1	Stability assessment of organic sulfur and organosulfate compounds in filter
2	samples for quantification by Fourier Transform-Infrared Spectroscopy
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Abstract

19 Organic sulfur and sulfate compounds, tracers for sources and atmospheric processes, are not currently measured in national monitoring networks such as the Interagency Monitoring of 20 21 Protected Visual Environments (IMPROVE). The goal of this paper is to begin to assess the 22 stability of organic sulfur and sulfate containing compounds on polytetrafluoroethylene (PTFE) filters and the suitability of Fourier-transform infrared (FT-IR) spectroscopy to measure these 23 compounds. Stability assessment is needed because PTFE samples collected by IMPROVE are 24 25 typically stored 6-9 months prior to analysis. For this study, two organosulfur compounds, methanesulfonic acid (MSA) and hydroxymethanesulfonate ion (HMS), and two organosulfate 26 27 compounds, methyl sulfate (MS) and 2-methyltetrol sulfate (2-MTS), are collected individually on PTFE filters. Gravimetric mass measurements are used to assess mass stability over time. FT-28 29 IR spectra are evaluated to assess the capability of measuring the compound from PTFE filters by assessing the compound stability or chemical changes over time. Ion chromatography (IC) and 30 31 Inductively Coupled Plasma Optical Emission Spectroscopy (ICP-OES) are used as an additional tool to assess stability or chemical changes over time. MS has the highest potential to be measured 32 33 by FT-IR in IMPROVE samples. For MS, a simple organosulfate, the mass changes are within 34 measurement uncertainty and FT-IR spectra indicate no compositional change over a 4-month period, suggesting MS can be measured using FT-IR. IC and ICP-OES support the conclusion 35 that MS is stable on the filter. However, for 2-MTS, the other organosulfate measured in this study, 36 37 spectral changes after a month on the filter suggests it decomposes into other organosulfates or an 38 inorganic sulfate. MSA in IMPROVE samples can be measured, but only as a lower bound, due to volatility off of the filter as indicated by FT-IR and gravimetry. FT-IR and IC both show that 39 MSA is not chemically changing over the course of the study. Measurements by all methods 40

	41	indicate HMS is unstable on PTFE filter and IC and FT-IR indicate that it likely converts to
	42	inorganic sulfate. Future work includes the evaluation of these compounds in as ambient aerosol
1	43	sample matrix to determine any differences in stability ₂ -and identifying interferences that could
	44	limit quantification and developing calibrations to measure the compounds or functional groups in
	45	ambient samples-

1. Introduction

48	Organic sulfur compounds exist in particulate form in the atmosphere and can be the result of
49	natural processes (e.g. marine sulfur, volcanic emissions) (Aneja and Cooper, 1989; Bates et al.,
50	1992) or activities of anthropogenic sources (e.g. combustion, sulfur-rich wastewaters, smelting)
51	(Grübler, 1998; Smith et al., 2011). Organic sulfur compounds can be categorized as
52	organosulfur compounds such as sulfones (RSO ₂) and sulfonic acids (RSO ₃ ⁻) having C-S bonds,
53	while organosulfates (ROSO3 ⁻) have a C-O-S bond in the structure (Song et al., 2019). Two
54	sulfonic acid compounds, methanesulfonic acid (MSA), a tracer for marine aerosol, and
55	hydroxymethanesulfonate ion (HMS), measured in high haze conditions, along with two
56	organosulfates compounds, methyl sulfate (MS) and 2-methyltetrol sulfate (2-MTS) were
57	selected for evaluation.
58	Methanesulfonic acid forms from photochemical oxidation of dimethyl sulfide (DMS)
59	(von Glasow and Crutzen, 2004; Kwong et al., 2018a). DMS is a naturally occurring sulfur
60	species produced by marine algae or phytoplankton and is an important precursor of sulfur
61	dioxide, non-sea salt inorganic sulfate and organosulfur compounds, including MSA (Barnes et
62	al., 1994; Hoffmann et al., 2016). This makes MSA a tracer for marine aerosol (Allen et al.,
63	1997; Becagli et al., 2013; Saltzman et al., 1986). Ion chromatography (IC) has been used to
64	measure MSA in ambient aerosol collected on PTFE filters (Amore et al., 2022) and nucleopore
65	filters (Allen et al., 2002). MSA has also been measured in water soluble fractions of ambient
66	aerosol using proton nuclear magnetic resonance (HNMR) (Decesari et al., 2000). Fourier-
67	transform infrared spectroscopy (FT-IR) has been used to characterize liquid and solid MSA in
68	the laboratory studies (Lee et al., 2019; Zhong and Parker, 2022; Chackalackal and Stafford,
69	1966) as has Raman spectroscopy (Zhong and Parker, 2022), but to the best of our knowledge,

neither FT-IR or Raman spectroscopy have been used to measure MSA in complex mixtures likeambient aerosol samples.

Hydroxymethanesulfonate (HMS), formed by sulfite and formaldehyde in aqueous phase, 72 73 is a strong acid that is stable at low pH (Seinfeld and Pandis, 2016) and is a tracer for aqueous processes (Chen et al., 2022). During severe winter haze in the North China Plain, HMS was 74 measured using real-time single particle mass spectrum instruments and filter-based IC methods 75 during periods of high SO2 and HCHO concentrations and low oxidant concentrations in 76 particles with high liquid water content (Ma et al., 2020). Very high concentrations of HMS 77 have been measured in Fairbanks, Alaska during pollution events in a cold, dark and humid 78 79 environment (Campbell et al., 2022) (Campbell et al., 2022). In the Interagency Monitoring of Protected Visual Environments (IMPROVE) Network, there is evidence of an ubiquitous 80 81 presence of HMS in ion chromatograms of samples collected at 150 sites in the United States (Moch et al., 2020). 82

However, HMS can be challenging to measure (Moch et al., 2018). Single particle mass 83 spectrometry techniques have identified m/z 111 as characteristic for HMS (Chapman et al., 84 1990; Lee et al., 2003; Song et al., 2019). However, methyl sulfate and other organic sulfur 85 compounds have the same characteristic m/z which makes quantifying HMS using mass 86 spectrometry challenging (Dovrou et al., 2019; Lee et al., 2003). High-resolution aerosol mass 87 spectrometry (HR-AMS) has been used to measure HMS and organosulfates, however the 88 majority of compounds mostly fragment into inorganic sulfate and a non-sulfur containing 89 90 organic fraction, leading to an underestimation of HMS and overestimation of inorganic sulfate (Dovrou et al., 2019; Song et al., 2019). HMS has been measured in field and laboratory studies 91 92 by IC (Dovrou et al., 2019; Campbell et al., 2022), however notable challenges have been

93	documented. HMS and sulfate are not fully resolved in all IC methods (Campbell et al., 2022)
94	(Campbell et al., 2022)-leading to poor resolution that can introduce error into the results for both
95	HMS and sulfate (Dovrou et al., 2019; Ma et al., 2020). In IC methods where HMS and sulfate
96	are well resolved, HMS and sulfite may be unresolved and co-elute with bisulfite (Moch et al.,
97	2018; Wei et al., 2020). Additionally, HMS may degrade to sulfite and formaldehyde at the high-
98	pH eluent used in IC (Moch et al., 2020). Degradation of both HMS and sulfite may occur in
99	aqueous solutions prior to analysis or in the column during analysis and it's suspected that some
100	of the sulfite oxidizes to sulfate in solution or in the column (Moch et al., 2020). The formation
101	of HMS in the atmosphere occurs at moderate pH and pH differences, between the filter and
102	atmospheric condition (e.g. cloud, fog, pH), can contribute to HMS sample mass loss off the
103	filter leading to an underestimation of HMS (Moch et al., 2020). With proper columns and eluent
104	composition, IC has been shown to separate HMS and sulfate peaks with only a small
105	underestimation of HMS due to sulfate conversion (Dovrou et al., 2019; Campbell et al., 2022).
106	At least one laboratory study (published in Japanese) has characterized HMS by FT-IR (Sato et
107	al., 1984) but FT-IR has not been used to measure HMS in ambient aerosol samples to the best of
108	our knowledge.
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108 109 110 111	our knowledge. Organosulfates are the most abundant form of organic sulfur compounds in atmospheric particles (Hettiyadura et al., 2015; Stone et al., 2012; Hawkins et al., 2010; Frossard et al., 2011; Olson et al., 2011). Organosulfates are secondary organic aerosols (SOA) from oxidation of
108 109 110 111 112	our knowledge. Organosulfates are the most abundant form of organic sulfur compounds in atmospheric particles (Hettiyadura et al., 2015; Stone et al., 2012; Hawkins et al., 2010; Frossard et al., 2011; Olson et al., 2011). Organosulfates are secondary organic aerosols (SOA) from oxidation of mostly biogenic, but also some anthropogenic volatile organic compounds, in the presence of

suggested to be tracers for SOA (Wang et al., 2021; Chen et al., 2021). Organosulfates have been

115 measured in ambient aerosol globally including at four sites in Asia where on average they

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116	contribute <1% of PM_{2.5}, 2.3% of organic carbon and 3.8% of total sulfate (Stone et al., 2012). In
117	Arctic haze aerosols in the spring, organosulfates contributed to 13% of organic matter (OM)
118	(Hansen et al., 2014) and contributed to OM at varying levels across the US, with higher levels
119	in summer (Chen et al., 2021). Most studies have used a liquid chromatography method coupled
120	to a mass spectrometer (LC-MS) for measuring organosulfates (Hettiyadura et al., 2015; Wang et
121	al., 2021). FT-IR has been used to measure total organosulfate functional groups (Hawkins et al.,
122	2010) and Raman (Lloyd and Dodgson, 1961; Bondy et al., 2018) and FT-IR (Lloyd et al., 1961)
123	have been utilized to characterize organosulfates in laboratory studies. FT-IR has been used to
124	measure the organosulfate functional group using peak fitting and showed that the organosulfate
125	functional group contributes up to 10% of organosulfate in the Aerctic region, when inorganic
126	sulfate concentrations are considered high (Frossard et al., 2011), and 4-8% of OM in the Pacific
127	marine boundary layer, during periods of high organic and sulfate concentrations (Hawkins et al.,
128	2010). While other studies showed little to no organosulfates, likely due to low sulfate
129	concentrations in Mexico City (Liu et al., 2009) and Bakersfield, CA_(Liu et al., 2012). (Liu et
130	al., 2012) (Liu et al., 2012).
131	Methyl sulfate is the smallest organosulfate (Kwong et al., 2018b) (Kwong et al., 2018b)
132	and measured mostly in trace amounts (Hettiyadura et al., 2017, 2015; Wang et al., 2021).
133	However, it is commercially available and therefore useful for laboratory studies. 2-Methyltetrol
134	sulfates are tracers for secondary organic aerosols (SOA) formation in atmospheric particles
135	derived from isoprene (Surratt et al., 2010; Chen et al., 2020) and one of the most abundant
136	organosulfates measured in ambient aerosol. In the eastern US, 2-MTS accounts for the highest
137	percentage summertime particulate organosulfate (11%) (Chen et al., 2021). In Centreville, AL,
138	2-MTS accounts for more than half of organosulfates during summer of 2013 (Hettiyadura et al.,

