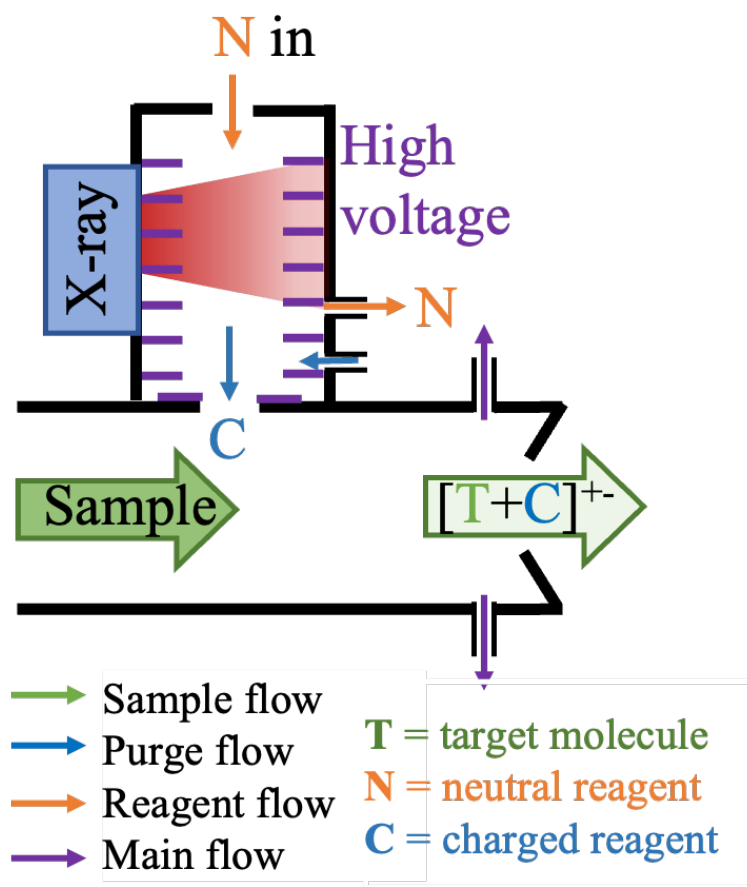



Figure A1. Schematic of an ionisation source of the MION2 inlet.

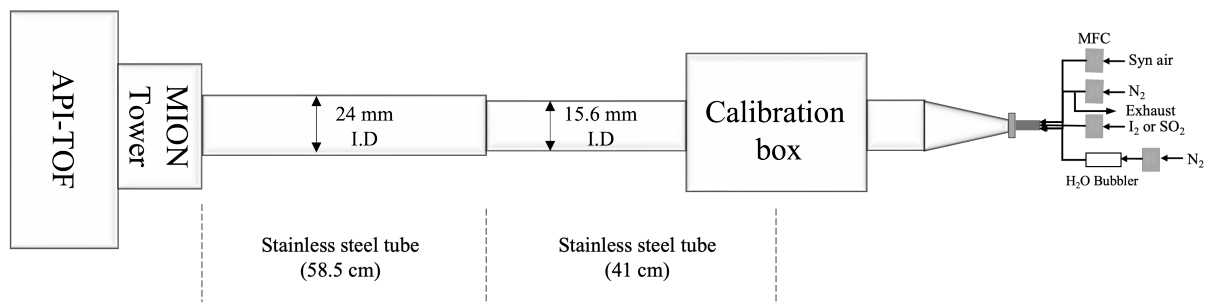



Figure A2. Schematic of a typical calibration experiment connecting the MION2 inlet (I.D. 24 mm) with the calibration source (I.D. 15.6 mm).

