



- 1 Field assessments on impact of CO₂ concentration fluctuations along with complex
- 2 terrain flows on the estimation of the net ecosystem exchange of temperate forests
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Abstract

The CO₂ storage (F_s) is the cumulation or depletion in CO₂ amount over a period 16 in an ecosystem. Along with the eddy-covariance flux and wind-stream advection of CO₂, it is a major term in the net ecosystem CO₂ exchange (NEE) equation and even 18 dominates in the equation under a stable atmospheric stratification while this equation 20 is used for forest ecosystems over complex terrains. However, estimating the Fs remains challenging due to the frequent gusts and random fluctuations in boundary-layer flows that arouse tremendous difficulties in catching the true trend of CO₂ changes for its 22 storage estimation from eddy-covariance along with the atmospheric profile techniques. 23 Using the measurements from Qingyuan Ker Towers equipped with NEE instrument systems separately covering mixed-broadleaf, oak, and larch forests towers in a 25 mountain watershed, this study investigates the gust periods and CO₂ fluctuation 26 magnitudes while examining their impact on F_s estimation in relation to the terrain complexity index (TCI). The gusts induce CO2 fluctuations at numerous periods of 1 to 10 min over two hours. Diurnal, seasonal, and spatial differences (P < 0.01) in the maximum amplitude of CO₂ fluctuations (A_{max}) ranges from 1.6 to 136.7 ppm and these 30 difference in a period (P_{max}) at the same significant level ranges 140 to 170 second. The 32 A_{max} and P_{max} are significantly correlated to the magnitude and random error of F_s with diurnal and seasonal differences. These correlations decrease as CO2 averaging time 33 34 windows becomes longer. To minimize the uncertainties of F_s, a constant [CO₂] 35 averaging time window for the F_s estimates is not ideal. Dynamic averaging time windows and a decision-level fusion model can reduce the potential underestimation of





- 37 F_s by 29%–33%, being equivalent to 1.9%–4.3% underestimation of the NEE for
- 38 temperate forests in complex terrains. The relative contribution of F_s to the 30-min NEE
- 39 observations ranged from 17% to 82% depending on wind speed and TCI. The study's
- 40 approach is notable as it incorporates TCI and utilizes three flux towers for replication,
- 41 making the findings relevant to similar regions with a single tower.
- 42 Keywords: Eddy covariance, complex terrain, carbon flux, storage term, carbon
- 43 dioxide concentration, random uncertainty

44 1 Introduction

- The accurate estimation of the net ecosystem exchange (NEE) of carbon dioxide
- 46 (CO₂) in forest ecosystems is crucial for a comprehensive understanding of the global
- 47 carbon cycle. The eddy covariance (EC) technique has been widely used in forest
- 48 ecosystems due to its capacity to directly measure the NEE while measurement
- 49 conditions satisfy the underlying theory. The EC technique is based on a simplified
- 50 mass conservation equation (after the Reynolds averaging), given by:

NEE =
$$\frac{1}{V_{m}} \int_{0}^{h} \left(\frac{\partial \overline{c}}{\partial t}\right) dz + \frac{1}{V_{m}} \left(\overline{w'c'}\right)_{h}$$

$$I \qquad \qquad II$$

$$+ \frac{1}{V_{m}} \int_{0}^{h} \left(\overline{w}(z) \frac{\partial \overline{c}}{\partial z} + \overline{c}(z) \frac{\partial \overline{w}}{\partial z}\right) dz, \qquad (1)$$

$$III a \qquad III b$$

$$+ \frac{1}{V_{m}} \int_{0}^{h} \left(\overline{u}(z) \frac{\partial \overline{c}}{\partial x} + \overline{v}(z) \frac{\partial \overline{c}}{\partial y}\right) dz$$

$$IV$$

- 52 where $V_{\rm m}$ is the volume of dry air in the control volume; c is the CO₂ mixing ratio; t is
- 53 the time; h is the measure height; u, v, and w denote the velocity components in the x,





y, and z directions, respectively; and an overbar denotes Reynolds averaging. This 54 55 equation conceptualizes the NEE within a control volume from the ground to the measurement height (h), while ignoring the horizontal turbulence term divergence 56 (Feigenwinter et al., 2004). In this equation, term I is the CO₂ storage (F_s) representing 57 58 the change in the average CO₂ concentration (hereafter [CO₂]). Terms II, IIIa, IIIb, and IV represent the vertical turbulent flux (Fc), the vertical advection, the interface vertical 59 60 mass advection, such as the evaporation process (Webb et al., 1980), and the horizontal 61 advection, respectively. 62 Most flux measurements typically lack the solutions for terms III and IV, and can only estimate the NEE by summing Fc and Fs, and even a significant number of sites 63 ignored the F_s. The F_s in the vertical gas column within a canopy can be substantial, 64 requiring attention in NEE estimates (Aubinet et al., 2000). The F_s contributes ~60% to 65 nocturnal turbulent flux underestimation in forest ecosystems with "ideal" topography 66 (Mchugh et al., 2017). Especially, during atmospherically stable periods such as the 67 early morning, sunset, and nighttime transitions, the Fs has a significant impact on the 68 69 NEE. For 30-min and annual forest ecosystem carbon flux measurements, ignoring Fs would underestimate the NEE (Zhang et al., 2010). The Fs value typically ranges from 70 -2 to -5 µmol m⁻² s⁻¹ in the early morning, and the F_s is about 1–3 µmol m⁻² s⁻¹ after 71 sunset for temperate forests. Neglecting the F_s value can also lead to a misunderstanding 72 73 of the CO₂ exchange processes, such as ecosystem respiration and photosynthesis, and their relationship with key control factors such as solar radiation, temperature, and 74 moisture (Mchugh et al., 2017). Therefore, it is imperative not to overlook F_s to ensure 75

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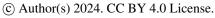
carbon flux of forest ecosystems in complex terrains or with heterogeneous underlying 78 surfaces remains an area of great interest. Topography complexity plays a complex role 79 80 in the transportation of momentum, energy, and mass in the atmospheric boundary layer, with direct impacts on the airflow patterns, spatiotemporal characteristics, and gas 81 82 concentration fluctuations (Sha et al., 2021; Finnigan et al., 2020). Differences in 83 airflow along the slope, lateral CO₂ discharge downhill, and spatiotemporal variations 84 in soil respiration result in the CO₂ outflow from slopes and valleys lagging behind the flat top of the mountain (De Araújo et al., 2010). At night, under stable atmospheric 85 stratification, cold air moves from the ground to the valley forest canopy and then flows 86 87 to low-lying areas, causing a "carbon pooling" effect. The gradient of [CO₂] below the 88 EC sensors fluctuates significantly, and the cold air discharge above the canopy reduces CO₂ storage, leading to an underestimation of forest ecosystem respiration (Yao et al., 89 2011; De Araújo et al., 2008; De Araújo et al., 2010). 90

more precise NEE estimates of forest ecosystems, particularly in complex terrains.

Despite the challenges inherent in monitoring forest conditions, understanding the

[CO₂] of the control volume at the beginning and the end of the EC averaging period (30 min or 1 h) and dividing by the EC averaging period (Finnigan, 2006). In practice, the F_s represents the integration of the time derivative of the vertically determined column-averaged [CO₂]. However, relying solely on tower-top measurements can lead to underestimation of F_s by up to 34% compared to the eight-level profile approach (Gu et al., 2012). The NEE magnitude with the F_s based on the two-min [CO₂] averaging

According to the theoretical definition, Fs estimates are derived by averaging the



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time window (instantaneous concentration approach) was found to be 5% higher than that of the 30-min-window-based F_s (averaging concentration approach), particularly during nighttime in the growing season (Wang et al., 2016). The effect of the F_s on the NEE of forest ecosystems increases with the increase of timescale, and the annual sum of the NEE obtained using the instantaneous concentration approach is higher than that obtained by averaging concentrations (Li et al., 2020). Most research has examined how vertical and horizontal gas concentration sampling point distribution affects the uncertainty in F_s estimation (Bjorkegren et al., 2015; Wang et al., 2016; Yang et al., 2007; Yang et al., 1999), with a small number of studies examining the effect of [CO₂] sampling frequency on the F_s (Finnigan, 2006; Heinesch et al., 2007). Certain studies have experimentally validated new concepts, such as correlating the gas sampling point concentration with the horizontal distribution (Nicolini et al., 2018). Some studies have approached the true value theoretically, such as through defining the control volume represented by flux measurements (Metzger, 2018; Xu et al., 2019). However, the number of complete column samples required to describe the column-averaged [CO₂] of each 30-min or 1-h F_s estimate is still undetermined. Previous studies have emphasized the significance of the Fs to the NEE and the influence of [CO₂] dynamics on F_s estimates in complex terrains. To overcome any disparities between sensors and obtain precise changes in the [CO₂] gradient above and below the forest canopy, individual gas analyzers are extensively utilized to measure [CO₂] levels vertically (Siebicke et al., 2011). However, a single gas analyzer introduces time delays when monitoring multi-point [CO₂] curves. Accurately determining the F_s