139	2017). In Shanghai, China (summer of 2015-2016, 2018-2019), 2-MTS was the most abundant
140	organosulfate (31%) of 29 organosulfates (Wang et al., 2021).
141	The IMPROVE network is a rural particulate matter monitoring network with ~ 165 sites
142	across the United States (http://vista.cira.colostate.edu/improve/). Polytetrafluoroethylene
143	(PTFE), nylon and quartz filters are used to collect PM _{2.5} every one in three days, have a field
144	latency period of up to 7 days and are analyzed by multiple analytical techniques. PTFE filters
145	are stored at room temperature and analyzed between 3 and 12 months after collection (typically
146	6 to 9 months) for PM mass, elements and filter-based light absorption. Recently, FT-IR
147	analysis, a non-destructive method, has been performed on IMPROVE samples to reproduce
148	routinely measured compositional data (Debus et al., 2022) and measure the functional group
149	composition of the organic fraction (Ruthenburg et al., 2014, Kamruzzaman et al., 2018).
150	Organic sulfur compounds and organosulfate functional groups were not measured in these
151	studies. FT-IR analysis has also been conducted on other networks and in chamber and field
152	studies to measure organic functional groups (Boris et al., 2021, 2019; Laurent and Allen, 2004;
153	
	Ruthenburg et al., 2014; Yazdani et al., 2022; Russell et al., 2011). The extracts of nylon filters
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154 155 156 157	Ruthenburg et al., 2014; Yazdani et al., 2022; Russell et al., 2011). The extracts of nylon filters have been analyzed by ion chromatography (IC) in the IMPROVE network for more than three decades to routinely measure the inorganic ions, sulfate, nitrate, chloride and nitrite. (http://vista.cira.colostate.edu/improve/wp-content/uploads/2020/02/1_Anion-Cation-Analysis- by-Ion-Chromatography-SOP-revision-7.pdf). Although not routinely measured in IMPROVE
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161 interfaced to electrospray ionization high-resolution quadrupole time-of-flight mass spectrometry 162 (HILIC-ESI-HR-QTOFMS). The goal of this paper is to assess the stability of four organic sulfur and sulfate 163 containing compounds on polytetrafluoroethylene (PTFE) filters and the suitability of Fourier-164 transform infrared (FT-IR) spectroscopy to measure these compounds The goal of this paper is 165 166 to assess the potential for using FT-IR analyses to measure organic sulfur compounds collected on PTFE filters in the IMPROVE network. Measuring organic sulfur compounds on a 167 continuous basis across the US would provide a rich data set to evaluate their sources, 168 169 concentration, seasonality and trends over time. Four organic sulfur compounds, two organosulfur compounds, methanesulfonic acid, hydroxymethanesulfonate, and two 170 organosulfates, methyl sulfate and 2-methyltetrol sulfate, are evaluated. For these compounds to 171 be measurable in IMPROVE by FT-IR, there must be minimal losses or other changes to the 172 compound during the latency period between collection and analysis (3 - 12 months), and there 173 174 must be minimal interferences in the spectra. To achieve this goal, each compound was dissolved in solution, aerosolized and collected on PTFE filters. Collected samples were 175 weighed and analyzed by FT-IR every few days for two months or more. Characterization of the 176 FT-IR spectra as well as changes (or lack therefore of) in the mass loading and spectra over time 177 indicate the potential for the compounds to be measured by FT-IR in IMPROVE samples. Filter 178 samples, extracted for IC analysis at different time points, indicate stability or chemical changes 179 in the compound on the filter, assists with interpreting gravimetric mass and FT-IR spectra 180

9

181

changes.

2. Materials and Methods

183	Two organic sulfur (C-S) compounds, MSA and HMS, and two organosulfates (C-O-S), MS and
184	2-MTS, were selected. The four compounds were selected for this study based on following
185	three criteria. The compound 1.) has been measured in atmospheric particulate matter and is of
186	interest to the atmospheric science community, 2.) is water soluble so it can be put into solution
187	for atomization, and 3.) is available in high purity form to minimize uncertainty in mass
188	measurement. Filter samples of the organic sulfur compounds were prepared for FT-IR,
189	gravimetry, IC and ICP-OES analyses by aerosolizing each compound individually and
190	collecting it on PTFE filters (Pall Corporation, 25 mm diameter). One set of filters was
191	generated for analysis by gravimetry and FT-IR at UC Davis and another set (or sets depending
192	on what was being evaluated) of filters were generated for analysis by IC and ICP-OES at
193	Research Triangle Institute (RTI) following gravimetric analysis at UC Davis. Analyses of
194	these laboratory filter samples were performed to characterize the compound within infrared
195	spectra and to determine the stability of these compounds over time.
196	2.1 Preparation of laboratory filter samples
197	Three commercially available standards were used for this study: HMS sodium salt
198	(>97% purity, TCI America), MSA (100% purity, Sigma Aldrich) and MS sodium salt (100 $\%$
199	purity, Sigma Aldrich). 2-methytetrol sulfate ammonium salt was synthesized following a
200	published method (Cui et al., 2018). Each compound was collected on PTFE filters by first
201	preparing an aqueous solution with a concentration of 0.005 M. For HMS, the solution was
202	acidified with hydrochloric acid (HCl) prior to aerosolization to obtain samples with
203	atmospherically relevant pH (pH 2), as pH plays a role in the stability of HMS. 2 μL of 1 M HCl
204	was added to HMS solution to obtain the final volume of 200 mL for aerosolization. Aerosols

205	were generated using an atomizer (Kamruzzaman et al., 2018; Ruthenburg et al., 2014) and dried
206	with a diffusion dryer (Model 3074B Filtered Air Supply, TSI Inc., St. Paul, MN) which
207	produces a high concentration of poly-disperse submicrometer sized particles allowing for short
208	collection times and adequately representing the expected response from particles of similar size
209	range in the atmosphere. Dry particles were collected on PTFE filters (Pall Corporation, 25 mm
210	diameter) using an IMPROVE sampler with varying collection times (40 to 720 s) at a flow rate
211	of 22.4 L/min.
212	2.1.1 Gravimetric mass determination
213	Filter mass, before and after particle collection, was measured using an ultra-
214	microbalance (XP2U, Mettler-Toledo, Columbus, OH) with 0.1 µg sensitivity. Ionizing
215	cartridges (Staticmaster® Model 2U500, Grand Island NY) housed on a flexible stand
216	(Staticmaster® Model BF2-1000, Grand Island NY) and Haug strip (Mettler Toledo 11140160,
217	Columbus, OH) were utilized to help eliminate static for more stable, accurate measurements.
218	Prior to particle collection, the mass of a filter was determined by the average of 5 mass
219	measurements taken on separate days. Only filters that weighed within measurements precision
220	for 25 mm filters (\pm 6 µg) for the 5 measurements were used. After particle collection, filters
221	were allowed to achieve equilibrium at room temperature for 24-hrs. Filters were weighed for
222	three consecutive days in the 1^{st} week, twice per week during $2^{nd} - 4^{th}$ weeks and once in the
223	weeks thereafter. Filters were stored at room temperature (21°C - 27 °C) and relative humidity
224	$(30\% \pm 10\%)$ to mimic the storage conditions for ambient IMPROVE Teflon filters. The
225	experiment was ended when the weights were stable for a month or more.

226 2.2 Infrared spectra collection and processing

227	FT-IR spectra of the filter samples of each compound were collected using a Tensor II FT-IR
228	spectrometer (Bruker Optics, Billerica, MA) with a liquid nitrogen cooled mercury cadmium
229	telluride (MCT) detector over the spectral range of 4 000–400 cm ^{-1} . Filters were placed in a
230	house-built sample chamber that is purged of water and CO2 (PureGas) for 4 min before spectra
231	acquisition_(Debus et al., 2019)(Debus et al., 2019) (Debus et al., 2019). Transmission mode
232	<u>m</u> Measurements were made using 512 scans for each filter at 4 cm ^{-1} resolution and ratioed to the
233	most recent (less than 1 h) background spectrum to obtain absorbance spectra using OPUS
234	software (Bruker Optics, Billerica, MA)(Debus et al., 2019) (Debus et al., 2019) .
235	To better visualize functional groups in the organosulfur compounds and minimize the
236	impact of the PTFE scattering and absorption on the spectra, several steps were taken
237	were baseline corrected from 1500 cm ⁻¹ to 500 cm ⁻¹ , using blank correction and smoothing
238	spline fitting (Kuzmiakova et al., 2016). (Kuzmiakova et al., 2016) The spectral region from
239	4000 cm^{-1} to 1500 cm^{-1} were baselined using an automated version of the Kuzmiakova et al.,
240	2016 smoothing spline process in AirSpec (Reggente et al., 2019). Regions with large PTFE
241	absorption (1300-1100 cm ⁻¹ , 700-600 cm ⁻¹ and 500-420 cm ⁻¹) were grayed out in spectra plots
242	and are not considered for peak identification. A baseline corrected spectrum of each compound
243	is shown in supplemental material Figure S1.
244	2.3 IC and ICP-OES sample analysis
245	PTFE filters of each organic sulfur and organosulfate sulfur compound were generated and
246	weighed at UC Davis prior to shipping the filters cold overnight to RTI for IC and ICP-OES
247	analysis. On the day the filters arrived at RTI, the filters were extracted in 50 ml of deionized