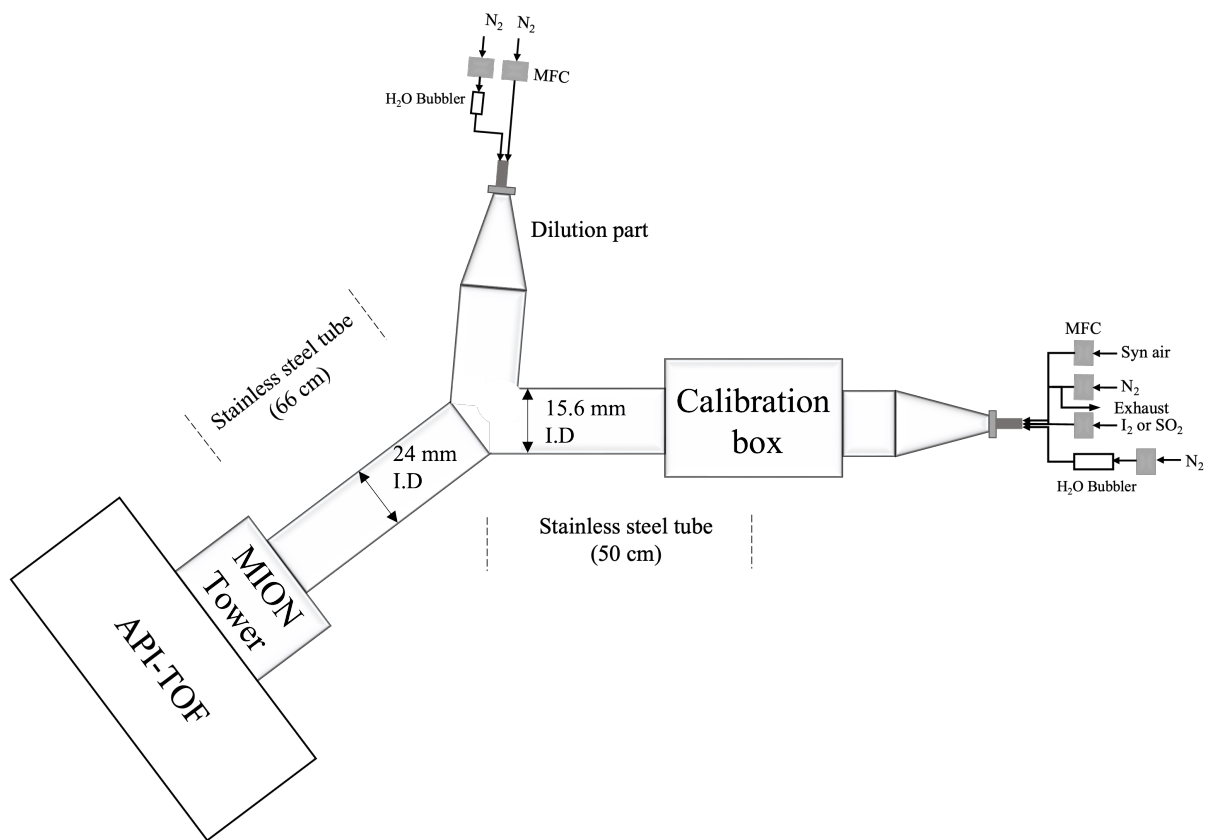



Figure A3. Schematic of the setup for examining the detection humidity effect of H_2SO_4, HOI and HO_2.

