

Preparation of hybrid calibrated absorption cross sections for a compact UV-DOAS measurement

Beni Adi Trisna^{1,2}, Sang Woo Kim^{1,3}, Yong-Doo Kim⁴, Miyeon Park^{1,2}, Seung-Nam Park⁵, Jeongsoon Lee^{1,2}

5

¹Greenhouse Gas Metrology team, Korea Research Institute of Standard and Science (KRISS), 267 Gajeong-ro, Yuseong-gu, Daejeon 34113, Republic of Korea

²Science of Measurement, University of Science and Technology (UST), 217 Gajeong-ro, Yuseong-gu, Daejeon 34113, Republic of Korea

³ 3 10 Sunkyungkwan University (SKKU), 2066 Seobu-ro, Jangan-gu, Suwon 16419, Republic of Korea 4Center for Gas Analysis, Korea Research Institute of Standard and Science (KRISS), 267 Gajeong-ro, Yuseonggu, Daejeon 34113, Republic of Korea ⁵National Centre for Standard Reference Data (NCSRD), Korea Research Institute of Standard and Science

(KRISS), Daejeon 34113, Republic of Korea

15

Correspondence to: Jeongsoon Lee (leejs@kriss.re.kr)

Abstract. A low-cost differential optical absorption spectroscopy (DOAS) spectrometer has been calibrated using a hybrid approach in which multiple absorption cross sections (ACSs) are measured and compared with previously obtained ACSs. The ACSs obtained from the measurement in this study are referred to as precise measurement

- 20 (PM) ACSs, while those obtained from the simulation are referred to as hybrid simulation (HS) ACSs. The compact UV grating spectrometer was used to measure PM ACSs at different pressures and temperatures ranging from 295.75 K to 298.15 K. The spectral range was 230-320 nm with a spectral resolution of 0.28-0.39 nm. Uncertainties were evaluated and traceable to SI units. The relative standard measurement uncertainty of the PM ACS for o-xylene, p-xylene, SO2, benzene, styrene and toluene were 3.1%, 3.1%, 3.1%, 3.0%, 3.0% and 2.9%.
- 25 respectively. This combined relative standard uncertainty includes contributions from optical path length (1.4%), CRM concentration (≤0.74%), absorbance repeatability (≤1.5%), pressure (0.78%) and temperature (2.1%). ACSs for reactive trace gases without CRMs were derived using the HS method. This involved convolving literature spectra with measured instrumental functions (IFs) and subsequently applying spectral fitting to achieve ACSs that are well-aligned with those that would be obtained if measured with this spectrometer. The uncertainty of the
- 30 HS ACSs results from the referenced ACSs, the determination of the IFs and the fitting procedures. The uncertainty of the referenced ACSs is taken directly from the literature and ranges from 0.05% for benzaldehyde and formaldehyde to 5% for oxygen. The uncertainty of the IFs is 0.67%, 0.74% and 0.65% in regions I, II and III respectively. The total uncertainty for HS ACS is estimated to be 0.25% for benzaldehyde, 1.01% for NO2, 0.95% for formaldehyde, 0.75% for p-cresol, 0.74% for m-xylene, 4.13% for phenol, 4.19% for ethylbenzene,
- 35 2.83% for ozone in region I, 3.51% for ozone in region II, 5.24% for oxygen in region I and 5.31% for oxygen in region II, taking into account the above uncertainty components. In a laboratory measurement of a gas mixture (benzene, toluene and ortho-xylene), the difference between the measurement values of the certified reference material (CRM) produced by the gravimetric method and the DOAS measurement is 8.5% for benzene, 1.6% for toluene and 4.9% for ortho-xylene. Therefore, the uncertainty of the DOAS system was estimated to be 15.2% for
- 40 benzene, 8.2% for toluene and 11.6% for o-xylene, taking into account the uncertainties of the ACS (\leq 3.1%), the

fitting procedure ($\leq 0.6\%$), the difference with the CRM value ($\leq 8.5\%$) and the path length (1.4%). In general, the system is ready for use in field measurements using the long-path DOAS technique.

List of Abbrevations and Variables

50

55

1 Introduction

- Toxic gases have been substantially released into the urban atmosphere since the rapid growth of industries and 65 mega-cities. These pollutants, including polycyclic aromatic hydrocarbons (PAHs) and volatile organic compounds (VOCs) such as benzene, toluene, sulfur dioxide (SO₂), styrene, and xylenes, originate from both indoor and outdoor environments. Major sources of these emissions include petrochemical industries, oil refineries, motor vehicles, and power plant smokestacks (Johansson et al., 2014; Liu et al., 2016; Skorokhod et al., 2017; Wei et al., 2017). Their active reactions in the atmosphere contribute to the formation of secondary
- 70 pollutants (Jenkin and Clemitshaw, 2000). Estimating emission inventories and understanding the photochemistry of these compounds are crucial, requiring continuous monitoring of gaseous components such as VOCs, SO₂, and NO₂. The use of compact monitoring devices enables the effective implementation of policies and actions to manage emissions, thereby promoting the restoration of clean and safe air in a cost-effective manner.
- Among numerous methods for measuring gaseous VOCs, SO2, and NO2, optical techniques such as 75 differential optical absorption spectroscopy (DOAS) (Platt and Stutz, 2008; Lee and Kim, 2003; Lee et al., 2002) and Fourier-transform infrared spectroscopy (FTIR) (Fally et al., 2009) has been widely used in monitoring stations. Absorption spectroscopy is particularly advantageous due to its ability to non-destructively quantify gas concentrations over a defined optical path length in the atmosphere, with minimal maintenance requirements. Beyond their economic benefits, small and lightweight spectrometers offer the additional advantage of easy
- 80 deployment for outdoor air quality monitoring. However, in terms of detector performance—such as sensitivity and pixel spacing—these compact spectrometers typically have lower performance compared to high-resolution spectrometers. Accordingly, studies of outdoor demonstrations using compact spectrometers are also widely reported with the growing interest in performance evaluation.
- To generate reliable air quality monitoring data, it is crucial to produce accurate ACSs. While the ideal 85 approach is to create ACSs using the same spectrometer intended for monitoring, an alternative method involves convoluting published ACSs. When using the published ACS, it is recommended to select high resolution ACSs to simulate ACSs by convoluting them with the instrument function (IF) of the spectrometer in use, as demonstrated by Fally et al. (2009). Typically, this convolution assumes that the IF of the spectrometer remains constant across the measurement wavelength range; however, in reality, the IF varies with wavelength. This 90 assumption can lead to significant discrepancies between measured and simulated ACSs. For instance, when comparing SO₂ absorption spectra obtained by convoluting spectral databases at various resolutions—ranging from low (0.16 to 0.49 nm) to high (FTIR: 16 cm⁻¹ and grating spectrometer: 0.05 to 0.11 nm)—with the IF, we observed a substantial difference of up to 62% (8.4×10^{-4}) in peak-to-peak optical density between the convoluted and measured spectra.
- 95 Since Pantos et al. (1978) measured the absolute ACSs of benzene in the temperature range of 293.15 to 300.15 K within the 135 to 270 nm region at resolutions of 0.08 and 0.25 nm, subsequent studies have reported ACSs in the 130–280 nm region at resolutions ranging from 0.065 to 0.2 nm (Hiraya and Shobatake, 1991; Suto et al., 1992). Additionally, ACSs for monocyclic aromatic hydrocarbons have been measured in the 220–290 nm region with resolutions between 0.11 and 0.2 nm (Milton et al., 1992; Trost et al., 1997; Etzkorn et al., 1999) and
- 100 in the UV region (238–270 nm) with improved spectral resolution of 1 cm⁻¹ (Fally et al., 2009). More recently, Klingbeil et al. (2007) reported the temperature dependence of ACSs from 298.15 to 773.15 K in the mid-IR region with a spectral resolution of 1 cm⁻¹ using FTIR. In addition to aromatic hydrocarbons, absolute ACSs of