248 water ($18M\Omega$ cm⁻¹, Millipore Milli-Q Darmstadt, Germany), sonicated for 30 minutes in an ice

249 bath and placed on a shaker table in a cold room for 8 hours prior to analysis. The PTFE filter 250 remained in the extraction vial for the duration of these experiments. IC analysis was performed on Dionex Thermo Scientific ICS-3000 and ICS-6000 (Sunnyvale, 251 CA) instruments using suppression and conductivity detection. For MSA, extracts were analyzed 252 using the AS19 analytical and AG19 guard columns (anion hydroxide method) for initial 253 extraction efficiency tests and AS28 analytical and AG28 guard columns (hydroxide method) for 254 the subsequent analyses to evaluate changes over time. HMS and MS extracts were analyzed 255 with AS12A analytical and AG12A guard columns (anion carbonate method), which has been 256 shown to provide sufficient separation of HMS and sulfate, but not separation of sulfite/bisulfite 257 and HMS (Dovrou et al., 2019). MS extracts were analyzed with AS12A analytical and AG12A 258 guard columns (anion carbonate method), the same method as HMS. An IC method for analyzing 259 260 2-MTS had not been developed and evaluated prior to this work. Eluent concentrations and flow rates were optimized for best separation of all ions of interest. 261 ICP-OES, used to measure total sulfur on the filter, was performed on a Thermo Scientific iCAP 262 263 7600 duo analyzer (Bremen, Germany). The ICP-OES was run in axial mode using a sprint valve and data were collected at 180.731 nm. The ICP-OES system was calibrated using the sulfate 264 265 calibration standards and validated using the sulfate calibration verification solutions described

266 below.

IC and ICP-OES systems were calibrated with calibration standards prepared via serial dilutions of single source stock standards using a primary source. A secondary source was used to prepare calibration verification solutions to validate the instrument calibration for all compounds except for 2-MTS, for which a second source standard was unavailable.

271	Primary and secondary sources of National Institute of Standards and Technology
272	(NIST)-traceable solutions were purchased and used to prepare calibration standards and
273	calibration verification solutions respectively, for sulfate analyses by both IC and ICP-OES.
274	When NIST-traceable solutions were unavailable, salts were used to prepare calibration
275	standards and calibration verification solutions for MSA, HMS, and MS. Vendor information for
276	primary and secondary sources are provided in Table S2 in supplemental material. Certified
277	American Chemical Society (ACS)-grade sodium carbonate (Na2CO3) obtained from Fisher
278	Scientific (Fairlawn, NJ) and sodium bicarbonate (NaHCO3) obtained from EMD Sciences
279	(Gibbstown, NJ) were used to prepare IC eluent when using anion carbonate methods for
280	analyses. Potassium hydroxide eluent generator cartridges purchased from Thermo Scientific
281	were used for eluent preparation for analyses conducted with anion hydroxide methods. NIST-
282	traceable, 1000 μ g/mL stock solutions of yttrium (Y) and cesium (Cs) obtained from High Purity
283	Standards (Charleston, SC) were used to for internal standard and ionization suppression,
284	respectively for ICP-OES measurements.
	2 Descrite and Discussion
285 286	5. Results and Discussion Methanesulfonic acid
287	Gravimetry
288	Mass changes, measured by gravimetry, for four methanesulfonic acid filter samples with masses
289	ranging from 64 μ g to 80 μ g are shown in Figure 1. Mass decreases steadily during the first
290	month to approximately 50% of the initial mass. During the second month of measurements, the

291 mass remains constant $(50 \pm 6\%)$.







297 FT-IR

Methanesulfonic acid (CH₃SO₃H) is composed of a methyl group attached to a sulfonic acid 298 [S(=O)2-OH], via a C-S bond. Methanesulfonic acid aerosols collected on PTFE filters have 299 300 peaks associated with CH₃, SO₃, S-OH and C-S bonds (Figure 2). Observed peaks (Figure 2) 301 can be ascribed to portions of the molecule based on previous FT-IR and Raman work (Lee et al., 2019; Zhong and Parker, 2022; Chackalackal and Stafford, 1966). The peaks at 1342 cm⁻¹ 302 and 536 cm⁻¹ arise from S=O bonds in MSA and are shifted compared to inorganic peaks at 1130 303 304 cm⁻¹, 620 cm⁻¹ (Larkin, 2018) or organic sulfate SO₄ peaks at ~1380 cm⁻¹ (Larkin, 2018; Lin-Vien et al., 1991). The peak at 895 cm⁻¹ is attributable to S-OH (Zhong and Parker, 2022) and 305 the peak at 766 cm⁻¹ is attributable to C-S (Lee et al., 2019). C-H peaks are observed at 3039 cm⁻ 306 ¹, 2951 cm⁻¹, 1414 cm⁻¹ and 987 cm⁻¹ (Chackalackal and Stafford, 1966). The broad peak at 3248 307

- 308 cm⁻¹ is suggested to be water as (Zeng et al., 2014) showed that this peak in MSA infrared
- 309 spectra increases with increasing RH. These peaks, particularly strong peaks, were similar to
- 310 spectral absorbance of MSA from reference spectra (Spectral Database for Organic
- 311 Compounds, SDBS, 2022), and Table S1 in Supplemental Materials compares the observed and
- 312 reference peaks.



Figure 2. Changes in the spectra of MSA over a 2- month period, denoted by number of elapsed

316 days. The shaded area indicates the absorbance regions of PTFE filter.

317 The MSA infrared peaks of SO₃ (1342 cm⁻¹), C–H (987 cm⁻¹), S–OH (895 cm⁻¹), C-S

 (766 cm^{-1}) , and S=O (536 cm⁻¹) decrease rapidly in the first 30 days, consistent with the decline

in mass during that time. The spectra suggest that MSA is volatilizing off the filter, even though this is inconsistent with the low vapor pressure of MSA (Knovel - Yaws' Critical Property Data for Chemical Engineers and Chemists - Table 12. Vapor Pressure - Organic Compounds, log P =A - B/(T + C), 2022), 0.00022 mmHg at 20°C). The three spectra obtained on days 30, 41 and 51 show only small decreases which is mostly consistent with the lack of mass changes during those days.

325 Not all peaks show consistent loss during the first month and little change during the second month. The C-H peaks at 3039 cm⁻¹ and 2942 cm⁻¹ behave slightly differently, reaching 326 stability (with some minor random variability) earlier (day 20) than most other peaks. The weak 327 peak at 1414 cm⁻¹ (ascribed to CH₃) increased and broadened to 1445 cm⁻¹ - 1400 cm⁻¹ after the 328 initial spectra was collected and remained fairly stable for the duration of the experiment. The 329 peak at 3250 cm⁻¹, increased rapidly followed by fluctuations, and then varies somewhat but 330 remained fairly consistent for the duration of the experiment, not unlike the behavior of the 1414 331 332 cm⁻¹ peak. One possible cause for these spectra changes in the spectra is water vapor condensing on the particles after collection. MSA is hygroscopic and although it effloresces at about RH = 333 50%, (Peng and Chan, 2001; Tang, 2020; Zeng et al., 2014), Zeng et al., 2014 shows that there is 334 some water associated with the particles below 50% RH, which is consistent with the FT-IR 335 spectra (Figure 3), particularly the 3248 cm⁻¹ peak. If water is indeed the cause, the change in 336 the 3248 cm⁻¹ peak can be explained by a very low RH in the particle generation and collection 337 system (lower initial peak) and a higher RH (RH = $30 \pm 10\%$) in the lab (increase in peak). The 338 change in the 1414 cm⁻¹ peak above 1400 cm⁻¹ behaves similarly and can be associated with OH 339 suggesting uptake of water after the initial FT-IR spectra was collected. 340

341	Another possibility is that ammonia is absorbing onto the MSA. Ammonium absorbs in
342	both the 3250 cm ⁻¹ and 1500-1400 cm ⁻¹ regions (Boer et al., 2007; Zawadowicz et al., 2015) and
343	when comparing MSA spectra to ammonium sulfate spectra, they show very similar peak
344	absorbance and shapes in these two regions suggesting. This suggests that the cause of these
345	peaks is ammonium (Supplemental Material, Figure S2).
346	A third possible cause for these changes is that MSA may be fragmenting into
347	formaldehyde (CH2O), that partitions into the gas phase, and sulfite (SO3) (Kwong et al., 2018a).
348	However, this would show a decrease in CHC-H peaks and a shift in the SOS=O peaks, neither
349	of which are observed in the spectra. The rapid decrease in peak height during the first month
350	and then little decrease or no trend during the second month, suggests that MSA is volatilizing
351	off the filter initially, but then has a slow decline, offset by increases in water or ammonium.
352	IC and ICP-OES
353	Twenty PTFE filters with MSA (14 μ g – 32 μ g; 60 μ g – 144 μ g) were shipped to RTI for
354	extraction and analysis by IC and ICP-OES. Each extract was analyzed by both IC and ICP-OES.
355	Recoveries of MSA from PTFE filters was $55 \pm 5\%$ for IC and $51 \pm 5\%$ for ICP-OES.
356	Calibration verification solution recoveries were 96 \pm 5% for IC and 101 \pm 3% for ICP-OES
357	analyses, suggesting that the lower recoveries from the PTFE filter are due to incomplete
358	extraction and/or losses occurring during shipment. To evaluate losses during shipping, six
359	samples were collected, weighed, and analyzed by FT-IR, shipped to RTI and then sent back to
360	UC Davis and analyzed with gravimetric and FT-IR analysis 9 days after initial measurements.
361	The mass loss (19 \pm 7%) during this period was similar to mass loss for filters that remained in
362	the UC Davis laboratory for ~8 days ($22 \pm 6\%$). Spectral changes in shipped filters were similar
363	to changes in spectra that occurred during the first week for filters that remained at UC Davis.