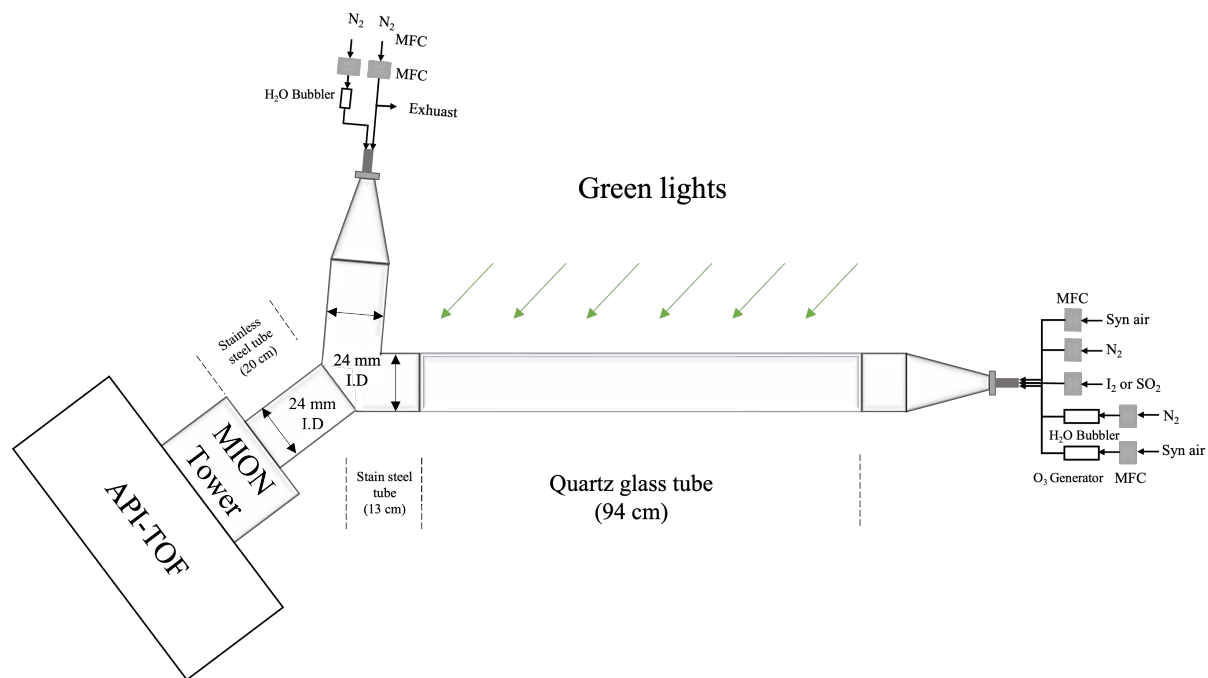



Figure A4. Schematic of the experimental setup for iodine chemistry experiments to produce higher concentrations of iodine oxides and oxoacids.