SO₂ have been reported at temperatures of 295 and 210 K in the 300-325 nm region with a resolution of 0.03 nm (Mcgee and Burris Jr, 1987), and at 295 K in the 250–370.37 nm region with resolutions of 2 and 16 cm⁻¹ using 105 a Fourier-transform spectrometer (Vandaele et al., 1994). Although the accuracy of ACSs has improved over time,

the uncertainties associated with these measurements have rarely been thoroughly assessed. In this study, we aimed to generate accurate ACSs either through precise measurement (PM) or hybrid simulation (HS) and to estimate their uncertainties, which are crucial for the application of compact spectrometers in air quality monitoring. We present an approach for generating accurate hybrid ACSs to ensure reliable results

110 in DOAS observations. The hybrid simulation method combines convolution and spectral fitting to the actual spectral pixels on the instrument detector array. After applying our method, the discrepancy between the hybrid and measured ACSs was reduced by up to 8.8% (2.7×10^{-4}) in the case of SO₂ absorption. The measurement uncertainties of ACSs have been thoroughly assessed, ensuring that the observations are traceable to the International System of Units (SI).

115 **2 Experimental details**

2.1 Instrumental setup for ACS measurement

Figure 1 provides a schematic overview of the ACS measurement setup. A xenon arc lamp (E7536, Hamamatsu, Shizuoka, Japan) was collimated using a lens with a focal length of 5 cm, allowing the beam to pass parallel through a gas cell that is 50 cm in length and 2.54 cm in diameter. ACS measurements were conducted by scanning

120 the CRM gas spectrum following the acquisition of a background spectrum. Each measurement was repeated five times to ensure accuracy and reproducibility.

First, the background spectrum was measured through an evacuated cell after the cell had been purged with N_2 gas at 10^{-1} mbar or less. Second, the CRM spectrum was obtained from a gas cell filled with CRM to the target pressure after several flushes, maintained at a specified temperature and pressure. The cell pressure was

- 125 measured with a relative standard uncertainty of 0.4% using a pressure transducer (MKS 626C13TAE, MA, USA), which was calibrated against an ultrasonic interferometer manometer at KRISS. Temperature measurements were conducted using a thermohygrometer (Testek, 303A, OH, USA) equipped with a PRT sensor probe attached to the cell surface. The PRT sensor was calibrated against a standard platinum resistance thermometer (Fluke-Hart Scientific, 5628, WA, USA), resulting in a measurement uncertainty of approximately 1%. CRM cylinders were
- 130 allowed to equilibrate in the laboratory for 3 to 5 days prior to measurement.

A Czerny-Turner type mini spectrometer (Avantes AVASpecs-ULS2048LTEC-USB2, Apeldoorn, Netherlands), which is both lightweight and compact, was employed. To obtain signals from 237 nm to 355 nm, the spectrometer was equipped with a 2400 line/mm grating, an entrance slit measuring 50 μ m wide × 1000 μ m high, and a charge-coupled device (CCD; Sony 2048-pixel CCD linear image sensor, ILX511, TYO, Japan). A

135 DUV lumogen-coated CCD was used to prevent second-order UV response effects and enhance sensitivity, and it was cooled to 278.15 K to reduce thermal noise and signal distortion. The spectra were acquired using integrated software provided by the spectrometer manufacturer and were processed using a custom fitting program to quantify the gas concentration in the cell.

2.2 Preparation of CRMs

140 All standard mixtures were prepared gravimetrically in accordance with ISO 6142: Gas Analysis–Preparation of Calibration Gas Mixtures–Part 1: Gravimetric Method for Class I Mixtures (ISO, 2015) (Guide, 2016), by diluting high-purity raw materials with high-purity nitrogen, as detailed in a previous study (Kang et al., 2021). A 10 L electro-polished aluminum cylinder (Luxfer, Nottingham, UK) was used to minimize adsorption and potential reactions on the cylinder surface or between components of the mixture. Before filling, each cylinder was 145 evacuated using a turbo pump (Varian TV301-NAV, USA) to a pressure of 10^{-6} bar while being heated to

approximately 340 K for two days to remove any residual moisture.

To accurately determine the mass of gas, two highly precise mass balances (Mettler-Toledo XP-26003L and AT201, Greifensee, Switzerland) were calibrated against Class E2 calibration weight reference standards (Metrology, 2004). After evacuation, a mass corresponding to the target molar fraction was injected via syringe 150 into a line connected to the evacuated cylinder, vaporized with an N₂ diluent, and transferred to the cylinder. The masses of the syringes before and after liquid loading, as well as the N_2 filling of the cylinder, were measured.