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measurements (Wang et al., 2016). The random error of the F_s estimates using one complete column sample is considerably high due to short-term [CO₂] fluctuations (Nicolini et al., 2018). The calculation of the F_s using time-averaged [CO₂] profiling leads to significant information loss at high frequency, resulting in a substantial underestimation bias. Furthermore, resource constraints in the measurement system, coupled with a lack of clear guidelines for estimating the F_s values and their associated uncertainties, create a significant gap between ideal initiatives and their implementation. These issues necessitate further efforts to characterize [CO₂] fluctuations across different sites and demonstrate the mechanisms influencing F_s magnitudes, uncertainties, and their contributions to NEE observations in complex terrains. Thus, this manuscript aims to bridge this gap by introducing a statistical method to estimate F_s values and their uncertainties. This paper employed an innovative EC site with three flux towers (Qingyuan-Ker Towers) to monitor three typical types of temperate forest stands located in complex terrains in northeastern China. This study introduces a decision-level fusion model based on weighing the underestimation bias and random error of the F_s to obtain more accurate results. The objectives of this study were to: 1) compare diurnal, seasonal, and spatial differences in [CO₂] fluctuations, F_s, and its uncertainty; 2) examine the variation in F_s uncertainty with different [CO₂] averaging time windows; and 3) investigate the response of F_s and its uncertainty to [CO₂] fluctuations, wind above the canopy, and terrain complexity, and quantify the impact of the Fs on the NEE estimates

estimates can be challenging due to the spatial and temporal resolution of [CO₂]

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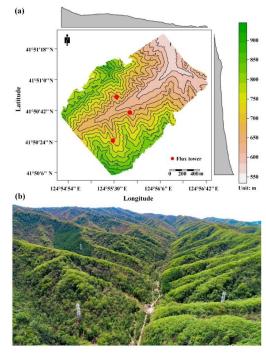


under these conditions.

2 Materials and methods

2.1 Study site and instrumental set-up

This study was conducted in temperate forests in a watershed based on the Ker towers (Zhu et al., 2021; Gao et al., 2020), situated in northeast China (41°50′N, 124°56′E). The region experiences a temperate continental monsoon climate, with an average annual temperature of 4.3 °C and annual rainfall of 758 mm from 2010 to 2021 (Li et al., 2023). The Ker towers consist of three 50-m-high EC towers (Fig. 1) that observe a mixed broadleaved forest (MBF), a Mongolian oak forest (MOF), and a Larch plantation forest (LPF).



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Fig. 1 Overview of the study area. The first map (a) depicts the topography of the study site, with





black curves indicating elevation contours, and marginal distributions represented as a gray graph, averaged over rows and columns. The second image (b) features an aerial photograph of the Qingyuan-Ker towers captured in the growing season (Gao et al., 2020).

The basic information regarding Ker towers in this study is presented in Table 1.

The closed-path EC system (EC310, Campbell Scientific Ltd., Logan, UT, USA), comprising a CSAT3 sonic anemometer and an EC155 closed-path infrared ray gas analyzer (IRGA), was employed to monitor the three-dimensional wind speed and CO₂/H₂O concentrations (10 Hz). The atmospheric profiling system (AP200, Campbell Scientific Ltd., Logan, UT, USA) was utilized to measure the CO₂/H₂O concentrations with eight height levels. Each level was measured for 15 s (with 10 s for the flushing of the manifold and 5 s for logging the average), leading to a measurement cycle of 2 min.

Table 1 Basic information of Ker towers

Forest	Mixed broad-	Mongolian oak	Larch plantation	
	leaved			
Experiment period	Jan 01, 2020-	Jan 01, 2020-	Jan 01, 2020-	
	Dec 31, 2021	Dec 31, 2021	Dec 31, 2021	
Elevation (m)	634	669	721	
Slope (°)	14.8 ± 2.1	19.1 ± 2.9	16.2 ± 5.3	
Canopy height (m)	21.5 ± 1.8	13.9 ± 0.6	19.5 ± 0.6	
Leaf area indices	3.0 ± 0.5	3.1 ± 0.8	3.9 ± 0.6	
Eddy covariance system	EC310	EC310	EC310	
Eddy covariance sensor	46	46	36	
height (m)				
Atmospheric profiling	AP200	AP200	AP200	
system				
Profile heights (m)	0.5, 2, 6, 11, 16,	0.5, 2, 6, 11, 16,	0.5, 2, 6, 11, 16,	
	21, 26, 36	21, 26, 36	21, 26, 36	





2.2 Calculation of storage flux

Averaging the [CO₂] in a time window was utilized to calculate the F_s values, in 168 addition to data on the air pressure, CO₂/H₂O molar fractions, and air temperature at 169 170 different heights above the ground surface (Finnigan, 2006; Montagnani et al., 2018; Xu et al., 2019). The molar mixing ratio and mass mixing ratio are conserved quantities 171 172 with the variation of air temperature, air pressure, and water vapor concentration, whereas the molar fraction is not. This study determined the F_s using the molar mixing 173 ratio obtained from CO₂/H₂O molar fraction observations, applying the ideal gas law 174 and Dolton's partial pressure law (Montagnani et al., 2009). The water vapor molar 175 mixing ratio (χ_v) in mmol mol⁻¹ is given by 176

$$\chi_v = \frac{c_v}{1 - c_v \times 10^{-3}},\tag{2}$$

- where c_v is the water vapor molar fraction in mmol mol⁻¹, and the CO₂ molar mixing
- 178 ratio (χ_c) in μ mol mol⁻¹ is given by

$$\chi_c = \frac{c_c}{1 - c_v \times 10^{-3}},\tag{3}$$

- where c_c is the CO₂ molar fraction in μ mol mol⁻¹.
- The dry air density $(\bar{\rho}_d)$ in mol m⁻³ is calculated as follows:

$$\bar{\rho}_d = \frac{\bar{P}}{(\bar{T} + 273.15) \times (R^* + \chi_v \times 10^{-3} \cdot R^* \cdot M_d/M_v)},$$
(4)

where R^* is the air gas constant (8.31441 Pa m³ K⁻¹ mol⁻¹), \bar{P} is the air pressure in Pa, and \bar{T} is the average air temperature in Celsius. M_d and M_v are the dry air and water vapor molar mass (18.015 g mol⁻¹), respectively. M_d is calculated from the CO₂ molar mixing ratio (Khélifa et al., 2007):





$$M_d = 28.9635 + M_c \cdot (\chi_c \times 10^{-6} - 0.0004),$$
 (5)

where M_c is the CO₂ molar mass (12.011 g mol⁻¹).

The F_s estimated from eight-level profiles are calculated as follows:

$$F_{s} = \bar{\rho}_{d} \int_{0}^{h} \frac{d\bar{\chi}_{c}}{dt} dz \doteq \bar{\rho}_{d} \sum_{i=1}^{8} \frac{\Delta \bar{\chi}_{c_{i}} \Delta h_{i}}{\Delta t}, \tag{6}$$

where $\bar{\chi}_c$ is the average CO₂ molar mixing ratio and Δh_i is the height represented by

each level. To ensure that F_s corresponds to F_c in time, the average [CO₂] at the start or

end moments (t) during a time window (τ min) is calculated as follows:

$$\bar{\chi}_{c_i} = \frac{2}{\tau} \sum_{t - \frac{\tau}{2} < t \le t + \frac{\tau}{2}} \chi_{c_i}(t). \tag{7}$$

190 2.3 Data analysis

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To evaluate the impact of [CO₂] fluctuations on F_s measurements and its corresponding uncertainty, empirical modal decomposition (EMD) and Fourier spectrum analysis were used to extract the period and amplitude of fluctuations in the high-frequency [CO₂] time series (10 Hz). EMD was used to decompose the [CO₂] time series into intrinsic mode functions based on local signal properties, which yield instantaneous frequencies as functions of time, allowing for the identification of embedded structures of eddies. EMD is applicable to non-linear and non-stationary processes (Huang et al., 1998). The period and amplitude of [CO₂] fluctuations above the forest canopies reflected the eddy size. Subsequently, the maximum period and amplitude of [CO₂] fluctuations in a short term (2h) was indicative of large eddies under the influence of gust.