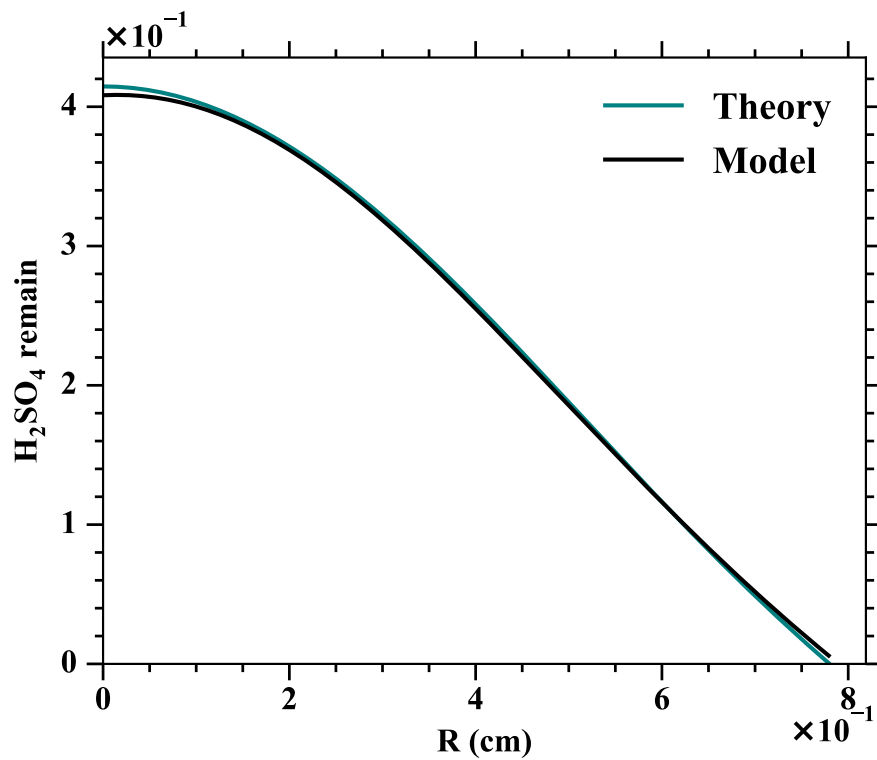



Figure A5. Comparison of the H_2SO_4 profiles at the outlet of a flow reactor. Theoretical values are predicted using Alonso et al. (2016) and the model results indicate the MARFORCE simulation. In both the theoretical prediction and the MARFORCE model, the tube length is assumed to be two meters, the inlet flow is set to 10 slpm and the diffusivity of H_2SO_4 is set to 0.088 $cm^2 s^{-1}$.

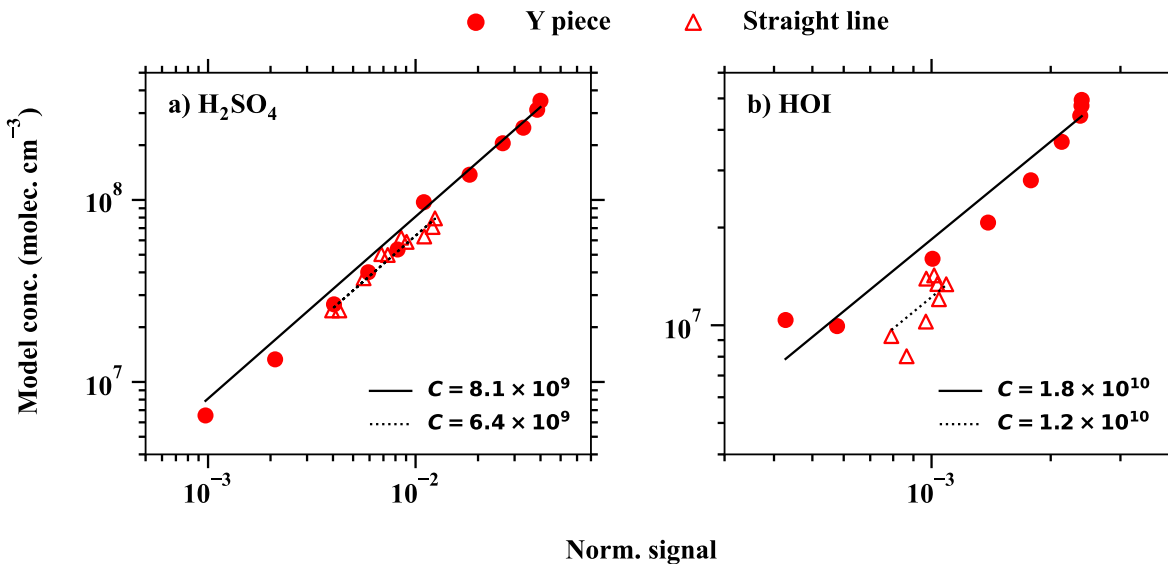



Figure A6. Comparing calibration experiments of a) H_2SO_4 and HOI with a straight tube (Figure A2) or additionally with a dilution flow (Figure A3). The difference in the calibration ~~coefficients~~ factors between the two experimental setups is the result of the less accurate representation of fluid dynamics when the dilution flow is added (Figure A3).