- During the weighing process, the laboratory was maintained at a temperature of 293 ± 1 K and a humidity of 50 \pm 10%. A detailed description of the CRM gas production process is available in previous studies (Kang et al., 2021; Kim et al., 2018).
- 155 Table 1 lists the concentrations and associated uncertainties of the CRMs used in this study. Benzene, toluene, ortho-xylene, para-xylene, styrene, and sulfur dioxide were prepared in a nitrogen balance, along with a mixture containing benzene, toluene, and ortho-xylene. The molar fractions of these compounds and their relative standard uncertainties, ranging from 0.45% to 0.74%.

2.3 Hybrid calibration of spectrometer's ACSs

160 **2.3.1 Instrument functions**

To produce accurate ACSs, the instrument function (IF) of the spectrometer must be determined. The IF, which characterizes the spectral features, can be obtained by measuring the emission line of a calibration lamp, such as a mercury (Hg) lamp (model: Pen-Ray 11SC, Analytik-jena). Hybrid simulation (HS) of ACSs (denoted as $A(\lambda)_{HS}$) involves converting the reference absorption spectrum from the literature (denoted as $A(\lambda)_{ref}$) using the

165 IF to simulate how the absorption spectrum would appear if measured by the specific spectrometer. This process, known as hybrid simulation (HS), is represented by the mathematical expression below (Platt and Stutz, 2008)

$$
|A(\lambda)_{meas} - A(\lambda)_{HS}| \to min,
$$
\nwhere\n
$$
\tag{1}
$$

 $A_{HS}(\lambda) = c \cdot (A_{ref}(\lambda) \otimes IF_{\lambda})_{\lambda_{this work} = a \cdot \lambda_{Lit} + b''}$ (2)

170 As a form of virtual spectrometer, the IF cannot perfectly replicate the spectrum dispersed by the monochromator onto the CCD array pixels, as shown in equation (2). To achieve ACSs that are well-matched to our spectrometer, the ACS spectrum is simulated independently for each wavelength region, using the IF specific to that region. Each regional IF captures the optical characteristics of the corresponding area, where the absorption lines of each gas are simulated. This approach accounts for the differing resolutions across the three regions. The 175 three regions were selected based on the significant variations in the BTX absorption peaks.

As shown in Figure 2, the full width at half maximum (FWHM) of the IFs measured in different regions varied significantly, a phenomenon commonly observed in compact grating-type spectrometers with short focal lengths (Palmer and Loewen, 2005). To address this, the measured spectra were processed by dividing the fitting region into smaller, arbitrary slices to minimize residuals. Given that the nonuniformity of the FWHM

180 substantially impacts the accuracy of gas concentration measurements, three distinct IFs were utilized in this study to mitigate this issue. Specifically, region I (238-249 nm) used the IF shown in Figure 2(a), region II (249-283 nm) used the IF depicted in Figure 2(b), and region III (283-305 nm) employed the IF illustrated in Figure 2(c). Each IF corresponds to a specific wavelength range.

2.3.2 Hybrid simulation (HS) of the ACS

- 185 The ACSs for oxygen, ozone, ethylbenzene, m-xylene, phenol, para-cresol, formaldehyde, and benzaldehyde were generated using HS method, with the with the categorization of the simulation range is summarized in Table 2. The reference spectra (Trost et al., 1997; Etzkorn et al., 1999; Meller and Moortgat, 2000; Voigt et al., 2001; Volkamer, 1996; Volkamer, 2001) were convoluted with the IFs determined in Section 2.3.1 using a custom processing tool developed in MATLAB software. The wavelength range was divided into three distinct regions,
- 190 each with its own IF. Although absorptions are distributed across all three regions and multiple species coexist within each region, species with prominent absorptions were grouped and fitted together. Notably, oxygen and ozone exhibit strong absorption across all regions. Accordingly, three separate ACSs for ozone were independently created for each of the three regions, as detailed in Table 2.
- The technical details of the reference ACSs, including cell length, concentration, wavelength range, 195 and measurement technique, and resolution are summarized in Table 3. The ACSs reported by B. Trost et al. (1997) and Etzkorn et al. (1999) are comparable to those presented in this study. Both studies utilized spectrometers with spectral resolutions of 0.146 nm and 0.11 nm (FWHM), respectively, and measured the ACSs for ethylbenzene, benzaldehyde, meta-xylene, and para-cresol using samples prepared by introducing liquids into a 19 mm diameter quartz cell. All ACSs listed in the table were measured with spectral resolutions ranging from 200 0.025 to 0.2 nm.

3 Results and discussions

3.1 Hybrid calibration of DOAS spectrometer

From the Lambert-Beer's law, the intensity *I(λ,L)* at the end of optical light path length *L* from a light intensity $I_0(\lambda, L)$ in the presence of gas species can be derived as

205
$$
I(\lambda, L) = I_0(\lambda, L) exp[-L(\sum \sigma_i(\lambda, p, T) \cdot c_i(l) + \varepsilon_R(\lambda, l) + \varepsilon_M(\lambda, l))],
$$
 (3)

where $\varepsilon_R(\lambda)$ is the extinction from air molecules (Rayleigh scattering) and $\varepsilon_M(\lambda)$ is the extinction from aerosols particles (Mie scattering). The spectral absorption of gas species *i* is designated by its ACS $\sigma_i(\lambda, p, T)$, which depends on the wavelength λ , pressure p , and temperature T , and by its position along the optical path length l . The total absorption in the path must to be obtained from both precise measurement ($\sigma_{PM}(\lambda)$) and hybrid

210 simulation ($\sigma_{HS}(\lambda)$). As a result, equation (3) can be modified to incorporate a hybrid combination of both measurement and simulation method, as follows:

$$
I(\lambda, L) = I_0(\lambda, L) exp[-L(\sum \sigma_{PM}(\lambda) \cdot c_i + \sigma_{HS}(\lambda) \cdot c_j + \varepsilon_R(\lambda) + \varepsilon_M(\lambda))],
$$
\n(4)

For the $\sigma_{HS}(\lambda)$, we have compared the simulated ACS with the measured in the work for some species, such as benzene, para-xylene, and SO₂. As shown in Figure 5, the ACS obtained from the HS method is not 215 significantly different from the PM value because it is determined through the optimization of *a*, *b*, and κ coefficients through solving NLS (Eq. 5) to equally fit the measured ACS.