364	To evaluate changes in mass and composition on PTFE filters over time, 6 filters ($42 \mu g - 63 \mu g$)
365	were shipped to RTI and extracted on days 0, 30, and 61. These filters had consistent recoveries
366	over time of 57 \pm 6% for IC and 55 \pm 6% for ICP-OES. The IC results indicate that MSA did
367	not change chemically during this time period, supporting the FT-IR spectral results. The mass
368	and spectral data indicate that the lower extraction efficiency was not due to loss of MSA off the
369	filter during shipping and suggest that the limitation is extraction efficiency. Despite low
370	extraction efficiencies and difference in behavior over time from FT-IR measurement, the FT-IR,
371	gravimetry and IC results suggest that a lower-bound of MSA can be measured on PTFE filters
372	in IMPROVE by FT-IR.
373	Hydroxymethanesulfonate
374	Gravimetry
375	Mass changes in three hydroxymethanesulfonate sodium salt (HOCH ₂ SO ₃ Na) filter samples with
376	mass loadings of \sim 62 µg per filter are shown (Figure 3). Mass decreases steadily for 1.3 months
377	to a maximum loss of 38% and then remains constant (38 \pm 3%) for the rest of the experiment
378	(4.1 months). Similar results were obtained for filters weighing approximately 30 μ g/filter of
379	HMS.







385 FT-IR

386 HMS (HOCH₂SO₃H) is a sulfonic acid compound with C-S bond, where the S bond is part of a

sulfonic acid group [S(=O)₂—OH], and the carbon is attached to an –OH functional group,

- similar to MSA except with the OH functional group attached to the carbon. The chemical used
- in our study has a sodium cation on the sulfonic acid group. HMS aerosol collected on PTFE
- filters have three infrared peaks (1094 cm⁻¹, 1041 cm⁻and¹, 611 cm⁻¹) between 1500 and 500 cm⁻¹
- 391 , although the peak at 611 cm⁻¹ is obscured by PTFE absorption, and O-H and C-H peaks







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Observed peaks at 1094 cm⁻¹ and 1041 cm⁻¹ are similar to the 1080 cm⁻¹ and 1040 cm⁻¹ peaks identified as S-O or S=O bonds in FT-IR spectra of Na HMS (Sato et al., 1984) and of HMS (Larkin, 2018; Shurvell, 2006). A weak band at 611 cm⁻¹ is like due to C-S (Lin-Vien et al., 1991; Sato et al., 1984) or S-O (Sato et al., 1984) but is obscured by PTFE absorbance. Above 1500 cm⁻¹ and similar to many aliphatic organic molecules, an O–H (broad peak centered near

403	3300 cm ⁻¹) and C-H (below 3000 cm ⁻¹) peaks were observed (Pavia et al., 2008; Shurvell, 2006).
404	Observed peaks are similar to spectral absorbance, although not all peaks in the reference spectra
405	(AIST: SDBS, 2022, See Table S1 in Supplemental Material) are observed in the measured
406	spectra. HMS has a very low vapor pressure (0.00000073 mmHg), (U.S. EPA. Comptox
407	Chemicals Dashboard, 2022) indicating that HMS should not volatilize off the filter.
408	All peaks below 1500 cm ⁻¹ decrease and are no longer visible by day 34, consistent with
409	the decline but not the extent of decline in mass as the HMS peaks are completely gone and the
410	mass has only decreased by 40%. This behavior is most clearly observed in the peak at 1041 cm ⁻
411	¹ , but also observed in 1094 cm ⁻¹ peak (Figure 4). Similar to the S-O/S=O peaks, C-H peaks
412	decline during the first 34 days and then completely disappear. The O-H peak, centered around
413	3300 cm ⁻¹ , disappears more slowly, but like the S-O and C-H peaks, is gone by the end of the
414	experiment. Counter balancing the loss of mass, a new peak becomes visible at 1003 cm ⁻¹ after
415	11 days and increases for the rest of the study. The peak at 1003 cm ⁻¹ is tentatively identified as
416	bisulfate (Boer et al., 2007; Krost and McClenny, 1994). The small peak centered around 3450
417	cm ⁻¹ becomes evident as the O-H peak disappears and may indicate the presence of condensed
418	water (Boer et al., 2007).
419	IC and ICP-OES
420 421	Sixteen HMS PTFE filters were analyzed by IC with recoveries of $65 \pm 4\%$ and calibration
422	verification solution recoveries of $94 \pm 5\%$. ICP-OES analysis was not performed on these
423	filters. Eight additional HMS filters (65 μ g – 100 μ g) were shipped to RTI and 2 filters were
424	extracted and analyzed by IC and ICP-OES on each of the following days 0, 30, 61, and 101.
425	The sulfur mass losses in IC (Figure S3, Supplemental Material) over the 101 days is ~60% loss
426	from the initial weighed mass or about 39% loss assuming a constant extraction efficiency of

427	65% for all samples and is in agreement with the 38% decrease in mass on the filter measured by
428	gravimetry. Similar results were obtained for ICP-OES. IC analysis (Figure 5) confirms the
429	samples are mostly HMS on day zero but over time the HMS is converted to sulfate (sulfate and
430	bisulfate are indistinguishable in IC), supporting assignment of the 1003 cm ⁻¹ infrared peak to
431	inorganic bisulfate. A small amount of the HMS may be converting to sulfate in solution or the
432	column, but the measured changes are much larger than what is expected due to that mechanism
433	alone. The small amount of HMS that is measured by IC on day 60 and 101 are near detection
434	limits for IC which corroborate the absence of HMS in the FT-IR spectra after two months. The
435	IC and FT-IR results both show a conversion of HMS into sulfate indicating that HMS is not
436	stable and cannot be quantified reliably on PTFE filters in IMPROVE by either FT-IR or IC. In
437	Moch et al. (2020), HMS did not degrade over time under cold storage conditions on nylon
438	filters from IMPROVE suggesting that storage or perhaps filter type may play an important role
439	in HMS degradation on filters.



440

Figure 5. Percentage of HMS and sulfate measured by IC. Eight PTFE filters were extracted andanalyzed in pairs over 101 days.

443 Methyl sulfate

444 Gravimetry

Mass measurements of methyl sulfate salt over a 4 -month period were within measurement 445 uncertainty for the three filters loaded with approximately 30 µg of methyl sulfate (the change 446 447 over time was indistinguishable from zero). For the 60 µg filters (Figure 6), the first 2 months 448 and the last two weeks mass change were within measurement uncertainty (+ 6 µg). However, for about a month, between day 70 and 100, much of the data (except day 91) is outside of 449 measurement uncertainty, indicating mass loss of between 10 and 20%. During this period only 450 451 one spectrum (day 79) was collected, and it does not support mass loss. Day 100 spectra, and all 452 spectra collected through the end of the study, for all three 60 ug loadings, filter samples

453 confirms no changes in the MS compound. Mass and spectral data indicate stability of mostly

15% 10% Change in mass of MS (ug) on PTFE filters 0 5% 0% -5% 0 -10% 6 0 0 0 6 -15% 60 ug 🖲 60 ug 0 -20% h ● 60 ug -25% 15 20 25 30 35 40 45 50 95 100 105 110 115 120 125 10 55 60 65 70 75 80 85 90 0 5 Time, days

454 stable MS on filters under ambient laboratory condition (24°C).





459 FT-IR

460	Methyl sulfate (CH ₃ SO ₄ H) is composed of a methyl (CH ₃) group attached to a sulfate (SO ₄)
461	group with C-O-S bond. The chemical used in this study is sodium methyl sulfate (CH ₃ SO ₄ Na).
462	Methyl sulfate aerosol collected on PTFE filters has two peaks between 4000 cm ⁻¹ and 2000 cm ⁻²
463	$^{\rm l}$ and eight peaks between 1500 and 500 cm $^{\rm -1}$ (Figure 7), similar to reference spectra in Table S1
464	(Spectral Database for Organic Compounds, SDBS, 2022). The doublets observed at 1020 cm ⁻¹ ,
465	1000 cm ⁻¹ and the nearly overlapping peaks at 795 cm ⁻¹ and 784 cm ⁻¹ are identified as S-O-C by
466	both FT-IR (Chihara, 1958; Lloyd et al., 1961; Lloyd and Dodgson, 1961; Segneanu et al., 2012;
467	Shurvell, 2006) and Raman(Okabayashi et al., 1974)(Okabayashi et al., 1974)(Okabayashi et
468	al., 1974). The S–O peaks from sulfate are observed around 1073 cm ⁻¹ and 591 cm ⁻¹ in the

469	spectra, similar to the previous study where peaks from 591 cm ⁻¹ to 593 cm ⁻¹ , and at 1063 cm ⁻¹
470	(solid) and 1081 cm ⁻¹ (solution) were ascribed to sulfate in potassium methyl sulfate (Chihara,
471	1958). There is a weak S=O peak at 1458 cm ⁻¹ (Segneanu et al., 2012). The O-H group
472	(assuming the Na ion was replaced by H for some of the molecules) at 3500 cm ⁻¹ is also fairly
473	weak. All collected spectra for one sample are shown, with the exception of day 91 data, which
474	appeared to be anomalous (Supplemental Material, Figure S4 all spectra including day 91).
475	The stability of the spectra over 4 months suggests that MS is stable when collected on a PTFE
476	filter. There were no major or consistent changes in peaks associated with S–O–C, S–O, Na-O
477	and C—H.
478	The pattern of change in the peak height of the 3500 cm ⁻¹ peak does not correlate with the
479	change in mass. Day 1 and day 30 spectra have smaller O-H peak intensities prior to mass
480	decline, compared to and the O-H peak in the spectrum from day 79, when mass had decreased
481	the mass is low, is higher than both. No consistent difference in the FT-IR spectra was observed
482	on day 79 (low gravimetric mass day) or the following days (days 105 - 120) suggesting that the
483	mass loss seen in the gravimetric data is erroneous. The high melting and boiling points of MS
484	are 96 °C and 298 °C, respectively and the low vapor pressure of 0.0038 mmHg (U.S. EPA.
485	Comptox Chemicals Dashboard, 2022) indicating that MS is not volatilizing off the filter.
486	•
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492 IC and ICP-OES