Due to the diurnal and seasonal variability of flux measurements, this study





- 203 defined the transition period and growing season. The solar elevation angle was used
- to define the transition period as 1-h before sunrise (sunset) to 2-h after sunrise (sunset).
- 205 The growing degree days (GDDs) were calculated using the base temperature (Tbase) to
- determine the beginning and end of the growing season, and the formula was as follows
- 207 (Mcmaster and Wilhelm, 1997):

$$GDD = \frac{1}{2}(T_{max} + T_{min}) - T_{base}, \tag{8}$$

- 208 where T_{base} is 6°C. Considering the persistent demand of temperature to support
- 209 vegetation growth, the fourth day of the first GDD greater than zero (less than zero)
- over a span of five consecutive days was defined as the starting (ending) time of the
- 211 growing season.
- The main data processing and analysis steps are outlined below:
- 213 1. EMD and Fourier spectrum analysis of [CO₂] high-frequency time series were
- used to extract the maximum amplitude (A_{max}) and corresponding period (P_{max}) of [CO₂]
- fluctuations every 2 h. The data were divided into two subsets based on P_{max} , with a
- 216 cut-off of 150 s.
- 2. CO₂ storage fluxes were calculated for different [CO₂] average time windows
- 218 (τ) , ranging from 4 to 28 min.
- 219 3. The standardized major axis (SMA) regression model (Warton et al., 2012) was
- used to compare the slope differences (bias) between $F_{s_{-}\tau}$ and $F_{s_{-}28}$ for different P_{max}
- 221 and the forest stands. The SMA model offers routines for comparing parameters a and
- b among groups for symmetric problems.
- 4. The normalized root mean square error (NRMSE) and slope were used to





- evaluate the relative error and bias between F_{s} τ and F_{s} 28. The NRMSE is calculated as
- 225 follows:

$$NRMSE = 100 \times \sqrt{\frac{\sum_{i=1}^{N} (F_{s_{-}\tau}^{(i)} - F_{s_{-}28}^{(i)})^{2}}{\sum_{i=1}^{N} (F_{s_{-}28}^{(i)} - \overline{F_{s_{-}28}})^{2}}}$$
(9)

- 5. The normalized weighting coefficient (w) of $F_{s_{-\tau}}$ was estimated based on the
- NRMSE and slope (Wang et al., 2020). The details are shown in Appendix A1. Then,
- 228 using the decision-level fusion model, F_{s_comb} was calculated as follows:

$$F_{s\ comb} = w_1^* \cdot F_{s\ 4} + w_2^* \cdot F_{s\ 8} + \dots + w_7^* \cdot F_{s\ 28}$$
 (10)

- 229 The decision-level fusion model automatically assigned weights to the F_s based on
- 230 different [CO₂] averaging time windows. Its purpose in this study was to balance the
- 231 relative error and bias of F_s estimates caused by [CO₂] sampling. The analysis was
- 232 performed using the EMD and smatr R packages (Warton et al., 2012; Huang et al.,
- 233 1998).
- 234 2.4 Uncertainty analysis
- To improve the accuracy of estimating the uncertainty of F_s using individual tower,
- 236 this work has made modifications to the 24-h difference method by extending the
- 237 sampling time windows and applying meteorological condition constraints (Hollinger
- and Richardson, 2005). This method trades time for space to estimate the uncertainty
- 239 of F_s. To determine the uncertainty of F_s, this study compared the observations at
- 240 moment *i* to the average of several observations during a similar period and with similar
- 241 meteorological conditions. The specific steps were as follows:
- 1. The average F_s was calculated in a certain time window (15 d) for the moment

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- interval (i–0.5 h, i+0.5 h) where moment i was located and where the meteorological conditions (such as the u*, air temperature, and sensible heat flux) were similar.
- 245 2. The difference between the F_s value corresponding to each moment i and \overline{F}_s 246 was calculated separately to obtain the residual sequence ε_s .
- 3. The standard deviation $\sigma(\varepsilon_s)$ related to ε_s for F_s was calculated in a certain time window (15 d) for the moment interval (i–0.5 h, i+0.5 h) where moment i is located and where the meteorological conditions (such as the u*, air temperature, and sensible heat flux) were similar.
- After estimating the uncertainty of F_s , this study extended the work conducted by Richardson et al. (2008) to analyze its relationship with the magnitude of flux measurements ($|F_s|$), [CO₂] fluctuations (A_{max} and P_{max}), wind speed (WS), and terrain complexity index (TCI). A comprehensible description of the TCI can be found in Appendix A2. This relationship can be approximated by using the following equation:

$$\sigma(\varepsilon_s) = \beta_0 + \sum_{i=1} \beta_i \cdot x_i, \tag{11}$$

where the nonzero intercept term β_0 indicates the size of the random uncertainty as x_i approaches 0, which varies with the observation site, with larger value of β_0 indicating greater uncertainty. The slope term β_i indicates the sensitivity of the size of the random uncertainty of x_i , with smaller β_i values indicating a probability distribution of uncertainty closer to white noise.





3 Results

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3.1 Characterization of [CO₂] fluctuation and F_s variations

The [CO₂] high-frequency time series above the forest canopies were decomposed 263 using EMD, followed by spectral analysis to extract the fluctuation period and 264 amplitude of [CO₂] at different time scales. As depicted in Fig. 2, it became evident that 265 the [CO₂] above the canopies displayed short-term fluctuations with periods ranging 266 from 1 to 10 min, and the amplitude of these fluctuations showed an increasing trend 267 268 with longer periods. This observation strongly suggested the presence of large eddies influenced by gusts above the canopies, and these eddies were responsible for the 269 increasing amplitude of [CO₂] fluctuations as their size increased. 270 271 To examine the spatio-temporal variations in large eddies, this study compared the A_{max} and P_{max} values above canopies across different forest stands. The analysis utilized 272 273 data from daytime, nighttime, and transition periods in both the growing and dormant seasons. The averages of A_{max} and P_{max} averages for the above-canopy [CO₂] in the 274 275 three forest stands ranged from 1.588 to 136.667 ppm and from 2.313 to 2.784 min, respectively (Table 2). Fig. 3 demonstrated significant seasonal and diurnal differences 276 (P < 0.01) in P_{max} , with higher values during daytime in the growing season, and lower 277 278 values during the daytime in the dormant season. Moreover, P_{max} was significantly 279 different (P < 0.01) among different forest stands during the same time period, with 280 MBF stand having the highest values, followed by the MOF, and the lowest values in the LPF. During the growing season, the A_{max} values were significantly higher than 281

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those during the dormant season, with both daytime and nighttime values also exhibiting significant differences (P < 0.01) among different forest stands. This observation provided evidence of significant spatio-temporal variability in large eddies influenced by gusts.

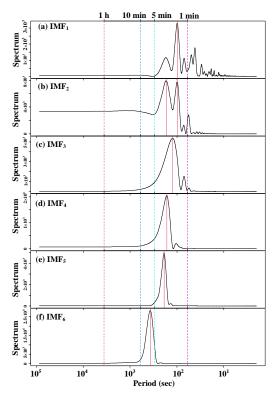


Fig. 2 Power spectral density of the intrinsic mode function (IMF) of above-canopy CO₂ concentrations in the Mongolian oak forest on July 2, 2020 (24 h).

Table 2 Mean of the A_{max} and P_{max} in different forest stands at different periods

37 ' 11	Tower	Growing season				Dormant season		
Variable	site	DT^1	NT^2	TP^3		DT	NT	TP
$A_{\rm max}^4$	MBF^6	57.932	139.667			2.219	5.212	4.944
	MOF^7	36.160	57.945	55.777		2.699	5.175	4.637
(ppm)	LPF^8	52.688	58.816	60.147		1.588	2.985	2.456
$P_{\rm max}^{5}$	MBF	154.563	167.024	164.824	1	158.449	151.428	158.121

amplitude of $F_{\rm s}$.