$$
\sigma_{HS}(\lambda) = \kappa \cdot (\sigma_{ref}(\lambda) \otimes IF_{\lambda})_{\lambda_{this \text{ work}} = a \cdot \lambda_{Lit} + b},\tag{5}
$$

where *a*, *b*, and κ are the NLS fit coefficients of wavelength squeezing, wavelength shifting, and intensity scaling for the convoluted ACS. We then applied the NLS fit coefficients for a set of gas species at regions (Region I to

220 III) to let them match to the measured through our spectrometer. Finally, we obtained the mathematical expression for a hybrid calibration of DOAS spectrometer.

$$
I'(\lambda, L) = I'_0(\lambda, L) exp\left[-L\left(\sum_{j=1}^{j=n} \sigma'_{PM}(\lambda) \cdot c_j + \sum_{k=1}^{k=n} \sigma'_{HS}(\lambda) \cdot c_k\right)\right],\tag{6}
$$

where the prime symbol (') indicates the differential quantities for the respective parameters, which is called hybrid calibration method to produce ACSs for this work.

225 **3.2 Determination of the precise measured ACSs,** $\sigma_{PM}(\lambda)$

be expressed as:

Figure 3 presents the raw spectra recorded by the spectrometer for both the background and sample measurements for toluene gas as a representative example. To ensure consistent illumination, we aligned and checked the uniformity of the xenon lamp beam and verified its uniformity at both the entry and exit points of the fiber optic cable. The background spectrum (black line) displays a lamp profile with a sharp peak around 291 nm. However,

230 by dividing the sample spectrum by the background spectrum, this approach effectively eliminates the lamp profile and spectral noise, enhancing the quality of the ACS measurements and mitigating stray light and internal variations from the xenon lamp.

The ACSs of benzene, toluene, styrene, sulfur dioxide, ortho-xylene, and para-xylene measured in this study are presented in Figure 4. Table 4 provides a comprehensive summary of all ACS data, including those 235 measured in this work and those reported by others, along with the corresponding measurement parameters, such as resolution, temperature, pressure, and the chemicals used. It is important to note that the spectrometer employed in this study inherently sacrifices some performance quality in exchange for its compact size. As previously discussed, the shape of the peaks is influenced by both optical and environmental factors, including optical resolution and the temperature and pressure within the gas cell, as detailed in Table 4. The intensities of the 240 measured ACSs differed by up to 47.6% from the ACSs derived by convolving the literature data with the instrument function (IF) previously obtained in Figure 2 (a-c). The apparent ACSs from these measurements can

$$
\sigma_{PM}(\lambda) = \frac{-\ln(A(\lambda)_{meas})^R \cdot T}{c_j p L} \bigg|_{\lambda = \lambda_0},\tag{7}
$$

In this context, $A(\lambda)$ represents the ratio of the measured transmittances with and without the gas species present 245 in the cell $\frac{I(\lambda)}{I_0(\lambda)}$, *L* is the optical path length, c_j is the mole fraction of the gas species, *p* is the total pressure in the cell, *T* is the temperature, and *R* is the universal gas constant. Although absorption due to electronic transitions is generally less sensitive to environmental parameters such as temperature compared to vibrational transitions (Grosch et al., 2010), the peak magnitudes in the ACS of Eq. (7) have been reported to be temperature-dependent. This effect has been observed not only in the ultraviolet but also in the infrared spectra, where higher temperatures 250 result in sharper and more pronounced peaks in the ACS of hydrocarbons (Klingbeil et al., 2007; Fally et al., 2009). These findings likely contribute to the differences observed between the ACS measurements in our study and those reported in previous research.

Unlike the earlier studies (Fally et al., 2009; Trost et al., 1997; Etzkorn et al., 1999), which primarily used methods involving the degassing of liquids into gas tubes, our study employed gaseous certified reference 255 materials (CRMs) under controlled conditions. The degassing of liquids into gas tubes has the disadvantage of a tendency for the resulting gases to easily condense and adsorb onto the walls of the apparatus, making it difficult to accurately quantify their concentration.

We evaluated the uncertainty in the measured ACS, considering factors as shown in Eq. (7). The total combined relative standard uncertainties *uc,r* of the ACS were calculated using the general law of uncertainty 260 propagation (BIPM, 2008), resulting in the following:

$$
u_{c,r}^2(\sigma_{PM}(\lambda)) = u_{L,r}^2 + u_{c_{j},r}^2 + u_{UV,r}^2 + u_{rep,r}^2 + u_{p,r}^2 + u_{T,r}^2,
$$
\n(8)

The relative standard uncertainty because of light path length u_{LT} contributed to 1.4% of the relative uncertainty. The relative standard uncertainty of CRM, $u_{c_j,r}$ obtained using the gravimetric technique vary depending on each gas species, and the values are summarized in Table 1. The relative standard uncertainty because of absorbance

- 265 has two sources: the stability of UV light source u_{IUVx} and the repeatability of absorbance u_{renx} . u_{IUVx} was estimated from the intensity stability of I_0 measurements for a day and the absorbance and $u_{rep,r}$ was calculated by considering the standard deviation of four ACS measurements. Upon combining these two uncertainties, the absorbance uncertainty values range between 1.0% (toluene) and 1.5% (o-xylene). The relative standard uncertainty because of pressure $u_{p,r}$ and temperature $u_{r,r}$ measurements contributed 0.78% and 2.1%, 270 respectively. The result shows that *uc,r* is 3.0% (benzene and styrene), 3.1% (o-xylene, p-xylene and SO2), and
- 2.9% for toluene. The main uncertainty component in the ACS calculation in this work is temperature, which contributes approximately 47-54% to the total uncertainty.

The detection limits presented in Table 4 were calculated based on a 1000 m light path length. The detection limit values (ppb·m·atm) were derived by applying the signal-to-noise ratio (Shrivastava and Gupta, 2011) to the following equation: $DL = \frac{3 * \overline{Noise}}{2D}$ 275 2011) to the following equation: $DL = \frac{s \times 1000 \text{ m}}{00 \text{ m}} c \cdot p \cdot L$. (Etzkorn et al., 1999) also noted that the minimum detectable optical density depends on the absorber concentration, visibility, and the field setup of LP-DOAS. In this study, the detection limits for each gas species ranged from 0.01 ppb (para-xylene) to 0.54 ppb (ortho-xylene), which is highly competitive given the low-resolution dispersion of the compact spectrometer.