Sixteen PTFE samples, eight with mass loadings between $38 \ \mu g - 57 \ \mu g$ and eight with higher mass loadings, between $98 \ \mu g - 118 \ \mu g$, were prepared for IC and ICP-OES analysis. Eight PTFE filters were extracted on day 0 and recoveries were $60 \pm 1\%$ for IC and $87 \pm 3\%$ for ICP-OES analyses. Recoveries of calibration verification solutions were $100 \pm 1\%$ for IC and $97 \pm$ 6% for ICP-OES. Additional filters were extracted on day 39 and 61 showed recoveries consistent with recoveries measured during the initial extracts, $59 \pm 2\%$ for IC and $86 \pm 4\%$ for

499	ICP-OES. The overall lower mass recovery is indicative that not all MS is extracted from the
500	filter and the lower recoveries by IC compared to ICP-OES, suggests that MS is converting to
501	another sulfur compound in solution. The consistency in recoveries over time indicates stability
502	of MS on PTFE over this time period, which is in agreement with the mass stability of as shown
503	by gravimetric over the course of the experiment. Unfortunately, filters were not extracted during
504	the time period when the gravimetric results show a small deviation from stability.
505	MS is stable on a PTFE filter as indicated by gravimetry, FT-IR and IC suggesting that at least
506	some atmospherically relevant organosulfates can be measured on PTFE filters in IMPROVE by
507	FT-IR.
508	2-Methyltetrol sulfate
509	Gravimetry
510	Mass changes in three 2-methyltetrol sulfate ($C_5H_{11}SO_7$) filter samples with concentrations from
511	55 μg to 57 μg are shown in Figure 8. Mass decreases during the first 23 days to 73% of the
512	initial mass ($25\% \pm 2\%$). No additional mass loss was observed after 25 days. Mass behavior
513	indicates a loss of 2-MTS on filters under ambient laboratory condition (24°C). Similar results

I



Figure 8. Change in mass of 2-MTS collected on PTFE filters over a 3-month period under
laboratory conditions. Dotted vertical lines indicate FT-IR analysis. Horizontal broken red line
indicates mass balance precision.

519 FT-IR

2-Methyltetrol sulfate ($C_5H_{11}SO_7$) is a branched compound with 3 units of –OH, one methyl group and a sulfate group. Like methyl sulfate, 2-methyltetrol sulfate is an organosulfate and has a C-O-S bond. 2-MTS collected on PTFE filters has broad, organic-related peaks between 4000 cm⁻¹ and 2500 cm⁻¹ and four peaks between 1500 and 500 cm⁻¹ (Figure 9). Observed peaks can be ascribed to functional groups with the molecule based on previous FT-IR and Raman work (Lloyd et al., 1961; Lloyd and Dodgson, 1961; Bondy et al., 2018; Fankhauser et al., 2022).



529 Figure 9. Changes in the spectra of 2-MTS over a 3-month period. The shaded area indicates the

⁵³⁰ absorbance regions of PTFE filter.

531	The observed peak at 1041 cm ⁻¹ is ascribed to S-O stretch (Bondy et al., 2018; Fankhauser et al.,
532	2022), as it is similar to the peak of MS at 1050 cm ⁻¹ (Lloyd et al., 1961; Lloyd and Dodgson,
533	1961). The doublet at 908 cm ⁻¹ and 898 cm ⁻¹ correspond to symmetric and asymmetric stretch of
534	S-O of 2-MTS_(Fankhauser et al., 2022) (Fankhauser et al., 2022). The weak peak at 827 cm ⁻¹ is
535	tentatively assigned to C-O-S stretch based on Raman spectra of 3-MTS_(Bondy et al.,
536	2018)(Bondy et al., 2018) (Bondy et al., 2018). The peak at 1446 cm ⁻¹ is tentatively assigned to
537	asymmetric S=O stretch based on density functional theory of FT-IR spectra of 2-MTS
538	(Fankhauser et al., 2022) (Fankhauser et al., 2022) and the assignment of S=O to peak at 1448
539	cm ⁻¹ in methyl sulfate (Bondy et al., 2018). However, this peak was suggested to be due to C-H
540	using density functional theory of Raman spectra of 2-methyltetrol sulfates (Bondy et al., 2018)
541	(Bondy et al., 2018) and commonly ascribed to CH_2 in organic molecules at ~1465 cm ⁻¹ (Pavia et
542	al., 2008). We decided to assign this to S=O because the sulfate related peaks are strong and the
543	C-H peaks are very weak in this molecule. In the higher frequency region, the very subtle peak at
544	2879 cm ⁻¹ is ascribed to C-H stretch and the large broad peaks at 3065 and cm ⁻¹ , 3210 cm ⁻¹ and
545	shoulder at 3426 cm ⁻¹ may be attributable to -OH stretch (Larkin, 2018; Shurvell, 2006; Bondy
546	et al., 2018; Fankhauser et al., 2022) or inorganic –NH stretch (Boer et al., 2007; Larkin, 2018).
547	Our tentative peaks assignments of ammonium at 3210 cm ⁻¹ and 3065 cm ⁻¹ and OH at 3426 cm ⁻¹
548	are based on changes in spectra discussed below.
549	Most of 2-MTS infrared peaks decreased and disappeared over time, consistent with the

decline in mass. The S-O peaks at 1041 cm⁻¹, 908 cm⁻¹ and 898 cm⁻¹ decreased and were gone by day 40. The weak C-O-S peak at 827 cm⁻¹ disappeared by day 26. The C-H region at 3000-2800 cm⁻¹ decreased, similar to the changes in S-O-C peaks. However, some peaks shifted or new peaks formed. New peaks (or possibly shifts from 1041 cm⁻¹ and 908 cm⁻¹ doublet) at 1066

554	cm ⁻¹ and at 987 cm ⁻¹ appear on day 40 and then increase slightly over time. The S=O peak at
555	1446 cm ⁻¹ behaves differently from other 2-MTS peaks and increases slightly until day 33 when
556	the peak height increased significantly and shifted to 1420 cm ⁻¹ where it continues to grow
557	increase over time. Two peaks at 3210 cm ⁻¹ and 3065 cm ⁻¹ increased whereas the region around
558	3450-3550 cm ⁻¹ decreased, indicating that the shoulder at 3450 cm ⁻¹ arises from a different bond
559	than the peaks at 3210 cm^{-1} and 3065 cm^{-1} .
560	The changes in the spectra indicate a change in the chemical composition on the filter. 2-MTS is
561	no longer present by mid-way through the experiment as evidenced by the disappearance of S-O,
562	C-O-S, C-H and O-H peaks. This result is supported by a study of ambient 2-MTS that showed
563	that the atmospheric lifetime of 2-MTS is to be about 16 days_(Chen et al., 2020) (Chen et al.,
564	2020). Intermediate oxidation products of 2-MTS transformation include other organosulfates
565	such as 2-methylglyceric acid organosulfate (C4H7SO7, MGOS) and glycolic acid organosulfate
566	(C ₂ H ₃ SO ₆ , GAS) (Zhao et al., 2020; Wei et al., 2020) and the final product is likely inorganic
567	(NH ₄) ₂ SO ₄ (Harrill, 2020; Zhao et al., 2020). Raman spectra of MGOS-(Bondy et al., 2018)
568	(Bondy et al., 2018)_indicate that the spectra at the end of the experiment could be 2-MGOS.

569 The increased peak at 3210 cm⁻¹ and 3065 cm⁻¹ could be associated with -OH stretch and CH₃ asymmetric stretch of 2-MGOS and the 1420 cm⁻¹ (C-H), 1066 cm⁻¹ (S-O) and 987 cm⁻¹ (SO4²⁻) 570 571 match the final spectra well. The carbonyl group in 2-MGOS is very weak in the Raman spectra and indistinguishable from a blank filter in the final FT-IR spectra in this project. Alternatively, 572 the final spectra could be ammonium sulfate as indicated by peaks at 3210 cm⁻¹, 3065 cm⁻¹ and 573 1420 cm⁻¹ which are indicative of ammonium, however, the 1066 cm⁻¹ and 987 cm⁻¹ peaks are 574 lower than typically observed for inorganic ammonium sulfate (Boer et al., 2007; Zawadowicz et 575 al., 2015) suggesting this is not inorganic sulfate. There are limited FT-IR or Raman spectra of 576

the many oxidation products of 2-MTS so definitive identification is not possible and it is likely 577 578 that there is a mixture of oxidation products present on the filter. 579 To further evaluate the possible compounds on the filter at the end of the experiment, mass loss calculations were performed using the molecular weight of 2-MTS and each product 580 (MGOS, GAS and ammonium sulfate). If 2-MTS is converted completely to MGOS, the mass 581 582 loss would be only 7%. If 2-MTS is converted completely to GAS, the mass loss would be 28% 583 and if 2-MTS is converted completely to ammonium sulfate the mass loss would be 39%. These values span the observed mass loss at 25% which suggesting that the compounds on the filter are 584 585 of an intermediate type and still in the organosulfate form not inorganic ammonium sulfate. IC and ICP-OES 586 587 Sixteen PTFE samples, eight with mass loading between $11-28 \ \mu g$ and eight with higher mass loadings, between 37 - 46 µg were extracted immediately upon receipt at RTI. The IC showed 588 low sensitivity to 2-MTS detection and results are not reported so no additional information was 589 590 available about what compounds the 2-MTS may have change into. Given that ICP-OES 591 measures sulfur and not individual compounds, the results from this method do not provide insight into chemical conversions on the filter but are briefly discussed in the supplemental 592 593 material (Figure S5).