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(s)	MOF	151.986	160.633	159.146	153.091	147.491	153.274			
	LPF	149.003	143.950	145.696	143.458	138.794	142.009			
1 DT represents daytime; 2 NT represents nighttime; 3 TP represents transition period. 4 $A_{\rm max}$										
represents	represents the maximum amplitude of short-term CO_2 concentration fluctuations; ⁵ P_{max} represents									
the corresp	onding per	riod of maxii	mum amplit	ude. 6 MBF re	epresents mix	ed broad-lea	ved forest; 7			
MOF repre	esents Mong	golian oak fo	orest; 8 LPF 1	represents Lar	ch plantation	forest.				
То е	estimate tl	he uncertai	nty of Fs	using an inc	dividual tov	ver, a com	prehensive			
analysis o	of its diur	nal and se	asonal dyn	amics, as w	rell as the fi	unctional r	elationship			
between 1	F_s and u^* ,	was neces	sary. Fig. 4	4 presented	significant o	liurnal var	iations and			
seasonal o	difference	s in F _s acro	ss the three	e forest stand	ls. During th	e growing	season, the			
median d	median diurnal variation of F_{s} for the three forest stands ranged from -2.960 to 2.647									
$\mu mol\ m^{-2}\ s^{-1},$ whereas during the dormant season, it ranged from -1.306 to $1.012\ \mu mol$										
$\mbox{m}^{-2} \mbox{ s}^{-1}.$ Comparing the amplitude of F_s diurnal variation among the three forest stands,										
MBF exh	MBF exhibited the largest amplitude during the growing season, while the amplitudes									
of the th	ree forest	stands we	re similar	during the	dormant sea	ison. Nota	bly, it was			
observed	that the ar	mplitudes f	for longer I	P _{max} values v	vere greater	than those	for shorter			
P _{max} valu	es. This	observation	n indicated	l that the la	arger the ed	ldies, the	greater the			

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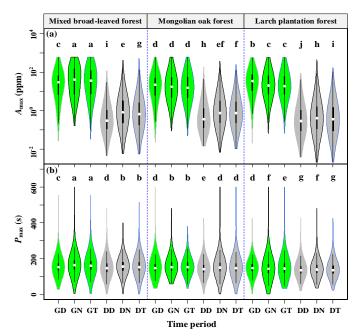


Fig. 3 Maximum amplitude (A_{max}) (a) and corresponding period (P_{max}) (b) of short-term CO₂ concentration fluctuations in different forest stands for seasonal and diurnal variations, where GD, GN, GT, DD, DN, and DT denote the growing season daytime, growing season nighttime, growing season transition period, dormant season daytime, dormant season nighttime, and dormant season transition period, respectively. Columns with different lowercase letters are significantly different (P < 0.05) according to Fisher's least significant difference test.

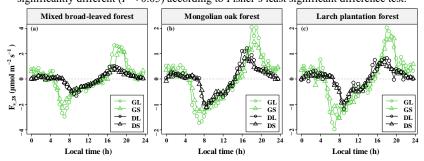


Fig. 4 Median diurnal variation of CO_2 storage flux (F_s) based on 28-min CO_2 concentration averaging time windows in the three forest stands during different seasons. GS indicates the growing season and a short period of maximum amplitude (P_{max}), GL indicates the growing season and a long P_{max} , DS indicates the dormant season and a short P_{max} , and DL indicates the dormant season and a long P_{max} .

Furthermore, a u^* threshold value was identified for the variation of F_s with u^*





during daytime in both the dormant and growing seasons (Fig. 5). When u^* fell below the u^* threshold, the magnitude of F_s ($|F_s|$) decreased with increasing u^* . Conversely, when u^* exceeded the u^* threshold, the $|F_s|$ tended to remain relatively constant. Notably, a maximum point for the $|F_s|$ was observed when the u^* was less than 0.5 m/s during the growing season, whereas not during the dormant season. This phenomenon was particularly evident during the nighttime and transition periods of the growing season, where $|F_s|$ exhibited an initial increase followed by a subsequent decrease with u^* . These observations strongly indicated that the effect of the turbulent mixing strength on the $|F_s|$ over complex terrains was nonlinear and exhibited diurnal and seasonal differences.

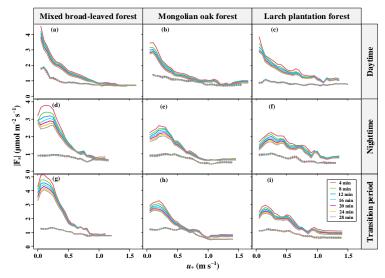


Fig. 5 Magnitudes of CO₂ storage flux (|F_s|) determined with different CO₂ concentration average time windows as a function of the friction velocity (*u**) and moving block averages from all 30-min data for the years 2020-2021. Dashed and solid lines donate the dormant and growing seasons, respectively.

3.2 Effect of [CO₂] fluctuations on the F_s and its uncertainty

To investigate the influence of the [CO₂] fluctuation periods on the error of F_s



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measurement, this study computed the diurnal average of the standard deviation $\sigma(\varepsilon_s)$ of the 30-min F_s uncertainty (ε_s) separately for different P_{max} values and the seasons. The overall distribution of ε_s showed a non-normal distribution with a high peak (kurtosis > 2 and P < 0.05, results presented in Supplementary Table 1–4). The daily variation curves of $\sigma(\varepsilon_s)$ at various [CO₂] averaging time windows are presented in Fig. 6. It was observed that the diurnal variation range of $\sigma(\varepsilon_s)$ was higher during the growing season compared to the dormant season, regardless of the P_{max} lengths, indicating a seasonal difference independent of the P_{max} . Additionally, during the growing season, both MBF and MOF demonstrated evident diurnal variation in $\sigma(\varepsilon_s)$, with the peak occurring at night and the trough during the daytime. The diurnal variation range of $\sigma(\varepsilon_s)$ varied across the three forest stands, with MBF exhibiting the largest amplitude. Furthermore, a significantly positive correlation was observed between $\sigma(\varepsilon_s)$ the $|F_s|$ (P < 0.01), with site, seasonal, and diurnal differences (Fig. 7). The relationship between these variables was characterized by intercepts and slopes that varied across different [CO₂] averaging time windows, ranging from 1.99 to 2.82 and from 0.24 to 0.28, respectively (results presented in the Supplementary Tables 5-6). Both decreased as the [CO₂] averaging time window increased, with the growing season exhibiting larger values compared to the dormant season (results shown in the Supplementary Tables 5–6). These findings suggested that increasing the [CO₂] averaging time window, results in a reduction of the random error in F_s, approaching a behavior similar to white noise.





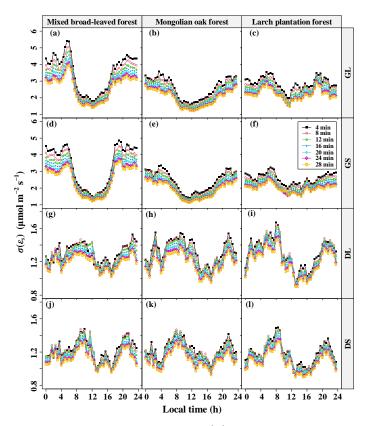


Fig. 6 Diurnal variations in the random uncertainty $(\sigma(\varepsilon_s))$ of CO₂ storage flux (F_s) errors (ε_s) at different CO₂ concentration ([CO₂]) averaging time windows and their seasonal differences, where GS indicates the growing season and a short period of maximum amplitude (P_{max}) of [CO₂] fluctuations, GL indicates the growing season and a long P_{max} , DS indicates the dormant season and a short P_{max} , and DL indicates the dormant season and a long P_{max} .

To assess the impact of [CO₂] fluctuations on the error and bias of F_s measurement, this study compared the NRMSE and slopes of F_s based on different [CO₂] averaging time windows, with reference to the baseline F_{s_2} 8, across various P_{max} values, time periods, and sites. As shown in Fig. 8, the NRMSE decreased and approached convergence as the [CO₂] averaging time windows increased. During both daytime and nighttime in the growing season, the NRMSE corresponding to longer P_{max} was greater than that corresponding to shorter P_{max} , while the opposite trend was observed during

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- 371 the dormant season. Additionally, the longer the [CO₂] averaging time window, the
- 372 greater the relative underestimation of F_s.

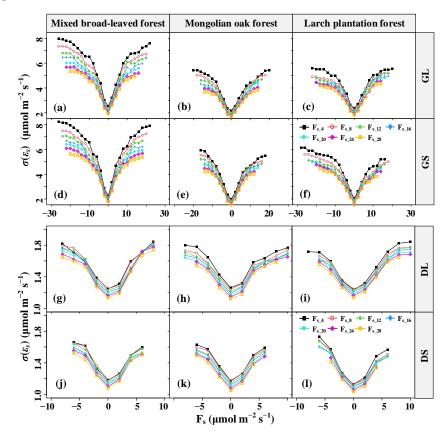


Fig. 7 Random uncertainty $\sigma(\varepsilon_s)$ of CO₂ storage flux (F_s) errors (ε_s) at different CO₂ concentration ([CO₂]) averaging time windows as a function of the F_s magnitude for mixed broadleaved forest, Mongolian oak forest, and Larch plantation forest during the growing and dormant seasons. GS indicates the growing season and a short period of maximum amplitude (P_{max}) of [CO₂] fluctuations, GL indicates the growing season and a long P_{max} , DS indicates the dormant

season and a short P_{max} , and DL indicates the dormant season and a long P_{max} .