3.3 ACS from hybrid simulation $\sigma_{HS}(\lambda)$

- 280 Figure 5 presents the ACSs from reference sources (blue line in the upper panel) and compares the ACS spectra obtained using the PM method (red line) with those from the HS method (black line) in the lower panel. The simulated ACSs were generated using parameters derived from the method outlined in Section 2.3. The HS procedure involved shifting, squeezing, linear multiplication, and convolution of the ACS spectra. The parameters applied in the non-linear least-squares (NLS) fit are summarized in Table 5.
-

285 When comparing the PM results to those processed with the HS method for various gas species, we found a strong agreement in the shape, position, and value of the ACS peaks, as shown in Figure 5 and summarized in Table 6. Notably, the discrepancy of up to 46% observed between the benzene ACS reported by Etzkorn et al. (1999) and this work was reduced to 0.8% after applying the HS method. The uncertainty in preparing ACSs using the HS method was found to be no more than 5.31%.

290 First-order fitting was performed by introducing shift and squeeze parameters to account for the nonuniformity of pixel spacing as a function of wavelength, resulting in a shift of the benzene peak position by 1 to 4 pixels (0.06 to 0.24 nm). Ultimately, applying the HS method led to an improved ACS, with peak positions matching within 1 pixel (0.06 nm).

Based on the successful application of the HS method, as shown in Table 6, the parameters derived for 295 benzene, para-xylene, and SO₂ (Table 5) were used to obtain the ACS for other species that could not be measured in the laboratory. The results for these additional gas species are presented in Figure 6, with details on the instrument function, peak location, and detection limits summarized in Table 7.

The detection limits for the gases measured in Table 7 were calculated for a path of 1,000 meters and were obtained by applying the same calculation method as for PM. From the simulated ACS spectra, the detection 300 limits of gas species range from 0.04 ppb (ozone in region I) to 0.80 ppb (nitrogen dioxide), which were

comparable to references of the Table 7.

The ACS intensity is valid only within a range close to the instrument function used in the convolution process. As demonstrated in Figure 6 for O2 and O3, the ACS intensity values vary in the simulation results due to their broad-band absorption characteristics and the use of three different instrument functions. Therefore, when 305 conducting measurements with LP-DOAS, it is essential to adjust the fitting region for the gas to align with the ACS obtained from the HS process, ensuring it closely matches the measured IF.

Eq. 9 is used to evaluate the uncertainty of the ACS preparation process using the HS method. The uncertainty of HS ACS comes from the ACS literature used $u_{\sigma_r(\lambda)}$, the uncertainty in determining IF $u_{IF_{convol}(\lambda),r}$, and the uncertainty due to the non-linear fitting process $u_{fit,r}$. The uncertainty from the literature is taken from

- 310 each literature as shown in Table 7, although in some cases, such as O_2 , the uncertainty value is not reported in detail, however, in the literature mentioned a pessimistic uncertainty value of 5%. Uncertainty due to IF is obtained from the standard deviation value of repeated measurements when making IF measurements repeated nine times. The result of this measurement is that the standard relative uncertainty in determining IF is 0.32%. Then this uncertainty value is propagated into the convolution and non-linear least squares fitting processes, so that each
- 315 uncertainty value due to IF is calculated for each gas species. Table 8 summarizes the uncertainty values for each component in the HS method. Finally, the uncertainty due to fitting error is taken from the standard deviation of the residual spectra in the NLS fitting process from region I to region III when determining the *a*, *b*, κ coefficients. The same method of determining fitting error was also used by Trisna et al (2023) and Yi Hongming et al (2018)

when performing NLS fitting for N₂O and CO₂ gases, respectively. Finally, after calculation, the total uncertainty 320 of HS ACS ranges from 0.25% (benzaldehyde) to 5.31% (oxygen 2).

 $u_{c,r}^2(\sigma_{HS}(\lambda)) = u_{\sigma_r(\lambda)}^2 + u_{IF_{convol},r}^2 + u_{fit,r}^2$

, (9)

3.4 Measurement of a gas mixture in the laboratory

To validate the PM ACSs, we measured the spectra of a VOC CRM mixture (cylinder D068005 in Table 1) in our laboratory and then fitted the PM ACSs to the measured spectrum. The PM ACSs used to build the database are 325 the same as those shown in Figure 4. Figure 7 presents the results of evaluating the gas mixtures in the lab. A

differential optical density (-σ′(λ)cL) was obtained after applying high-pass filtering, logarithmic transformation, and low-pass filtering to the $\frac{l(\lambda)}{l_0(\lambda)}$. We then fit the ACS to the mixture spectra using non-linear least-squares (NLS) fitting, applying the Levenberg and Marquardt method along with the steepest descent method (Press et al., 1988). The ACS of benzene, toluene, and o-xylene were then fitted subsequently with the mixture spectrum in the 330 wavelength interval between 240 and 280 nm in region II.

Table 9 summarizes the values obtained from the DOAS measurement alongside the corresponding CRM values and their standard uncertainties. The differences between the two range from 1.6% for toluene to 8.5% for benzene. The uncertainty in determining the concentration (*c*) in the optical density equation (*−σ′(λ)cL*) arises from three factors: the uncertainty of the PM ACS $u_{c,r}(\sigma_{PM}(\lambda))$, the path length $u_{L,r}$, and the fitting method $u_{fit,r}$.

335 Due to the limitations of the fitting process, minimizing the cost function still leaves a residual value (ideally zero). Therefore, following the general law of uncertainty propagation (BIPM, 2008), the relative total standard uncertainty can be expressed by the following equation:

$$
u_{c,r}^2(fv) = u_{c,r}^2(\sigma_{PM}(\lambda)) + u_{fit,r}^2 + u_{L,r}^2,
$$
\n(10)

The uncertainty from the PM ACSs was calculated using Eq. (8) from Section 3.2. The ACS uncertainty 340 contributes significantly to the DOAS measurements, making up 92.2% for o-xylene and 95.1% for both benzene and toluene. This shows that accurately measuring the ACS is crucial for reducing overall uncertainty. The path length contributes about 4.7% to the uncertainty for o-xylene and 4.9% for benzene and toluene. The fitting uncertainty is minimal, being almost negligible for benzene and toluene, and about 3.1% for o-xylene. This fitting uncertainty was determined by comparing optical density and peak-to-peak residual values (Platt and Stutz, 2008).