594

4. Conclusions

The stability and therefore potential for FT-IR to measure organosulfur and organosulfates collected on PTFE filters varies by compound. MS has the highest potential to be measured on PTFE filters in IMPROVE samples by FT-IR, due to its minimal mass change and no spectral changes. Consistent recoveries by IC and ICP-OES over multiple months of analysis support the conclusion that MS is stable on the filters. Consistent results from analysis at UC

600	Davis and RTI suggest robustness to storage, shipping and handling conditions. MS is one of
601	many organosulfates observed in the atmosphere and not necessarily representative of
602	organosulates in general as indicated by 2-MTS. Gravimetric mass suggests some (30%) mass
603	loss from 2-MTS samples on the PTFE filter, over a three-month period. FT-IR suggests that 2-
604	MTS is unstable on PTFE filter and changing into different compound(s), likely still an
605	organosulfate. FT-IR and gravimetry show that MSA can be measured from PTFE filters but
606	due to volatility off the filters a lower bound of MSA is measured (i.e., less than the amount of
607	MSA in the atmosphere). IC further confirmed that MSA did not chemically change while on the
608	filter. Infrared peaks in HMS spectra mid-way through the experiment indicate that HMS is not
609	stable on PTFE filters and likely converts to bisulfate. IC indicates that HMS is changing to
610	(bi)sulfate over time. Further investigations of measurements by FT-IR on PTFE of other
611	organosulfates are warranted to evaluate the extent to which the organosulfate functional group
612	can be quantified from IMPROVE PTFE filters. Further work to determine the stability and
613	ability to measure these compounds in aerosol mixtures as found in ambient samples is needed
614	before confidently using FT-IR on IMPROVE samples to measure organic sulfur compounds.
615	
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617	Funding: AMD, TD; Investigation: MBA, MDeB, LH, KL; Project Administration: AMD;
618	Supervision: AMD, TD; Writing First Draft: MBA, TD; Writing Revisions and Editing: MBA,
619	ST, TD, MDeB, LH, KL AMD; software: ST, MA; Visualization: MA
620	

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623

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5. References

- 631 Spectral Database for Organic Compounds, SDBS: https://sdbs.db.aist.go.jp/sdbs/cgi-
- 632 bin/direct_frame_top.cgi, last access: 22 February 2022.

- sulphate, nitrate and related species over the temperate South Pacific, Atmospheric Environment, 31,
 191–205, https://doi.org/10.1016/1352-2310(96)00194-X, 1997.
- Allen, A. G., Davison, B. M., James, J. D., Robertson, L., Harrison, R. M., and Hewitt, C. N.: Influence of
- 637 Transport over a Mountain Ridge on the Chemical Composition of Marine Aerosols during the ACE-2
- Hillcloud Experiment, Journal of Atmospheric Chemistry, 41, 83–107,
- 639 https://doi.org/10.1023/A:1013868729960, 2002.
- 640 Amore, A., Giardi, F., Becagli, S., Caiazzo, L., Mazzola, M., Severi, M., and Traversi, R.: Source
- 641 apportionment of sulphate in the High Arctic by a 10 yr-long record from Gruvebadet Observatory (Ny-
- 642 Ålesund, Svalbard Islands), Atmospheric Environment, 270, 118890,
- 643 https://doi.org/10.1016/j.atmosenv.2021.118890, 2022.
- Aneja, V. P. and Cooper, W. J.: Biogenic Sulfur Emissions, in: Biogenic Sulfur in the Environment, vol. 393,
 American Chemical Society, 2–13, https://doi.org/10.1021/bk-1989-0393.ch001, 1989.
- Barnes, I., Becker, K. H., and Mihalopoulos, N.: An FTIR product study of the photooxidation of dimethyl
 disulfide, Journal of Atmospheric Chemistry, 18, 267–289, https://doi.org/10.1007/BF00696783, 1994.
- 648 Bates, T. S., Lamb, B. K., Guenther, A., Dignon, J., and Stoiber, R. E.: Sulfur emissions to the atmosphere
- 649 from natural sources, Journal of Atmospheric Chemistry, 14, 315–337,
- 650 https://doi.org/10.1007/BF00115242, 1992.
- 651 Becagli, S., Lazzara, L., Fani, F., Marchese, C., Traversi, R., Severi, M., di Sarra, A., Sferlazzo, D.,
- 652 Piacentino, S., Bommarito, C., Dayan, U., and Udisti, R.: Relationship between methanesulfonate (MS-)
- 653 in atmospheric particulate and remotely sensed phytoplankton activity in oligo-mesotrophic central
- 654 Mediterranean Sea, Atmospheric Environment, 79, 681–688,
- 655 https://doi.org/10.1016/j.atmosenv.2013.07.032, 2013.
- 656 Boer, G. J., Sokolik, I. N., and Martin, S. T.: Infrared optical constants of aqueous sulfate-nitrate-
- ammonium multi-component tropospheric aerosols from attenuated total reflectance measurements—
 Part I: Results and analysis of spectral absorbing features, Journal of Quantitative Spectroscopy and
- 659 Radiative Transfer, 108, 17–38, https://doi.org/10.1016/j.jqsrt.2007.02.017, 2007.
- 660 Bondy, A. L., Craig, R. L., Zhang, Z., Gold, A., Surratt, J. D., and Ault, A. P.: Isoprene-Derived
- 661 Organosulfates: Vibrational Mode Analysis by Raman Spectroscopy, Acidity-Dependent Spectral Modes,
- and Observation in Individual Atmospheric Particles, The Journal of Physical Chemistry A, 122, 303–315,
 https://doi.org/10.1021/acs.jpca.7b10587, 2018.
- 664 Boris, A. J., Takahama, S., Weakley, A. T., Debus, B. M., Fredrickson, C. D., Esparza-Sanchez, M., Burki, C.,
- Reggente, M., Shaw, S. L., Edgerton, E. S., and Dillner, A. M.: Quantifying organic matter and functional groups in particulate matter filter samples from the southeastern United States – Part 1: Methods,

630

Allen, A. G., Dick, A. L., and Davison, B. M.: Sources of atmospheric methanesulphonate, non-sea-salt

- Atmospheric Measurement Techniques, 12, 5391–5415, https://doi.org/10.5194/amt-12-5391-2019,
 2019.
- 669 Boris, A. J., Takahama, S., Weakley, A. T., Debus, B. M., Shaw, S. L., Edgerton, E. S., Joo, T., Ng, N. L., and
- 670 Dillner, A. M.: Quantifying organic matter and functional groups in particulate matter filter samples from
- the southeastern United States Part 2: Spatiotemporal trends, Atmospheric Measurement Techniques,
 14, 4355–4374, https://doi.org/10.5194/amt-14-4355-2021, 2021.
- 673 Campbell, J. R., Battaglia, M. Jr., Dingilian, K., Cesler-Maloney, M., St Clair, J. M., Hanisco, T. F., Robinson,
- 674 E., DeCarlo, P., Simpson, W., Nenes, A., Weber, R. J., and Mao, J.: Source and Chemistry of
- Hydroxymethanesulfonate (HMS) in Fairbanks, Alaska, Environmental Science & Technology, 56, 7657–
 7667, https://doi.org/10.1021/acs.est.2c00410, 2022.
- Chackalackal, S. M. and Stafford, F. E.: Infrared Spectra of Methane-, Fluoro-, and Chlorosulfonic Acids,
 Journal of the American Chemical Society, 88, 4815–4819, https://doi.org/10.1021/ja00973a010, 1966.
- 679 Chapman, E. G., Barinaga, C. J., Udseth, H. R., and Smith, R. D.: Confirmation and quantitation of
- 680 hydroxymethanesulfonate in precipitation by electrospray ionization-tandem mass spectrometry,
- Atmospheric Environment. Part A. General Topics, 24, 2951–2957, https://doi.org/10.1016/0960-
- 682 1686(90)90475-3, 1990.
- 683 Chen, C., Zhang, Z., Wei, L., Qiu, Y., Xu, W., Song, S., Sun, J., Li, Z., Chen, Y., Ma, N., Xu, W., Pan, X., Fu, P., 684 and Sun, Y.: The importance of hydroxymethanesulfonate (HMS) in winter haze episodes in North China
- 685 Plain, Environmental Research, 211, 113093, https://doi.org/10.1016/j.envres.2022.113093, 2022.
- 686 Chen, Y., Zhang, Y., Lambe, A. T., Xu, R., Lei, Z., Olson, N. E., Zhang, Z., Szalkowski, T., Cui, T., Vizuete, W.,
- 687 Gold, A., Turpin, B. J., Ault, A. P., Chan, M. N., and Surratt, J. D.: Heterogeneous Hydroxyl Radical
- 688 Oxidation of Isoprene-Epoxydiol-Derived Methyltetrol Sulfates: Plausible Formation Mechanisms of
- Previously Unexplained Organosulfates in Ambient Fine Aerosols, Environmental Science & Technology
 Letters, https://doi.org/10.1021/acs.estlett.0c00276, 2020.
- 691 Chen, Y., Dombek, T., Hand, J., Zhang, Z., Gold, A., Ault, A. P., Levine, K. E., and Surratt, J. D.: Seasonal
- 692 Contribution of Isoprene-Derived Organosulfates to Total Water-Soluble Fine Particulate Organic Sulfur 693 in the United States, ACS Earth and Space Chemistry, 5, 2419–2432,
- 694 https://doi.org/10.1021/acsearthspacechem.1c00102, 2021.
- 695 Chihara, G.: Measurement of Infrared Absorption Spectra by Absorption on Japanese Hand-made Paper
- and Its Application to Paper Chromatography., Chemical & Pharmaceutical Bulletin, 6, 143–147,
- 697 https://doi.org/10.1248/cpb.6.143, 1958.
- Debus, B., Takahama, S., Weakley, A. T., Seibert, K., and Dillner, A. M.: Long-Term Strategy for Assessing
 Carbonaceous Particulate Matter Concentrations from Multiple Fourier Transform Infrared (FT-IR)
 Instruments: Influence of Spectral Dissimilarities on Multivariate Calibration Performance, Applied
 Spectroscopy, 73, 271–283, https://doi.org/10.1177/0003702818804574, 2019.
- 702 Debus, B., Weakley, A. T., Takahama, S., George, K. M., Amiri-Farahani, A., Schichtel, B., Copeland, S.,
- 703 Wexler, A. S., and Dillner, A. M.: Quantification of major particulate matter species from a single filter
- 704 type using infrared spectroscopy application to a large-scale monitoring network, Atmospheric
- 705 Measurement Techniques, 15, 2685–2702, https://doi.org/10.5194/amt-15-2685-2022, 2022.