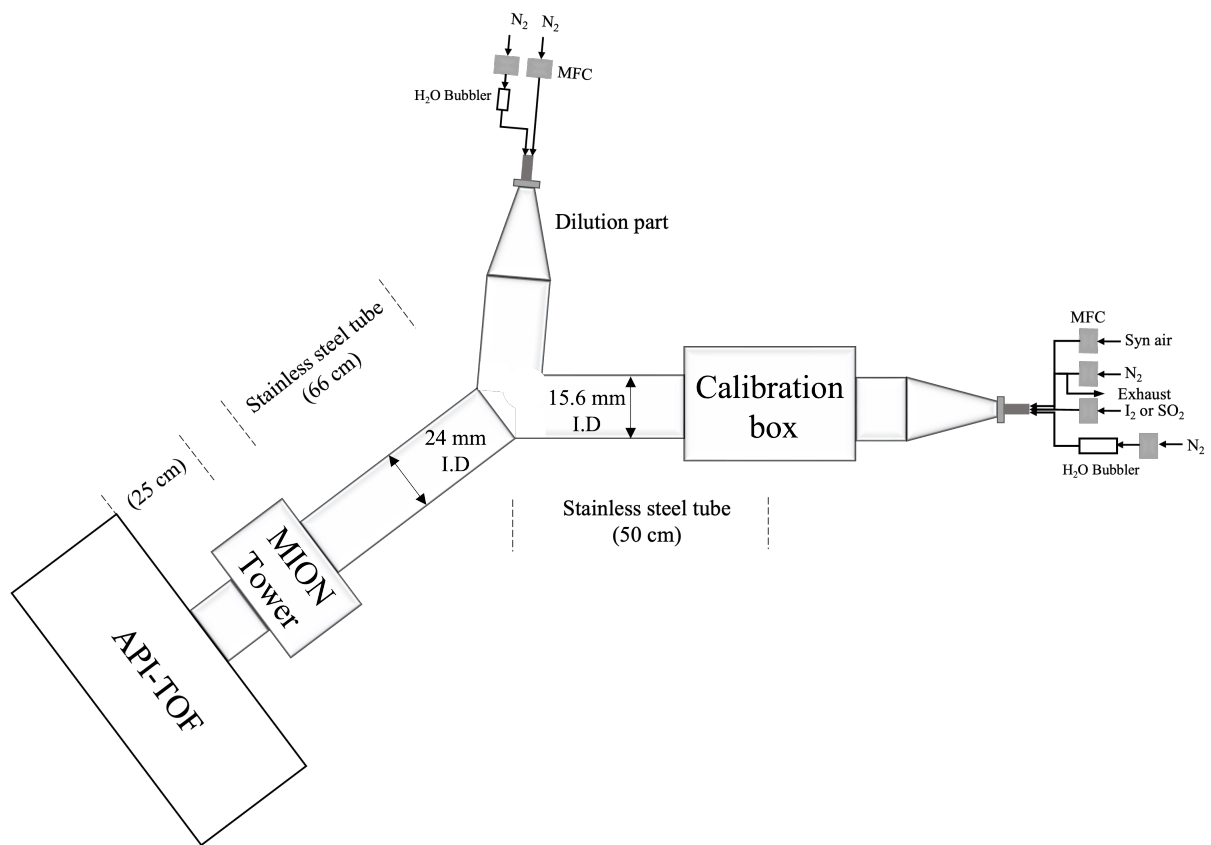



Figure A7. Schematic of the setup for H_2SO_4, HOI and HO_2 calibration experiment with the tower 2. The difference between this setup and the one shown in Figure A3 is that the position of the MION2 tower is changed from tower 1 to tower 2.

~~geometry optimised at the ωB97X-D/aug-cc-pVTZ-PP level of theory at 298.15 K. Color coding: Iodine = purple, oxygen = red, hydrogen = white, bromine = brown.~~