345 Table 9 shows that DOAS measurements, with an expanded uncertainty of 6-7%, deviate from CRM values by 1.6-8.5%, with the largest difference being 8.5% for benzene. These differences are likely due to adsorption or reactions on surfaces inside the gas cell and connecting lines. Specifically, benzene's sticky nature (Mitchell et al., 2017; Mori, 1987) can cause large discrepancies between DOAS measurements and CRMs. Additionally, before the actual DOAS measurement, we usually repeat the sample injection, vacuuming, and

350 nitrogen flushing two to three times. This process can naturally passivate the cell surface and connections, which may lead to more gas being present than indicated by the pressure, resulting in DOAS values that are higher than CRM values.

4 Conclusion

In this study, we successfully applied a hybrid calibration method for a long-path DOAS (LP-DOAS) system used 355 in atmospheric gas observation. This hybrid calibration involved two methods: direct precise measurement of ACSs, known as the PM method, and simulation of ACSs using a virtual instrument based on the IFs of the spectrometer, followed by a NLS fit, referred to as the HS method.

We successfully measured ACSs for benzene, toluene, styrene, sulfur dioxide, ortho-xylene, and paraxylene. The standard uncertainties for these measurements were 3.1% for ortho-xylene, para-xylene, and sulfur

- 360 dioxide, 3.0% for benzene and styrene, and 2.9% for toluene. The main sources of uncertainty in the ACS measurements were the light path length (1.4%), the concentration of calibration reference materials (≤0.74%), absorbance determination (≤1.5%), pressure (0.78%), and temperature (2.1%). For ACSs obtained via the HS method, the uncertainty was ≤5.31%, with the accuracy improved by dividing the convolution area into three separate parts.
- 365 To validate the reliability of the prepared ACSs, we measured a gas mixture of benzene, toluene, and ortho-xylene, prepared gravimetrically, using the spectrometer. The results, obtained through spectrum fitting, showed discrepancies between DOAS retrieval and CRM values of 8.5% for benzene, 1.6% for toluene, and 4.9% for ortho-xylene. The uncertainty components for the DOAS retrieval included the determination of ACSs $(\leq 3.1\%)$, the fitting procedure $(\leq 0.6\%)$, and the path length (1.4%). The determination of ACSs significantly
- 370 influenced the accuracy of DOAS measurements, contributing 92.2% to the uncertainty for ortho-xylene and 95.1% for benzene and toluene. This underscores the importance of accurate ACS measurement for reducing overall uncertainty in the DOAS measurement.

Controlled laboratory tests confirmed that the measured gas concentrations using the prepared ACSs were accurate within 15%. These tests also showed that the measurement values can be linked to SI units, ensuring 375 that data collected during measurement campaigns are reliable and trustworthy.

Acknowledgements

This work was funded by the Korea Meteorological Administration Research and Development Program under Grant (KMI2022-01511). The funding agencies had no role in study design; in the collection, analysis, and interpretation of data; in the writing of the report; and in the decision to submit the paper for publication.

380 **Data availability**

Data are available at Zenodo at 10.5281/zenodo.13677342 (Beni, 2024).

Conflict of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this article.

385 **Author contributions**

- BAT performed measurement, analyzed the data, and wrote the manuscript; SK and MP helped BAT during CRM preparation; YDK prepared gas CRM for ACSs measurement; SP provided technical and managerial supports;
- 390 JL wrote the manuscript, conceived, and supervise the experiment;

References

Beni, A. T., Kim, S., Kim, Y-D., Park, M., Park, SN., and Lee, J.: Preparation of hybrid calibrated absorption cross sections for a compact UV-DOAS measurement (data set), Zenodo, 10.5281/zenodo.13677342, 2024.

BIPM, I., IFCC, I., IUPAC, I., and ISO, O.: Evaluation of measurement data—guide for the expression of 395 uncertainty in measurement. JCGM 100: 2008, Citado en las, 167, 2008.

Etzkorn, T., Klotz, B., Sørensen, S., Patroescu, I. V., Barnes, I., Becker, K. H., and Platt, U.: Gas-phase absorption cross sections of 24 monocyclic aromatic hydrocarbons in the UV and IR spectral ranges, Atmospheric Environment, 33, 525-540, 1999.

Fally, S., Carleer, M., and Vandaele, A. C.: UV Fourier transform absorption cross sections of benzene, toluene, 400 meta-, ortho-, and para-xylene, Journal of Quantitative Spectroscopy and Radiative Transfer, 110, 766-782, 2009.

Grosch, A., Beushausen, V., Wackerbarth, H., Thiele, O., and Berg, T.: Temperature-and pressure-dependent midinfrared absorption cross sections of gaseous hydrocarbons, Applied optics, 49, 196-203, 2010.

GUIDE, I.: Guide 6142: 2015 (En)-Gas analysis-preparation of calibration gas mixtures-part 1: Gravimetric method for class I mixtures. 2015, Geneva, Switzerland. Available in:< https://www. iso. org/o. Accessed on, 13, 405 2016.

Hiraya, A. and Shobatake, K.: Direct absorption spectra of jet-cooled benzene in 130–260 nm, The Journal of chemical physics, 94, 7700-7706, 1991.

Jenkin, M. E. and Clemitshaw, K. C.: Ozone and other secondary photochemical pollutants: chemical processes governing their formation in the planetary boundary layer, Atmospheric Environment, 34, 2499-2527, 2000.

410 Johansson, J. K., Mellqvist, J., Samuelsson, J., Offerle, B., Lefer, B., Rappenglück, B., Flynn, J., and Yarwood, G.: Emission measurements of alkenes, alkanes, SO2, and NO2 from stationary sources in Southeast Texas over a 5 year period using SOF and mobile DOAS, Journal of Geophysical Research: Atmospheres, 119, 1973-1991, 2014.