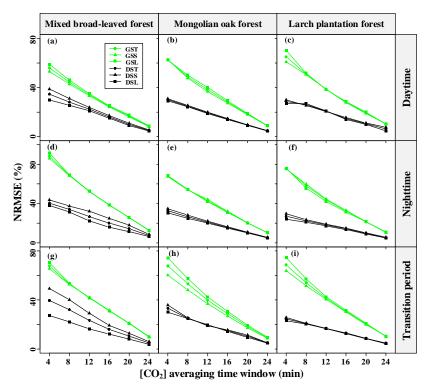


Fig. 8 Seasonal and diurnal differences in the normalized root mean square error (NRMSE) of CO_2 storage flux (F_s) versus the respective F_{s_28} values for different CO_2 concentration ([CO_2]) averaging time windows. GST indicates the growing season and does not distinguish the period of maximum amplitude (P_{max}) of [CO_2] fluctuations, GSS indicates the growing season and a short P_{max} , GSL indicates the growing season and a long P_{max} , DST indicates the dormant season and does not distinguish P_{max} , DSS indicates the dormant season and a short P_{max} , and DSL indicates the dormant season and a long P_{max} .

The comparison of slopes between F_{s_4} and F_{s_28} in the three forest stands revealed interesting patterns, as depicted in Fig. 9. During the growing season, the slopes corresponding to the shorter P_{max} of [CO₂] fluctuations were consistently lower than those for the longer P_{max} , indicating that the effect of P_{max} on F_{s} uncertainty decreased with increasing [CO₂] averaging time windows. However, for the MBF stand (Fig. 9d and Fig. 9g), the slopes corresponding to the shorter P_{max} of [CO₂] fluctuations during the dormant season nighttime were actually greater than those for the longer P_{max} ,

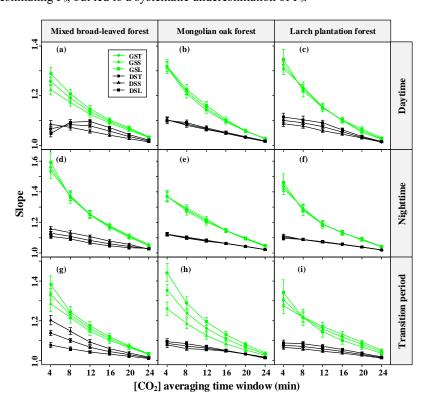
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primarily due to diurnal variations in the daily dynamics of F_s . Overall, the influence of P_{max} on F_s uncertainty decreased with increasing [CO₂] averaging time windows. This suggested that averaging [CO₂] reduced the effect of gusts on the random uncertainty in estimating F_s , but led to a systematic underestimation of F_s .



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Fig. 9 Seasonal and diurnal differences in the slope of CO_2 storage flux (F_s) versus the F_{s_28} for the different CO_2 concentration ([CO_2]) averaging time windows. GST indicates the growing season and does not distinguish the period of maximum amplitude (P_{max}) cases, GSS indicates the growing season and a short P_{max} , GSL indicates the growing season and a long P_{max} , DST indicates the dormant season and does not distinguish P_{max} , DSS indicates the dormant season and a short P_{max} , and DSL indicates the dormant season and a long P_{max} .

To analyze the effect of [CO_2] fluctuations on $|F_s|$ in complex terrains, this study

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developed a multiple linear regression model, considering the interaction effects of

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wind speed and terrain complexity on |Fs|, as shown in Fig. 10. Amax exhibited a





significant positive correlation with $|F_s|$ in all time periods (P < 0.05). Conversely, P_{max} showed a significant negative correlation with $|F_s|$ during the dormant season daytime, the growing season daytime, and the transition periods (P < 0.05). Additionally, their correlation coefficient decreased with increasing τ . In Fig. 10d and Fig. 10e, a u^* threshold was observed during the growing season nighttime. When the u^* was below the threshold, higher TCI values resulted in smaller $|F_s|$; whereas when the u^* was above the threshold, higher TCI values led to larger $|F_s|$. During the growing season nighttime and transition periods, u^* showed a significant negative correlation (P < 0.05) with $|F_s|$, and the correlation coefficient decreased with increasing TCI values.

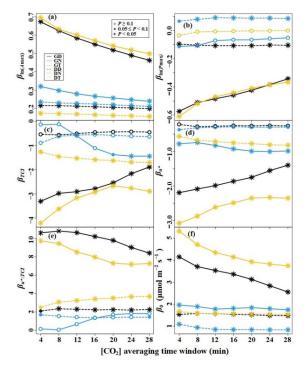
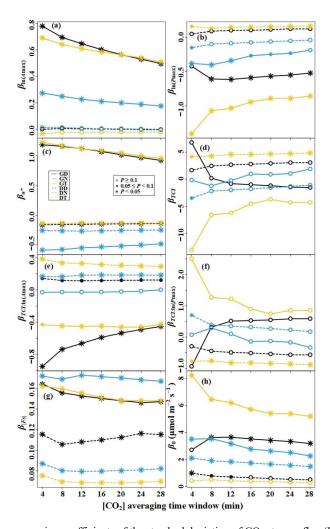


Fig. 10 Linear regression coefficients of the CO₂ storage flux (F_s) magnitude—driving factors relationships for the seven CO₂ concentration ([CO₂]) averaging time windows. u^* : friction velocity; TCI: terrain complexity index; A_{max} : maximum amplitude of [CO₂] fluctuations; P_{max} : corresponding period of maximum amplitude.





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Fig. 11 Linear regression coefficients of the standard deviation of CO_2 storage flux (F_s)—driving factors relationships determined with Eq. (11) for the seven CO_2 concentration ([CO_2]) averaging time windows. u^* : friction velocity; TCI: terrain complexity index; A_{max} : maximum amplitude of [CO_2] fluctuations; P_{max} : corresponding period of maximum amplitude.

As evident from Fig. 11a and Fig. 11e, the A_{max} exhibited a significant positive correlation (P < 0.05) with $\sigma(\varepsilon_s)$ during both the dormant season's nighttime and the growing season. Throughout the transition period of the growing season, P_{max} displayed a significant negative correlation with $\sigma(\varepsilon_s)$ (P < 0.05). During the transition period of the dormant season, a TCI threshold was observed, with P_{max} showing a significant

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a significantly negative correlation (P < 0.05) with $\sigma(\varepsilon_s)$ when the TCI exceeded the 434 threshold (Fig. 11b and Fig. 11f). The u^* showed a significantly negative correlation 435 with $\sigma(\varepsilon_s)$ during the daytime and transition periods of the growing season (P < 0.05), 436 while in other time periods, u^* was significantly positively correlated with $\sigma(\varepsilon_s)$ (P <437 0.05). The $|F_s|$ demonstrated a significant positive correlation with $\sigma(\varepsilon_s)$ (P < 0.05) in 438 439 all time periods, with its correlation coefficient being greater during the growing season 440 than during the dormant season. These observations suggested that the effect of turbulent mixing on the magnitude of Fs and its uncertainty was regulated by terrain 441 complexity. 442 3.3 Effect of CO₂ storage fluxes uncertainty on NEE observations 443 The 30-min F_s comb was obtained by weighing the bias and random error of F_s using 444 different [CO₂] averaging time windows and P_{max} values. This study then focused on 445 the magnitude of F_s comb in relation to the F_c magnitude and its diurnal, seasonal, and 446 site variations. To assess the significance of Fs in NEE observations, the relative 447 contribution ratio of F_{s_comb} magnitude ($|F_{s_comb}|/(|F_c|+|F_{s_comb}|)$) was employed. The 448 449 $|F_{s \text{ comb}}|/(|F_{c}|+|F_{s \text{ comb}}|)$ showed a decreasing trend to convergence with increasing u^* 450 (Fig. 12). On average, the $|F_{s_comb}|/(|F_c|+|F_{s_comb}|)$ ranged from 17.2% to 82.0%, with a 451 higher value during the dormant season compared to the growing season. This indicated that as turbulence intensity increased, the contribution of Fs to the NEE in forests 452

positive correlation (P < 0.05) with $\sigma(\varepsilon_s)$ when the TCI was below the threshold, and

decreased to a constant value. Nevertheless, even under strong turbulence intensity, F_s





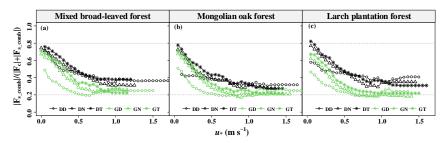


Fig. 12 Relative contribution ratio of the CO₂ storage flux magnitude (|F_{s_comb}|/(|F_c|+|F_{s_comb}|)) determined by decision-level fusion model as a function of the friction velocity (*u**) moving block averages from all 30-min data for the years 2020–2021. GD represents the growing season's daytime; GN represents the growing season's nighttime; GT represents the growing season's transition period; DD represents the dormant season's daytime; DN represents the dormant

season's nighttime; DT represents the dormant season's transition period.