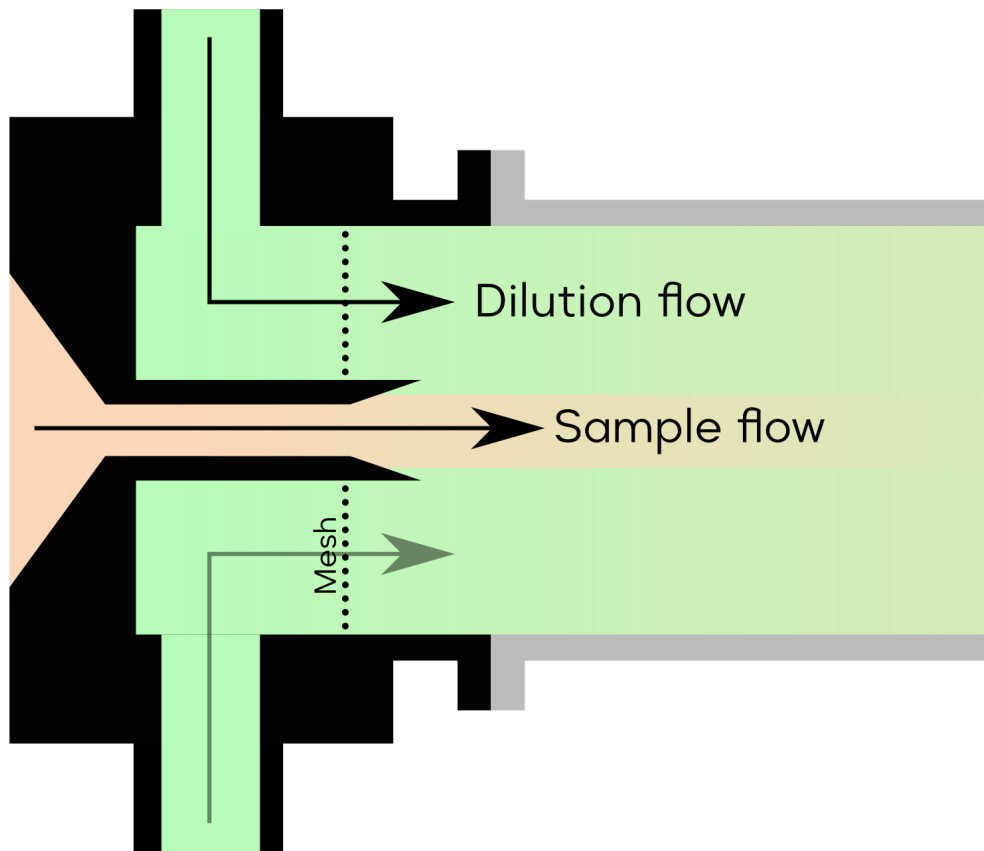



Figure A8. The configuration of the core-sampling device (Karsa Ltd.) which is used for adjusting the sheath and sample flows. This core-sampling piece features three ports for the dilution flows which pass through a mesh and further mixed with the sample flow.

Table A1. Chemical reactions and the reaction rate coefficients used for H_2SO_4 and HOI calibration experiments