- 706 Decesari, S., Facchini, M. C., Fuzzi, S., and Tagliavini, E.: Characterization of water-soluble organic
- 707 compounds in atmospheric aerosol: A new approach, Journal of Geophysical Research: Atmospheres,
- 708 105, 1481–1489, https://doi.org/10.1029/1999JD900950, 2000.
- 709 Dovrou, E., Lim, C. Y., Canagaratna, M. R., Kroll, J. H., Worsnop, D. R., and Keutsch, F. N.: Measurement
- techniques for identifying and quantifying hydroxymethanesulfonate (HMS) in an aqueous matrix and
 particulate matter using aerosol mass spectrometry and ion chromatography, Atmospheric
- 712 Measurement Techniques, 12, 5303–5315, https://doi.org/10.5194/amt-12-5303-2019, 2019.
- 713 Fankhauser, A. M., Lei, Z., Daley, K. R., Xiao, Y., Zhang, Z., Gold, A., Ault, B. S., Surratt, J. D., and Ault, A.
- 714 P.: Acidity-Dependent Atmospheric Organosulfate Structures and Spectra: Exploration of Protonation
- State Effects via Raman and Infrared Spectroscopies Combined with Density Functional Theory, The
 Journal of Physical Chemistry A, 126, 5974–5984, https://doi.org/10.1021/acs.jpca.2c04548, 2022.
- 717 Frossard, A. A., Shaw, P. M., Russell, L. M., Kroll, J. H., Canagaratna, M. R., Worsnop, D. R., Quinn, P. K.,
- and Bates, T. S.: Springtime Arctic haze contributions of submicron organic particles from European and
 Asian combustion sources, Journal of Geophysical Research: Atmospheres, 116,
- 720 https://doi.org/10.1029/2010JD015178, 2011.
- von Glasow, R. and Crutzen, P. J.: Model study of multiphase DMS oxidation with a focus on halogens,
 Atmospheric Chemistry and Physics, 4, 589–608, https://doi.org/10.5194/acp-4-589-2004, 2004.
- Grübler, A.: A Review of Global and Regional Sulfur Emission Scenarios, Mitigation and Adaptation
 Strategies for Global Change, 3, 383–418, https://doi.org/10.1023/A:1009651624257, 1998.
- 725 Hansen, A. M. K., Kristensen, K., Nguyen, Q. T., Zare, A., Cozzi, F., Nøjgaard, J. K., Skov, H., Brandt, J.,
- 726 Christensen, J. H., Ström, J., Tunved, P., Krejci, R., and Glasius, M.: Organosulfates and organic acids in
- Arctic aerosols: speciation, annual variation and concentration levels, Atmospheric Chemistry & Physics,
 14, 7807–7823, https://doi.org/10.5194/acp-14-7807-2014, 2014.
- Harrill, A. J.: Aqueous-Phase Processing of 2-Methyltetrol Sulfates by Hydroxyl Radical Oxidation in Fog
 and Cloud Water Mimics: Implications for Isoprene-Derived Secondary Organic Aerosol, 2020.
- 731 Hawkins, L. N., Russell, L. M., Covert, D. S., Quinn, P. K., and Bates, T. S.: Carboxylic acids, sulfates, and
- 732 organosulfates in processed continental organic aerosol over the southeast Pacific Ocean during
- 733 VOCALS-REx 2008, Journal of Geophysical Research: Atmospheres, 115,
- 734 https://doi.org/10.1029/2009JD013276, 2010.
- 735 Hettiyadura, A. P. S., Stone, E. A., Kundu, S., Baker, Z., Geddes, E., Richards, K., and Humphry, T.:
- 736 Determination of atmospheric organosulfates using HILIC chromatography with MS detection,
- Atmospheric Measurement Techniques, 8, 2347–2358, https://doi.org/10.5194/amt-8-2347-2015, 2015.
- 738 Hettiyadura, A. P. S., Jayarathne, T., Baumann, K., Goldstein, A. H., de Gouw, J. A., Koss, A., Keutsch, F.
- N., Skog, K., and Stone, E. A.: Qualitative and quantitative analysis of atmospheric organosulfates in
- Centreville, Alabama, Atmospheric Chemistry and Physics, 17, 1343–1359, https://doi.org/10.5194/acp 17-1343-2017, 2017.

- 742 Hoffmann, E. H., Tilgner, A., Schrödner, R., Bräuer, P., Wolke, R., and Herrmann, H.: An advanced
- 743 modeling study on the impacts and atmospheric implications of multiphase dimethyl sulfide chemistry,
- Proceedings of the National Academy of Sciences, 113, 11776–11781, 2016.
- 745 Kamruzzaman, M., Takahama, S., and Dillner, A. M.: Quantification of amine functional groups and their
- 746 influence on OM/OC in the IMPROVE network, Atmospheric Environment, 172, 124–132,
- 747 https://doi.org/10.1016/j.atmosenv.2017.10.053, 2018.
- Knovel Yaws' Critical Property Data for Chemical Engineers and Chemists Table 12. Vapor Pressure Organic Compounds, log P = A B/(T + C):
- 750 https://app.knovel.com/kn/resources/kt009ZN2S3/kpYCPDCECD/eptble/itable?b-
- 751 =&columns=1%2C2%2C3%2C6%2C4%2C5%2C10%2C11%2C13%2C14%2C12%2C7%2C8%2C9, last
- 752 access: 7 September 2022.
- 753 Krost, K. J. and McClenny, W. A.: FT-IR Transmission Spectroscopy for Quantitation of Ammonium
- Bisulfate in Fine-Particulate Matter Collected on Teflon® Filters, Applied Spectroscopy, 48, 702–705,
 1994.
- 756 Kuzmiakova, A., Dillner, A. M., and Takahama, S.: An automated baseline correction protocol for infrared
- 757 spectra of atmospheric aerosols collected on polytetrafluoroethylene (Teflon) filters, Atmospheric
- 758 Measurement Techniques, 9, 2615–2631, https://doi.org/10.5194/amt-9-2615-2016, 2016.
- 759 Kwong, K. C., Chim, M. M., Hoffmann, E. H., Tilgner, A., Herrmann, H., Davies, J. F., Wilson, K. R., and
- Chan, M. N.: Chemical Transformation of Methanesulfonic Acid and Sodium Methanesulfonate through
 Heterogeneous OH Oxidation, ACS Earth and Space Chemistry, 2, 895–903,
- 762 https://doi.org/10.1021/acsearthspacechem.8b00072, 2018a.
- Kwong, K. C., Chim, M. M., Davies, J. F., Wilson, K. R., and Chan, M. N.: Importance of sulfate radical
 anion formation and chemistry in heterogeneous OH oxidation of sodium methyl sulfate, the smallest
 organosulfate, Atmospheric Chemistry and Physics (Online), 18, https://doi.org/10.5194/acp-18-28092018, 2018b.
- Larkin, P. J.: Chapter 6 IR and Raman Spectra–Structure Correlations: Characteristic Group Frequencies,
 in: Infrared and Raman Spectroscopy (Second Edition), edited by: Larkin, P. J., Elsevier, 85–134,
 https://doi.org/10.1016/B978-0-12-804162-8.00006-9, 2018.
- 770 Laurent, J.-P. and Allen, D. T.: Size Distributions of Organic Functional Groups in Ambient Aerosol
- 771 Collected in Houston, Texas Special Issue of *Aerosol Science and Technology* on Findings from the Fine
- 772 Particulate Matter Supersites Program, Aerosol Science and Technology, 38, 82–91,
- 773 https://doi.org/10.1080/02786820390229561, 2004.
- Lee, J.-K., Lee, J.-S., Ahn, Y.-S., and Kang, G.-H.: Restoring the Reactivity of Organic Acid Solution Used for
 Silver Recovery from Solar Cells by Fractional Distillation, Sustainability, 11, 3659,
- 776 https://doi.org/10.3390/su11133659, 2019.
- 277 Lee, S.-H., Murphy, D. M., Thomson, D. S., and Middlebrook, A. M.: Nitrate and oxidized organic ions in
- single particle mass spectra during the 1999 Atlanta Supersite Project, Journal of Geophysical Research:
 Atmospheres, 108, SOS 5-1-SOS 5-8, https://doi.org/10.1029/2001JD001455, 2003.

- 780 Lin-Vien, D., Colthup, N. B., Fateley, W. G., and Grasselli, J. G.: CHAPTER 14 Organic Sulfur Compounds,
- 781 in: The Handbook of Infrared and Raman Characteristic Frequencies of Organic Molecules, edited by:
- 782 Lin-Vien, D., Colthup, N. B., Fateley, W. G., and Grasselli, J. G., Academic Press, San Diego, 225–250,
- 783 https://doi.org/10.1016/B978-0-08-057116-4.50020-1, 1991.

784 Liu, S., Takahama, S., Russell, L. M., Gilardoni, S., and Baumgardner, D.: Oxygenated organic functional

- 785 groups and their sources in single and submicron organic particles in MILAGRO 2006 campaign,
- 786 Atmospheric Chemistry & Physics, 9, 6849–6863, https://doi.org/10.5194/acp-9-6849-2009, 2009.