As indicated in Table 3, both P_{max} and TCI exhibited a significant positive correlation with $|F_{\text{s_comb}}|/(|F_{\text{c}}|+|F_{\text{s_comb}}|)$ (P < 0.05), while both A_{max} and WS showed a significant negative correlation with $|F_{\text{s_comb}}|/(|F_{\text{c}}|+|F_{\text{s_comb}}|)$ (P < 0.05). Notably, seasonal variations in correlation coefficients were observed. The correlation between the WS and $|F_{\text{s_comb}}|/(|F_{\text{c}}|+|F_{\text{s_comb}}|)$ was more pronounced during both the dormant season's transition period and the growing season, and it decreased with increasing TCI values during the dormant season's daytime and nighttime.

Table 4 presented a comprehensive comparison of diurnal, seasonal, and site differences in slope (with intercept terms forced to zero) and NRMSE between F_{s_comb} and F_{s_28} . The slopes for the three forest stands ranged from 28.6% to 33.3%, with higher slopes observed during the growing season compared to the dormant season. This suggested that the F_{s_28} was underestimated by 28.6%–33.3% compared to the F_{s_comb} . The NRMSE of F_{s_comb} versus the F_{s_28} in the three forest stands ranged from 59.2% to 67.2%.

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Table 3 Linear regression coefficients of the relative contribution ratio of F_{s_comb} magnitudes to NEE observations (|F_{s_comb}|/(|F_c|+|F_{s_comb}|)) —driving factors relationships for the six time periods.

Time	$oldsymbol{eta}_0$	$ln(P_{max})^7$	$ln(A_{max})^8$	u^{*9}	TCI^{10}	u∗:TCI	R^2
period							
Total	0.292	0.048	-0.037	-0.334	0.790	-1.018	0.278
	***	***	***	***	***	***	***
GD^1	0.299	0.016	-0.041	-0.183	-0.293	0.239	0.158
	***		***	***	*		***
GN^2	0.370	0.029	-0.023	-0.386	-0.038	0.081	0.103
	***		***	***			***
GT^3	0.161	0.060	-0.014	-0.182	1.056	-1.754	0.186
		***	***		***		***
DD^4	0.393	0.011	-0.020	-0.154	0.306	-0.153	0.040
	***		***	*			***
DN^5	0.661	0.012	-0.026	-0.443	-0.035	0.399	0.088
	***		***	***			***
DT^6	0.495	0.017	-0.036	-0.294	0.564	-0.852	0.149
	***		***	***			***

¹ GD represents the growing season's daytime; ² GN represents the growing season's nighttime;

To evaluate the impact of F_s comb on NEE_{obs} ($F_c + F_s$), we further evaluated the

slope (with intercept terms forced to zero) and NRMSE of F_c + F_{s_comb} compared to F_c

+ F_s ₂₈, as presented in Table 5. The slopes for the three forest stands indicated that the

NEEobs with Fs 28 was underestimated by 1.9%-4.3% compared to the NEEobs with

489 F_{s_comb} , with the order MBF > LPF > MOF. The underestimation of NEE_{obs} was higher

490 (6.4%–15.5%) during the growing season nighttime and transition period but much

 $^{^{3}}$ GT represents the growing season's transition period; 4 DD represents the dormant season's

daytime; ⁵ DN represents the dormant season's nighttime; ⁶ DT represents the dormant season's

transition period. $^{7}A_{max}$: maximum amplitude; $^{8}P_{max}$: corresponding period of maximum amplitude.

⁴⁸³ 9 u_* : friction velocity; 10 TCI: terrain complexity index; *** represents P < 0.001; ** represents P < 0.01; * represents P < 0.05.





lower (< 1%) during the growing season daytime. The NRMSE of NEE_{obs} with the F_{s_comb} versus the F_{s_28} in the three forest stands ranged from 16.0% to 25.4%. The analysis suggested that combining the F_s values based on different averaging [CO₂] time windows in the decision-level fusion model could successfully weigh potential underestimation bias and random uncertainties.

Table 4 Statistical inference of major axis regression for the F_{s_28} calculated by combining multiple [CO₂] averaging time windows F_{s_comb} and the 28-min averaging window-based F_s

Forest	Time	N	R^2	Slope	95% CI		NRMSE %
	period						
MBF^7	Total	28061	0.729	1.333	1.323	1.342	67.2
	GD^1	5726	0.813	1.248	1.234	1.262	55.1
	GN^2	3640	0.643	1.535	1.503	1.567	85.8
	GT^3	3092	0.725	1.311	1.287	1.337	67.2
	DD^4	4906	0.914	1.064	1.054	1.075	32.9
	DN^5	6791	0.900	1.099	1.089	1.109	35.0
	DT^6	3906	0.948	1.068	1.059	1.077	25.0
MOF^8	Total	28817	0.783	1.286	1.279	1.294	59.2
	GD	5886	0.781	1.281	1.266	1.296	61.0
	GN	3799	0.752	1.338	1.318	1.358	65.6
	GT	3190	0.751	1.350	1.327	1.373	66.8
	DD	5236	0.931	1.108	1.098	1.118	31.1
	DN	6669	0.934	1.110	1.103	1.118	29.3
	DT	4037	0.951	1.079	1.070	1.088	24.7
LPF9	Total	24273	0.763	1.287	1.278	1.296	61.4
	GD	4659	0.732	1.335	1.314	1.357	68.2
	GN	2995	0.733	1.444	1.418	1.472	73.7
	GT	2544	0.737	1.250	1.225	1.28	63.1
	DD	4231	0.937	1.104	1.094	1.113	29.3
	DN	6288	0.952	1.094	1.088	1.100	24.8
	DT	3556	0.963	1.071	1.064	1.079	21.2

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¹ GD represents the growing season's daytime; ² GN represents the growing season's nighttime; ³ GT represents the growing season's transition period; ⁴ DD represents the dormant season's daytime; ⁵ DN represents the dormant season's nighttime; ⁶ DT represents the dormant season's transition period. ⁷ MBF represents mixed broad-leaved forest; ⁸ MOF represents Mongolian oak forest; ⁹ LPF represents Larch plantation forest.

Table 5 Statistical inference of major axis regression for the $F_c+F_{s_28}$ (NEE_{obs}) calculated by combining multiple [CO₂] averaging time windows ($F_c+F_{s_comb}$) and the 28-min averaging window-based NEE_{obs}

Forest	Time	N	R^2	Slope	95%	. CI	NRMSE %
	period						
MBF^7	Total	28061	0.942	1.043	1.040	1.046	25.4
	GD^1	5726	0.986	1.010	1.007	1.014	15.9
	GN^2	3640	0.828	1.155	1.138	1.172	50.8
	GT^3	3092	0.839	1.136	1.119	1.154	46.3
	DD^4	4906	0.984	1.010	1.007	1.014	12.8
	DN^5	6791	0.949	1.040	1.034	1.045	23.9
	DT^6	3906	0.949	1.050	1.043	1.058	23.9
MOF ⁸	Total	28817	0.976	1.019	1.017	1.020	16.0
	GD	5886	0.993	1.005	1.002	1.007	11.7
	GN	3799	0.908	1.078	1.067	1.089	39.1
	GT	3190	0.892	1.097	1.084	1.110	38.3
	DD	5236	0.993	1.010	1.008	1.012	8.6
	DN	6669	0.974	1.033	1.029	1.037	17.1
	DT	4037	0.972	1.022	1.016	1.027	17.7
LPF^9	Total	24273	0.969	1.024	1.021	1.026	18.1
	GD	4659	0.984	1.012	1.008	1.016	17.0
	GN	2995	0.938	1.062	1.052	1.072	31.9
	GT	2544	0.891	1.064	1.050	1.079	38.1
	DD	4231	0.989	1.017	1.014	1.020	10.9
	DN	6288	0.979	1.035	1.031	1.038	15.5
	DT	3556	0.980	1.030	1.025	1.035	14.8

¹ GD represents the growing season daytime; ² GN represents the growing season nighttime; ³

GT represents the growing season transition period; ⁴ DD represents the dormant season daytime; ⁵

DN represents the dormant season nighttime; ⁶ DT represents the dormant season transition period.





MBF represents mixed broad-leaved forest; ⁸ MOF represents Mongolian oak forest; ⁹ LPF
 represents Larch plantation forest.