Chemical reactions	Reaction rate coefficients
H_2SO_4 calibration:	
1. $OH + SO_2 \Longrightarrow HSO_3$	[a]$1.32 \times 10^{-12} \times (Temp/300)^{-0.7}$
2. $OH + HO_2 \Longrightarrow H_2O + O_2$	[b]$4.8 \times 10^{-11} \times \exp(250/Temp)$
3. $HO_2 + HO_2 \Longrightarrow H_2O_2$	[b]$(2.2 \times 10^{-13} \times \exp(600/Temp)+$ $1.9 \times 10^{-33} \times M \times \exp(980/Temp)) \times KMT06$
4. $OH + OH \Longrightarrow H_2O_2$	[c]$2 \times 6.9 \times 10^{-31} \times (Temp/300)^{-0.8} \times p/(1.38 \times 10^{-23})/Temp/10^6$
4.5. $OH + OH \Longrightarrow H_2O$	[b]$6.2 \times 10^{-14} \times (Temp/298)^{2.6} \times \exp(945/Temp)$
6. $HSO_3 + O_2 \Longrightarrow HO_2 + SO_3$	[b]$1.3 \times 10^{-12} \times \exp(-330/Temp)$
7. $SO_3 + 2H_2O \Longrightarrow H_2SO_4$	[b]$3.9 \times 10^{-41} \times \exp(6830.6/Temp)$
HOI calibration:	
1. $IO + IO \Longrightarrow I + I$	[d]$0.11 \times 5.4 \times 10^{-11} \times \exp(180/Temp)$
2. $IO + IO \Longrightarrow OIO + I$	[d]$0.38 \times 5.4 \times 10^{-11} \times \exp(180/Temp)$
3. $IO + IO \Longrightarrow I_2O_2$	[d]$0.45 \times 5.4 \times 10^{-11} \times \exp(180/Temp)$
4. $I_2 + OH \Longrightarrow HOI + I$	[e]2.1×10^{-10}
5. $IO + OIO \Longrightarrow I_2O_3$	[f]$w1a \times \exp(w2a \times Temp)$
6. $OIO + OIO \Longrightarrow I_2O_4$	[f]$w1b \times \exp(w2b \times Temp)$
7. $IO + OH \Longrightarrow HO_2 + I$	[g]1.0×10^{-10}
8. $HI + OH \Longrightarrow H_2O + I$	[b]$1.6 \times 10^{-11} \times \exp(440/Temp)$
9. $HOI + OH \Longrightarrow H_2O + IO$	[h]2.0×10^{-13}
10. $I + HO_2 \Longrightarrow HI + O_2$	[i]$1.47 \times 10^{-11} \times \exp(-1090/Temp)$
11. $IO + HO_2 \Longrightarrow HOI + O_2$	[b]$1.4 \times 10^{-11} \times \exp(540/Temp)$
12. $OH + OH \Longrightarrow H_2O_2$	[c]$2 \times 6.9 \times 10^{-31} \times (Temp/300)^{-0.8} \times p/(1.38 \times 10^{-23})/Temp/10^6$
13. $OH + OH \Longrightarrow H_2O$	[b]$6.2 \times 10^{-14} \times (Temp/298)^{2.6} \times \exp(945/Temp)$
14. $OH + HO_2 \Longrightarrow H_2O + O_2$	[b]$4.8 \times 10^{-11} \times \exp(250/Temp)$
15. $HO_2 + HO_2 \Longrightarrow H_2O_2$	[b]$2.2 \times 10^{-13} \times KMT06 \times \exp(600/Temp)+$ $1.9 \times 10^{-33} \times M \times KMT06 \times \exp(980/Temp)$

[a]Wine et al. (1984); [b]Atkinson et al. (2004); [c]Zellner et al. (1988); [d]Bloss et al. (2001); [e]Gilles et al. (1999); [f]Saiz-Lopez et al. (2014); [g]Bösch (2003); [h]Chameides and Davis (1980); [i]Jenkin et al. (1990).

$KMT06 = 1 + (1.4 \times 10^{-21} \times \exp(2200/Temp) \times [H_2O])$, $[H_2O]$ is the absolute water concentration. M is the total number of molecules in the atmosphere. p is the pressure.

$w1a = 4.7 \times 10^{-10} - 1.4 \times 10^{-5} \times \exp(-0.75 \times p/1.62265) + 5.51868 \times 10^{-10} \times \exp(-0.75 \times p/199.328)$;

$w2a = -0.00331 - 0.00514 \times \exp(-0.75 \times p/325.68711) - 0.00444 \times \exp(-0.75 \times p/40.81609)$;

$w1b = 1.166 \times 10^{-9} - 7.796 \times 10^{-10} \times \exp(-0.75 \times p/22.093) + 1.038 \times 10^{-9} \times \exp(-0.75 \times p/568.154)$;

$w2b = -0.00813 - 0.00382 \times \exp(-0.75 \times p/45.57591) - 0.00643 \times \exp(-0.75 \times p/417.95061)$.

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Competing interests. Paxton Juuti and Jyri Mikkilä work for Karsa, Ltd. Finland. Juha Kangasluoma works partially for Karsa, Ltd. Finland

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