Kang, J. H., Kim, Y. D., Lee, J., and Lee, S.: Development of primary reference gas mixtures of 18 volatile organic 415 compounds in hazardous air pollutants (5 nmol/mol level) and their analytical methods, Analytical Science and Technology, 34, 202-211, 2021.

Kim, M. E., Kang, J. H., Kim, Y. D., Lee, D. S., and Lee, S.: Development of accurate dimethyl sulphide primary standard gas mixtures at low nanomole per mole levels in high-pressure aluminium cylinders for ambient measurements, Metrologia, 55, 158, 2018.

420 Klingbeil, A. E., Jeffries, J. B., and Hanson, R. K.: Temperature-dependent mid-IR absorption spectra of gaseous hydrocarbons, Journal of Quantitative Spectroscopy and Radiative Transfer, 107, 407-420, 2007.

Lee, J. S. and Kim, Y. J.: Extinction measurement using a differential optical absorption spectrometer system, JOURNAL-KOREAN PHYSICAL SOCIETY, 42, 731-734, 2003.

Lee, J. S., Kuk, B. J., and Kim, Y. J.: Development of a differential optical absorption spectroscopy (DOAS) 425 system for the detection of atmospheric trace gas species; NO2, SO2, and O3, Journal of the Korean Physical Society, 41, 693-698, 2002.

Liu, J., Mauzerall, D. L., Chen, Q., Zhang, Q., Song, Y., Peng, W., Klimont, Z., Qiu, X., Zhang, S., and Hu, M.: Air pollutant emissions from Chinese households: A major and underappreciated ambient pollution source, Proceedings of the National Academy of Sciences, 113, 7756-7761, 2016.

430 McGee, T. J. and Burris Jr, J.: SO2 absorption cross sections in the near UV, Journal of Quantitative Spectroscopy and Radiative Transfer, 37, 165-182, 1987.

Meller, R. and Moortgat, G. K.: Temperature dependence of the absorption cross sections of formaldehyde between 223 and 323 K in the wavelength range 225–375 nm, Journal of Geophysical Research: Atmospheres, 105, 7089-7101, 2000.

435 Metrology, I. O. o. L.: Weights of classes E1, E2, F1, F2, M1, M1–2, M2, M2–3 and M3, Part 1: Metrological and technical requirements, 2004.

Milton, M., Woods, P., Jolliffe, B., Swann, N., and McIlveen, T.: Measurements of toluene and other aromatic hydrocarbons by differential-absorption lidar in the near-ultraviolet, Applied Physics B, 55, 41-45, 1992.

Mitchell, H. T., Smith, M. K., Blelloch, N. D., Van Citters, D. W., and Mirica, K. A.: Polycyclic aromatic 440 hydrocarbons as sublimable adhesives, Chemistry of Materials, 29, 2788-2793, 2017.

Mori, S.: Adsorption of benzene on the fresh steel surface formed by cutting under high vacuum, Applied surface science, 27, 401-410, 1987.

Palmer, C. and Loewen, E. G.: Diffraction grating handbook, 2005.

Pantos, E., Philis, J., and Bolovinos, A.: The extinction coefficient of benzene vapor in the region 4.6 to 36 eV, 445 Journal of Molecular Spectroscopy, 72, 36-43, 1978.

Platt, U. and Stutz, J.: Differential optical absorption spectroscopy, Physics of Earth and Space Environments, Springer Berlin Heidelberg, Berlin, Heidelberg, doi, 10, 978-973, 2008.

Press, W. H., Teukolsky, S. A., Vetterling, W. T., and Flannery, B. P.: Numerical recipes in C, 1988.

Shrivastava, A. and Gupta, V. B.: Methods for the determination of limit of detection and limit of quantitation of 450 the analytical methods, Chron. Young Sci, 2, 21-25, 2011.

Skorokhod, A. I., Berezina, E. V., Moiseenko, K. B., Elansky, N. F., and Belikov, I. B.: Benzene and toluene in the surface air of northern Eurasia from TROICA-12 campaign along the Trans-Siberian Railway, Atmospheric Chemistry and Physics, 17, 5501-5514, 2017.

Stutz, J., Kim, E., Platt, U., Bruno, P., Perrino, C., and Febo, A.: UV‐visible absorption cross sections of nitrous 455 acid, Journal of Geophysical Research: Atmospheres, 105, 14585-14592, 2000.

Suto, M., Wang, X., Shan, J., and Lee, L.: Quantitative photoabsorption and fluorescence spectroscopy of benzene, naphthalene, and some derivatives at 106–295 nm, Journal of Quantitative Spectroscopy and Radiative Transfer, 48, 79-89, 1992.

Trisna, B. A., Park, S., Park, I., Lee, J., and Lim, J. S.: Measurement report: Radiative efficiencies of (CF 3) 2 460 CFCN, CF 3 OCFCF 2, and CF 3 OCF 2 CF 3, Atmospheric Chemistry and Physics, 23, 4489-4500, 2023.

Trost, B., Stutz, J., and Platt, U.: UV-absorption cross sections of a series of monocyclic aromatic compounds, Atmospheric Environment, 31, 3999-4008, 1997.

Vandaele, A. C., Simon, P. C., Guilmot, J. M., Carleer, M., and Colin, R.: SO2 absorption cross section measurement in the UV using a Fourier transform spectrometer, Journal of Geophysical Research: Atmospheres, 465 99, 25599-25605, 1994.

Voigt, S., Orphal, J., Bogumil, K., and Burrows, J.: The temperature dependence (203–293 K) of the absorption cross sections of O3 in the 230–850 nm region measured by Fourier-transform spectroscopy, Journal of Photochemistry and Photobiology A: Chemistry, 143, 1-9, 2001.

Volkamer, R.: Absorption von sauerstoff im Herzberg I system und anwendung auf aromatenmessungen am 470 EUropean PHOto REactor (EUPHORE), Diploma Thesis, Institut fuer Umweltphysik, University of Heidelberg, 1996.

Volkamer, R.: A DOAS study on the oxidation mechanism of aromatic hydrocarbons under simulated atmospheric conditions, 2001.

Wei, X., Lyu, S., Yu, Y., Wang, Z., Liu, H., Pan, D., and Chen, J.: Phylloremediation of air pollutants: exploiting 475 the potential of plant leaves and leaf-associated microbes, Frontiers in plant science, 8, 1318, 2017.