787 Liu, S., Ahlm, L., Day, D. A., Russell, L. M., Zhao, Y., Gentner, D. R., Weber, R. J., Goldstein, A. H., Jaoui,

- 788 M., Offenberg, J. H., Kleindienst, T. E., Rubitschun, C., Surratt, J. D., Sheesley, R. J., and Scheller, S.:
- 789 Secondary organic aerosol formation from fossil fuel sources contribute majority of summertime organic
- 790 mass at Bakersfield, Journal of Geophysical Research: Atmospheres, 117,
- 791 https://doi.org/10.1029/2012JD018170, 2012.
- Lloyd, A. G. and Dodgson, K. S.: Infrared studies on sulphate esters. II. Monosaccharide sulphates,
 Biochim Biophys Acta, 46, 116–120, https://doi.org/10.1016/0006-3002(61)90653-9, 1961.
- 794 Lloyd, A. G., Dodgson, K. S., Price, R. G., and Rose, F. A.: Infrared studies on sulphate esters. I.
- Polysaccharide sulphates, Biochim Biophys Acta, 46, 108–115, https://doi.org/10.1016/0006 3002(61)90652-7, 1961.
- 797 Ma, T., Furutani, H., Duan, F., Kimoto, T., Jiang, J., Zhang, Q., Xu, X., Wang, Y., Gao, J., Geng, G., Li, M.,
- Song, S., Ma, Y., Che, F., Wang, J., Zhu, L., Huang, T., Toyoda, M., and He, K.: Contribution of
- hydroxymethanesulfonate (HMS) to severe winter haze in the North China Plain, Atmospheric Chemistry
 and Physics, 20, 5887–5897, https://doi.org/10.5194/acp-20-5887-2020, 2020.
- 801 Moch, J. M., Dovrou, E., Mickley, L. J., Keutsch, F. N., Cheng, Y., Jacob, D. J., Jiang, J., Li, M., Munger, J.
- 802 W., Qiao, X., and Zhang, Q.: Contribution of Hydroxymethane Sulfonate to Ambient Particulate Matter: A
- Potential Explanation for High Particulate Sulfur During Severe Winter Haze in Beijing, Geophysical
 Research Letters, 45, 11,969-11,979, https://doi.org/10.1029/2018GL079309, 2018.
- 805 Moch, J. M., Dovrou, E., Mickley, L. J., Keutsch, F. N., Liu, Z., Wang, Y., Dombek, T. L., Kuwata, M.,
- 806 Budisulistiorini, S. H., Yang, L., Decesari, S., Paglione, M., Alexander, B., Shao, J., Munger, J. W., and
- Jacob, D. J.: Global Importance of Hydroxymethanesulfonate in Ambient Particulate Matter: Implications
 for Air Quality, Journal of Geophysical Research: Atmospheres, 125, e2020JD032706,
- 809 https://doi.org/10.1029/2020JD032706, 2020.
- 810 Okabayashi, H., Okuyama, M., Kitagawa, T., and Miyazawa, T.: The Raman Spectra and Molecular
- 811 Conformations of Surfactants in Aqueous Solution and Crystalline States, Bulletin of the Chemical
- 812 Society of Japan, 47, 1075–1077, https://doi.org/10.1246/bcsj.47.1075, 1974.
- 813 Olson, C. N., Galloway, M. M., Yu, G., Hedman, C. J., Lockett, M. R., Yoon, T., Stone, E. A., Smith, L. M.,
- and Keutsch, F. N.: Hydroxycarboxylic Acid-Derived Organosulfates: Synthesis, Stability, and
- 815 Quantification in Ambient Aerosol, Environmental Science & Technology, 45, 6468–6474,
- 816 https://doi.org/10.1021/es201039p, 2011.
- Pavia, D. L., Lampman, G. M., Kriz, G. S., and Vyvyan, J. A.: Introduction to Spectroscopy, Cengage
 Learning, 745 pp., 2008.

- 819 Peng, C. and Chan, C. K.: The water cycles of water-soluble organic salts of atmospheric importance, 820 Atmospheric Environment, 35, 1183–1192, https://doi.org/10.1016/S1352-2310(00)00426-X, 2001.
- 821 Reggente, M., Höhn, R., and Takahama, S.: An open platform for Aerosol InfraRed Spectroscopy analysis
- 822 - AIRSpec, Atmospheric Measurement Techniques, 12, 2313-2329, https://doi.org/10.5194/amt-12-
- 823 2313-2019, 2019.
- 824 Russell, L. M., Bahadur, R., and Ziemann, P. J.: Identifying organic aerosol sources by comparing
- 825 functional group composition in chamber and atmospheric particles, Proceedings of the National Academy of Sciences, 108, 3516–3521, https://doi.org/10.1073/pnas.1006461108, 2011. 826
- 827 Ruthenburg, T. C., Perlin, P. C., Liu, V., McDade, C. E., and Dillner, A. M.: Determination of organic matter
- 828 and organic matter to organic carbon ratios by infrared spectroscopy with application to selected sites in the IMPROVE network, Atmospheric Environment, 86, 47-57, 829
- 830 https://doi.org/10.1016/j.atmosenv.2013.12.034, 2014.
- Saltzman, E. S., Savoie, D. L., Prospero, J. M., and Zika, R. G.: Methanesulfonic acid and non-sea-salt 831
- 832 sulfate in pacific air: Regional and seasonal variations, Journal of Atmospheric Chemistry, 4, 227-240, https://doi.org/10.1007/BF00052002, 1986. 833
- 834 Sato, S., Higuchi, S., and Tanaka, S.: Structural Examinations of "Sodium FormaldehydeSulfoxylate" by
- Infrared and Raman Spectroscopy, Nippon Kagaku Kaishi, 1984, 1151–1157, 835
- 836 https://doi.org/10.1246/nikkashi.1984.1151, 1984.
- 837 Segneanu, A. E., Gozescu, I., Dabici, A., Sfirloaga, P., and Szabadai, Z.: Organic Compounds FT-IR 838 Spectroscopy, IntechOpen, https://doi.org/10.5772/50183, 2012.
- 839 Seinfeld, J. H. and Pandis, S. N.: Atmospheric Chemistry and Physics: From Air Pollution to Climate 840 Change, Wiley, 2016.
- 841 Shurvell, H. F.: Spectra- Structure Correlations in the Mid- and Far-Infrared, in: Handbook of Vibrational Spectroscopy, American Cancer Society, https://doi.org/10.1002/0470027320.s4101, 2006. 842
- 843 Smith, S. J., van Aardenne, J., Klimont, Z., Andres, R. J., Volke, A., and Delgado Arias, S.: Anthropogenic 844 sulfur dioxide emissions: 1850–2005, Atmospheric Chemistry and Physics, 11, 1101–1116,
- 845 https://doi.org/10.5194/acp-11-1101-2011, 2011.
- 846 Song, S., Gao, M., Xu, W., Sun, Y., Worsnop, D. R., Jayne, J. T., Zhang, Y., Zhu, L., Li, M., Zhou, Z., Cheng,
- 847 C., Lv, Y., Wang, Y., Peng, W., Xu, X., Lin, N., Wang, Y., Wang, S., Munger, J. W., Jacob, D. J., and McElroy, 848 M. B.: Possible heterogeneous chemistry of hydroxymethanesulfonate (HMS) in northern China winter
- haze, Atmospheric Chemistry and Physics, 19, 1357–1371, https://doi.org/10.5194/acp-19-1357-2019, 849 850 2019.
- 851 Stone, E. A., Yang, L., Yu, L. E., and Rupakheti, M.: Characterization of organosulfates in atmospheric aerosols at Four Asian locations, Atmospheric Environment, 47, 323–329, 852
- 853 https://doi.org/10.1016/j.atmosenv.2011.10.058, 2012.
- 854 Surratt, J. D., Chan, A. W. H., Eddingsaas, N. C., Chan, M., Loza, C. L., Kwan, A. J., Hersey, S. P., Flagan, R. C., Wennberg, P. O., and Seinfeld, J. H.: Reactive intermediates revealed in secondary organic aerosol 855

- formation from isoprene, Proceedings of the National Academy of Sciences, 107, 6640–6645,
 https://doi.org/10.1073/pnas.0911114107, 2010.
- Tang, K.: Chemical Diversity and Biochemical Transformation of Biogenic Organic Sulfur in the Ocean, Frontiers in Marine Science, 7, 2020.
- 860 U.S. EPA. Comptox Chemicals Dashboard:
- https://comptox.epa.gov/dashboard/chemical/details/DTXSID80805075 (accessed November 14, 2022)
 Methyl 10,10-diethoxydec-2-ene-4,6,8-triynoate, 2022.
- 863 Wang, Y., Zhao, Y., Wang, Y., Yu, J.-Z., Shao, J., Liu, P., Zhu, W., Cheng, Z., Li, Z., Yan, N., and Xiao, H.:
- 864 Organosulfates in atmospheric aerosols in Shanghai, China: seasonal and interannual variability, origin,
- and formation mechanisms, Atmospheric Chemistry and Physics, 21, 2959–2980,
- 866 https://doi.org/10.5194/acp-21-2959-2021, 2021.

867 Wei, L., Fu, P., Chen, X., An, N., Yue, S., Ren, H., Zhao, W., Xie, Q., Sun, Y., Zhu, Q.-F., Wang, Z., and Feng,

- 868 Y.-Q.: Quantitative Determination of Hydroxymethanesulfonate (HMS) Using Ion Chromatography and
- 869 UHPLC-LTQ-Orbitrap Mass Spectrometry: A Missing Source of Sulfur during Haze Episodes in Beijing,
- Environmental Science & Technology Letters, 7, 701–707, https://doi.org/10.1021/acs.estlett.0c00528,
 2020.
- Yazdani, A., Dudani, N., Takahama, S., Bertrand, A., Prévôt, A. S. H., El Haddad, I., and Dillner, A. M.:
- Fragment ion-functional group relationships in organic aerosols using aerosol mass spectrometry and
 mid-infrared spectroscopy, Atmospheric Measurement Techniques, 15, 2857–2874,
- 875 https://doi.org/10.5194/amt-15-2857-2022, 2022.
- Zawadowicz, M. A., Proud, S. R., Seppalainen, S. S., and Cziczo, D. J.: Hygroscopic and phase separation
 properties of ammonium sulfate/organics/water ternary solutions, Atmospheric Chemistry and Physics,
- 878 15, 8975–8986, https://doi.org/10.5194/acp-15-8975-2015, 2015.
- Zeng, G., Kelley, J., Kish, J. D., and Liu, Y.: Temperature-Dependent Deliquescent and Efflorescent
 Properties of Methanesulfonate Sodium Studied by ATR-FTIR Spectroscopy, The Journal of Physical
- 881 Chemistry A, 118, 583–591, https://doi.org/10.1021/jp405896y, 2014.
- 882 Zhao, X., Shi, X., Ma, X., Zuo, C., Wang, H., Xu, F., Sun, Y., and Zhang, Q.: 2-Methyltetrol sulfate ester-
- initiated nucleation mechanism enhanced by common nucleation precursors: A theory study, Science of
 The Total Environment, 723, 137987, https://doi.org/10.1016/j.scitotenv.2020.137987, 2020.
- Zhong, L. and Parker, S. F.: Structure and vibrational spectroscopy of methanesulfonic acid, Royal Society
 of Open Science, 5, 181363, https://doi.org/10.1098/rsos.181363, 2022.
- 887