4 Discussion

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4.1 Short-term [CO₂] fluctuations above the forest canopy and F_s estimates in complex terrains

This study found that short-term fluctuations of [CO₂] above the canopy exhibited a range of 1 to 10 min (Fig. 2). These fluctuations were characterized by an average P_{max} ranging from 2.313 to 2.784 min (Table 2). Our results are in line with previous research using wavelet analysis, which reported fluctuation periods of [CO2] within and above the forest canopy to be between 14 and 116 s (Cava et al., 2004). Their observations of the canopy waves during periods of extreme atmospheric stability (when $z/L \gg 1$) exhibited a dominant period of 1–2 min, consistent with our findings. The period of [CO₂] fluctuations was found to be predominantly influenced by turbulent fluxes and the residence time of CO₂ within the canopy. This indicated a potential correlation between P_{max} and the residence time of CO₂ within the canopy. Fuentes et al. (2006) employed a Lagrangian model and calculated the residence time of air parcels released near the ground and canopy, finding values ranging from 3 to 10 min and from 1 to 10 min, respectively. Similarly, Edburg et al. (2011) used the standard deviation of [CO₂] averages to determine CO₂ residence time at different locations, including the ground, within the canopy, and in their gas mixtures, yielding values of 8.6, 3.6, and 5.6 min, respectively. The results of these simulation experiments are obtained from our study, further supporting the association between [CO2] fluctuations above the forest

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canopy and CO2 residence time.

residence time; in flat terrains, the air parcel residence time correlate with u^* (Gerken et al., 2017), and an increase in vegetation leaf area leads to longer residence times when turbulence is not fully penetrative. During the growing season, forests exhibit higher leaf area index and canopy densities compared to the dormant season, resulting in longer P_{max} of short-term [CO₂] fluctuations above the canopy (Fig. 3). Additionally, at night, stable atmospheric conditions lead to longer residence times due to suppressed turbulent mixing, resulting in relatively long nighttime P_{max} values compared to daytime and transition periods (Fig. 3). Complex terrains introduce multiple factors that influence [CO₂] fluctuations, including gravity-induced waves, drainage, and advection. These contribute to uncertainties in estimating F_s. During nighttime, long-wave radiation emitted from the valley soil surface leads to the cooling and downslope acceleration of air near the soil surface due to gravity, potentially causing katabatic flow. As inertia-driven upslope winds are halted by katabatic acceleration, a local shallow drainage flow is established, reaching a quasi-equilibrium state approximately 1.5 h after sunset (Nadeau et al., 2013). Under stable atmospheric conditions, even gentle slopes (around 1°) can generate strong gravity-driven waves (Belušić and Mahrt, 2012). Consequently, advection may complicate the interpretation of nighttime EC measurements at certain relatively gentle sites, but this complexity is not evident during daytime measurements (Leuning et al., 2008). Advection plays a role in depleting the CO₂ accumulated within the canopy,

Tree density and canopy structure also play a role in influencing the air parcel

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resulting in lower F_s fluxes and establishing an inverse relationship between storage and advection (Van Gorsel et al., 2011). The occurrence of larger F_s values for long P_{max} values suggests weaker advection compared to short P_{max} values (Fig. 4). In our study, we observed that the F_s magnitude was relatively large during nighttime and transition periods, while it was smaller during daytime (Fig. 4), which is consistent with the findings reported by Wang et al. (2016). During nighttime and transition periods in a closed canopy, the turbulent coupling state above and below the canopy gradually decouples, eventually reaching complete decoupling as the u^* decreases (Fig. 5). However, this decoupling does not lead to stable stratification within the canopy. Despite the occurrence of decoupling and advection in the closed canopy, waves are unlikely to exist within the canopy itself (Van Gorsel et al., 2011). As a result, a consistent trend in the variation of F_s with τ is observed across the three forest stands during the growing season, independent of P_{max} (Fig. 9). Conversely, in an open canopy where waves are present, the observations of F_s become more complex. This complexity could be the primary reason why the variation of F_s with [CO₂] averaging time windows differs between the three forest stands for short P_{max} values during the dormant season daytime (Fig. 9). The presence of waves introduces additional variability in the measurements, leading to differences in Fs estimates based on different [CO2] averaging time windows in these particular conditions. 4.2 Uncertainty in forest ecosystem F_s measurement in complex terrains

Previous studies have highlighted the significant the random uncertainty of Fs in

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an open-canopy forest approximately 0.9 umol m⁻² s⁻¹, compared to the measured change in F_s of 0.3 µmol m⁻² s⁻¹ and the estimated NEE of 6.0 µmol m⁻² s⁻¹ (Van Gorsel et al., 2009). In the current study, we found that the uncertainty of F_s estimates was close to 2 µmol m⁻² s⁻¹ and 1 µmol m⁻² s⁻¹ for the growing and dormant seasons, respectively, as F_s approached zero (Fig. 7). The result for the dormant season was consistent with the previous findings. However, the estimation method employed in our study, comparing observations at two similar moments and ambient conditions, is susceptible to environment changes and flux footprint variability, potentially leading to an overestimation of the total random uncertainty in Fs. The random uncertainty of F_s shares similarities with NEE estimation. For example, the magnitude of F_s measurements is positively correlated with the standard deviation of random uncertainty in Fs. Additionally, the overall distribution of Fs measurements exhibits a non-Gaussian distribution with a high peak, aligning with the statistical properties of NEE uncertainty (Richardson et al., 2006; Richardson et al., 2008). Various factors contribute to the uncertainty in F_s estimates, including flux measurement footprint variations, sampling frequency, spatial sampling resolution of CO₂/H₂O concentrations, and instrumental measurement accuracy. The uncertainty arising from variations in the flux measurement footprint is considerable, typically on the order of tens of percentages, which is an order of magnitude higher than typical sensor errors (Metzger, 2018). The AP200 atmospheric profiling system used in this study has an accuracy of ± 0.5 µmol mol⁻¹ and ± 0.1 mmol mol⁻¹ for CO₂ and H₂O concentration measurements, respectively (Montagnani et al., 2018). Efforts to reduce





random errors in [CO2] originating from pressure fluctuations include adding buffer 598 599 volumes before IRGA pumping tests (Marcolla et al., 2014). The AP200 adopts buffer volumes that are fully mixed during gas extraction and performs a weighted average of 600 601 [CO₂] instantaneous measurements to minimize the sampling error for each level's 602 [CO₂] measurement (Cescatti et al., 2016). The Fs estimates can be influenced by singular eddies that penetrate inside the 603 604 canopy (Finnigan, 2006). Accurate calculation of F_s requires considering the period of 605 [CO₂] fluctuations with the eddy coherence structure. The spectral energy of the F_s time 606 series is primarily concentrated between 0.001 and 0.2 Hz (500 and 5 s, respectively). However, even with sampling frequencies of 2 Hz and below, significantly lower Fs 607 values are obtained (Bjorkegren et al., 2015). The Nyquist-Shannon sampling theorem 608 609 dictates that accurate measurements of [CO₂] require a sampling period no longer than 610 half the period of [CO₂] fluctuations. Consequently, to monitor short-term changes in [CO₂], measurements must be taken over a period no longer than half of the period 611 corresponding to the maximum amplitude (or major energy) of [CO₂] fluctuations. In 612 613 this study, the average P_{max} for [CO₂] fluctuations fell within the range of 2.313–2.784 min (Table 2). Therefore, it is crucial to ensure that the sampling period for [CO₂] does 614 not exceed 1.256 to 1.392 min, which corresponds to half the average P_{max} range. 615 Monitoring fluctuations of P_{max} for less than 4 min during a 2-min monitoring period 616 617 of [CO₂] presents a significant challenge. This is a primary reason that the systematic bias and random error in F_s estimate with a single profile system are irreconcilable 618 (Wang et al., 2016). Short-term [CO₂] fluctuations are mainly influenced by boundary 619