Yi, H., Liu, Q., Gameson, L., Fleisher, A. J., and Hodges, J. T.: High-accuracy 12C16O2 line intensities in the 2 µm wavelength region measured by frequency-stabilized cavity ring-down spectroscopy, Journal of Quantitative Spectroscopy and Radiative Transfer, 206, 367-377, 2018.

485

490

Figures and Tables

495

Figure 1: Experimental setup of the ACSs measurements in the laboratory. A quartz cylinder tube of 50 cm in length was used to obtain the ACS.

505

510

515

Figure 2: Instrument function of a dispersive spectrometer at three wavelength regions obtained from Hg lamp (model: Pen-Ray 11SC, Analytik-jena): (a) region I: at the center wavelength $x_c = 248.08$ nm (FWHM = 0.39 nm); (b) region II: at the center wavelength $x_c = 265.28$ nm (FWHM = 0.35 nm); and (c) region III: at the center 520 wavelength $x_c = 302.33$ nm (FWHM = 0.28 nm).

525

540

Figure 3: Raw measured spectra for toluene, a representative gas. The black line shows the background spectrum *I0***, and the blue line represents the sample spectrum** *I***. The sample spectrum was taken at an ambient temperature of** 545 **295.8 K and a pressure of 224.86 Torr.**

550

555

Figure 4: The apparent UV absorption cross sections (ACSs) of benzene, toluene, styrene, sulfur dioxide, orthoxylene, and para-xylene, measured in the laboratory, are shown by the black line. The relative expanded 565 **uncertainties of the measurements are indicated by the upper and lower bounds, represented by the blue and red lines, respectively. These uncertainties are as follows: 6.2% for ortho-xylene and sulfur dioxide, 6.0% for benzene, 5.8% for toluene, 6.0% for styrene, and 6.2% for para-xylene, at a 95% confidence level with a coverage factor** $k = 2$ **.**

570

575

Figure 5: The comparison between ACSs obtained using the precise measurement (PM) method and the hybrid simulation (HS) method is shown. The reference ACSs σr(λ) are represented by the blue lines and were sourced from various literatures: benzene and para-xylene (T. Etzkorn et al., 1999) and SO₂ (Vandaele et al., 1994). After convoluting 585 **these reference spectra with the respective IFs for region I (benzene), region II (para-xylene), and region III (SO₂), a nonlinear least-squares (NLS) fit was applied to match the dispersion of pixels in the precise lab-measured ACSs σPM(λ), represented by the red lines. This process results in the final hybrid simulation ACSs σHS(λ), depicted by the black lines. The multiplicative factor from the least-squares fit addresses discrepancies arising from differences in spectrometer resolution, aligning the ACS values from the literature more closely with the measured ACSs.**

590

595

600

605

610

620

625

Figure 6: Absolute UV absorption cross sections (ACSs) of oxygen, ozone, ethylbenzene, phenol, meta-xylene, paracresol, formaldehyde, nitrogen dioxide, and benzaldehyde obtained from HS method. The colored line of the graphs indicates each instrument function used for convolution: red (248 nm), black (265 nm), and blue (302 nm).

630

640 **Figure 7: This example demonstrates the evaluation of spectra measured in the lab. Each gas spectrum was sequentially subtracted from the mixture, with the black line representing the mixture spectrum and the blue line showing the fitting result multiplied by the ACS. The remaining residual spectra are displayed at the bottom, with an** average residual value of 1.2×10^{-4} (peak-to-peak).

645

650

655

660

670 **Table 1 Summary of gas CRMs used in this study.**

675

685

690 **Table 2 Summary of target gas species for HS of ACSs considering different IFs in three regions.**

^aObtained from full width half maximum (FWHM) of mercury lines measurement

Table 3 Technical data from ACS referenced for simulation.

 a^aD = deuterium lamp, X = xenon arc lamp, G = grating, FT = Fourier-transform spectrometer, C:CCD detector, $PD = photodiode array, T: tungsten/halogen lamp$

700 **Table 4 Comparison of the ACSs and detection limits obtained in this work with data from the literature.**

The detection limits, as concentration detected at 1km of light path length, are computed 3 times noise of the spectrometer used.

^aThe differential cross section value after high-pass filtering.

^bThe reported pressures for all gas species are the total pressure (calibrated) with N_2 as the gas balance. The

705 resolutions of this work for ACS measurement of benzene are 0.39 nm; toluene, ortho- and para-xylenes are 0.35 nm; styrene and sulfur dioxide are 0.28 nm.

^cThe resolution reported by Etzkorn et al. (1999) is 0.146 nm.

 d The resolution reported by Trost et al. (1997) is 0.11 nm.

^eThe resolution reported by Vandaele et al. (1994) is 2 cm-1 .

Ē, ÷,

Table 5 Parameters of the NLS fit determined through comparing the convolved spectrum with the PM spectrum. These parameters were then used to optimize the dispersion of pixels for other simulated gas species not measured in the lab.

Wavelength shifting and squeezing were found by applying optimization process to the convoluted literature

725 ACS σ_r with the reference measured ACS $\sigma_{This work}$ using the following relation: σ_{HS} =

κxspline (λ_r (convolution), σ_r (convolution), $a + bx\lambda_{PM}$). Where a cost function $\chi^2 = \sigma_{PM} - \sigma_{HS}$ is minimized.

730

j,

Table 6 Summary of among the ACSs the obtained after applying HS (convolution and nonlinear least-squares fitting) and the measured ACS.

^aThe data was from convolution of ACS from (Etzkorn et al., 1999) with IF of the spectrometer

735 ^bThe data was from convolution of ACS from (Vandaele et al., 1994) with IF of the spectrometer

^cThe percentage differences between the HS ACS σ_{HS} and the PM ACS σ_{PM} were calculated using the relation = $|\sigma_{HS}-\sigma_{PM}|$ s ормі x100%
_{брм}

740

755 **Table 7 Summary of HS ACSs and their detection limits. Simulated by applying the NLS fit parameters to the convolved spectrum. Detection limits were calculated assuming a path length of 1,000 m.**

760 **Table 8 Uncertainty budget table for HS ACS proposed in this study.**

770

775 **Table 9 Comparison of concentration obtained from fitting results using the DOAS technique and the gravimetric technique.**