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layer turbulence, and sampling errors in incomplete fluctuation cycles will be superimposed with the real advection flux (anisotropy) dispersion in complex terrains (Van Gorsel et al., 2011). This substantially increases the random uncertainty in F_s based on shorter [CO2] averaging time windows (Fig. 6 and Fig. 8). As a result, the deviation of NEE estimates from the actual value expands. In complex terrains, the bidirectional airflow within forests along slopes can cause the decoupling of soil CO₂ fluxes from EC measurements above the forest canopy (Feigenwinter et al., 2008; Aubinet et al., 2003), leading to significant errors in CO₂ flux measurements. Forest soil serves as the primary source of CO₂ gas and regions of high flux over complex terrains act like chimneys, transporting air parcels from the soil surface within forests (Chen et al., 2019). In situations where turbulence is not welldeveloped, and CO₂ mixing is inadequate, the trend of F_s with turbulence intensity aligns with that of advective fluxes, which is opposite to that of turbulent fluxes (Mchugh et al., 2017). The temporal dynamics and amplitudes of F_s changes are influenced by topography complexity and wind conditions above the forest canopy (Fig. 10). Locations with more complex and sloping topography at the flux tower are more likely to generate advective fluxes that may not be easily observed at a single point. Estimating landscape CO2 fluxes in complex terrains solely based on measurements from a single flux tower can introduce significant errors and biases that are not acceptable. The magnitude of these errors in F_s estimates is dependent on the height of the forest canopy and the endogenous source/sink (Chen et al., 2020). To mitigate errors and biases associated with estimating Fs in complex terrains, we





employed a regression modeling approach using the decision-level fusion model. This method involves computing a weighted average of F_s based on different [CO₂] averaging time windows, effectively reducing errors and biases in the estimation of F_s (see Table 5). In fact, from the definition of storage flux, it can be seen that weighting the storage flux is essentially weighting the [CO₂] in the average time window, which means replacing spatial sequences with temporal sequences for weighting. The weighting coefficients used to construct the model were based on the relative errors and biases of F_s estimation, with the weighting coefficient decreasing as the represented moment's length increased. To obtain more accurate estimates of forest ecosystem F_s in complex terrains, further research should focus on understanding the spatiotemporal patterns and dynamics of [CO₂].

5 Conclusions

This study investigated the impact of short-term [CO₂] fluctuations on the estimation of F_s in temperate forest ecosystems within complex terrains. Additionally, it examined the F_s uncertainty and the contribution of the F_s to NEE using data from three flux towers. To enhance F_s uncertainty estimation, statistical sampling techniques were applied based on the individual tower approach.

The results highlighted the significance of considering multiple time windows for averaging [CO₂] when estimating F_s , as [CO₂] above the forest canopies exhibited fluctuations with periods ranging from 1 to 10 minutes. Diurnal, seasonal, and spatial

variations were observed in the amplitude and periodicity of [CO₂] fluctuations,

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highlighting the need for thoughtful sampling strategies. The use of individual gas analyzers to sample the CO₂ in the control volume was inadequate, leading to systematic biases and random errors in the F_s estimates. Increasing [CO₂] averaging time windows mitigated the effect of [CO₂] fluctuations on F_s estimates, reducing both magnitude and uncertainty. The study also revealed that the uncertainty of Fs followed a non-normal distribution, with its standard deviation positively correlated with F_s magnitude, which has important implications for quality control. To improve F_s estimation, a decisionlevel fusion model was introduced, integrating F_s estimates from multiple [CO₂] averaging time windows, effectively reducing the impact of short-term [CO2] fluctuations while considering underestimation bias and random errors. The contribution of Fs to NEE exhibited diurnal, seasonal, and spatial variations associated with u^* , contributing to the NEE observations at rates ranging from 17.2% to 82.0% depending on the turbulent mixing and terrain complexity. The influence of terrain complexity on the relationship between [CO₂] fluctuations, turbulent mixing, and the contribution of Fs to NEE was also evident. The findings from the three flux towers allowed for the generalization of these results beyond the study site. These insights provide crucial scientific support for the practical application of the eddy covariance technique and advance our understanding of accurately estimating NEE in forest

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Appendix A

ecosystems in complex terrains.





- 685 A.1 the weight parameters of the decision-level fusion model
- For each 30-min CO₂ storage flux (F_s) estimate based on the CO₂ concentration
- 687 ([CO₂]) averaging time window (τ), the weight in the decision-level fusion model can
- be obtained by weighting the random uncertainty and bias of $F_{s_{\underline{\tau}}}$.
- The weight of the random uncertainty for the $F_{s_{-\tau}}$ is expressed as follows:

$$w_{\tau} = \frac{1/\sigma(\varepsilon_{\tau})}{\sum_{i} 1/\sigma(\varepsilon_{\tau})},\tag{A.1}$$

- where $\sigma(\varepsilon_{\tau})$ is the random uncertainty of the $F_{s_{\tau}}$, qualified as the standard deviation.
- The weight of the bias for the $F_{s \tau}$ is expressed as follows:

$$W_{\tau} = \frac{K_{\tau}}{\sum_{i} K_{i}},\tag{A.2}$$

- where K_{τ} is the slope between the $F_{s_{-}\tau}$ and $F_{s_{-}28}$.
- Ultimately, the weight of the $F_{s_{-}\tau}$ in the decision-level fusion model can be
- 694 calculated using the following equation:

$$w_{\tau}^* = rw_{\tau} + (1 - r)W_{\tau},\tag{A.3}$$

- where r represents the proportion of the weight of random uncertainty.
- 696 A.2 Complex terrain index
- This study employed a novel descriptor called the terrain complexity index (*TCI*)
- 698 to quantify the complexity of the three-dimensional terrain. For a given unit area, the
- 699 TCI equation can be expressed as follows:

$$TCI = (1 - P_d \cos \alpha)(1 - Z_d^{-1})(D_f - 2)^{H/\ln{(12)}},$$
 (A.4)

- 700 where, P_d represented the volume of terrain above the lowest elevation of an area unit
- 701 (V_u) divided by the product of its largest vertically projected area (S_v) and the edge





length of the side of the area unit (d), expressed as $P_d = V_u/(S_v d)$; P_d was defined to 702 be one when the S_v is zero. α_d indicated the slope of the area unit. Z_d denoted the 703 terrain roughness, which defined as the ratio of the terrain surface area to the projected 704 horizontal plane. D_f is the fractal dimension of terrain surface area, which ranged from 705 706 2 to 3 and described the complexity in spatially self-similar structure of the local surface within the area unit and the area unit surface (B. B. Mandelbrot, 1967; Taud and Parrot, 707 708 2005). Employing terrain surface area, the box-counting method is used to estimate 709 fractal dimension of unit area. H denoted the Shannon-Wiener index and expressed as 710 $H = -\sum_{i=1}^{n} P_i \ln(P_i)$, capturing the uniformity of the spatial distribution of the pixel aspects within the area unit. When the aspect of each pixel was divided into 30° 711 segments, P_i denoted the proportion of the i^{th} type of pixel aspects within the area unit 712 713 and n was the total number of pixel aspect types within the area unit.

To quantify the terrain complexity of the underlying surface around the flux towers, we computed the quartiles of *TCI* for all area units within a sector (divided by 30°) with a radius of 380 m. A weighted geometric mean was employed to construct *TCIs*, which describe the statistical distribution of *TCI* of the sector. The *TCIs* represents the topographic complexity of the sector and are calculated using the following equation:

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$$TCI_{s} = (TCI_{5}TCI_{25}TCI_{50}TCI_{75}TCI_{95})^{1/5}$$
 (A.5)

720 where *TCI*₅, *TCI*₂₅, *TCI*₅₀, *TCI*₇₅, and *TCI*₉₅ are the quartiles of 5%, 25%, 50%, 75%,

and 95%, respectively. The *TCIs* values range from 0 to 1, with higher values indicating

722 greater terrain complexity.

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723 Data availability. Data used in this paper are available at the Science Data Bank





- 724 (https://www.scidb.cn/en/s/7ZfQZv) or upon request to the corresponding author. 725 Author contributions. DT developed the manuscript; JZ was responsible for conceptualizing the idea and designing the research study; TG substantially structured 726 the manuscript; FY contributed to the data collection process; YZ helped in the design 727 728 and preparation of the figures and tables; XZ and BY revised the manuscript. Competing interests. The authors declare that they have no known competing 729 730 financial interests or personal relationships that could have appeared to influence the 731 work reported in this paper. 732 Acknowledgments. We are grateful to Qingyuan Forest CERN, Chinese Academy of Sciences/Qingyuan Forest, National Observation and Research Station, Liaoning 733 Province, China for providing forest sites, instrument systems, and logistic supports. 734 735 Financial support. This research was financially supported by the National Natural 736 Science Foundation of China (No. 32192435), the China Postdoctoral Science Foundation (No. 2023M733672), and the Postdoctoral Research Startup Foundation 737 of Liaoning Province of China (No. 2022-BS-022). 738 739 Reference 740 Aubinet, M., Heinesch, B., and Yernaux, M.: Horizontal and Vertical CO₂ Advection In A Sloping Forest, 741 Boundary-Layer Meteorology, 108, 397-417, 10.1023/a:1024168428135, 2003. 742 Aubinet, M., Grelle, A., Ibrom, A., Rannik, Ü., Moncrieff, J., Foken, T., Kowalski, A. S., Martin, P. H.,
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